FINAL REPORT

Evaluation of Perchlorate Sources in the Rialto-Colton and Chino California Subbasins using Chlorine and Oxygen Isotope Ratio Analysis

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List of Acronyms and Abbreviations

‰ per mil

 δ , Δ delta, relative difference of isotope ratios

Ag silver

AgCl silver chloride AgNO₃ silver nitrate

AMS accelerator mass spectrometry

BaSO₄ barium sulfate bls below land surface

C carbon

12C carbon-12

13C carbon-13

°C degrees Celsius

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CF continuous-flow

CF-IRMS continuous-flow isotope ratio mass spectrometry

CH₃Cl methyl chloride CH_3I methyl iodide chloride Clchlorine Cl ³⁵Cl chlorine-35 ³⁶Cl chlorine-36 ³⁷Cl chlorine-37 ClO_3 chlorate ClO_4 perchlorate centimeters cm

CO carbon monoxide CO₂ carbon dioxide DI dual-inlet

DI-IRMS dual-inlet isotope ratio mass spectrometry

DIW deionized water
DO dissolved oxygen
DoD Department of Defense

EIGL Environmental Isotope Geochemistry Laboratory

ESTCP Environmental Security Technology Certification Program

 $\begin{array}{ll} Fe & iron \\ Fe^{3+} & Iron (III) \end{array}$

FeCl₃ iron (III) chloride FeCl₄ tetrachloroferrate

GIS geographic information system

²H deuterium ³H tritium

HCl hydrochloric acid

HCO₃ bicarbonate
He helium
HNO₃ nitric acid
H₂O water

H₂O₂ hydrogen peroxide

hr hours

IAEA International Atomic Energy Agency

IC ion chromatography

IRMS isotope-ratio mass-spectrometry

ITRC Interstate Technology & Regulatory Council

IX ion exchange k one thousand KCl potassium chloride KClO₄ potassium perchlorate

kg kilogram

km² square kilometer KNO₃ potassium nitrate KOH potassium hydroxide

L liter

LPM liters per minute µg microgram

μg/L microgram per liter

 $\begin{array}{ll} \mu mol & micromole \\ M & molar \\ M\Omega & megaohm \end{array}$

MADEP Massachusetts Department of Environmental Protection

MCL maximum contaminant level

mg milligram
mL milliliter
mm millimeter

MRGB Middle Rio Grande Basin MVSL Mid-Valley Sanitary Landfill

m/z mass to charge ratio

N nitrogen or normal or number of entities

N₂ nitrogen gas NaCl sodium chloride

NADP National Atmospheric Deposition Program

NaNO₃ sodium nitrate ng nanogram

NH₄ClO₄ ammonium perchlorate

 N_2 nitrogen gas NO_2 nitrogen dioxide

NO₃ nitrate

NWIS National Water Information System

O oxygen oxygen-16

 $\begin{array}{ccc} ^{17}O & oxygen-17 \\ ^{18}O & oxygen-18 \\ O_2 & oxygen gas \\ O_3 & ozone \\ pCi & picoCuries \end{array}$

pCi/L picoCuries per liter per mil (%) part per thousand (× 10⁻³)

PRIME Purdue Rare Isotope Measurement Laboratory

QA/QC Quality Assurance/Quality Control

R ratio

RFF rockets, fireworks, and flares

rg reference gas

S sulfur

SARWQCB Santa Ana Regional Water Quality Control Board

SERDP Strategic Environmental Research and Development Program

SHP Southern High Plains

SIMS secondary ion mass spectrometry
SLAP Standard Light Antarctic Precipitation

SMOC Standard Mean Ocean Chloride

SO₄²- sulfate

TAC Technical Advisory Committee

TPA tetrapentylammonium

TPABr tetrapentylammonium bromide TPAClO₄ tetrapentylammonium perchlorate UIC University of Illinois at Chicago

U.S. United States

USDA United States Department of Agriculture

USEPA United States Environmental Protection Agency

USGS U.S. Geological Survey

VSMOW Vienna Standard Mean Ocean Water

wt weight yr year

Abstract

Perchlorate (ClO₄⁻) in groundwater can be from synthetic or natural sources, the latter of which include both historical application of imported nitrate fertilizers from the Atacama Desert of Chile and naturally deposited ClO₄ that forms atmospherically and accumulates in arid regions such as the southwestern US. The objective of this study was to evaluate the use of isotopic data to distinguish sources of ClO₄ in groundwater in a specific region of the Rialto-Colton and Chino, CA groundwater subbasins (Study Area). This region includes two groundwater ClO₄ plumes emanating from known military/industrial source areas, and a larger area outside of these plumes having measurable ClO₄⁻. Perchlorate extracted from wells in this region was analyzed for chlorine and oxygen stable isotope ratios (δ^{37} Cl, δ^{18} O, δ^{17} O) and radioactive chlorine-36 (³⁶Cl) isotopic abundance, along with other geochemical, isotopic, and hydrogeologic Isotope data indicate synthetic, Atacama, and indigenous natural ClO₄ were present in the Study Area. Stable isotope data from nearly all sampled wells within the contours of the two characterized plumes, including those located in a perched zone and within the regional groundwater aquifer, were consistent with a dominant synthetic ClO₄ source. In wells downgradient from the synthetic plumes and in the Chino subbasin to the southwest, isotopic data indicate the dominant source of ClO₄ largely was Atacama, presumably from historical application of nitrate fertilizer in this region. Past agricultural land use and historical records are consistent with this source being present in groundwater. The 36 Cl and δ^{18} O data indicate that wells having predominantly synthetic or Atacama ClO₄⁻ also commonly contained small fractions of indigenous natural ClO₄⁻. The indigenous ClO₄ was most evident isotopically in wells having the lowest overall ClO₄ concentrations (< 1 µg/L), consistent with its occurrence as a low-level background constituent in the region. A small subset of wells outside the contours of the two synthetic plumes, including an upgradient well, had characteristics indicating small amounts of synthetic ClO₄ mixed with one or both of the natural source types. Hydrogeologic data indicate synthetic ClO₄ in the upgradient well may be from a source other than the identified plume sources, but it is not known whether this source might also be present in other wells at low concentrations. The stable isotope and ³⁶Cl data provided relatively unambiguous discrimination of synthetic and Atacama ClO₄ sources in most wells having relatively high concentrations, providing regional perspective on anthropogenic ClO_4^- contamination in the Rialto-Colton and Chino subbasins. Where indigenous natural ClO_4^- was indicated as a substantial component, total ClO_4^- concentrations were low and concentrations of anthropogenic components were near background levels.

1.0 Introduction

The contamination of groundwater in the United States by perchlorate (ClO₄-) has become an issue of national concern. It was once assumed that ClO₄ contamination of groundwater largely resulted from historical testing and disposal practices by the military, the aerospace and ordnance industries, and ClO₄ manufacturers. However, during the past decade it has become apparent that widespread agricultural use of nitrate fertilizer from the Atacama Desert of Chile (Urbansky et al., 2001a,b; Böhlke et al., 2005; Böhlke et al., 2009; Sturchio et al., 2011, 2012, 2014) as well as mobilization of indigenous natural ClO₄ (Dasgupta et al., 2005, 2006; Rao et al, 2007) may also contribute to groundwater contamination. Isotopic techniques have been developed over the past decade to characterize different sources of ClO₄⁻ in the environment. The objective of this study was to evaluate sources of ClO₄ in groundwater in a specific region of the Rialto-Colton and Chino, CA subbasins (Study Area) using chlorine and oxygen stable isotope ratio analysis and chlorine-36 (³⁶Cl) radioactive isotope analysis. In conjunction with ClO₄ isotopic analysis, additional geochemical and hydrogeological data were collected in this region and evaluated. This evaluation included (1) water level mapping to evaluate the flow of groundwater, and the potential for ClO₄ transport in the study area; (2) water chemistry analyses, including major ion composition, ClO₄⁻ concentrations, stable isotope ratios of hydrogen, oxygen, nitrogen, and sulfur in water, nitrate, and sulfate, groundwater age dating parameters, and dissolved gases; and (3) coupled well-bore flow and depth-dependent water quality data for several groundwater supply wells. The geochemical and hydrogeological data, along with interpretations, were published previously (Izbicki et al., 2014). That paper provides the hydrogeologic framework to support the isotopic study described herein.

1.1 Perchlorate Sources

The following sections summarize information about the principal sources of ClO_4^- in the environment.

1.1.1 Synthetic Perchlorate

Ammonium perchlorate (NH₄ClO₄) has been used for several decades in the United States (U.S.) as the primary oxidant in a variety of solid rocket propellants and explosives produced for military and aerospace applications. More than 100 varieties of military rocket motors contain NH₄ClO₄ (Cunniff et al., 2006). Potassium perchlorate (KClO₄) is also used for many military applications. Besides military propellants and explosives, a variety of commercial products contain synthetic ClO₄⁻ either intentionally or as a manufacturing byproduct, including fireworks, matches, air bags, chlorine bleach, safety flares, perchloric acid, and chlorate herbicides (Trumpholt et al., 2005; Aziz et al., 2006, 2008).

Based on United States Environmental Protection Agency (USEPA) estimates, approximately 4 × 10⁸ kg (400,000 metric tons) of synthetic ClO₄⁻ has been produced in the U.S. since the 1950's, with an average production rate of 7.1 × 10⁶ kg/yr from 1951-1997 (Dasgupta et al., 2006). Historical testing and disposal practices at some military installations, ClO₄⁻ production facilities, and aerospace sites have resulted in substantial contamination of soils and groundwater at these locations. Many groundwater plumes, which are often extensive and have ClO₄⁻ concentrations up to several hundred milligrams per liter (mg/L), have been identified and are subject to site assessment and(or) remediation (e.g., Hatzinger, 2005; ITRC, 2008). However, the contribution of other commercial products as non-point or small point sources of synthetic ClO₄⁻ is more difficult to assess and quantify, although road flares, blasting agents, and fireworks have been indicated as sources of ClO₄⁻ in some groundwaters (e.g., MADEP, 2007; Munster et al., 2008; Munster, 2008; Böhlke et al., 2009; Munster and Hanson, 2009). Further information on synthetic ClO₄⁻ sources can be found in previously published reviews (Trumpholt et al., 2005; ITRC, 2008; Aziz et al., 2006).

1.1.2 Atacama Nitrate Deposits and Fertilizer

Natural ClO₄⁻ has long been known to co-occur with sodium nitrate (NaNO₃) in surficial deposits in the Atacama Desert of Chile at an average concentration of around 0.1 % (by mass) of the total soluble salt, with concentrations as high as 6.8 % reported (Schilt, 1979; Ericksen, 1981, 1983; Dasgupta et al., 2006). These NO₃⁻ deposits,

sometimes referred to as "nitrate caliche", were widely used in the U.S. during the first half of the 20^{th} century as a source of inorganic nitrogen fertilizer. According to the California Department of Agriculture, more than 477,000 metric tons of imported Atacama NO_3^- was used in California as fertilizer between 1923 and 1998 (California Department of Food and Agriculture, 1999). The final ClO_4^- concentration of processed Atacama NO_3^- fertilizer prior to 2002 was variable, and likely ranged from ~1,500 to 10,000 mg/kg (i.e., 0.15 to 1.0 wt. %) based upon analysis of historical samples (Eldridge et al., 2000; Urbansky et al, 2001a,b; Dasgupta et al., 2006). Changes in manufacturing processes after 2002 were reported to reduce the final ClO_4^- concentration to ≤ 100 mg/kg (0.01%) (Dasgupta et al., 2006). Thus, apart from synthetic sources, past application of Atacama NO_3^- fertilizer provides a potential source of ClO_4^- in groundwater and drinking water in the U.S.

Based on a reasonable assumption of ~2,000 mg ClO₄-/kg Atacama fertilizer (0.2%) (Dasgupta et al., 2006), one metric ton of processed Atacama fertilizer (prior to 2002) contained about 2 kg of ClO₄-; enough to contaminate more than 300 million liters of groundwater to above the current California maximum contaminant level (MCL) of 6 μg/L, assuming no losses to biodegradation or other removal processes. Additional information on Atacama NO₃- fertilizers as a source of ClO₄- can be found in published papers (Ericksen, 1981, 1983; Bao and Gu, 2004; Dasgupta et al., 2006; Jackson et al., 2006; Böhlke et al., 2009; Sturchio et al., 2011, 2012, 2014). Some other fertilizer materials, including plant products such as kelp, have been reported to contain ClO₄-, but the concentrations in these materials are generally orders of magnitude lower than those historically present in Atacama NO₃-, and they are considered less likely to be of widespread environmental significance (Dasgupta et al., 2006; Böhlke et al., 2009).

1.1.3 Indigenous Natural Perchlorate in the US

Natural ClO₄⁻ that is not associated with fertilizers from the Atacama has also been detected in soils, groundwaters, and mineral deposits collected from arid regions in the western U.S., including groundwater underlying an area of 155,000 km² in the Southern High Plains (SHP) of Texas and New Mexico (Jackson et al., 2004, 2005, 2006; Dasgupta et al., 2005; Rajagopolan et al., 2006), in the Middle Rio Grande Basin of New

Mexico (Plummer et al., 2006), and also in the northwestern U.S. (Hatzinger et al., 2013). Natural ClO₄⁻ also was reported in the Great Lakes (Poghosyan et al., 2014). Based on groundwater dating, some New Mexico samples (having ClO₄⁻ concentrations ranging from ~ 0.12 to 1.8 μg/L) were recharged many thousands of years before present, indicating pre-anthropogenic (natural) ClO₄⁻ sources and accumulation processes (Plummer et al., 2006). Natural ClO₄⁻ is hypothesized to form in the atmosphere through photochemical reactions (Murphy and Thomson, 2000; Bao and Gu, 2004; Dasgupta et al., 2005; Kang et al., 2008; Sturchio et al., 2009) and to reach Earth's surface via precipitation or dry deposition. This hypothesis was supported by ubiquitous occurrence of ClO₄⁻ in precipitation from National Atmospheric Deposition Program (NADP) collectors at 26 sites across the contiguous U.S., Puerto Rico, and Alaska over a 3-yr period at concentrations ranging from <5 ng/L to 102 ng/L (n=1578) (Rajagopalan et al., 2009).

In arid regions, such as parts of the southwestern U.S., ClO₄⁻ in wet and dry deposition can accumulate with time in the vadose zone, along with other deposited salts. Rao et al. (2007) detected ClO₄⁻ in the vadose zone at several sites in the southwestern U.S. with an average accumulated mass per unit area of 408 ± 88 g/ha. The ClO₄⁻ concentrations were correlated (r = 0.59-0.99) with those of meteoric Cl⁻ accumulated over the last 6 to 100 kyr, indicating that these anions accumulated together via similar processes. Indigenous natural ClO₄⁻ also is associated with natural surficial NO₃⁻ deposits in the vicinity of Death Valley, California at concentrations ranging from 0.25 to 23 mg/kg (Jackson et al., 2010; Lybrand et al., 2013). Similarly, ClO₄⁻ concentrations as high as 1.1 mg/kg were reported in surficial deposits in the Dry Valleys region of Antarctica (Kounaves et al., 2010; Jackson et al., 2013).

When unsaturated-zone salt accumulations in arid environments become subject to large-scale irrigation, such as in the SHP and other agricultural regions of the western U.S., accumulated ClO₄⁻ can be mobilized and transported to the water table. Such agricultural mobilization is hypothesized to account for high ClO₄⁻ concentrations in groundwater in West Texas, which has been widely irrigated for several decades for production of cotton and other crops (Rajagopalan et al., 2006). Irrigation also may be

associated with transport of ClO₄⁻ to groundwater from crops treated with Atacama nitrate fertilizer (Böhlke et al., 2009; Sturchio et al., 2014)

1.2 Location and Characteristics of the Study Area

Release of synthetic ClO₄ from military and industrial sources in the northern region of the Rialto-Colton subbasin of San Bernadino County, CA has impacted a number of municipal supply wells. Previously identified sources include two sites: (1) a 160-acre industrial site formerly occupied by several different industries, and also known as the "Rockets, Fireworks, and Flares" (RFF) site (USEPA, 2014), and (2) properties adjacent to the current Mid-Valley Sanitary Landfill (MVSL) operated by San Bernardino County, including the Rialto Ammunition Backup Storage Point, a site of several World War II era bunkers that were later used for the manufacture, transport, and/or disposal of fireworks, flares, explosives and other potentially hazardous materials (SAIC, 2004). This site is known as the "Mid Valley Sanitary Landfill/Bunker Site" (Geologic Associates, 2003; SARWQCB, 2005; Woolfenden, 2007) or the "Former Bunker Area". The distribution of ClO₄ concentrations in wells indicates two parallel plumes emanating from those sites (Figure 1.1). As defined by previous studies, the mapped plume from the 160-acre site is parallel to the Rialto-Colton fault and extends at least 6 km downgradient of the site in the southeast direction of groundwater flow (GeoLogic Associates, 2013). Recent USEPA reports suggest that this plume may reach as far as 9 km downgradient, but the distal extent is not yet known (USEPA, 2014). The mapped plume from the Bunker Site/MSVL extends ~ 3 km downgradient, parallel to the RFF site plume.

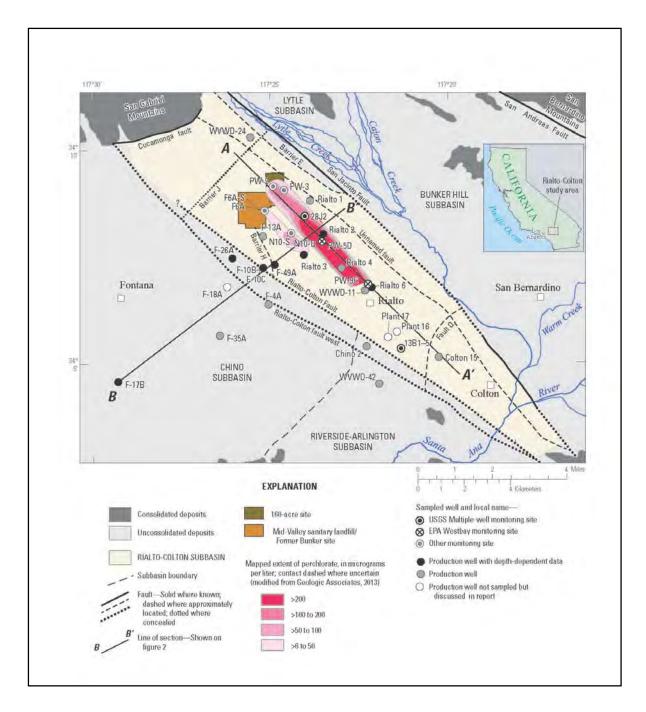


Figure 1.1. Overhead view of the Study Area with the two previously identified source areas and the associated perchlorate plumes delineated. Figure from Izbicki et al. (2014). This plume map was generated previously and modified from Geologic Associates (2013). The map does not include concentration data from this study, the most recent USEPA site sampling (USEPA, 2014) or Izbicki et al. (2014).

The extent of high-level ClO_4^- contamination (e.g., > 50 µg/L) emanating as plumes from the two military/industrial sites is reasonably well defined. However, ClO₄⁻ has also been detected in groundwater wells throughout the Study Area at lower concentrations (~ 1 to 20 µg/L). Perchlorate has been measured in groundwater wells to the southeast of the two known plumes in the general direction of groundwater flow, such as Colton 15 and 13B1-5 (RHSW-5), and in wells to the south of the sources but outside the hypothesized groundwater flow path, including a number of wells positioned to the south of the Rialto-Colton Fault in the Chino subbasin, such as F-35A, F-4A, F-17B and others (Figure 1.1). Several wells in the Chino subbasin positioned between the Rialto-Colton Fault and the postulated Rialto-Colton Fault West (Paulinski, 2012), such as F-26A and Chino 2 (Figure 1.1), also have measurable ClO₄, albeit at very low concentrations (< 2 µg/L). The general pattern of low-level ClO₄ contamination in groundwater throughout the Study Area (apart from the plumes emanating from the two source areas) suggests the possibility of additional sources of ClO₄⁻ in the area, possibly including ClO₄ derived from historical Atacama fertilizer application, natural indigenous ClO₄, and secondary synthetic sources from flares, fireworks, blasting, or others. The focus of this project was to use stable isotope ratio analysis and ³⁶Cl analysis to evaluate the major source(s) of ClO₄- throughout the Study Area. A separate report provides information on groundwater flow and additional hydrogeochemical characteristics in the Study Area (Izbicki et al., 2014).

1.3 Isotopic Analyses of ClO₄-

The two elements comprising the ClO₄⁻ molecule (i.e., Cl and O) each have multiple stable isotopes. Oxygen has three stable isotopes (¹⁶O, ¹⁷O, and ¹⁸O), which have molar abundances (mole fractions) of approximately 99.8 %, 0.04 %, and 0.20 %, respectively (Coplen et al., 2002). Chlorine has two stable isotopes (³⁵Cl and ³⁷Cl), with molar abundances of approximately 75.8 % and 24.2 %, respectively (Coplen et al., 2002).

Stable isotopic compositions are measured and reported as relative differences of isotope ratios (in either moles or numbers of atoms) between two substances, one of which is generally an international measurement standard. For oxygen:

[Eq. 1.1]
$$\delta^{18}O = R(^{18}O/^{16}O)_{\text{sample}}/R(^{18}O/^{16}O)_{\text{standard}} - 1$$

[Eq. 1.2]
$$\delta^{17}O = R(^{17}O/^{16}O)_{\text{sample}}/R(^{17}O/^{16}O)_{\text{standard}} - 1,$$

where the standard in both equations is Vienna Standard Mean Ocean Water (VSMOW). R = molar ratio. Values of $\delta^{18}O$ and $\delta^{17}O$ are reported in parts per thousand (per mil, or ‰). By international convention, the $\delta^{18}O$ scale is defined by two reference materials, with VSMOW at 0.0 ‰ and Standard Light Antarctic Precipitation (SLAP) at a value of 55.5 ‰ (Gonfiantini, 1978; Coplen, 1994).

Variations in $R(^{17}\text{O}/^{16}\text{O})$ and $R(^{18}\text{O}/^{16}\text{O})$ caused by most physical-chemical fractionation processes on Earth are related systematically by the relative differences in the masses of the isotopes. Such "mass-dependent" variations can vary slightly for different processes and they can be described in various ways (e.g., Thiemens, 1999; Miller, 2002; Angert et al., 2004; Assonov and Brenninkmeijer, 2005). For this project:

[Eq. 1.3]
$$(1 + \delta^{17}O) = (1 + \delta^{18}O)^{\lambda},$$

with $\lambda \approx 0.525$ (Miller, 2002; Böhlke et al., 2005). Departures from mass-dependent O-isotope variation are important features of some materials, including some natural ClO₄⁻. Departures from mass-dependent O-isotope variation in ClO₄⁻ are described in this report as deviations from the relation given in Equation 1.3:

[Eq. 1.4]
$$\Delta^{17}O = [(1 + \delta^{17}O) / (1 + \delta^{18}O)^{0.525}] - 1.$$

Values of Δ^{17} O are reported in parts per thousand (per mil, or ‰). Alternative definitions of Δ^{17} O used to describe ClO_4^- isotope data can yield slightly different Δ^{17} O values for the same measured δ^{18} O and δ^{17} O values; those differences range from about 0.0 to 0.2 ‰ for groundwater data summarized in this report.

For reporting chlorine stable isotope ratios:

[Eq. 1.5]
$$\delta^{37}\text{Cl} = R(^{37}\text{Cl})^{35}\text{Cl})_{\text{sample}} / R(^{37}\text{Cl})^{35}\text{Cl})_{\text{standard}} - 1$$

where the international measurement standard is Standard Mean Ocean Chloride (SMOC). Values of δ^{37} Cl typically are reported in parts per thousand (per mil, or ‰).

Chlorine also has a long-lived radioactive isotope (³⁶Cl) with a half-life of ~ 301,000 yr. The relative abundance of ³⁶Cl is small (typically <10⁻¹⁵ to 10⁻¹² relative to the stable Cl isotopes) but it can be useful for studying origins of chloride and Cl-bearing compounds (Phillips, 2000). Techniques to analyze the stable isotopes of both chlorine and oxygen in the ClO₄⁻ molecule, as well as its ³⁶Cl isotopic abundance, have been developed and used to characterize a variety of natural and man-made ClO₄⁻ samples, including Atacama nitrate fertilizers and caliche deposits, a wide array of synthetic ClO₄⁻ types and ClO₄⁻-containing products (e.g., laboratory reagents, fireworks, flares, gunpowder, military sources), and samples from the southwest US that contain natural indigenous ClO₄⁻ (Bao and Gu, 2004; Böhlke et al., 2005, 2009; Jackson et al., 2010; Gu et al., 2011; Sturchio et al., 2006, 2009, 2012; Hatzinger et al., 2011, 2013). A summary of these techniques as applied to ClO₄⁻ isotopic analysis of Study Area groundwater samples is provided in Section 2.

1.3.1 Stable Isotope Ratios of Cl and O in Synthetic ClO₄

Synthetic ClO_4^- is synthesized electrochemically using NaCl and H_2O (Schumacher, 1960). Samples from a variety of different synthetic sources, including laboratory reagents, commercial manufacturers, and ClO_4^- derived from road flares, fireworks, Pyrodex gunpowder, and chlorate herbicides, were analyzed for Cl and O stable isotope ratios. Additional details on these samples are provided in Bao and Gu (2004), Böhlke et al. (2005), Sturchio et al. (2006, 2011), and Hatzinger et al. (2011, 2013). The $\delta^{37}Cl$ values of all synthetic ClO_4^- samples published to date group within a range from -3 ‰ to + 2 ‰. These values are similar to those reported for common industrial sources of NaCl, such as halite from Phanerozoic bedded marine evaporites, which has a mean $\delta^{37}Cl$ value of 0.0 ± 0.9 ‰ (Eastoe et al., 2007). The electrochemical synthesis of ClO_4^- is nearly stoichiometric for Cl^- (i.e., nearly all of the added Cl^- is converted to ClO_4^-), so the Cl isotope ratio in the ClO_4^- product is similar to that in the starting NaCl. The published $\delta^{18}O$ values of synthetic ClO_4^- vary over a wider range than those for $\delta^{37}Cl$, from -25 ‰

to -13 ‰ (Figure 1.2). It has been hypothesized that the δ^{18} O values of the synthetic ClO_4^- samples generally reflect the δ^{18} O in the H_2 O used for production modified by O isotopic fractionation during ClO_4^- synthesis (Sturchio et al., 2006). In contrast to δ^{18} O, Δ^{17} O values of all synthetic ClO_4^- samples analyzed to date are indistinguishable (0.0 ± 0.1 ‰), indicating that there is negligible mass-independent isotopic fractionation of O during ClO_4^- synthesis (Sturchio et al., 2006).

1.3.2 Stable Isotope Ratios of Cl and O in Atacama ClO₄

Current data from isotopic analyses reveal that the $^{37}\text{Cl}/^{35}\text{Cl}$ isotope ratio ($\delta^{37}\text{Cl}$ value) of naturally-occurring ClO_4^- from Chile is consistently and significantly lower than that of synthetic ClO_4^- from all other sources tested (Figure 1.2). In fact, the reported $\delta^{37}\text{Cl}$ values for Atacama ClO_4^- are the lowest for any common substance on Earth (Coplen et al., 2002). Based on samples analyzed to date, which include soils, caliche deposits and a groundwater sample from the Atacama Desert (blue diamonds in Figure 1.2), and samples of commercial Atacama nitrate fertilizer (green diamonds in Figure 1.2), the $\delta^{37}\text{Cl}$ values range from -19 to -9 ‰ (Bao and Gu, 2004; Böhlke et al., 2005; Sturchio et al., 2006, 2012; Böhlke et al., 2009; Jackson et al., 2010; Hatzinger et al., 2011, 2013; unpublished results from SERDP Project ER-1435 for a subset of the Atacama soil samples). The $^{18}\text{O}/^{16}\text{O}$ isotope ratio ($\delta^{18}\text{O}$ value) of Atacama natural ClO_4^- spans a wide range, from -25 to -2 ‰, which overlaps that of synthetic ClO_4^- (Figure 1.2).

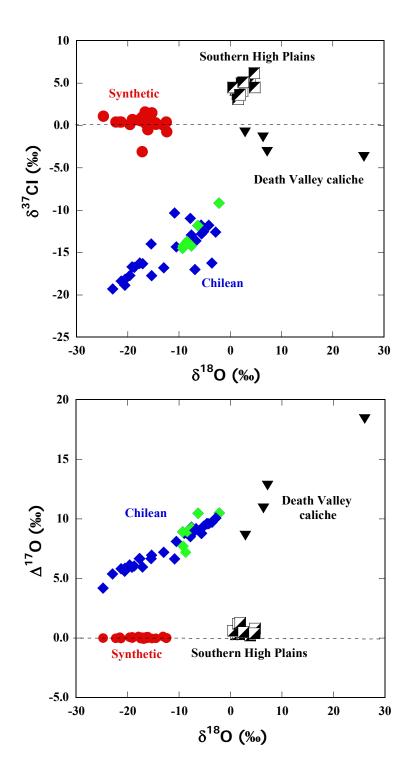


Figure 1.2. Comparison of δ^{37} Cl versus δ^{18} O (top panel) and Δ^{17} O versus δ^{18} O (bottom panel) in natural indigenous ClO₄⁻ of the US Southern High Plains (black and white squares) and Death Valley caliche (black triangles), natural Atacama ClO₄⁻, (blue diamonds for caliche and groundwater and green diamonds for nitrate fertilizer samples) and synthetic ClO₄⁻ (red circles). Sources of data are provided in the accompanying text.

One important isotopic difference between synthetic ClO_4^- and natural ClO_4^- from the Atacama Desert is revealed by analysis of ^{17}O abundances. Synthetic ClO_4^- samples have $\Delta^{17}O$ values near 0, consistent with ClO_4^- production from brine by electrolysis. In contrast, $\Delta^{17}O$ values of Atacama ClO_4^- samples analyzed to date range from + 4 to +11 ‰, with commercial nitrate fertilizer ClO_4^- samples being in the upper end of this range (Figure 1.2). Elevated $\Delta^{17}O$ values in Atacama ClO_4^- were first measured by Bao and Gu (2004), who noted that the ^{17}O enrichment is consistent with an atmospheric formation mechanism. The data indicate that oxidation of volatile Cl species by ozone (O₃) (which is known to have elevated $\Delta^{17}O$ values; Johnson et al., 2000) in the upper atmosphere may be responsible for the initial production of this ClO_4^- (Bao and Gu, 2004; Rao et al., 2010).

1.3.3 Stable Isotope Ratios of Cl and O in Indigenous Natural ClO₄

As previously discussed, research conducted during the past decade has revealed that natural ClO₄⁻ occurs in environments other than the Atacama Desert of Chile. Most significantly for forensic studies in the U.S., indigenous natural ClO₄⁻ has been widely detected in the southwestern U.S., as evidenced through various surveys of ClO₄⁻ in surface soils, vadose-zone profiles, wet and dry deposition, and groundwater in New Mexico, Texas, California and elsewhere in this region (Dasgupta et al., 2006; Plummer et al., 2006; Rajagopalan et al., 2006, 2009; Rao et al., 2007; Parker et al., 2008; Jackson et al., 2010; Lybrand et al., 2013). Recent detections of ClO₄⁻ in Antarctic dry valley soils and lakes, and also on the surface of Mars, have caused additional interest in natural ClO₄⁻ (Hecht et al. 2009; Ming et al. 2010; Catling et al. 2010; Kounaves et al. 2010; Jackson et al., 2013).

The stable isotopic composition of indigenous natural ClO₄⁻ from several locations and environments (vadose zone, surface caliche deposits, groundwater) in the southwestern U.S. has recently been reported (Jackson et al., 2010; Hatzinger et al., 2013). Groundwater ClO₄⁻ samples were obtained from the Southern High Plains (SHP; including one sample from the adjacent rolling plains) of western Texas and eastern New Mexico and from the Middle Rio Grande Basin (MRGB) of central New Mexico. A single sample was obtained from a natural subsurface accumulation of salts within

unsaturated sub-soils at the Range Ecology Research Site at Texas Tech University. Lastly, ClO₄⁻ was collected from near-surface caliche-type salt deposits on clay hills at four locations in the Death Valley region of the Mojave Desert, CA. Clay-hills caliche salts in this area were studied previously because of their unusually high NO₃⁻ concentrations, which resemble those in the Atacama Desert (Ericksen et al., 1983; Böhlke et al., 1997; Lybrand et al., 2013).

All of the indigenous samples collected from groundwater and vadose soils in the SHP and from groundwater in the MRGB were similar isotopically, despite the large areal extent over which they were collected, with δ^{37} Cl values ranging from +3 to +5 %, δ^{18} O values ranging from 0 to +4 ‰, and Δ^{17} O values ranging from +0.3 to +1.3 ‰. The data indicate that ClO₄- from the SHP and MRGB regions of Texas and New Mexico is consistently different from both Atacama ClO₄ and synthetic ClO₄ when all relevant stable isotopic abundances are considered (¹⁶O, ¹⁷O, ¹⁸O, ³⁵Cl and ³⁷Cl) (Figure 1.2). Similar isotopic characteristics to those of the SHP and MRGB samples were also recently observed for presumably indigenous ClO₄- from the U.S. Great Lakes (lake water) and from the Umatilla Basin, Oregon (groundwater), although the Δ^{17} O values were slightly higher for some of these samples, ranging to +2.9 % in Umatilla and +2.7 ‰ in the Great Lakes (Hatzinger et al., 2013; Poghosyan et al., 2014). The smaller, but significantly positive, Δ^{17} O values of indigenous groundwater ClO₄ in comparison to Atacama ClO₄-, could indicate either (1) indigenous groundwater ClO₄- is formed predominantly by a different mechanism than Atacama ClO₄-, or (2) the indigenous ClO₄initially had higher Δ^{17} O but was affected by post depositional O exchange with groundwater.

ClO₄⁻ samples from the Death Valley caliche deposits, which have ClO₄⁻ concentrations ranging from 0.25 to 1.7 mg/kg (about 1-3 orders of magnitude lower than the Atacama deposits), have isotopic characteristics that differ from those of Atacama and synthetic ClO₄⁻, and from those of SHP and MRGB samples. In comparison to the SHP samples, the Death Valley samples have lower δ^{37} Cl values (from -0.8 to -3.7 ‰) and much higher Δ^{17} O values (+8.6 to +18.4 ‰). As is apparent from Figure 1.2, Δ^{17} O values of Death Valley samples are similar to or, in some instances, higher than those of the Atacama samples, consistent with an important component of atmospheric origin and

relative lack of post depositional exchange of O with groundwater. The SHP/MRGB and Death Valley samples can be considered together as "U.S. indigenous sources" and, even though there are substantial ranges in the individual isotope values, this indigenous grouping remains isotopically distinct from synthetic and Atacama ClO₄⁻.

Given present data, it is not yet possible to fully explain the origin of the observed variations in the stable isotopic composition of natural ClO₄⁻ sources. The data permit the interpretation that natural ClO₄⁻ may have more than one formation mechanism, there may be global variations in the isotopic compositions of precursor compounds, and it may be subject to isotopic modification in the terrestrial environment. Resolving these issues would contribute to understanding atmospheric Cl chemistry, as well as the veracity of the isotopic approach for quantifying ClO₄⁻ sources in the environment. Nevertheless, despite uncertainty about processes responsible for some of the isotopic variations, it appears that natural ClO₄⁻ indigenous to the southwestern U.S. is distinguishable from synthetic ClO₄⁻ and from imported Atacama ClO₄⁻ on the basis of isotopic composition. These differences in isotopic composition may find important applications in resolving questions of ClO₄⁻ source apportionment for affected water supplies, such as those in the Study Area.

1.4 Abundance of ³⁶Cl in Synthetic and Natural Perchlorate

The abundance of the radioactive isotope 36 Cl (reported as atom fraction, or mole fraction, designated as 36 Cl/Cl) has proven to be another important tool for distinguishing ClO₄⁻ sources. Chlorine-36 is present naturally in Cl⁻ in groundwater in the U.S., with 36 Cl/Cl values ranging from $\sim 10 \times 10^{-15}$ near coasts to as high as $\sim 1700 \times 10^{-15}$ in the central Rocky Mountains (Bentley et al., 1986; Phillips, 2000; Davis et al., 2003). Isotopic abundances of 36 Cl measured in 35 different ClO₄⁻ samples from synthetic, Atacama, and southwestern U.S. sources ranged over more than four orders of magnitude (Figure 1.3 and Sturchio et al., 2009). Synthetic ClO₄⁻ samples were characterized by relatively low values of 36 Cl/Cl from 0×10^{-15} to 40×10^{-15} (Sturchio et al., 2009). These values are consistent with Cl sources such as the geologically ancient halite-rich evaporite deposits (e.g., salt domes, bedded salts) from which large amounts of NaCl are mined commercially in the form of rock salt.

In contrast to synthetic samples, all indigenous natural ClO₄ samples from the southwestern U.S. that were tested (including some of the SHP and MRGB groundwater samples and Death Valley deposits described in the previous section and shown in Figure 1.2 with differing δ^{37} Cl and Δ^{17} O values) had unusually high 36 Cl/Cl values, ranging from $3,130 \times 10^{-15}$ to $28,800 \times 10^{-15}$. Groundwater samples with indigenous ClO₄ (and some with mixed indigenous/synthetic ClO₄⁻) from eastern Oregon similarly had elevated ³⁶Cl/Cl values $(4,530 - 15,900 \times 10^{-15})$; Hatzinger et al., 2011) as did samples collected from the U.S. Great Lakes, where the isotopic characteristics of ClO₄- (which were similar to the southwestern U.S.) suggested a dominantly indigenous source (7,400 - 71,200× 10⁻¹⁵). The presence of bomb-generated ³⁶Cl from nuclear tests in the mid 1950's (Phillips, 2000; Davis et al., 2003) could not be ruled out for samples from southwestern U.S. groundwater having detectable tritium, and apparently it is present in indigenous ClO₄ in the upper Great Lakes, which still contain much of the water that was present in the 1950s and 1960s (Poghosyan et al., 2014). However, 36 Cl/Cl values as high as $8,400 \times 10^{-15}$ were measured in ClO₄ from old groundwater in New Mexico with estimated recharge ages of > 5,000 yr (Plummer et al., 2006; Sturchio et al., 2009). High concentrations of ³⁶Cl in some ClO₄⁻ samples may point toward the stratosphere, rather than the troposphere, as an important area of atmospheric ClO₄⁻ formation (Sturchio et al., 2009).

Natural ClO₄⁻ from Atacama deposits had 36 Cl/Cl ratios from 22×10^{-15} to 590×10^{-15} , much lower than any of the natural samples from the southwestern U.S., Great Lakes, or Umatilla area. The Atacama samples may have had higher initial ³⁶Cl/Cl ratios that decreased via radioactive decay. This could be consistent with a relatively long history of Atacama ClO₄⁻ accumulation, as hyper-arid conditions in this region have persisted for at least 3 to 8 million years (Alpers and Brimhall, 1988; Hartley and Chong, 2002), which is 10 or more times the 301,000-yr half-life of ³⁶Cl. The ³⁶Cl/Cl ratios in Cl⁻ from the Atacama Desert are similar to those of the coexisting ClO₄ as shown in Figure 1.3, which indicates that ³⁶Cl in these samples may be near radioactive equilibrium with its environment. The accumulation time of ClO₄ in the arid southwestern U.S. appears to have been much shorter (of the order of 10⁴ yr) than that in the Atacama Desert, and natural ClO₄ accumulated from the atmosphere during the past 10⁴ years would still possess most of its initial ³⁶Cl activity (Jackson et al., 2010). More importantly for forensic studies of ClO₄, ³⁶Cl abundances in combination with stable isotope ratios of O and Cl help differentiate synthetic, Atacama, and indigenous U.S. ClO₄ sources (Sturchio et al., 2009; Hatzinger et al., 2011).

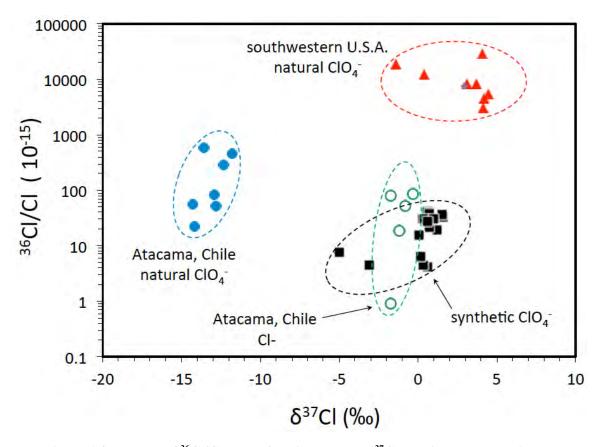


Figure 1.3. Values of 36 Cl/Cl (mole fraction) versus δ^{37} Cl (‰) in representative samples of synthetic ClO₄⁻ reagents and products, natural ClO₄⁻ extracted from soil and groundwater from the Atacama Desert, Chile, and natural ClO₄⁻ extracted from groundwater and soil from the southwestern U.S. (modified from Sturchio et al., 2009).

2.0 Materials and Methods

2.1 Selection of Wells for Groundwater Sample Collection

In the early phase of this project, U.S. Geological Survey (USGS) personnel in San Diego, CA compiled a GIS database of groundwater wells in the Study Area using information from numerous sources, including the California Department of Public Health, USGS National Water Information System, U.S. Environmental Protection Agency (EPA), County of San Bernardino, and various local water agencies and consultants. From this database, a preliminary group of 27 groundwater wells in the Rialto-Colton and Chino subbasins was selected by the project team for isotopic sampling, and 16 wells were selected as possible alternates. North-south and east-west transects across the two synthetic plumes were used as a basis for well selection. These transects included wells within each of the two defined plume areas, and wells that were upgradient (northwest), downgradient (southeast), and crossgradient (northeast and southwest) of the mapped plume areas. Wells were also selected on each side of, and in close proximity to, the Rialto-Colton Fault, and between the Rialto-Colton Fault and the Rialto-Colton Fault-West. Wells selected for sample collection were reviewed by a Technical Advisory Committee (TAC) established for the project which included local water purveyors, regulatory agencies, and other stakeholders. Substitutions of wells were made at the suggestion of the TAC and in the field, where necessary to meet the objectives of this study, when initial well selections were precluded based on well conditions and/or well accessibility, low ClO₄ concentrations, or other issues observed in the field.

A total of 27 wells (17 production wells, 8 monitoring wells, and 2 Westbay wells) were selected and sampled for ClO₄⁻ isotopes (Figure 2.1). In addition, in conjunction with coupled well bore-flow and depth-dependent water quality sampling (Izbicki et al., 2014), depth-dependent ClO₄⁻ stable isotope samples were collected from five of the production wells (Rialto 2, Rialto 3, Rialto 6, F-17B, F-26A). Additional details on the study area hydrogeology and well characteristics are provided in a previously published paper (Izbicki et al., 2014).

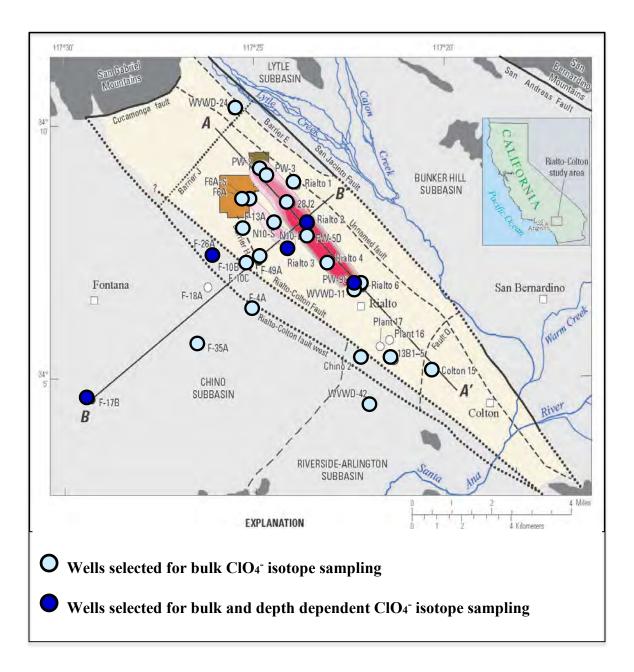


Figure 2.1. Wells selected for bulk discharge ClO₄ isotope sampling (light blue symbols) and depth-dependent ClO₄ isotope sampling (dark blue symbols). Depth-dependent well bore flow and water quality analysis was also performed on wells F-10C and F-49A (see Izbicki et al., 2014). The map does not include concentration data from this study, the most recent USEPA site sampling (USEPA, 2014) or Izbicki et al. (2014). See Figure 1.1 for explanation of map.

2.2 Groundwater Sampling Procedure

Approximately 20 µmol of pure ClO₄ salt (as KClO₄, prepared as described in Section 2.3) is normally sufficient to obtain duplicate stable isotope ratio measurements by IRMS for both O and Cl in the ClO₄-ion, as well as a measurement of ³⁶Cl isotopic However, because of potential losses during sample collection and abundance. purification (Section 2.3), and allowing sufficient sample for additional replicate analyses, it is generally desirable to collect samples containing at least 100 µmol (i.e. 10 mg) of ClO₄⁻ from each source. For groundwaters having low ClO₄⁻ concentrations (i.e., < 5 μg/L), shipping enough water to obtain 10 mg of ClO₄ is impractical. Instead, small columns containing about 100 mL (~ 60 g dry wt) of Purolite A-530E ClO₄-specific anion exchange resin (IX resin) have been developed to preconcentrate ClO₄- from large volumes of water in the field. Water is passed through these columns, and the ClO₄ is trapped by the IX resin. Although many hours may be required for sample collection in some cases, the columns are capable of trapping 10 mg of ClO₄ on a small volume of resin. The ClO₄ is then extracted from the resin and purified prior to IRMS analyses, as described in Section 2.3.

For this project, the USGS performed all groundwater sampling, as described elsewhere (Izbicki et al., 2014). Resin columns were used for sample collection from all wells, except the Westbay wells for which aqueous ClO_4^- concentrations were high enough for water to be collected and shipped directly to the laboratory without using the IX column collection technique. All IX columns for sampling were prepared at the Environmental Isotope Geochemistry Laboratory (EIGL) of the University of Illinois at Chicago (UIC) and sent directly to the USGS with unique identification numbers. In many instances, two or more (up to 8) columns were set up at a single well to collect enough ClO_4^- for stable isotope analysis from low-concentration waters, and in some cases to provide independent replicate samples for isotopic analysis so that method reproducibility could be evaluated. In some cases, groundwater was passed through columns for as long as two weeks to accumulate sufficient ClO_4^- for analysis. The original goal of this project was to conduct duplicate analyses (i.e, process and analyze independent sample columns) on 25% of the wells sampled. This goal was exceeded, as

replicate samples were obtained for 10 of the 33 well locations (wells plus depthdependent samples).

Most production wells were sampled from the surface discharge of the existing well pump. Depth-dependent samples were collected from selected depths within five production wells under pumping conditions (see Izbicki et al., 2014). Sample depths were identified on the basis of available geologic and geophysical logs and unpumped and pumped well-bore flow logs collected as part of the study. Samples from those wells also were collected from the surface discharge of a temporary pump installed within the well after the production pump was removed. Monitoring wells were sampled using a positive displacement gas-reciprocating pump (Bennett pump) capable of lifting water from depths greater than 200 m below land surface (bls). Production and monitoring wells were purged to remove at least three casing volumes prior to sample collection. Field parameters (pH, temperature, and specific conductance) were monitored during purging using a thermometer and portable meters. After field parameters stabilized, groundwater samples were collected for measurement of major ion concentrations (including ClO₄⁻ and NO₃⁻), groundwater dating parameters, dissolved gases, and δ¹⁸O and δ^2 H in water, as described in Izbicki et al. (2014). Samples were collected from Westbay installations using equipment designed by the manufacturer for this purpose.

At the conclusion of all required water quality sampling, one (or more) ion exchange column(s) was connected to the groundwater discharge from each well pump. The flow to the columns was generally set at 1 to 2.5 liters per minute (LPM). Periodically, samples were collected from the influent and effluent of the columns to evaluate stability of influent concentration with time, to detect column breakthrough of ClO₄-, and to estimate the total amount of ClO₄- trapped on each column. After a predetermined time, ranging from a few hours to as long as two weeks (based on flow rate and starting ClO₄- concentration in each well), the IX columns were removed from the each well, sealed in Zip-Loc type bags, placed at 4°C (or on ice), and shipped to the UIC EIGL for ClO₄- extraction and purification.

2.3 Sample Purification and Isotopic Analysis

2.3.1 Sample purification

The preparation of ClO₄⁻ present on the IX resin for chlorine and oxygen isotope ratio analysis involved two main procedures prior to IRMS: (1) elution and recovery of ClO₄⁻ from the resin and (2) separation of ClO₄⁻ from other materials trapped on the IX resin, including a variety of anions and organics. The sample elution and purification procedures were conducted by laboratory technicians under the supervision of Dr. Neil Sturchio at UIC EIGL. Briefly, the key steps in the extraction and purification method were as follows: (1) the resin was washed ultrasonically with deionized water (DIW) and flushed with several pore volumes of 4M HCl to remove adsorbed SO₄²⁻, NO₃⁻, HCO₃⁻, and some of the humics, but not ClO₄⁻ which is more strongly held by the bifunctional resin than most other anions or organics; (2) ClO₄⁻ was eluted from the resin bed using 1 M FeCl₃-4M HCl solution (Gu et al., 2001, 2011); (3) eluted ClO₄⁻ was purified by a series of cation exchange, oxidation, and evaporation steps; (4) KOH was added to precipitate KClO₄ for isotopic analysis; and (5) KClO₄ crystals were washed with methanol and their purity was measured by ion chromatography (IC).

Tetrachloroferrate (FeCl₄⁻) ions are present in the 1 M FeCl₃ and 4 M HCl eluant solution and these ions remove ClO₄⁻ from the Purolite A530E resin (Gu et al., 2001). The FeCl₃-HCl eluent containing dissolved ClO₄⁻ was diluted to convert FeCl₄⁻ to cationic Fe³⁺ species, then it was passed through a large glass chromatography column packed with AG-50W-X12 resin to remove Fe. The eluent from this column was collected and evaporated on a hot plate, with addition of several mL concentrated H₂O₂ to oxidize organic compounds, until the remaining solution volume was reduced to about 50 mL. The evaporation produces an azeotropic mixture (~6 M HCl) so that most of the excess HCl was removed by evaporation. To separate residual NO₃⁻ from ClO₄⁻, the concentrated sample was reloaded onto 1 mL of A530E resin, flushed with 4 M HCl, reeluted with FeCl₃-HCl, and the resulting eluent evaporated to <5 mL. The remaining excess dissolved Cl⁻ (HCl) was then removed by passing the sample through an OnGuard II Ag Sample Pretreatment Cartridge (Dionex, Sunnyvale, California). OnGuard II H (cation-exchange hydronium form resin) cartridges were then used to remove any dissolved Ag. Residual organics were removed by oxidation with H₂O₂ or by Strata

SDB-L (Phenomenex, Torrance, California) solid phase extractant.

The ClO₄⁻ in solution was then precipitated with an excess of KOH to produce KClO₄. Residual KOH was removed from KClO₄ by rinsing the residue with methanol. The KClO₄ crystals were then collected by filtration and rinsed with a few drops of 90 % methanol (by volume, in water). The purified KClO₄ was then dissolved in DIW and a small aliquot was analyzed by IC to assess purity, with a target of <1% anionic impurities. If necessary to achieve target purity, samples were further treated by addition of tetrapentylammonium bromide (TPABr) to precipitate ClO₄⁻ as TPAClO₄, followed by ethanol washing to remove excess TPABr, followed by reaction of TPAClO₄ with KOHmethanol solution to reprecipitate KClO₄ (Poghosyan et al., 2014). The final KClO₄ product was rinsed with ethanol and redissolved in DIW for IC analysis of purity. Mass balance of O during sample decomposition for O isotope analysis was evaluated subsequently to further assess sample purity (i.e., to determine if the sample yielded more or less O than expected from pure KClO₄). In some samples, presence of C or N was also evaluated as part of the IRMS analysis (See Section 2.3.2).

2.3.2 Analysis of Oxygen Isotope Ratios in Perchlorate by IRMS

After the KClO₄ was purified according to the procedures described in the previous section, it was analyzed for O and Cl isotopic composition by IRMS according to the procedures described in this section for O and in Section 2.3.3 for Cl. Analysis of O isotopes in ClO₄⁻ was conducted by three different methods when sufficient sample was available. These methods were (1) Off-line conversion to O₂ with dual-inlet isotoperatio mass spectrometry using a liquid nitrogen trap (O₂-DI-IRMS(+N₂)); (2) Off-line conversion to O₂ with dual-inlet isotope-ratio mass spectrometry without a liquid nitrogen trap (O₂-DI-IRMS((-N₂)); and (3) On-line conversion to CO, with continuous-flow isotope-ratio mass spectrometry (CO-CF-IRMS). The O₂-DI-IRMS method yielded values of δ^{18} O, δ^{17} O, and Δ^{17} O in O₂ gas. A liquid N₂ trap (+N₂) minimized potential interferences caused by condensable gases (e.g., CO₂ and H₂O) entering the IRMS with the O₂ sample; thus O₂-DI-IRMS(+N₂) was the primary method applied to all samples. Performed without a liquid N₂ trap (-N₂), this method permitted condensable gases to enter the IRMS where they could be ionized and measured. The O₂-DI-IRMS(-N₂)

method was used to evaluate contaminants and co-products of ClO_4^- decomposition. The CO-CF-IRMS method yielded values of $\delta^{18}O$, but not $\delta^{17}O$ or $\Delta^{17}O$. This method included measurement of the relative amount of N_2 as an indication of potential contamination by air or N-O salts such as NO_3^- and it was used selectively as an independent test for major problems with the O_2 methods. The methods are described in more detail in sections 2.3.2.1 and 2.3.2.2.

All samples were analyzed at least once by O₂-DI-IRMS(+N₂), and many samples were analyzed multiple times. The data from this technique were used for comparisons provided in the *Results and Discussion* section, and generally for the stable isotope data published previously by this group (e.g., Böhlke et al., 2009; Jackson et al., 2010). Most samples were also analyzed by O₂-DI-IRMS(-N₂) to determine approximate relative susceptibility of O isotopes to C contamination effects. Relatively few samples were analyzed by CO-CFIRMS, as this method was not compatible with Cl isotopic analysis. Data from all three techniques are summarized in Section 3.4 (Table 3.7; Figure 3.10).

2.3.2.1 Analysis of $\delta^{18}O$ and $\delta^{17}O$ by Off-Line Conversion to O_2 , with Dual-Inlet IRMS (O2-DI-IRMS)

To perform DI-IRMS on O_2 derived from ClO_4 , aliquots of pure KClO₄ were weighed into quartz glass tubes (mass equivalent to 2.5 µmol of ClO_4 per tube). The tubes were evacuated and sealed with a torch and then baked at 650 °C for 20 min to produce O_2 gas from the ClO_4 . The tubes were broken manually in an evacuated tube cracker, and the O_2 expanded into a liquid N_2 cold trap for 1 min to remove traces of condensable gases (O_2 -DI-IRMS (+ N_2)) or no N_2 trap was used (O_2 -DI-IRMS (- N_2)) if the effects of condensable gases were under study. The O_2 was then admitted to an IRMS and analyzed in dual-inlet mode against an O_2 reference gas from a tank by monitoring m/z 32 ($^{16}O^{16}O$), 33 ($^{17}O^{16}O$), and 34 ($^{18}O^{16}O$, plus an insignificant contribution from $^{17}O^{17}O$). Typical reproducibility of $\delta^{18}O$ measurements by O_2 -DI-IRMS on multiple aliquots of KClO₄ (typically three to four) in a given batch is generally around $\pm 0.2\%$ or better for the reference material USGS38 (Hatzinger et al., 2011).

2.3.2.2 Analysis of $\delta^{18}O$ by On-Line Conversion to CO, with Continuous-Flow IRMS (CO-CF-IRMS)

To perform CF-IRMS on CO derived from ClO_4^- , aliquots of pure KClO₄ were weighed into silver foil cups (mass equivalent to 2 µmol of ClO_4^- per cup). The loaded cups were dropped automatically from a He-flushed carousel into a graphite crucible in a glassy carbon reactor at a nominal (gauge) temperature of 1325°C to produce CO gas from the ClO_4^- . The CO was transferred in He carrier gas through a molecular-sieve gas chromatograph to an IRMS and analyzed in continuous-flow mode by monitoring peaks at m/z 28 ($^{12}C^{16}O$) and 30 ($^{12}C^{18}O$, plus an insignificant contribution from $^{13}C^{17}O$). Typical reproducibility of $\delta^{18}O$ measurements by CO-CF-IRMS on multiple aliquots (typically four to eight) in a given batch is generally around $\pm 0.2\%$ or better for USGS37 and $\pm 0.2\%$ or better for USGS38 (Hatzinger et al., 2011).

2.3.3 Analysis of Chlorine Isotope Ratios in Perchlorate by IRMS

The analysis of Cl stable isotope ratios was conducted on the KCl produced by thermal decomposition of KClO₄ as described in Section 2.3.2.2 for production of O₂. The Cl⁻ derived from ClO₄⁻ was converted to methyl chloride (CH₃Cl) gas (Eggenkamp, 1994; Holt et al., 1997), which was then analyzed by IRMS according to the procedures described in this section. The methods and calibrations of Cl isotopic analyses from ClO₄⁻ have been summarized in several papers and book chapters as described previously, and additional details on method procedures and QA/QC are given in Hatzinger et al. (2011).

2.3.3.1 Analysis of $\delta^{37}Cl$ by Off-Line Conversion to Methyl Chloride, with Dual-Inlet IRMS (CH₃Cl-DI-IRMS)

Chlorine isotopic analyses were performed on samples of KCl from decomposition of ClO_4^- salts as described above for preparation of O_2 for isotopic analysis (Section 2.3.2). KCl residue in a decomposition tube was dissolved using 10 mL of warm 18.2 M Ω deionized water. The dissolved alkali halide residue was transferred into a 50-mL polypropylene conical tube and acidified with 100 μ L concentrated nitric acid (HNO₃). This solution was then heated to 80 °C and an excess of silver nitrate

(AgNO₃) was added as described in Eggenkamp (1994). Silver chloride (AgCl) precipitates were then allowed to ripen in a dark cabinet for ~ 24 hr. The AgCl solids were then centrifuged, the supernatant was removed, and 0.03 M HNO₃ was used to rinse the solids three times. Solids were then transferred into a Pyrex combustion tube (20 cm x 9 mm) and dried in a darkened vacuum oven at 80 °C. After the sample was dry, the combustion tube was evacuated and CH₃I was cryogenically transferred into the tube which was then sealed and baked for 2 h at 300 °C as described in Holt et al. (1997). The resulting CH₃Cl was purified using gas chromatography, cryo-concentrated, and then admitted to the IRMS and analyzed in dual-inlet mode by monitoring peaks at m/z 52 (12 Cl 14 H₃ 37 Cl) and 50 (12 Cl 14 H₃ 35 Cl). Typical reproducibility of δ ³⁷Cl measurements on multiple aliquots in a given batch are generally \pm 0.2‰ or better for USGS37 and \pm 0.3‰ or better for USGS38 (Hatzinger et al., 2011).

2.3.4 Standards for Oxygen and Chlorine Stable Isotopic Analysis

Equations 1.1 and 1.2 are expanded to permit routine calibration of ClO₄⁻ analyses using a pair of ClO₄⁻ isotopic reference materials (USGS37 and USGS38) (see description in Hatzinger et al., 2011) with contrasting isotopic compositions on the VSMOW-SLAP scale, a process commonly referred to as "normalization":

$$\begin{split} [\text{Eq. 2.1}] & \delta^{18} O_{i/\text{VSMOW}} = \delta^{18} O_{37/\text{VSMOW}} + \\ & [\delta^{18} O_{i/\text{rg}} - \delta^{18} O_{37/\text{rg}}]_{\text{meas.}} \cdot [\delta^{18} O_{38/\text{VSMOW}} - \delta^{18} O_{37/\text{VSMOW}}] \, / \, [\delta^{18} O_{38/\text{rg}} - \delta^{18} O_{37/\text{rg}}]_{\text{meas.}} \end{split}$$

$$\begin{split} \text{[Eq. 2.2]} \qquad & \delta^{17} O_{i/\text{VSMOW}} \, = \, \delta^{17} O_{37/\text{VSMOW}} \, + \\ & \left[\delta^{17} O_{i/\text{rg}} - \delta^{17} O_{37/\text{rg}} \right]_{meas.} \cdot \left[\delta^{17} O_{38/\text{VSMOW}} - \delta^{17} O_{37/\text{VSMOW}} \right] / \left[\delta^{17} O_{38/\text{rg}} - \delta^{17} O_{37/\text{rg}} \right]_{meas.}, \end{split}$$

where 37 and 38 refer to the ClO₄⁻ isotopic reference materials USGS37 and USGS38, and rg is an internal laboratory reference gas (either CO or O₂) against which all samples and reference materials are analyzed in the mass spectrometer during a single batch of analyses.

The isotopic reference materials consist of reagent-grade KClO₄ salts that were prepared specifically for calibration of ClO₄⁻ isotopic analyses. The δ ¹⁸O scale is based

on CO-CF-IRMS analyses of the CIO₄⁻ isotopic reference materials against international H₂O, NO₃⁻, and SO₄²- isotopic reference materials as described by Böhlke et al. (2003), and all data are referenced to the conventional VSMOW-SLAP scale (Gonfiantini, 1978; Coplen, 1994). For δ^{18} O, the secondary calibration values used to generate provisional ClO₄⁻ data with respect to VSMOW are -27.9 % for USGS34 (KNO₃), +25.6 % for IAEA-N3 (KNO₃), +57.5 % for USGS35 (NaNO₃), and +8.6 % for NBS 127 (BaSO₄) (Böhlke et al., 2003). The δ^{17} O scale for ClO₄⁻ is provisionally based on the assumption that the normal reagent KClO₄ reference material (USGS37) has $R(^{17}\text{O}/^{16}\text{O})$ and $R(^{18}\text{O}/^{16}\text{O})$ values that are related to those of VSMOW by mass-dependent processes ($\Delta^{17}\text{O} = 0$ as defined by Equation 1.4; Böhlke et al., 2005). Perchlorate calibration values used for data normalization in this study are consistent with those reported previously (Böhlke et al., 2009): for USGS37, $\delta^{18}\text{O} = -17.00$ % and $\delta^{17}\text{O} = -8.96$ %; for USGS38, $\delta^{18}\text{O} = +52.5$ % and $\delta^{17}\text{O} = +102.5$ %. For USGS37, our $\Delta^{17}\text{O}$ value of 0.0 % defined by Equation 1.4 is in agreement with a mean value of -0.11 ± 0.06 % reported for synthetic ClO₄⁻ using an alternative $\Delta^{17}\text{O}$ definition (Bao and Gu, 2004).

The most widely-used Cl isotope reference material is chloride prepared from seawater which has uniform δ^{37} Cl to within \pm 0.08 ‰ (Godon et al., 2004). Routine calibration of ClO₄⁻ isotopic analyses was also conducted by using the pair of ClO₄⁻ isotopic reference materials (USGS37 and USGS38) as follows:

$$\begin{split} \text{[Eq. 2.3]} \qquad \qquad & \delta^{37}\text{Cl}_{\text{ i/SMOC}} = \delta^{37}\text{Cl}_{\text{ 37/SMOC}} + \\ & [\delta^{37}\text{Cl}_{\text{ i/rg}} - \delta^{37}\text{Cl}_{\text{ 37/rg}}]_{\text{meas.}} \cdot \left[\delta^{37}\text{Cl}_{\text{ 38/SMOC}} - \delta^{37}\text{Cl}_{\text{ 37/SMOC}}\right] / \left[\delta^{37}\text{Cl}_{\text{ 38/rg}} - \delta^{37}\text{Cl}_{\text{ 37/rg}}\right]_{\text{meas.}} \end{split}$$

where, as with O isotope analysis, 37 and 38 refer to the perchlorate isotopic reference materials USGS37 and USG38, and rg is an internal laboratory reference gas (CH₃Cl) against which all samples and reference materials are analyzed in the mass spectrometer during a single batch of analyses. Perchlorate calibration values used in this study are (Böhlke et al., 2009): for USGS37, δ^{37} Cl = +0.6 %; for USGS38, δ^{37} Cl = -87.2 %.

2.3.5 Analysis of ³⁶Cl in ClO₄

Analysis of ³⁶Cl in ClO₄⁻ was performed by accelerator mass spectrometry (AMS) using Cl⁻ derived from ClO₄⁻. The ³⁶Cl analyses on ClO₄⁻ collected during this project were performed at the Purdue Rare Isotope Measurement Laboratory (PRIME) at Purdue University (www.physics.purdue.edu/ primelab). The procedure used to produce Cl⁻ for AMS is as described in Section 2.3.3 of this document to the point of AgCl precipitation, washing, and drying of crystals (i.e., prior to the reaction with CH₃I). A portion of the sample prepared to this step for Cl stable isotopic analysis was saved for ³⁶Cl analysis. The AgCl was subsequently re-dissolved and the Cl⁻ purified twice by anion chromatography (using a method developed by the PRIME Lab at Purdue University; http://www.physics.purdue.edu/primelab/AMSQAQC/chemProc004.pdf) to ensure removal of trace amounts of S that might cause isobaric interference at mass 36. Purified Cl⁻ is then re-precipitated as AgCl for AMS measurement. Analysis of seawater Cl⁻ provides a reference datum of ³⁶Cl/Cl with a value of 0.5 × 10⁻¹⁵ (Argento et al., 2010).

3.0 Results and Discussion

3.1 ClO₄- Concentrations in Groundwater Wells

Measured concentrations of ClO_4^- in Study Area groundwater ranged from 0.3 $\mu g/L$ to 1,150 $\mu g/L$ in wells that were sampled for ClO_4^- isotopes (Figure 3.1; Table 3.1). With the exception of Well F-17B in the Chino subbasin, all of the wells with concentrations above the California MCL of 6 $\mu g/L$ were located within the two mapped ClO_4^- plumes (Figure 3.1). In general, the average influent concentrations from the IX columns used to collect ClO_4^- for isotopic analysis were similar to the individual samples taken for ClO_4^- analysis (Table 3.1). A statistical analysis of the variation of ClO_4^- concentration entering the columns with time is provided in Izbicki et al., (2014). There was no apparent correlation between the average ClO_4^- concentration and the variability in that concentration.

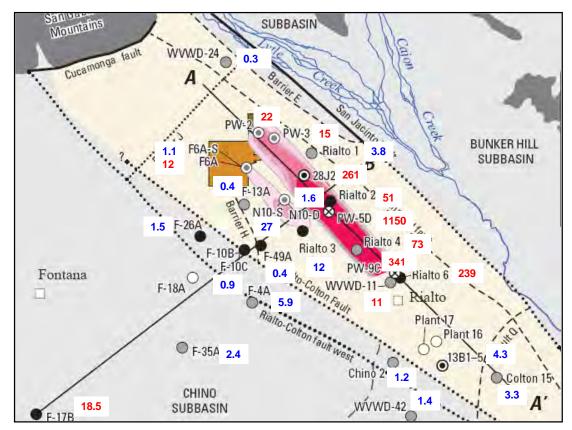


Figure 3.1. Concentration of ClO_4^- in groundwater wells during stable isotope sampling. Values provided are the average influent concentration to IX columns used for groundwater sampling except for F-10C, F-49A, PW-5D and PW-9C, where individual analyses are provided. Concentrations above 6 μ g/L are shown in red, and those below 6 μ g/L are shown in blue. See Figure 1.1 for explanation of the map.

3.2 Isotope Results for ClO₄-

The Cl and O stable isotope results for ClO_4^- in Study Area well samples are provided in dual isotope plots in comparison to previous source data for synthetic, Atacama, and selected indigenous natural ClO_4^- samples (SHP, MRGB, and Death Valley caliche) in Figures 3.2 and 3.3. The isotopic values ($\delta^{37}Cl$, $\delta^{18}O$ $\Delta^{17}Cl$, and $^{36}Cl/Cl$) are plotted with respect to the inverse of the measured ClO_4^- concentrations in Figure 3.4 to illustrate potential mixing and dilution trends. Average isotope values and ClO_4^- concentrations in the column influent are provided in Table 3.1.

In general, it appears that ClO_4^- from all three proposed sources was present in varying proportions in Study Area groundwater (Figures 3.2-3.4). Samples having relatively high ClO_4^- concentrations (values of inverse ClO_4^- concentration < 1, i.e. ClO_4^- concentrations > 1 μ g/L) generally plot near a single dominant ClO_4^- source, either synthetic or Atacama (Figure 3.4), whereas samples having lower concentrations commonly indicate mixtures with varying fractions of indigenous natural ClO_4^- . These relations are consistent with the presence of a low background concentration of indigenous ClO_4^- , toward which the isotopic composition of ClO_4^- converges as fractions of the other two ClO_4^- sources approach zero (Figure 3.4).

To facilitate discussion of results, hypothetical mixing zones among three potential ClO_4^- endmembers (synthetic, Atacama, indigenous) are plotted with the Study Area stable isotope data ($\delta^{37}Cl$ vs $\delta^{18}O$ and $\Delta^{17}Cl$ vs $\delta^{18}O$) and with ^{36}Cl data ($^{36}Cl/Cl$ vs $\delta^{37}Cl$) in Figure 3.5. A similar three-endmember mixing approach was also recently used to determine the dominant source of ClO_4^- in groundwater of the nearby location of Pomona, CA, using $\delta^{37}Cl$, $\delta^{18}O$, and $\Delta^{17}O$ data (^{36}Cl values were not reported) (Sturchio et al., 2014). Characteristics of three ClO_4^- endmembers for the Study Area groundwaters were determined as follows: (1) The Atacama and synthetic endmembers were obtained by separate linear least-squares regressions of $\delta^{37}Cl$, $\delta^{18}O$, and $\Delta^{17}O$ values vs. inverse ClO_4^- concentrations. This procedure is based on the assumption that all samples may have varying amounts of background ClO_4^- , and the Atacama and synthetic endmembers are most likely to dominate at high concentrations (Sturchio et al., 2014). The best-fit regression parameters (slope and intercept, with 2-sigma errors) for samples having ClO_4^- concentrations exceeding 1 μ g/L were determined using the regression function of Microsoft Excel. This concentration cutoff was based on the fact that samples in this

concentration range could clearly be seen as having either a dominantly Atacama or dominantly synthetic source, based on all three stable isotope values (see Figure 3.4). (2) For the indigenous endmember, no sample of ClO₄⁻ was obtained from the Study Area that was clearly of "pure" indigenous origin, so published data for the SHP and MRGB groundwater ClO₄⁻ samples from Sturchio et al. (2006) and Jackson et al. (2010) were used to define its stable isotopic composition (Sturchio et al., 2014). This represents one possible endmember choice, and other alternatives are also discussed below. (3) Because of relatively large variability in measured ³⁶Cl/Cl ratios (particularly for samples relatively enriched in the indigenous endmember), in contrast with the stable isotope ratios, regression analysis did not yield precise estimates of ³⁶Cl/Cl ratios in the dominant endmembers from the Study Area. Therefore, medians and ranges of published values were used to represent ³⁶Cl/Cl ratios of all three endmembers in mixing calculations (Sturchio et al., 2009; Jackson et al., 2010; Poghosyan et al., 2014). The endmember values used for the analysis are provided in Table 3.2 (mean and 2-sigma errors).

The δ^{37} Cl vs. δ^{18} O, Δ^{17} O vs. δ^{18} O, and 36 Cl/Cl vs. δ^{37} Cl plots of Study Area sample data, in comparison with the three ClO₄⁻ source endmembers defined above, indicate that the proposed three-component mixing is capable of accounting for the observed variations in isotopic composition in nearly all samples, given uncertainties and potential variability of endmember characteristics (Figure 3.5). The mixing regions between the end members (shown as boxes with each side being the limit of 2-sigma error) are indicated by the gray lines in Figure 3.5. The mixing results in Figure 3.5, as well as the Δ^{17} O vs 1/concentration plot in Figure 3.4, indicate that the Δ^{17} O range of the proposed indigenous Study Area endmember is similar to that in the SHP, MRGB, eastern Oregon and Great Lakes samples (i.e., +0.3 to +2.9 ‰) rather than the much higher range observed for Death Valley caliche samples (i.e., +8.6 to +18.4 ‰) (Jackson et al., 2010; Hatzinger et al., 2013; Poghosyan et al., 2014).

To estimate mixing fractions of each of the three endmembers in the Study Area samples, the average values of all stable isotope ratio analyses for each well (from Table 3.1) along with endmember mean values and standard deviations (from Table 3.2) were used as input for the spreadsheet program *IsoError1_04.xls* [described in Phillips and Gregg (2001) and in the instructions accompanying the spreadsheet]. These spreadsheet calculations yielded quantitative endmember fractions for each sample, based on

assumptions given above, accounting for sample measurement uncertainties and population sizes of endmembers. Mixing fractions were calculated with 2-sigma error estimates from three pairs of isotope values, δ^{37} Cl vs. δ^{18} O, Δ^{17} O vs. δ^{18} O, and δ^{18} O vs. δ^{37} Cl. The 95 % confidence limits (maximum and minimum mixing fractions) from these calculations are provided in Table 3.3 to Table 3.5. Mixing fractions calculated from δ^{37} Cl vs. δ^{18} O generally agree with those calculated from Δ^{17} O vs. δ^{18} O; there is a cluster of samples near the Atacama endmember, a cluster of samples near the synthetic endmember, and a few intermediate mixed samples (that are also the samples with the lowest total ClO₄⁻ concentrations). Some calculated maximum values of mixing fractions at a 95 % confidence interval exceeded 1 (with a maximum value of 1.1) and a few of the minimum values were negative at a 95 % confidence limit (with a minimum value of -0.2), reflecting uncertainties in the assumed endmember characteristics. The dominant sources indicated by the 36 Cl/Cl vs δ^{37} Cl mixing analysis were generally consistent with those determined using the stable isotope values (Table 3.5). The data also indicated minor fractions of indigenous ClO₄ in many of the samples as is apparent in Figure 3.5 and discussed in more detail later in this section.

Measured values of δ^{18} O, Δ^{17} O, and δ^{37} Cl were consistent with those of synthetic ClO₄⁻ sources for all sampled wells located in the perched aquifer within the two mapped ClO₄⁻ plumes (28J-2, F-6, F6A-S, N-10S) (Figures 3.3-3.5; Tables 3.1, 3.3, 3.4). Similarly, isotopic data from wells in the regional aquifer within the mapped plumes also were consistent with a dominantly synthetic source (with minimum mixing fractions generally ranging from 60 – 90% and maximum fractions at 100 to 110 % using a 95 % confidence interval), including PW-2, PW-3, Rialto 2, Rialto 2 (880'), PW-5D, Rialto 3, Rialto 3 (670'), PW-9C, Rialto 6, Rialto 6 (480'), Rialto 6 (560'), and Rialto 4) (Table 3.3, 3.4). All of these samples had mean δ³⁷Cl values between -1.0 and +1.5 ‰ and Δ¹⁷O values between -0.1 and + 0.2 ‰, consistent with dominantly synthetic ClO₄⁻. The δ¹⁸O values varied from -12.5 to -20.3 ‰, also within the range previously reported for synthetic ClO₄⁻ sources. However, the relatively wide range in δ¹⁸O for these samples also appears to reflect mixing with small fractions of natural indigenous ClO₄⁻ (see next paragraph) in addition to possible variation in synthetic source isotopic ratios.

Table 3.1. Perchlorate concentrations and Cl and O isotope data for samples collected from wells in the Study Area, June 2010 to February 2012.

11								
well	well		$\delta^{37}\text{CI}$	$\delta^{18}O$	$\Delta^{17}O$	³⁶ CI		Perchlorate
number	name	Date	(‰)	(‰)	(‰)	³⁶ CI/Cl x 10 ⁻¹⁵	(μg/L)	(μg/L)
							single	column influent
	Wells in the ner	ched aguifer in th	e Rialto-Colto	nn subhasin w	ithin the man	ned nlume		
1N/5W-28J2	28J-2	07/01/10	0.8 ± 0.1	-18.3 ± 0.1	-0.1±0.0	102+1	285	261+5
1N/5W-29H1	F6	07/20/10	-0.5	-13.5	0.1	1040 (20)	8.8	12+2
1N/5W-29H3	F6A-S	07/22/10	-0.9	-15.5 ± 0.3	0.2 ± 0.1	NA	0.7	1.1+0.4
1N/5W-33B2	N-10S	07/21/10	1.0 ± 0.0	-12.5 ± 1.7	0.0 ± 0.1	504 (12)	33	27+5
11,511 5552		gional aquifer in th					- 55	_
1N/5W-21N2	PW-2	08/11/10	1.4 ± 0.1	-16.9 ± 0.4	0.0 ± 0.1	118 (4)	22	21+0
1N/5W-21N2	PW-3	08/10/10	1.1 ± 0.2	-18.0 ± 0.2	0.0±0.1 0.1±0.1	192 (17)	15	15 ±0
1N/5W-33B1	N-10D	05/19/11	0.5	-14.1 ± 0.5	0.2 ± 0.0	3640 (200)	1.9	1.6+0.2
1N/5W-33B1	Rialto 02	02/25/12	0.2	-20.3 ± 0.2	0.0±0.0	NA	51	51 ±1
1N/5W-34B2	Rialto 02 (800')	02/25/12	0.3	-20.1 ± 0.1	0.0±0.0	NA NA	35	35 ± 1
1N/5W-34G4	PW-5D	08/04/10	0.7 ± 0.4	-17.6 ± 0.6	0.0±0.1	82 (9)	1150	NA
1N/5W-34M1	Rialto 03	03/15/11	0.6 ± 0.3	-15.2 ± 0.1	0.0 ± 0.2	687 (14)	11	12 ±0
1N/5W-34M1	Rialto 03 (670')	03/15/11	0.6	-14.4 ± 0.1	0.0 ± 0.0	502 (23)	11	11+0
1S/5W-02B3	PW-9C	08/05/10	1.3	-16.1 ± 0.3	0.0 ± 0.1	127 (18)	341	NA NA
1S/5W-02G1	Rialto 06	03/11/11	1.3 ± 0.3	-16.9 ± 0.2	0.1 ± 0.1	86 (4)	226	239+13
1S/5W-02G1	Rialto 06 (480')	03/11/11	1.0	-17.1 ± 0.0	-0.1 ± 0.1	194 (30)	198	206 ± 13
1S/5W-02G1	Rialto 06 (560')	03/11/11	1.1	-15.7 ± 0.2	0.0 ± 0.1	310 (70)	109	115 ± 13
1S/5W-03A1	Rialto 04	10/20/10	1.1 ± 0.1	-17.6 ± 0.2	0.0 ± 0.2	131 (3)	72	73 <u>+</u> 2
	'	Wells	outside the ma	pped plume			ı	-
1N/5W-17K2	WVWD 24	06/29/10	1.5 ± 0.4	-8.4 ± 0.3	1.2 ± 0.1	6021 (110)	0.3*	0.3 <u>+</u> 0.0j
1N/5W-27D1	Rialto 01	05/18/11	-0.2 ± 0.1	-15.9 ± 0.2	0.2 ± 0.0	2660 ± 127	3.7	3.8 <u>+</u> 0.5
1N/5W-33N1	F-49A	04/23/11	-2.4 ± 0.0	-8.5 ± 1.0	2.8 ± 0.2	28,900 (1600)	0.4*	NA
1S/4W-18G1	Colton 15	07/08/10	-12.3 ± 0.7	-4.6 ± 0.1	8.6 ± 0.1	1590 (40)	3.2	3.3 <u>+</u> 0.1
1S/5W-02K1	WVWD 11	06/30/10	-0.5 ± 0.4	-15.3 ± 0.6	1.0 ± 0.2	66 (3)	9.7	11 <u>+</u> 1
1S/5W-05A5	F-10C	04/18/11	-4.3 ± 0.2	-9.8 ± 0.5	3.4 ± 0.1	10,200 (400)	0.9*	NA
1S/5W-13B5	13B1-5 (RHSW 5)	08/12/10	-9.8	-7.1 ± 0.1	6.9 ± 0.0	2210 (160)	4.4	4.2 <u>+</u> 0.1
1N/5W-32A1	F-13A	05/11/11	-5.5 ± 0.2	-4.8 ± 0.0	4.7 ± 0.2	24,200 (1200)	0.4*	0.4 <u>+</u> 0.0j
•		Wells in th	e Chino (or Arli	ngton) subbasins				
1S/5W-23A1	WVWD 42	07/08/10	-10.3 ± 0.1	-7.0 ± 0.0	7.1 ± 0.1	1785 <u>+</u> 134	1.4	1.4 <u>+</u> 0.0
1N/5W-32N1	F-26A	12/12/10	-9.7	-8.3 ± 0.0	6.1 ± 0.0	5450 (130)	2.0	1.5 <u>+</u> 0.2
1N/5W-32N1	F-26A (800')	12/12/10	-9.9	-8.4 ± 0.1	6.2 ± 0.0	5140 (90)	1.6	1.5 <u>+</u> 0.1
1S/5W-04N1	F-4A	07/28/10	-11.8 ± 0.3	-7.9 ± 0.2	7.6 ± 0.1	1380 (30)	5.8	5.9 <u>+</u> 1.7
1S/5W-04N1	F-4A (J613; 2007 sample)	03/09/07	-11.9	-8.0	7.6	NA	12	NA
1S/5W-07R1	F-35A	07/27/10	-11.1	-9.4 ± 0.1	6.4 ± 0.1	415 (15)	0.6	2.4 <u>+</u> 0.4
1S/5W-14B1	Chino 02	07/22/10	-11.1	-6.8 ± 0.2	7.4 ± 0.0	780 (40)	2.1	1.2 <u>+</u> 0.8
1S/6W-23D2	F-17B	06/17/10	-13.5 ± 0.0	-6.6 ± 0.3	8.2 ± 0.1	91 <u>+</u> 4	22	19 <u>+</u> 0.7
1S/6W-23D2	F-17B (580')	06/17/10	-13.0 ± 0.2	-8.0 ± 0.2	8.1 ± 0.1	100 (4)	20	18 <u>+</u> 1

NA - data not available

Means and standard deviations are provided for all individual analyses per well. When individual values are provided samples were analyzed once.

 $[\]delta^{18}\text{O}$ and $\Delta^{17}\text{O}$ were analyzed by O_2-DI-IRMS(+N_2). See Equation 1.4 for definition of $\Delta^{17}\text{O}.$

j - estmated value between the MDL and PQL by IC $\,$

^{*} Analysis by IC-MS at Texas Tech University

[&]quot;Column influent" samples were collected intermittently over a period of hours to days while isotope samples were being loaded onto columns.

[&]quot;Single" samples were collected once when samples for other chemical and isotopic analyses were collected (Izbicki et al., 2014).

 $For \ ^{36}Cl, values in parentheses are \ laboratoy reported uncertainty for individual samples while \ +/- \ values are standard \ deviations for multiple analyses.$

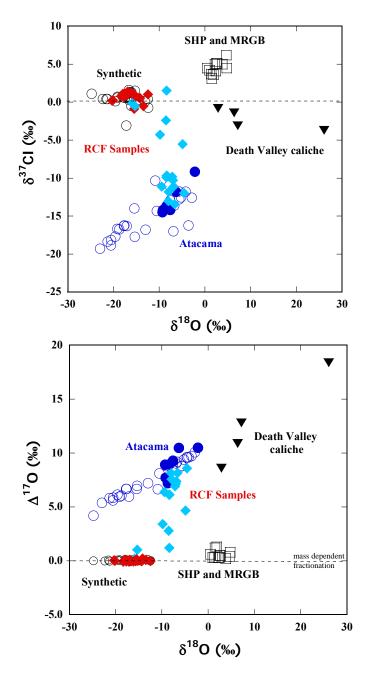


Figure 3.2. Comparison of δ³⁷Cl versus δ¹⁸O (top panel) and Δ¹⁷O versus δ¹⁸O (bottom panel) in ClO₄⁻ from Study Area groundwater samples (red closed diamonds indicate wells located within the mapped ClO₄⁻ plume areas and light blue closed diamonds indicate wells located outside the mapped plume areas; Figure 1.1; Table 3.1; Izbicki et al., 2014) with data for synthetic ClO₄⁻ (open black circles), natural Atacama ClO₄⁻ (open blue circles for field samples and closed blue circles for Atacama nitrate fertilizer samples), and indigenous natural ClO₄⁻ of the southwestern U.S. including Southern High Plains (SHP) and Middle Rio Grande Basin (MRGB) groundwater (black open squares) and Death Valley caliche (black triangles). ClO₄⁻ isotope data from Bao and Gu., 2004; Böhlke et al., 2005; Sturchio et al., 2006, 2012; Böhlke et al., 2009; Jackson et al., 2010; Hatzinger et al., 2011, 2013; unpublished results from SERDP Project ER-1435 for a subset of Atacama field samples. See Section 1.3 for source details.

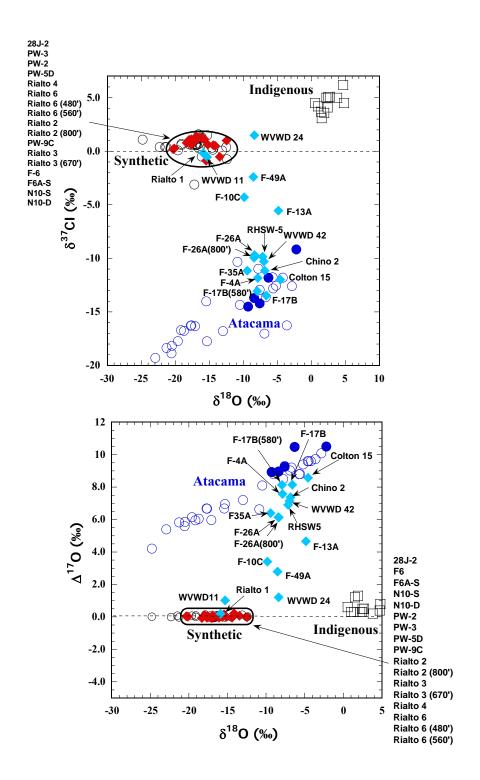


Figure 3.3. Comparison of δ^{37} Cl versus δ^{18} O (top panel) and Δ^{17} O versus δ^{18} O (bottom panel) in ClO₄⁻ from Study Area groundwater samples within the mapped plume areas (red closed diamonds with well designations) and outside the mapped plume areas (light blue closed diamonds) with source data (symbols as described in the caption to Figure 3.2). Sources of ClO₄⁻ isotope data for Atacama, Synthetic and Indigenous as described in Figure 3.2.

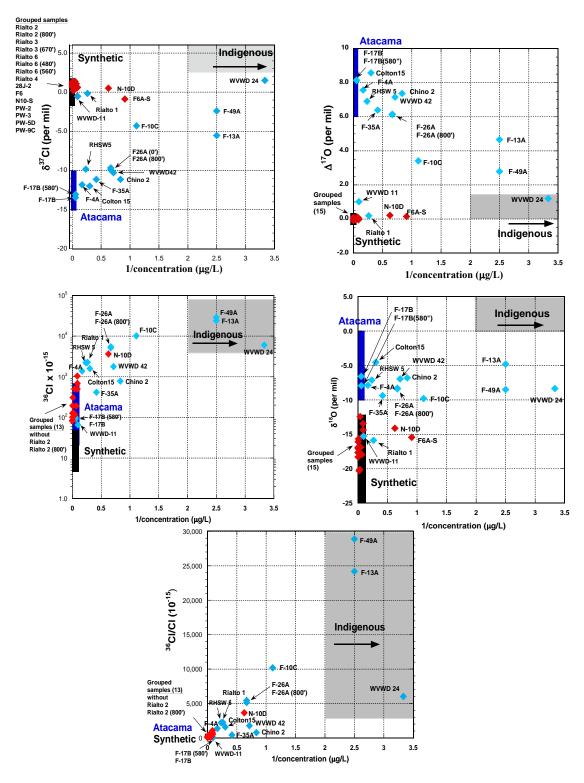


Figure 3.4. Comparison of $\delta^{37}\text{Cl}$ versus $1/\text{ClO}_4^-$ concentration (top left panel), $\Delta^{17}\text{O}$ versus $1/\text{ClO}_4^-$ concentration (top right panel), ${}^{36}\text{Cl/Cl}$ (x 10^{-15}) on log scale versus $1/\text{ClO}_4^-$ concentration (middle left panel), $\delta^{18}\text{O}$ versus $1/\text{ClO}_4^-$ concentration (middle right panel) and ${}^{36}\text{Cl/Cl}$ (x 10^{-15}) on linear scale versus $1/\text{ClO}_4^-$ concentration (bottom middle panel), with symbols as described in the caption to Figure 3.2. The solid bars (including the wide gray bar for indigenous ClO_4^-) represent typical values for the different sources (see text for explanation). The concentration of indigenous ClO_4^- may be lower than indicated on each panel as represented by the arrow. Grouped samples include the list in the upper left unless otherwise noted.

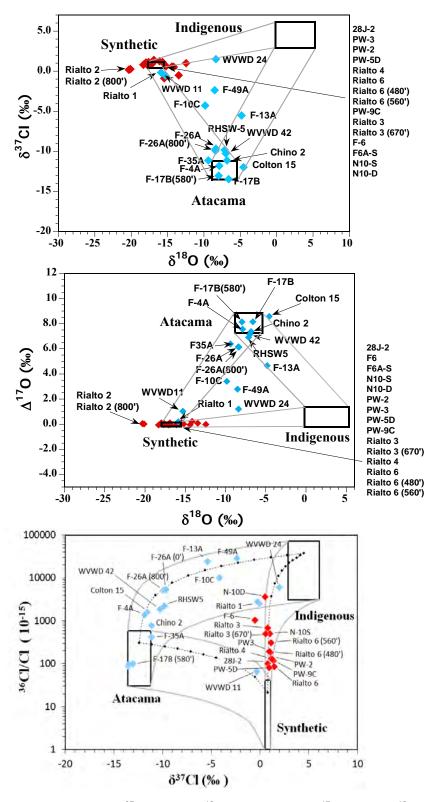


Figure 3.5. Comparison of δ^{37} Cl versus δ^{18} O (top panel); Δ^{17} O versus δ^{18} O (middle panel) and δ^{37} Cl versus δ^{16} Cl/Cl (bottom panel) in ClO₄- from Study Area groundwater samples (symbols as in Figure 3.2). The sides of each box represent the endmember value \pm 2-sigma error and the gray lines represent mixing zones between the endmember regions. The endmember values were determined as detailed in the accompanying text and are provided in Table 3.2. The dotted lines in the bottom panel are hypothetical mixing curves between the three endmembers with tick marks shown every 10% as a function of increasing 36 Cl/Cl ratio in the mixture (note log scale of y-axis).

Table 3.2. Isotopic compositions of hypothetical endmembers used for analysis of dominant sources in Study Area groundwater. The values in parentheses are 2-sigma errors based on regression analysis of local data (δ^{37} Cl, δ^{18} O, Δ^{17} O for Atacama and synthetic endmembers) or based on published data (δ^{37} Cl, δ^{18} O, Δ^{17} O for the indigenous endmember and 36 Cl/Cl for all endmembers). See text for explanation and sources of data.

Isotope parameter	Atacama ClO ₄ -	Synthetic ClO ₄ -	Indigenous ClO ₄ -
δ ³⁷ Cl (‰)	-12.4 (1.2)	0.8 (0.3)	4.5 (1.6)
δ ¹⁸ O (‰)	-7.2 (1.7)	-16.7 (1.1)	2.5 (2.8)
△ ¹⁷ O (‰)	8.1 (0.8)	0.1 (0.1)	0.6 (0.8)
³⁶ Cl/Cl x 10 ⁻¹⁵	310 (280)	21 (20)	37,200 (34,000)

Table 3.3. Mixing fractions of Atacama, Synthetic, and Indigenous perchlorate based on $\delta^{37}Cl$ vs $\delta^{18}O$ stable isotope data for samples collected from wells in the Study Area, June 2010 to February 2012. Values provided are the 95% confidence limits calculated using IsoError1_04.xls with as described in Phillips and Gregg, (2001). Endmember values are given in Table 3.2.

State	Local	Atacar	ma	Synt	hetic	Indigenous		
well	well	mixing fra	action	mixing fraction		mixing fraction		
number	name	95% confider		95% confidence limits		95% confidence limits		
		maximum	minimum	maximum	minimum	maximum	minimum	
	We	Ils in the perched	aguifer in the	Rialto-Colton	subbasin wit	hin the mapped	plume	
1N/5W-28J2	28J-2	0.2	0.0	1.0	0.8	0.2	0.0	
1N/5W-29H1	F6	0.3	-0.1	1.1	0.5	0.3	-0.1	
1N/5W-29H3	F6A-S	0.3	-0.1	1.0	0.6	0.2	0.0	
1N/5W-33B2	N-10S	0.2	0.0	1.1	0.5	0.4	0.0	
	We	ells in the Regiona	l aquifer in th	e Rialto-Coltor	subbasin wit	hin the mapped	l plume	
1N/5W-21N2	PW-2	0.1	-0.1	1.1	0.7	0.2	0.0	
1N/5W-21P2	PW-3	0.2	0.0	1.0	0.8	0.2	0.0	
1N/5W-33B1	N-10D	0.2	0.0	1.1	0.5	0.3	-0.1	
1N/5W-34B2	Rialto 02	0.2	0.0	1.1	0.9	0.0	-0.2	
1N/5W-34B2	Rialto 02 (800')	0.2	0.0	1.1	0.9	0.0	-0.2	
1N/5W-34G4	PW-5D	0.2	0.0	1.0	0.8	0.2	0.0	
1N/5W-34M1	Rialto 03	0.2	0.0	1.0	0.6	0.3	-0.1	
1N/5W-34M1	Rialto 03 (670')	0.2	0.0	1.0	0.6	0.3	-0.1	
1S/5W-02B3	PW-9C	0.2	0.0	1.1	0.7	0.3	-0.1	
1S/5W-02G1	Rialto 06	0.2	0.0	1.1	0.7	0.2	0.0	
1S/5W-02G1	Rialto 06 (480')	0.2	0.0	1.1	0.7	0.2	0.0	
1S/5W-02G1	Rialto 06 (560')	0.2	0.0	1.1	0.7	0.3	-0.1	
1S/5W-03A1	Rialto 04	0.2	0.0	1.0	0.8	0.2	0.0	
		We	ells outside th	e mapped plu	me		•	
1N/5W-17K2	WVWD 24	0.3	-0.1	0.8	0.2	0.5	0.3	
1N/5W-27D1	Rialto 01	0.3	-0.1	1.0	0.6	0.2	0.0	
1N/5W-33N1	F-49A	0.5	0.1	0.7	0.1	0.5	0.1	
1S/4W-18G1	Colton 15	1.1	0.9	0.1	-0.1	0.1	-0.1	
1S/5W-02K1	WVWD 11	0.2	0.0	1.0	0.6	0.3	-0.1	
1S/5W-05A5	F-10C	0.6	0.2	0.7	0.1	0.4	0.0	
1S/5W-13B5	13B1-5 (RHSW 5)	0.9	0.7	0.2	0.0	0.2	0.0	
1N/5W-32A1	F-13A	0.7	0.5	0.2	0.0	0.4	0.2	
		Wells in	the Chino (or	Arlington) sul	<u>obasins</u>		,	
1N/5W-32N1	F-26A	0.9	0.7	0.3	0.1	0.1	-0.1	
1N/5W-32N1	F-26A (800')	0.9	0.7	0.3	0.1	0.1	-0.1	
1S/5W-04N1	F-4A	1.1	0.9	0.2	0.0	0.1	-0.1	
1S/5W-04N1	F-4A (J613; 2007 sample)	1.1	0.9	0.2	0.0	0.1	-0.1	
1S/5W-07R1	F-35A	1.0	0.8	0.3	0.1	0.1	-0.1	
1S/5W-14B1	Chino 02	1.0	0.8	0.2	0.0	0.2	0.0	
1S/6W-23D2	F-17B	1.1	0.9	0.1	-0.1	0.1	-0.1	
1S/6W-23D2	F-17B (580')	1.1	0.9	0.1	-0.1	0.1	-0.1	
	WVWD 42	1.0	0.8	0.2	0.0	0.2	0.0	

Table 3.4. Mixing fractions of Atacama, Synthetic, and Indigenous perchlorate based on $\Delta^{17}O$ vs $\delta^{18}O$ stable isotope data for samples collected from wells in the Study Area, June 2010 to February 2012. Values provided are the 95% confidence limits calculated using IsoError1_04.xls as described in Phillips and Gregg, (2001). Endmember values are given in Table 3.2.

State	Local	Atacar	na	Synthetic		Indigenous		
well	well	mixing fra	ction	mixing fraction		mixing fraction		
number	name	95% confiden		95% confid		95% confidence limits		
		maximum	minimum	maximum	minimum	maximum	minimum	
	We	Ils in the perched		Rialto-Colton	subbasin wit	hin the mapped	plume	
1N/5W-28J2	28J-2	0.2	0.0	1.0	0.8	0.2	0.0	
1N/5W-29H1	F6	0.2	0.0	1.0	0.6	0.5	-0.1	
1N/5W-29H3	F6A-S	0.2	0.0	1.1	0.7	0.3	-0.1	
1N/5W-33B2	N-10S	0.2	0.0	1.1	0.5	0.5	-0.1	
	We	Ils in the Regional	aquifer in th	e Rialto-Coltor	subbasin wit	hin the mapped	plume	
1N/5W-21N2	PW-2	0.2	0.0	1.1	0.7	0.3	-0.1	
1N/5W-21P2	PW-3	0.2	0.0	1.0	0.8	0.2	0.0	
1N/5W-33B1	N-10D	0.2	0.0	1.0	0.6	0.4	0.0	
1N/5W-34B2	Rialto 02	0.2	0.0	1.1	0.9	0.0	-0.2	
1N/5W-34B2	Rialto 02 (800')	0.2	0.0	1.1	0.9	0.0	-0.2	
1N/5W-34G4	PW-5D	0.2	0.0	1.0	0.8	0.2	0.0	
1N/5W-34M1	Rialto 03	0.2	0.0	1.0	0.6	0.4	0.0	
1N/5W-34M1	Rialto 03 (670')	0.2	0.0	1.0	0.6	0.4	0.0	
1S/5W-02B3	PW-9C	0.2	0.0	1.0	0.8	0.3	-0.1	
1S/5W-02G1	Rialto 06	0.2	0.0	1.0	0.6	0.3	-0.1	
1S/5W-02G1	Rialto 06 (480')	0.2	0.0	1.0	0.8	0.3	-0.1	
1S/5W-02G1	Rialto 06 (560')	0.2	0.0	1.0	0.6	0.3	-0.1	
1S/5W-03A1	Rialto 04	0.2	0.0	1.0	0.8	0.2	0.0	
		<u>We</u>	lls outside th	e mapped plu	me_			
1N/5W-17K2	WVWD 24	0.4	0.0	0.6	0.4	0.5	0.3	
1N/5W-27D1	Rialto 01	0.2	0.0	1.0	0.6	0.3	-0.1	
1N/5W-33N1	F-49A	0.5	0.1	0.5	0.3	0.6	0.0	
1S/4W-18G1	Colton 15	1.1	0.9	0.1	-0.1	0.2	0.0	
1S/5W-02K1	WVWD 11	0.4	0.0	1.0	0.6	0.3	-0.1	
1S/5W-05A5	F-10C	0.7	0.1	0.5	0.3	0.4	0.0	
1S/5W-13B5	13B1-5 (RHSW 5)	0.9	0.7	0.2	0.0	0.2	0.0	
1N/5W-32A1	F-13A	0.9	0.3	0.2	0.0	0.4	0.2	
		Wells in	the Chino (or	Arlington) sub	<u>obasins</u>			
1N/5W-32N1	F-26A	0.9	0.7	0.3	0.1	0.2	0.0	
1N/5W-32N1	F-26A (800')	0.9	0.7	0.3	0.1	0.2	0.0	
1S/5W-04N1	F-4A	1.1	0.9	0.2	0.0	0.1	-0.1	
1S/5W-04N1	F-4A (J613; 2007 sample)	1.1	0.9	0.2	0.0	0.1	-0.1	
1S/5W-07R1	F-35A	1.0	0.8	0.3	0.1	0.1	-0.1	
1S/5W-14B1	Chino 02	1.0	0.8	0.1	-0.1	0.2	0.0	
1S/6W-23D2	F-17B	1.1	0.9	0.1	-0.1	0.1	-0.1	
1S/6W-23D2	F-17B (580')	1.1	0.9	0.1	-0.1	0.1	-0.1	
1S/5W-23A1	WVWD 42	1.0	0.8	0.1	-0.1	0.2	0.0	

Table 3.5. Mixing fractions of Atacama, Synthetic, and Indigenous perchlorate based on 36 Cl/Cl abundance vs δ^{37} Cl stable isotope data for samples collected from wells in the Study Area, June 2010 to February 2012. Values provided are the 95% confidence limits calculated using IsoError1_04.xls as described in Phillips and Gregg, (2001). Endmember values are given in Table 3.2.

State	Local	Atacar	na	Synthetic		Indigenous		
well	well	mixing fra	ction	mixing	mixing fraction		mixing fraction	
number	name	95% confiden	ce limits	95% confidence limits		95% confidence limits		
		maximum	minimum	maximum	minimum	maximum	minimum	
	We	lls in the perched	aquifer in the	Rialto-Colton	subbasin wit	hin the mapped	plume	
1N/5W-28J2	28J-2	0.2	0.0	1.0	0.8	0.002	0.001	
1N/5W-29H1	F6	0.3	0.0	1.0	0.7	0.03	0.02	
1N/5W-29H3	F6A-S	NA*	NA	NA	NA	NA	NA	
1N/5W-33B2	N-10S	0.2	0.0	1.0	0.8	0.02	0.01	
	We	Ils in the Regional	aquifer in th	e Rialto-Coltor	subbasin wit	hin the mapped	l plume	
1N/5W-21N2	PW-2	0.2	0.0	1.0	0.8	0.004	0.002	
1N/5W-21P2	PW-3	0.2	0.0	1.0	0.8	0.006	0.003	
1N/5W-33B1	N-10D	0.3	0.0	0.9	0.8	0.12	0.07	
1N/5W-34B2	Rialto 02	NA	NA	NA	NA	NA	NA	
1N/5W-34B2	Rialto 02 (800')	NA	NA	NA	NA	NA	NA	
1N/5W-34G4	PW-5D	0.2	0.0	1.0	0.8	0.002	0.001	
1N/5W-34M1	Rialto 03	0.2	0.0	1.0	0.8	0.02	0.01	
1N/5W-34M1	Rialto 03 (670')	0.2	0.0	1.0	0.8	0.02	0.01	
1S/5W-02B3	PW-9C	0.2	0.0	1.0	0.8	0.005	0.002	
1S/5W-02G1	Rialto 06	0.2	0.0	1.0	0.8	0.003	0.001	
1S/5W-02G1	Rialto 06 (480')	0.2	0.0	1.0	0.8	0.007	0.002	
1S/5W-02G1	Rialto 06 (560')	0.2	0.0	1.0	0.8	0.02	0.00	
1S/5W-03A1	Rialto 04	0.2	0.0	1.0	0.8	0.004	0.002	
		<u>We</u>	lls outside th	e mapped plu	me			
1N/5W-17K2	WVWD 24	0.2	0.0	0.9	0.8	0.2	0.1	
1N/5W-27D1	Rialto 01	0.3	0.0	0.9	0.8	0.09	0.05	
1N/5W-33N1	F-49A	0.5	0.4	0.0	-0.2	1.0	0.6	
1S/4W-18G1	Colton 15	1.0	0.9	0.06	0.02	0.04	0.03	
1S/5W-02K1	WVWD 11	0.3	0.0	1.0	0.7	0.001	0.000	
1S/5W-05A5	F-10C	0.6	0.4	0.4	0.2	0.3	0.2	
1S/5W-13B5	13B1-5 (RHSW 5)	0.9	0.7	0.2	0.1	0.07	0.04	
1S/5W-23A1	F13A	0.7	0.6	0.0	-0.1	0.8	0.5	
	-	Wells in	the Chino (or	Arlington) sul	obasins .		-	
1N/5W-32A1	WVWD 42	0.9	0.8	0.2	0.0	0.05	0.03	
1N/5W-32N1	F-26A	0.9	0.7	0.1	0.0	0.2	0.1	
1N/5W-32N1	F-26A (800')	0.9	0.8	0.1	0.0	0.2	0.1	
1S/5W-04N1	F-4A	1.0	0.9	0.1	0.0	0.04	0.02	
1S/5W-04N1	F-4A (J613; 2007 sample)	NA	NA	NA	NA	NA	NA	
1S/5W-07R1	F-35A	1.0	0.8	0.2	0.0	0.01	0.00	
1S/5W-14B1	Chino 02	1.0	0.8	0.2	0.0	0.02	0.01	
1S/6W-23D2	F-17B	1.1	1.0	0.0	-0.1	0.00	-0.01	
1S/6W-23D2	F-17B (580')	1.05	0.97	0.03	-0.04	0.00	-0.01	
*NA - data not available.								

Some data from wells within the mapped contaminant plume indicate minor components (0 % to ~ 10 %) of natural indigenous ClO_4 , based upon the values of $^{36}Cl/Cl$ in the samples compared to synthetic ClO_4 sources (Table 3.5). Minor indigenous components were difficult to resolve in the stable isotope endmember mixing model results, because these fractions were small compared to the uncertainties associated with the mixing calculations (Tables 3.3 and 3.4). However, hypothetical mixing curves for $^{36}Cl/Cl$ and $\delta^{37}Cl$ with endpoints in synthetic, Atacama, and natural indigenous ClO_4 in Figure 3.5 (bottom panel) highlight an apparent trend of increasing indigenous ClO_4 in the data array from PW5D (0 %) to N10-D (~ 10 %). $\delta^{18}O$ values slightly higher than the proposed synthetic endmember (with a $\delta^{18}O$ value of -16.7 \pm 1.1‰) are also consistent with varying fractions of indigenous natural ClO_4 in some of the same samples. This general relation is apparent for samples within the mapped plume areas for which both $\delta^{18}O$ and $\delta^{36}Cl/Cl$ values are available (Figure 3.6).

Within the mapped plume areas, samples with larger apparent fractions of indigenous ClO_4^- based on elevated $\delta^{18}O$ and ${}^{36}Cl/Cl$ ratios (Figure 3.6) tended also to have lower total ClO₄⁻ concentrations (Figure 3.7), consistent with varying fractions of a low-concentration background source. According to a regional statistical study (Fram and Belitz, 2011), in the absence of anthropogenic sources or localized flushing of ClO₄⁻ from the vadose zone by irrigation or other artificial recharge processes, concentrations of indigenous natural ClO₄ in groundwater in arid/semiarid areas of the southwestern U.S. are likely to be less than 1 µg/L. Where accumulated salts are flushed from the vadose zone, natural ClO₄ concentrations can be much higher, sometimes exceeding 4 $\mu g/L$ and occasionally reaching > 50 $\mu g/L$, as reported for irrigated agricultural regions of the SHP (Jackson et al., 2005; Rajagopalan et al., 2006; Rao et al., 2007). Some of the variability observed in Figures 3.6 and 3.7 may reflect differing concentrations of natural background ClO₄ in these wells and variability in synthetic and indigenous endmember isotope values; nevertheless, the general relations are consistent with relatively low background concentrations (<0.5 μg/L). The sample from N-10D, with total ClO₄⁻ concentration of 1.6 µg/L, had a larger apparent fraction of indigenous natural ClO₄ (~ 10 %, based on its ³⁶Cl/Cl ratio) than most other wells in the mapped plume area, possibly indicating this sample included groundwater from beneath the contaminant plume, which would be consistent with the deeper screened interval from which the sample was collected.

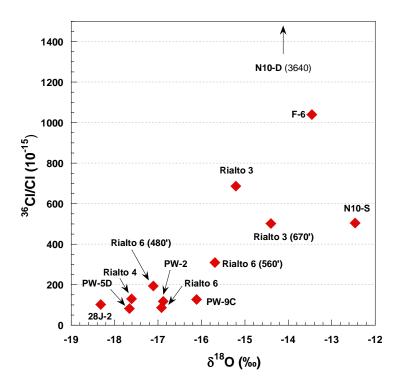


Figure 3.6. Comparison of δ^{18} O values (‰) and 36 Cl/Cl ratios for wells located within the mapped ClO₄- plumes areas for which both values are available.

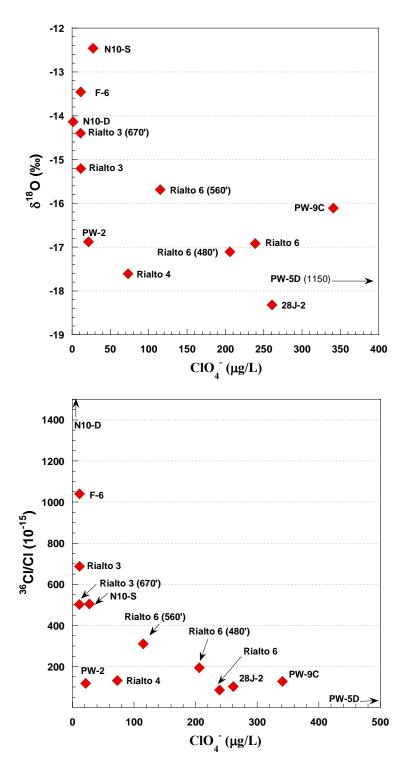


Figure 3.7. Comparison of $\delta^{18}O$ value (‰) versus ClO_4^- concentration (top panel) and $^{36}Cl/Cl$ ratio vs ClO_4^- concentration (bottom panel) for all wells located within the mapped ClO_4^- plume areas for which both values are available.

Downgradient from the two mapped plumes in the southeastern direction, parallel to the Rialto-Colton Fault, the dominant source of ClO₄ appears to shift from synthetic to Atacama, presumably because of past fertilizer use in the recharge areas for these groundwaters. The ClO₄ in WVWD 11 appeared to be predominantly synthetic (> 60 % based on mixing fraction analysis), but with a significant fraction of Atacama ClO₄ based upon its Δ^{17} O value of ~ +1 ‰, compared to < +0.2 ‰ for all other samples collected in the mapped plume area. WVWD 11 also had a lower δ^{37} Cl value (-0.5 %) than most of the wells that appeared to have predominantly synthetic ClO_4^- (average + 0.6 \pm 0.6 %). Farther downgradient, wells 13B1-5 (RHSW-5) and Colton 15 appeared to have had ClO_4^- that was predominantly Atacama (> 70 %), based upon their low $\delta^{37}Cl$ values (-9.8 and -12.3 ‰, respectively) and elevated Δ^{17} O values (+6.9 and +8.6 ‰, respectively). The combination of low δ^{37} Cl values and high Δ^{17} O values in these samples is characteristic of Atacama-derived ClO₄ (Böhlke et al., 2005, 2009; Jackson et al., 2010; Sturchio et al., 2011, 2012, 2014). The non-Atacama ClO₄ in these samples (RHSW-5, Colton 15) appears to be largely indigenous, based upon elevated ³⁶Cl/Cl values and other correlations.

Samples from the Chino subbasin, southwest of the Rialto-Colton Fault, including those from wells WVWD 42, F-17B, F-35A, F-4A, Chino 2, and F-26A, also had isotopic characteristics indicating predominantly Atacama ClO_4^- (≥ 70 % based on mixing fraction calculations), including elevated values of $\Delta^{17}O$ (+ 6.1 to + 8.2 %) and low values of $\delta^{37}Cl$ (-9.7 to -13.5 %) (Figures 3.3-3.5, Tables 3.3 and 3.4). The finding of ClO_4^- derived from Atacama fertilizer in the Chino subbasin is consistent with data from previous studies examining ClO_4^- sources in the western region of this subbasin using the stable isotope approach (Sturchio et al., 2008, 2011, 2014). Similarly, Atacama ClO_4^- was recently identified in a number of wells in Riverside, CA in the vicinity of the San Bernardino Airport to the southeast of the Study Area using the stable isotope approach Sturchio et al., (2011, 2012). In the latter case, Atacama ClO_4^- was identified in wells outside a plume of synthetic ClO_4^- . Minor components of indigenous natural ClO_4^- also were indicated in some of the Riverside wells based on the stable isotope values.

The detection of Atacama ClO₄⁻ in the Study Area on both sides of the Rialto-Colton Fault and across the Chino subbasin is consistent with the past agricultural history of this area, and with anecdotal evidence of the application of Atacama nitrate fertilizer

throughout this region (Sturchio et al., 2008, 2011, 2012, 2014). A 1949 land use map showing agricultural regions in the Chino Basin, and covering a substantial portion of the Study Area, is provided in Figure 3.8. This map shows the locations of irrigated vineyards, and irrigated and non-irrigated citrus, within the Study Area. Based on the history of this region, it is likely that these areas were in agriculture for at least a few decades before this 1949 map, when application of Atacama fertilizer was common. Additional historical citrus production along Lytle Creek that is not shown on this figure has also been documented (Geologic Associates, 2002). Peak citrus production in the Chino Basin occurred during the period 1930-1950 (USDA, 2013).

Well WVWD 24 was initially selected for sampling as a possible representative of uncontaminated ("background") conditions, based upon its location far upgradient of the two ClO₄⁻ plume source areas, its low ClO₄⁻ concentration (~ 0.3 μg/L), and its potential for recharge from the San Gabriel Mountains (Izbicki et al., 2014). It was presumed that wells in this part of the Study Area were most likely to contain local indigenous natural ClO₄⁻, rather than synthetic or Atacama ClO₄⁻. Three additional wells (F-27A, F-33A, and F-42A) located to the north of WVWD 24 were considered for background sampling, but ClO₄⁻ concentrations in these wells, all of which were monitoring wells, were considered to be too low to obtain adequate ClO₄⁻ for isotopic analysis (0.10, 0.11, and 0.14 μg/L, respectively).

Based on our endmember selections for mixing analysis, isotope data for ClO_4^- from WVWD 24 indicated that this well contained a mixture composed primarily of indigenous natural ClO_4^- and synthetic ClO_4^- (Figures 3.3-3.5, Table 3.3 and 3.4). The sample had a $\delta^{18}O$ value of -8.4, which is between the estimated endmember values of synthetic ClO_4^- (-16.7 ‰) and indigenous natural ClO_4^- (+2.5 ‰) listed in Table 3.2. The WVWD 24 sample had a $\delta^{17}O$ value of +1.5 ‰, which is within the range reported for indigenous natural ClO_4^- in the southwest US (Jackson et al., 2010) and slightly lower than that for eastern Oregon (Hatzinger et al., 2012) or the Great Lakes (Poghosyan et al., 2014). Finally, the $^{36}Cl/Cl$ ratio of WVWD 24 (6021 x $^{10}O_4^-$) placed it within the lower end of the range reported for indigenous natural ClO_4^- (Sturchio et al., 2009; Poghosyan et al., 2014). All of these observations could be consistent with a synthetic-indigenous mixture in WVWD 24. Tritium (3H) was detected in this well at 10.6 pCi/L (Izbicki et al., 2014), indicating that it contained at least some modern (post-bomb) recharge water.

Because WVWD-24 is located upgradient from the identified source areas of the two mapped ClO₄⁻ plumes (Figure 3.1) it is unlikely to have acquired synthetic ClO₄⁻ from those source areas by groundwater flow (Izbicki et al., 2014). If synthetic ClO₄⁻ was present in WVWD 24, as indicated by the isotope data, it may have originated from other commercial sources, such as road flares, fireworks, or blasting explosives, each of which could have contributed to groundwater contamination in this area, as they apparently have elsewhere in the U.S. (e.g., Böhlke et al., 2009; Munster and Hanson, 2009; Munster, 2008; Munster et al., 2008; Aziz et al., 2007).

An alternative hypothesis is that the sample from WVWD 24 represents a local indigenous natural component that is different from the one proposed in the mixing analysis. Additional regional sampling would be required to further evaluate this hypothesis. Moreover, whether ClO₄⁻ from WVWD 24 was entirely or only partly natural (i.e., indigenous mixed with synthetic), it could represent a regional background ClO₄⁻ that is present in varying proportions in other wells in the Study Area that are not within the mapped plumes.

Three wells not discussed above, F-13A, F-49A, and F-10C, all had low ClO₄⁻ concentrations ranging from 0.4 to 0.9 µg/L and appeared to represent mixed ClO₄⁻ sources. In map view (Figure 3.1), well F-13A is located along Barrier H, to the south of the 160 Acre Site, well F-49A is located to the east of Barrier H, and well F-10C is located east of the Rialto-Colton Fault and west of Barrier H. Wells F-10C and F-49A were observed to have perched groundwater cascading downward and mixing with deeper water in the well casing (Izbicki et al., 2014). The perched groundwater may have carried ClO₄ that was not present in deeper groundwater in these wells. Evaluating all three stable isotopes (δ^{18} O, δ^{37} Cl, Δ^{17} O), and assuming the characteristics of Atacama, synthetic, and indigenous end members as previously described, the ClO₄ in well F-13A would appear to be composed largely of Atacama and indigenous ClO₄, whereas wells F10-C and F49-A would appear to have ClO₄ from all three endmember sources, with no individual component dominating the mixture (Tables 3.3 and 3.4). If isotope data from WVWD 24 represented a widespread mixed (indigenous + synthetic) background ClO₄⁻ type, then data from F10-C and F49-A could be consistent with mixtures of Atacama and that mixed background type.

Relations between 36 Cl/Cl and δ^{37} Cl values (Figure 3.5, bottom panel) generally support the observations above for these wells, but would appear to indicate somewhat larger fractions of indigenous ClO₄ and smaller fractions of synthetic ClO₄. These apparent differences could be due to variability associated with the ³⁶Cl/Cl ratio of natural ClO₄. Alternatively, it could indicate contributions of anomalously ³⁶Cl-enriched indigenous ClO₄ from the brief period of thermonuclear bomb tests in the western Pacific (1952-1964). Poghosyan et al. (2014) propose that ClO₄- having ³⁶Cl/Cl ratios around 1,000,000 x 10⁻¹⁵ was widely deposited during that time. Most of the samples analyzed in the current study, with the exception of bulk discharge from well F-17B, had measurable ³H (Izbicki et al., 2014), indicating at least a portion of the water precipitated after 1952. Any bomb-pulse ClO₄ contribution (if present) would likely be most evident in the lowest-concentration samples (i.e., F-10C, F-13A, F-49A, and WVWD 24); these are the samples having the highest ³⁶Cl/Cl ratios in the present study. A potential bombpulse contribution is partially accounted for in the indigenous endmember used for mixing calculations (Table 3.2) by inclusion of data from the Great Lakes (Poghosyan et al. 2014). However, it is possible that the ³⁶Cl/Cl ratio was actually higher than that assumed for our mixing calculations or that the specific bomb-pulse contribution varied on a well by well basis in relation to varying local recharge conditions.

The origin of the apparent synthetic ClO₄⁻ component in wells F-10C and F-49A (based on stable isotope data) is uncertain, as it contributed a maximum of only about 0.6 and 0.2 μg/L, respectively, to the mixtures in the wells based on stable isotope mixing fraction estimates (Table 3.3 and Table 3.4). Sources of such low concentrations could include upgradient background groundwater (as in WVWD 24), plume-related or non-plume-related ClO₄⁻ in perched groundwater, or other local contamination. Alternatively, if the ClO₄⁻ in WVWD 24 was isotopically representative of natural indigenous ClO₄⁻ in the Study Area, then the apparent synthetic components in wells F-10C and F-49A would be reduced or eliminated, as these wells fall reasonably in a mixing line between WVWD 24 and the Atacama endmember based on stable isotope values (Figure 3.5). In this case, however, the ³⁶Cl/Cl ratios in these two samples would be expected to fall between that of WVWD 24 and the Atacama endpoint, rather than being higher, particularly in the case of well F-49A.

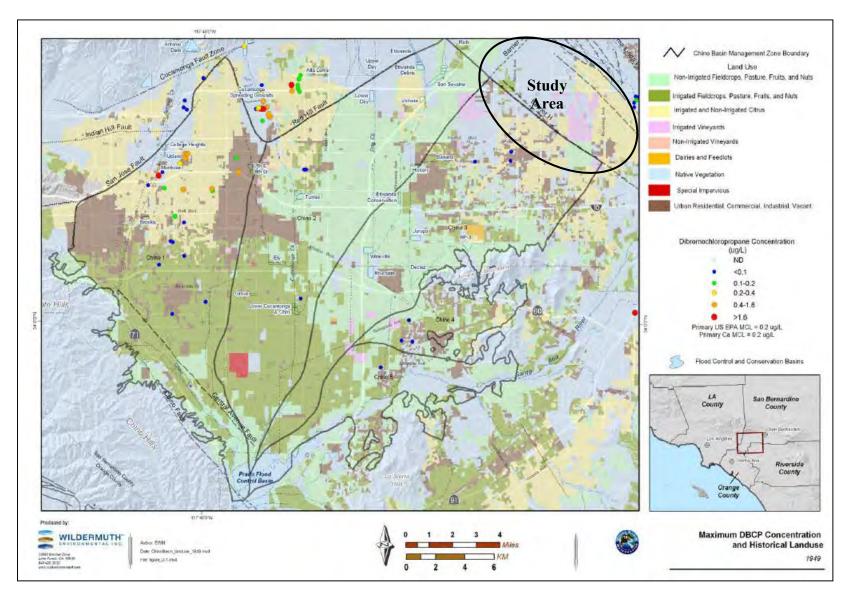


Figure 3.8. Map of land use characteristics in 1949 in the Chino Basin Management Zone. The Study Area is shown in the top right corner. Figure from Sturchio et al. (2008).

3.3. Depth-Dependent Isotope Samples

Depth-dependent groundwater sampling and isotopic analysis was conducted on five of the production wells in the Study Area: F-17B, F-26A, Rialto 2, Rialto 3, and Rialto 6. This work was conducted as part of the coupled well-bore flow and depth dependent water quality data analysis described by Izbicki et al. (2014). The primary objective of sampling ClO₄⁻ isotopes with depth in these wells was to determine whether different sources were entering the wells at different depth intervals in the aquifer(s) (e.g., synthetic ClO₄⁻ at the top of the well screen and indigenous ClO₄⁻ at a deeper screened interval).

Overall, the results showed that the source(s) of ClO₄ collected from the bulk discharge of each of the wells was similar to that collected from a deep interval within the well, or in the case of Rialto 6, from two different depth intervals (Table 3.1; Figure 3.9). There were small differences in δ^{18} O between F-17B (580') and F-17B (bulk discharge) and between Rialto 6 (480') and Rialto 6 (560'), but these differences did not indicate substantial differences in ClO₄ source fractions. In general, the major-ion composition of the wells with depth also was similar to the surface discharge from each well, with the exception of dissolved oxygen (DO), which generally decreased with depth. Biological degradation of ClO₄ and associated fractionation of Cl and O isotopes in ClO₄ are possible under anoxic conditions, particularly if organic carbon or other electron donors are available for growth of perchlorate-reducing bacteria (e.g., Sturchio et al., 2007). However, depth-dependent discharge data indicate deeper groundwater with low DO may not contribute substantially to the total discharge from the wells (Izbicki et al., 2014). No evidence of ClO₄ isotopic fractionation was apparent in the depth-dependent samples. Similarly, no evidence of NO₃⁻ isotopic fractionation that might be associated with NO₃⁻ reduction was apparent in the same set of samples (Izbicki et al., 2014). For wells F-26A, F-17B, and Rialto 6, both the deep sampled interval and the bulk discharge had high DO concentrations (> 8 mg/L; Izbicki et al., 2014), indicating the bulk of the sampled groundwater was from the upper oxic part of the aquifer. The DO concentration was low for Rialto 3 at 670' (0.2 mg/L) but there was no evidence for biological fractionation of ClO₄ isotopes in this sample. Dissolved oxygen data were not available for Rialto 2 with depth.

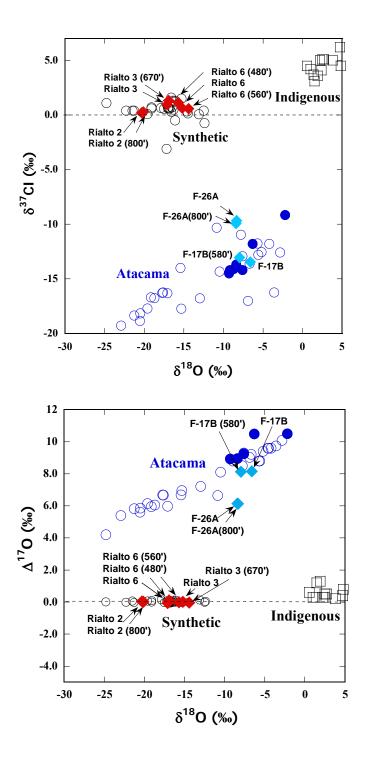


Figure 3.9. Comparison of $\delta^{37}\text{Cl}$ versus $\delta^{18}\text{O}$ (top panel) and $\Delta^{17}\text{O}$ versus $\delta^{18}\text{O}$ (bottom panel) in ClO_4^- from Study Area groundwater depth-dependent samples (symbols as in Figure 3.2) with published data for indigenous natural ClO_4^- of the SHP and MRGB (black open squares), natural Atacama ClO_4^- , (open blue circles for field samples and closed blue circles for Atacama nitrate fertilizer samples) and synthetic ClO_4^- (open black circles). Sources of ClO_4^- isotope data for Atacama, Synthetic and Indigenous as described in Figure 3.2.

3.4 Data Quality and Reproducibility

As described in the Methods section, several different procedures were implemented to ensure data quality and reproducibility. These procedures included (1) collection of replicate columns from ~ 30 % of the wells to establish column to column variability; (2) analysis of purified samples multiple times when large enough quantities were available to establish analytical variability; (3) routine analysis of ClO_4^- isotopic reference materials for consistency of IRMS data calibration; (4) analysis of $\delta^{18}O$ and $\delta^{17}O$ by O_2 -DI-IRMS(- N_2) in addition to O_2 -DI-IRMS(+ N_2) to evaluate levels of contaminants and co-products of ClO_4^- decomposition; (5) analysis of select samples by by CO-CF-IRMS for independent quantification of $\delta^{18}O$; and (6) preparation and analysis of ClO_4^- isotopic reference materials loaded onto IX columns to assess effects of routine sample processing on stable isotope values.

3.4.1 Results from Replicate Columns

The data from individual replicate columns are provided in Table 3.6. For $\delta^{18}O$ and $\Delta^{17}O$, differences between replicate columns were ≤ 0.5 % for all samples tested. For $\delta^{37}Cl$ differences in replicates were ≤ 1.0 %. For well F-17B, four IX columns were processed and analyzed independently, and the differences among replicates were ≤ 0.2 % for $\delta^{37}Cl$, ≤ 0.4 % for $\delta^{18}O$ and ≤ 0.2 % for $\Delta^{17}O$. Similarly, for the four replicate columns from F-17B, $^{36}Cl/Cl$ ratios varied by $< 10 \times 10^{-15}$.

3.4.2 δ^{18} O from O₂-DI-IRMS(+N₂), O₂-DI-IRMS(-N₂), and CO-CFIRMS

All samples were analyzed by O_2 -DI-IRMS(+ N_2) and those results are the reported values (Table 3.1). Most samples also were analyzed by O_2 -DI-IRMS(- N_2) to evaluate contaminants and co-products of ClO_4 decomposition such as CO_2 . A subset of samples also were analyzed independently by CO-CF-IRMS, which includes measurement of the relative amount of N_2 as an indication of potential contamination by air or N-O salts such as NO_3 . For most samples, $\delta^{18}O$ values from O_2 -DI-IRMS(+ N_2) were slightly lower than $\delta^{18}O$ values from CO-CF-IRMS and O_2 -DI-IRMS(- N_2) (Figure 3.10; Table 3.7). Several different processes could have contributed to differences between results of these methods including the following: (1) isotopic fractionation of O_2 by partial oxidation of reduced C in the samples (both O_2 -DI-IRMS methods); (2)

isotopically anomalous O₂ produced by fragmentation of CO₂ in the IRMS (O₂-DI-IRMS(-N₂)); (3) contamination by O with elevated δ^{18} O (all three methods); (4) isotopic exchange between O₂ and CO₂ in the reaction tubes (both O₂-DI-IRMS methods); and (5) analytical artifact of elevated CO₂ in the ion source (O₂-DI-IRMS(-N₂)). Linear fits through the δ^{18} O data have R² values of 0.993-0.994 and intercepts between 0.4 and 0.6 % indicating data from different methods are well correlated. Values of Δ^{17} O were indistinguishable between the two O₂-DI-IRMS methods, with R² of 0.999 and intercept of 0.0 %. To illustrate these effects in context of ClO₄- source identification, data from all three methods are plotted together with ClO₄ source data in Figure 3.11. In most cases, the symbols for isotopic values from each well by each of the three methods overlap. More importantly, the interpretation of sources for all of the wells is the same irrespective of the analytical technique used to measure $\delta^{18}O$ and/or $\delta^{17}O$. Data reported in Table 3.1 (from O₂-DI-IRMS(+N₂)) represent approximately 100 analyses of separate sample aliquots. A total of three analyses of this type (one from Rialto 4 and two from WVWD 42) were eliminated as outliers (small aliquots, highly fractionated); larger aliquots of those samples subsequently were purified and analyzed. All other analyses were included in the averages.

3.4.3 Perchlorate Extraction and Purification – Method Tests for QA/QC

Quality Assurance/Quality Control (QA/QC) tests were conducted using the USGS37 and USGS38 KClO₄ reference materials to evaluate the extent of Cl and/or O isotopic fractionation caused by the laboratory extraction and purification methods. The methods and results from these tests are presented in Appendix A. Here, isotope data are given for samples subjected to extraction and purification procedures after normalization to analyses of pure reference materials in order to isolate the effects of sample preparation. In summary, values of δ^{18} O, Δ^{17} O, and δ^{37} Cl of USGS37 that had been loaded on IX columns and purified by typical procedures at UIC were within the stated analytical uncertainties of ± 0.2 , ± 0.2 , and ± 0.1 ‰, respectively, of the assumed reference values for USGS37. For USGS38, which was intentionally synthesized with unnaturally elevated values of δ^{18} O (+52.5 ‰) and Δ^{17} O (+73.3 ‰) and low value of δ^{37} Cl (-87.2 ‰), deviations of -2.1, -1.9 ‰, and +0.6 ‰, respectively, were observed for δ^{18} O, Δ^{17} O, and

 δ^{37} Cl. Contraction of the apparent O isotope scale (factor of approximately 0.97) is consistent with the presence of small quantities of oxygen-bearing impurities with more normal δ^{18} O and δ^{17} O values in the prepared samples. The observed deviations in the USGS38 tests, when spread over the entire range of the calibration between USGS37 and USGS38, would not cause significant changes to δ values within the ranges of the samples analyzed in this study.

3.4.4 Secondary Ion Mass Spectrometry (SIMS)

An independent analysis of a subset of the purified Study Area ClO₄⁻ samples was conducted by SIMS at the Caltech Microanalysis Center (John Eiler, unpublished data). An advantage of the SIMs method is that much smaller quantities of ClO₄⁻ are required for a SIMS analysis than for traditional IRMS. However, the method has not been validated for complete stable isotope analysis of Cl and O isotopes in ClO₄⁻, and is considered experimental. The data from the SIMS analysis are compared to those from IRMS for split samples in Appendix B.

Table 3.6. Perchlorate Cl and O isotope data for replicate sample columns from wells in the Study Area, June 2012 to February 2012.

State	Local						Column			
well	well		$\delta^{37}\text{CI}$	$\delta^{18}O$	$\Delta^{17}O$	³⁶ Cl	identification			
number	name	Date	(‰)	(‰)	(‰)	³⁶ CI/CI x 10 ⁻¹⁵	number			
	Wells in the perched aquifer in the Rialto-Colton subbasin within the mapped plume									
1N/5W-28J2	28J-2	07/01/10	0.8 ± 0.1	-18.2	-0.1	101 (5)	242			
1N/5W-28J2	28J-2	07/01/10	0.8	-18.4	-0.1	103 (6)	248			
	Wells in the Regional a	quifer in the Rialt	o-Colton subl	basin within th	e mapped plu	<u>ıme</u>				
1N/5W-21N2	PW-2	08/11/10	1.4 ± 0.2	-16.9 ±0.3	0.0 ± 0.2	118 (4)	230			
1N/5W-21N2	PW-2	08/11/10	1.4	-16.8 ±0.5	-0.1 ± 0.1	NA	231			
1S/5W-03A1	Rialto 04	10/20/10	1.1±0.0	-17.7 ± 0.1	0.0 ± 0.2	131 (3)	269			
1S/5W-03A1	Rialto 04	10/20/10	1.1 ± 0.1	-17.5 ± 0.2	0.0 ± 0.3	NA	277			
		Wells outside	the mapped plu	ıme_	Į.					
1N/5W-17K2	WVWD 24	06/29/10	2.0	-8.7	1.1	6021 (110)	223+224+250+251			
1N/5W-17K2	WVWD 24	06/29/10	1.3 ± 0.0	-8.2 ± 0.2	1.3 ± 0.2	NA	259+260+261+262			
1N/5W-27D1	Rialto 01	05/18/11	0.0 ± 0.0	-16.1 ± 0.0	0.2 ± 0.0	2570 (50)	324			
1N/5W-27D1	Rialto 01	05/18/11	-0.3 ± 0.1	-15.7 ± 0.0	0.2 ± 0.1	2750 (40)	325			
1S/4W-18G1	Colton 15	07/08/10	-12.6 ± 0.4	-4.6 ± 0.1	8.6 ± 0.1	NA	245			
1S/4W-18G1	Colton 15	07/08/10	-11.6	-4.5 ± 0.0	8.6 ± 0.1	1590 (40)	246			
1S/5W-02K1	WVWD11	06/30/10	-0.3 ± 0.1	-15.4 ± 0.0	1.0 ± 0.2	66 (3)	232			
1S/5W-02K1	WVWD11	06/30/10	-0.7 ± 0.4	-15.2 ± 0.7	1.1 ± 0.2	NA	240			
1N/5W-32A1	F-13A	05/11/11	-5.4	-4.9	4.4	NA	320			
1N/5W-32A1	F-13A	05/11/11	-5.7	-4.8 ± 0.0	4.8 ± 0.1	24200 (1200)	322			
		Wells in the Chino	(or Arlington) s	<u>ubbasins</u>						
1S/5W-23A1	WVWD 42	07/08/10	-10.3	NA	NA	1690 (110)	225			
1S/5W-23A1	WVWD 42	07/08/10	-10.3	NA	NA	1880 (90)	226			
1S/5W-23A1	WVWD 42	07/08/10	-10.4	-7.0 ± 0.0	7.1 ± 0.1	NA	225 + 226			
1S/6W-23D2	F-17B	06/17/10	-13.4	-6.5 ± 0.2	8.2 ± 0.1	92 (6)	238			
1S/6W-23D2	F-17B	06/17/10	-13.5	-6.5 ± 0.3	8.2 ± 0.1	88 (5)	239			
1S/6W-23D2	F-17B	06/17/10	-13.5	-6.6 ± 0.3	8.0 ± 0.0	88 (6)	243			
1S/6W-23D2	F-17B	06/17/10	-13.6 ± 0.2	-6.9 ± 0.3	8.2 ± 0.1	96 (6)	244			
1S/5W-04N1	F-4A	07/28/10	-11.8 ± 0.3	-7.9 ± 0.2	7.6 ± 0.1	1380 (30)	255			
1S/5W-04N1	F-4A (J613; 2007 sample)	03/09/07	-11.9	-8.0	7.6	NA	mpg31			

NA - data not available.

Means and standard deviations are provided for replicate runs on individual columns. When individual values are provided, samples were analyzed once. For WVWD-24, columns 223, 224, 250, 251 were combined into a single sample and 259, 250, 261, 262 were combined into a second sample.

The two samples were then treated independently.

 $Columns~225~\&~226~from~WVWD~42~were~independently~analyzed~for~\delta^{37}Cl~and~^{36}Cl~but~combined~form~\delta^{18}O~and~\Delta^{17}O~analysis.$

For ³⁶Cl, values in parentheses are laboratory reported unceratinty for individual samples.

Table 3.7. Perchlorate isotope data using three different methods of O stable isotope analysis for samples collected from wells in the Study Area, June 2012 to February 2012.

State	Local			O ₂ -DI-II	RMS(+N ₂)	O ₂ -DI-IR	MS(-N ₂)	CO-CFIRMS				
well	well		δ ³⁷ Cl	$\delta^{18}O$	Δ ¹⁷ O	δ ¹⁸ O	Δ ¹⁷ O					
		Data	-	(‰)	(‰)		(%)					
number	name	Date	(‰)	(‰)	(‰)	(‰)	(‰)	(‰)				
Wells in the perched aquifer in the Rialto-Colton subbasin within the mapped plume												
1N/5W-28J2	28J-2	07/01/10	0.8 ± 0.1	-18.3 ± 0.1	-0.1 ± 0.0	-18.0 ± 0.9	0.2 ± 0.0	-17.4 ± 0.1				
1N/5W-29H1	F6	07/20/10	-0.5	-13.5	0.1	-13.3	0.0	NA				
1N/5W-29H3	F6A-S	07/22/10	-0.9	-15.5 ± 0.3	0.2 ± 0.1	NA	NA	NA				
1N/5W-33B2	N-10S	07/21/10	1.0 ± 0.0	-12.5 ± 1.7	0.0 ± 0.1	-11.6 ± 1.4	0.0 ± 0.1	NA				
	_We	ells in the Regiona	l aquifer in th	e Rialto-Coltor	<u>ı subbasin wit</u>	hin the mapped	plume					
1N/5W-21N2	PW-2	08/11/10	1.4 ± 0.1	-16.9 ± 0.4	0.0 ± 0.1	16.0 ± 0.3	0.0 ± 0.2	15.8 ± 0.4				
1N/5W-21P2	PW-3	08/10/10	1.1 ± 0.2	-18.0 ± 0.2	0.1 ± 0.1	-17.6 ± 0.4	0.1 ± 0.1	-17.1 ± 0.4				
1N/5W-33B1	N-10D	05/19/11	0.5	-14.1 ± 0.5	0.2 ± 0.0	-13.9	0.1	NA				
1N/5W-34B2	Rialto 02	02/25/12	0.2	-20.3 ± 0.2	0.0 ± 0.0	-19.5	-0.1	NA				
1N/5W-34B2	Rialto 02 (800')	02/25/12	0.3	-20.1 ± 0.1	0.0 ± 0.0	-19.9	-0.2	NA				
1N/5W-34G4	PW-5D	08/04/10	0.7 ± 0.4	-17.6 ± 0.6	0.0 ± 0.1	NA	NA	NA				
1N/5W-34M1	Rialto 03	03/15/11	0.6 ± 0.3	-15.2 ± 0.1	0.0 ± 0.2	-15.1 ± 0.1	0.1 ± 0.1	-14.6 ± 0.2				
1N/5W-34M1	Rialto 03 (670')	03/15/11	0.6	-14.4 ± 0.1	0.0 ± 0.0	-14.4	0.0	NA				
1S/5W-02B3	PW-9C	08/05/10	1.3	-16.1 ± 0.3	0.0 ± 0.1	-15.5	0.0	NA				
1S/5W-02G1	Rialto 06	03/11/11	1.3 ± 0.3	-16.9 ± 0.2	0.1 ± 0.1	-16.5 ± 0.1	0.2±0.1	-16.2 ± 0.2				
1S/5W-02G1	Rialto 06 (480')	03/11/11	1.0	-17.1 ± 0.0	-0.1 ± 0.1	-16.9	0.0	NA				
1S/5W-02G1	Rialto 06 (560')	03/11/11	1.1	-15.7 ± 0.2	0.0 ± 0.1	-15.0	0.0	-14.2 ± 0.2				
1S/5W-03A1	Rialto 04	10/20/10	1.1 ± 0.1	-17.6 ± 0.2	0.0 ± 0.2	-16.9 ± 0.1	0.1 ± 0.1	NA				
		<u>We</u>	lls outside th	e mapped plu	me_							
1N/5W-17K2	WVWD 24	06/29/10	1.5 ± 0.4	-8.4 ± 0.3	1.2 ± 0.1	-7.0	1.2	-6.8 ± 0.3				
1N/5W-27D1	Rialto 01	05/18/11	-0.2 ± 0.1	-15.9 ± 0.2	0.2 ± 0.0	-15.5 ± 0.5	0.2 ± 0.1	-15.3 ± 0.4				
1N/5W-33N1	F-49A	04/23/11	-2.4 ± 0.0	-8.5 ± 1.0	2.8 ± 0.2	-8.9	3.0	NA				
1S/4W-18G1	Colton 15	07/08/10	-12.3 ± 0.7	-4.6 ± 0.1	8.6 ± 0.1	-4.1 ± 0.2	8.6 ± 0.1	-4.0 ± 0.2				
1S/5W-02K1	WVWD 11	06/30/10	-0.5 ± 0.4	-15.3 ± 0.6	1.0 ± 0.2	-15.2 ± 0.2	1.1 ± 0.2	-15.0 ± 0.1				
1S/5W-05A5	F-10C	04/18/11	-4.3 ± 0.2	-9.8 ± 0.5	3.4 ± 0.1	-9.3	3.6	NA				
1S/5W-13B5	13B1-5 (RHSW 5)	08/12/10	-9.8	-7.1 ± 0.1	6.9 ± 0.0	-7.1	6.9	NA				
1N/5W-32A1	F-13A	05/11/11	-5.5 ± 0.2	-4.8 ± 0.0	4.7 ± 0.2	-4.8 ± 0.3	4.5 ± 0.1	NA				
		Wells in	the Chino (or	Arlington) sul	<u>obasins</u>							
1S/5W-23A1	WVWD 42	07/08/10	-10.3 ± 0.1	-7.0 ± 0.0	7.1 ± 0.1	-5.7 ± 0.7	7.1 ± 0.1	NA				
1N/5W-32N1	F-26A	12/12/10	-9.7	-8.3 ± 0.0	6.1 ± 0.0	-7.9	6.2	NA				
1N/5W-32N1	F-26A (800')	12/12/10	-9.9	-8.4 ± 0.1	6.2 ± 0.0	-8.0	6.3	NA				
1S/5W-04N1	F-4A	07/28/10	-11.8 ± 0.3	-7.9 ± 0.2	7.6 ± 0.1	-7.5 ± 0.2	7.6 ± 0.0	-7.3 ± 0.0				
1S/5W-04N1	F-4A (J613; 2007 sample)	03/09/07	-11.9	-8.0	7.6	-7.4	7.7	NA				
1S/5W-07R1	F-35A	07/27/10	-11.1	-9.4 ± 0.1	6.4 ± 0.1	-9.0	6.7	NA				
1S/5W-14B1	Chino 02	07/22/10	-11.1	-6.8 ± 0.2	7.4 ± 0.0	-6.5	7.5	NA				
1S/6W-23D2	F-17B	06/17/10	-13.5 ± 0.0	-6.6 ± 0.3	8.2 ± 0.1	-6.6 ± 0.6	8.3 ± 0.1	-5.9 ± 0.4				
1S/6W-23D2	F-17B (580')	06/17/10	-13.0 ± 0.2	-8.0 ± 0.2	8.1 ± 0.1	-7.5 ± 0.0	8.1 ± 0.0	-7.7 ± 0.1				
								•				

NA - data not available.

Means and standard deviations are provided for all individual analyses per well. When individual values are provided samples were analyzed once.

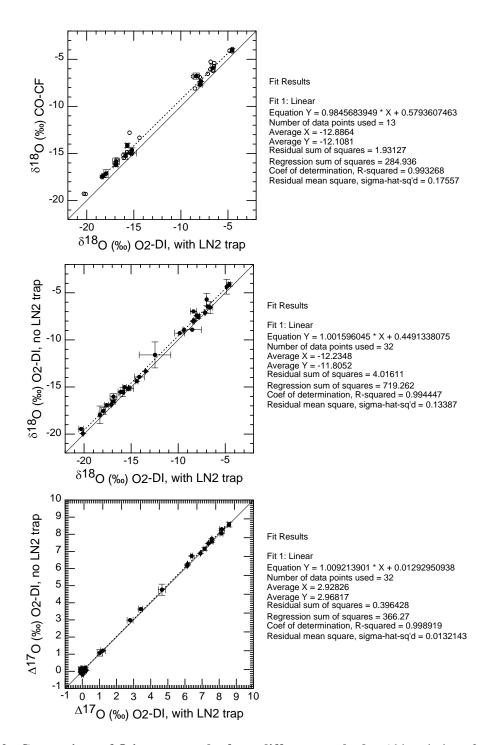


Figure 3.10. Comparison of O isotope results from different methods. Abbreviations for methods are: "CO-CF" = Conversion of O in ClO_4 sample to CO with isotopic analysis of CO by continuous-flow (CF) isotope-ratio mass spectrometry (IRMS); "O2-DI" = Conversion of O in ClO_4 to O_2 with isotopic analysis by dual-inlet (DI) IRMS; "with LN2 trap" = liquid N_2 trap in place to remove condensable gases (e.g., CO_2 , H_2O) at the inlet to the mass spectrometer; "no LN2 trap" = liquid N_2 trap removed, admitting condensable gases into the mass spectrometer. Solid lines indicate 1:1 slope (perfect agreement). Dashed lines are linear fits to the data. Open symbols in the upper panel represent samples prepared with TPA and not pre-baked, potentially containing excess C or N (not included in the fit).

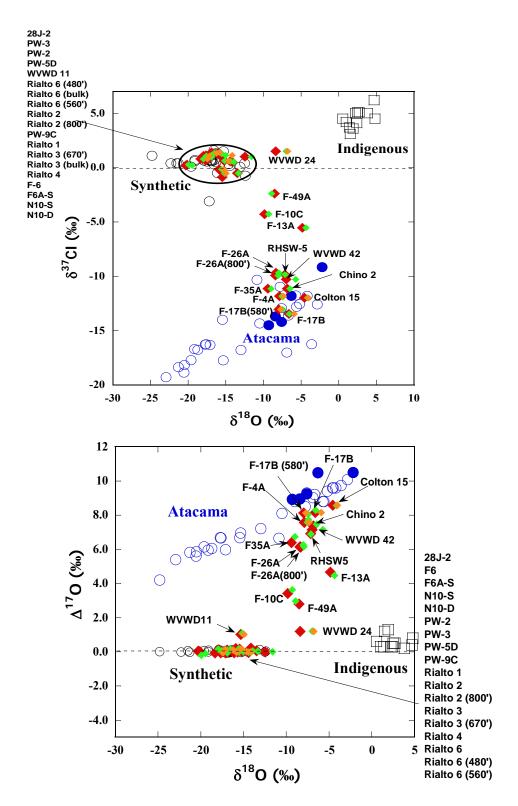


Figure 3.11. Comparison of $\delta^{37}\text{Cl}$ versus $\delta^{18}\text{O}$ (top panel) and $\Delta^{17}\text{O}$ versus $\delta^{18}\text{O}$ (bottom panel) in ClO_4 - from individual Study Area groundwater samples analyzed for O isotopes by $\text{O}_2\text{-DI-IRMS}(+\text{N}_2)$ (closed red diamonds), $\text{O}_2\text{-DI-IRMS}(-\text{N}_2)$ (closed green diamonds), or CO-CFIRMS (closed orange diamonds) with published data for natural indigenous ClO_4 - of the US (open black squares), natural Atacama ClO_4 -, (open blue circles and closed blue circles for actual fertilizer samples) and synthetic ClO_4 - (open black circles). Sources of ClO_4 - isotope data for Atacama, Synthetic and Indigenous as described in Figure 3.2.

3.5 Summary and Conclusions

Samples were collected from wells throughout the Study Area to evaluate Cl and O stable isotope ratios and ³⁶Cl abundances in ClO₄⁻. The dominant source(s) of ClO₄⁻ in each of these samples was evaluated based on the isotopic data. The data indicate that synthetic, Atacama, and indigenous natural ClO₄ were present in groundwater in the Study Area. In general, data for samples with relatively high ClO₄⁻ concentrations were consistent with either synthetic or Atacama ClO₄ being dominant, whereas data for samples with low concentrations were more likely to indicate mixtures including substantial indigenous natural ClO₄ fractions. The stable isotope data from nearly all samples from wells within two mapped ClO₄ contaminant plume areas, including those located in the perched aquifer (28J-2, F6, F6A-S, N-10S) and those in the regional aquifer (PW-2, PW-3, Rialto 2, Rialto 2 (880'), PW-5D, Rialto 3, Rialto 3 (670'), PW-9C, Rialto 6, Rialto 6 (480'), Rialto 6 (560'), and Rialto 4) were consistent with a dominant synthetic ClO₄ source in groundwater. The ³⁶Cl and δ¹⁸O data indicate that some of these samples also contained smaller fractions of indigenous natural ClO₄-. Downgradient from the mapped plume areas at wells 13B1-5 and Colton 15, and across the Rialto-Colton Fault to the southwest (Chino 2, WVWD 42, F-4A, F-35A, F-17B, F-26A, F-26A (800')), the dominant source of ClO₄ in analyzed samples was Atacama, presumably from historical application of fertilizers imported from the Atacama Desert of Chile. Past agricultural land use in these areas is consistent with the use of such fertilizers.

A sample from an upgradient well (WVWD 24) appeared to contain a mixture of synthetic and indigenous ClO_4^- based upon all isotopic measures. This sample indicated that other synthetic sources may be present in the Study Area besides those originating from the two mapped plume source locations. There is also a possibility that the ClO_4^- in this sample represented a pure indigenous natural source, with isotopic characteristics (primarily $\delta^{18}O$) that differ from the assumed values; however, the presence of tritium in this sample indicates a synthetic component cannot be ruled out. Interpretation of dominant sources in most samples did not differ substantially if one assumes that WVWD 24 represents a pure indigenous source, although this assumption is likely to affect the estimated fractions of indigenous ClO_4^- in mixing calculations.

Isotopic data for samples from wells outside the mapped plume areas with very low ClO_4^- concentrations (F-13A, F-49A, and F-10C) indicated mixed ClO_4^- sources. ClO_4^- from F-13A apparently was composed largely of Atacama and indigenous ClO_4^- , with a relatively large fraction of indigenous ClO_4^- . A small fraction of synthetic ClO_4^- was also possible based upon the $\delta^{18}O$ value. Based on the stable isotope values, ClO_4^- from F-10C and F-49A appeared to be composed of mixtures of all three sources. However, relations between $^{36}Cl/Cl$ and $\delta^{37}Cl$ values could also indicate mixtures of indigenous and Atacama ClO_4^- , with no apparent contribution from synthetic ClO_4^- . Uncertainties in the interpretation of ClO_4^- sources in these low-concentration samples are related in part to uncertainty and possibly local variability in the isotopic characteristics of indigenous natural ClO_4^- .

Depth-dependent sampling was conducted in several production wells to determine if different sources of ClO₄⁻ were present in these wells at different depth intervals (e.g., synthetic ClO₄⁻ in shallow groundwater and indigenous ClO₄⁻ in deep groundwater). The isotopic data generally were similar between the bulk discharge of the well and the depth dependent sample(s), indicating little or no variation in major ClO₄⁻ source(s) with depth in these wells.

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APPENDIX APerchlorate Extraction and Purification – Method Tests for QA/QC

A total of 27 separate Quality Assurance/Quality Control (QA/QC) tests were performed to evaluate the extent of Cl and O isotopic fractionation caused by the laboratory extraction and purification methods. These tests were done in addition to the field replication tests described in Section 3.4.1. The QA/QC tests were performed by the same analysts who processed all of the field samples at UIC (Abelardo Beloso, Jr. and Linnea Heraty). The perchlorates used in these QA/QC tests were the USGS37 and USGS38 KClO₄ reference materials. A few of the tests included potential interfering compounds, such as chlorate (ClO₃⁻) and organics, to measure their effects on ClO₄⁻ isotope ratios.

Twelve tests involved the entire procedure: loading the Purolite A530E ion exchange resin column (i.e., field column) with two liters of a KClO₄ solution (10 mg/L ClO₄⁻), followed by extraction and purification of the ClO₄⁻ from each column according to the procedures used for the field samples. These tests of the entire procedure used the following initial solutions: (1) four columns prepared by loading pure KClO₄ solutions made from deionized water plus USGS37 (two columns) or USGS38 (two columns); (2) four columns loaded with KClO₄ solutions made from USGS37 and USGS38 plus added ClO₃⁻ from reagent NaClO₃ (two columns made with USGS37 and two columns made with USGS38, both having ~1,000 mg/L ClO₃⁻); and (3) four columns loaded with KClO₄ solutions made from USGS37 and USGS38 plus added humic and fulvic acid from commercially available extracts (two columns made with USGS37 and two columns made with USGS38, each having ~50 mg/L humic + fulvic acid). The outcome of these tests for the six column tests having ClO₄⁻ from USGS37 resulted in little isotopic fractionation (Table A-1). The mean normalized values of δ^{18} O, Δ^{17} O, and δ^{37} Cl were within the stated analytical uncertainties of ± 0.2 , ± 0.2 , and ± 0.1 per mil, respectively, for the calibration values of USGS37. For the six column tests having ClO₄⁻ from USGS38, which was intentionally synthesized with elevated values of $\delta^{18}O$ (+52.5 %) and Δ^{17} O (+73.3 ‰), and a low value of δ^{37} Cl (-87.2 ‰), the mean normalized values of δ^{18} O, Δ^{17} O, and δ^{37} Cl had deviations of -1.8, -1.7, and +0.6 per mil, respectively. The estimated analytical uncertainties for USGS-38 are ± 0.5 and ± 0.3 for $\delta^{18}O$ and $\delta^{37}Cl$. The relatively large deviations for the O isotopes in the USGS38 tests may reflect the presence of small amounts of O-bearing impurities which are isotopically light relative to O in USGS38. If these impurities were present in the USGS37 method test samples, they did not significantly affect the resulting values of $\delta^{18}O$ and $\Delta^{17}O$. The deviation of the mean normalized $\delta^{37}Cl$ value for the USGS38 method tests, +0.6 per mil, is about two standard deviations higher than the accepted value and could be caused by small amounts of gaseous impurities in the CH₃Cl used for Cl isotope ratio measurements. However, the observed deviations in the O and Cl isotope ratios of the USGS38 tests, when spread over the entire range of the calibration between USGS37 and USGS38, would not cause significant changes to δ values within the ranges of the samples analyzed in this study.

Other QA/QC tests were done to examine the effects of specific steps in the purification procedure, after the extraction from the Purolite A530E resin (Table A-2). Compared with the deviations described above for the entire extraction and purification procedure, the post-extraction steps caused deviations that were comparable in direction and not significantly different in magnitude from those caused by the entire procedure. This implies that most of the observed isotopic effects associated with extraction and purification of perchlorate samples collected on Purolite A530E resin occur during the post-extraction purification steps, in which there are opportunities for addition of small amounts of O-bearing impurities (or C-bearing impurities that may become oxidized to CO₂ during the high-temperature KClO₄ decomposition step) during a number of adsorption, precipitation, and washing steps.

Table A-1. Evaluation of fractionation of Cl and O isotopes during extraction and purification of perchlorate from ion exchange columns.

Test ID	date (mm/yr)	mg CIO4	δ ³⁷ CI	δ ¹⁸ Ο	Δ ¹⁷ O	D37	D18	D17	description of test
UIC-EX1-37C	8/13	3.4	NA	-17.0	0.0	NA	0.0	0.0	USGS-37 solution in DIW, 10 mg/L, 2 L solution passed through column, entire procedure including TPA-CIO4 precipitation
UIC-EX1-37D	8/13	3.8	0.9	-17.7	0.0	0.3	-0.6	0.0	USGS-37 solution in DIW, 10 mg/L, 2 L solution passed through column, entire procedure including TPA-CIO4 precipitation
UIC-EX1-38C	8/13	3.3	-86.8	50.6	71.6	0.4	-1.9	-1.7	USGS-38 solution in DIW, 10 mg/L, 2 L solution passed through column, entire procedure including TPA-CIO4 precipitation
UIC-EX1-38D	8/13	2.8	-86.4	49.3	70.6	0.8	-3.2	-2.7	USGS-38 solution in DIW, 10 mg/L, 2 L solution passed through column, entire procedure including TPA-CIO4 precipitation
UIC-EX2-37C	8/13	2.9	0.8	-16.8	0.0	0.2	0.2	0.0	USGS-37 solution in DIW, 10 mg/L, plus 1000 mg/L chlorate, 2 L solution passed through column, entire procedure including TPA-CIO4 precipitation
UIC-EX2-37D	8/13	2.8	0.8	-16.6	-0.1	0.2	0.4	-0.1	USGS-37 solution in DIW, 10 mg/L, plus 1000 mg/L chlorate, 2 L solution passed through column, entire procedure including TPA-CIO4 precipitation
UIC-EX2-38C	8/13	3.2	-86.5	51.2	72.0	0.7	-1.3	-1.3	USGS-38 solution in DIW, 10 mg/L, plus 1000 mg/L chlorate, 2 L solution passed through column, entire procedure including TPA-CIO4 precipitation
UIC-EX2-38D	8/13	2.9	-86.7	50.8	71.7	0.5	-1.7	-1.6	USGS-38 solution in DIW, 10 mg/L, plus 1000 mg/L chlorate, 2 L solution passed through column, entire procedure including TPA-CIO4 precipitation
UIC-EX3-37C	8/13	4.2	0.9	-16.9	0.0	0.3	0.1	0.0	USGS-37 solution in DIW, 10 mg/L, plus 50 mg/L humic and fulvic acid, 2 L solution passed through column, entire procedure including TPA-CIO4 precipitation
UIC-EX3-37D	8/13	3.8	NA	-17.1	0.1	NA	-0.1	0.1	USGS-37 solution in DIW, 10 mg/L, plus 50 mg/L humic and fulvic acid, 2 L solution passed through column, entire procedure including TPA-CIO4 precipitation
UIC-EX3-38C	8/13	2.7	-86.7	50.5	71.5	0.5	-2.0	-1.8	USGS-38 solution in DIW, 10 mg/L, plus 50 mg/L humic and fulvic acid, 2 L solution passed through column, entire procedure including TPA-CIO4 precipitation
UIC-EX3-38D	8/13	4.2	NA	51.6	72.3	NA	-1.0	-1.0	USGS-38 solution in DIW, 10 mg/L, plus 50 mg/L humic and fulvic acid, 2 L solution passed through column, entire procedure including TPA-CIO4 precipitation

0.0 0.0 0.4 0.0 stdev USGS-38 -1.8 -1.7 mean

D37, D18, and D17 are deviations from reference values.

NA – data not available.

Reference values of $\delta^{18}O$, $\Delta^{17}O$, and $\delta^{37}Cl$ for USGS37 are -17.0 %, 0.0 %, and 0.6%, respectively. Reference values of $\delta^{18}O$, $\Delta^{17}O$, and $\delta^{37}Cl$ for USGS38 are 52.5 %, 73.3 %, and -87.2 %, respectively. See Equation 1.4 for definition of $\Delta^{17}O$.

Table A-2. Evaluation of fractionation of Cl and O isotopes during specific steps of perchlorate purification and precipitation.

Test ID	date (mm/yr)	mg CIO ₄	δ ³⁷ CI	δ ¹⁸ Ο	Δ ¹⁷ Ο	D37	D18	D17	description of test
RSIL-4 (125A)	11/11	3.1	0.6	-16.9	0.0	0.0	0.1	0.0	USGS-37 stock solution evaporated into combustion tube
RSIL-4 (125B)	11/11	3.1	0.7	-16.6	-0.1	0.1	0.4	-0.1	USGS-37 stock solution evaporated into combustion tube
RSIL-5 (132A)	11/11	3.1	-86.8	51.9	72.3	0.4	-0.6	-1.0	USGS-38 stock solution evaporated into combustion tube
RSIL-5 (132B)	11/11	3.1	-86.2	51.9	72.4	1.0	-0.6	-0.9	USGS-38 stock solution evaporated into combustion tube
RSIL-4 (81410AA)	11/11	1.7	0.7	-16.5	0.1	0.1	0.5	0.1	USGS-37 stock solution through SPE, precipitation, washing, dissolved in DIW and evaporated into combustion tube
RSIL-4 (81410AB)	11/11	1.5	0.6	-16.0	0.0	0.0	1.0	0.0	USGS-37 stock solution through SPE, precipitation, washing, dissolved in DIW and evaporated into combustion tube
RSIL-5 (92010AA)	11/11	2.1	-87.2	52.1	72.6	0.0	-0.4	-0.7	USGS-38 stock solution through SPE, precipitation, washing, dissolved in DIW and evaporated into combustion tube
RSIL-5 (92010AB)	11/11	2.0	-87.2	51.6	71.8	0.0	-0.9	-1.5	USGS-38 stock solution through SPE, precipitation, washing, dissolved in DIW and evaporated into combustion tube
RSIL-4 (1 mg ppt)	12/12	0.8	0.4	-16.4	-0.2	-0.2	0.6	-0.2	USGS-37 stock solution through TPA-CIO4 precipitation, conversion to KCIO4, evaporation into combustion tube, bake 10 min @ 400 C
RSIL-4 (3 mg ppt)	12/12	2.7	0.6	-16.6	-0.1	0.0	0.4	-0.1	USGS-37 stock solution through TPA-CIO4 precipitation, conversion to KCIO4, evaporation into combustion tube, bake 10 min @ 400 C
RSIL-5 (1 mg ppt)	12/12	0.8	NA	51.2	70.5	NA	-1.3	-2.8	USGS-38 stock solution through TPA-CIO4 precipitation, conversion to KCIO4, evaporation into combustion tube, bake 10 min @ 400 C
RSIL-5 (3 mg ppt)	12/12	2.7	-86.4	51.0	71.8	0.8	-1.5	-1.5	USGS-38 stock solution through TPA-CIO4 precipitation, conversion to KCIO4, evaporation into combustion tube, bake 10 min @ 400 C
UIC RSIL-4	7/13	2.5	1.3	-17.4	0.0	0.7	-0.4	0.0	USGS-37 stock solution through TPA-CIO4 precipitation, conversion to KCIO4, evaporation into combustion tube, bake 10 min @ 400 C
UIC-RSIL-4-Aug-2013	8/13	2.5	0.9	-17.3	-0.1	0.3	-0.3	-0.1	USGS-37 stock solution through TPA-CIO4 precipitation, conversion to KCIO4, evaporation into combustion tube, bake 10 min @ 400 C
UIC-RSIL-5-Aug-2013	8/13	2.4	-87.0	51.2	72.3	0.2	-1.3	-1.0	USGS-38 stock solution through TPA-CIO4 precipitation, conversion to KCIO4, evaporation into combustion tube, bake 10 min @ 400 C

D37, D18, and D17 are deviations from reference values.

NA – data not available.

Reference values of $\delta^{18}O$, $\Delta^{17}O$, and $\delta^{37}Cl$ for USGS37 are -17.0 %, 0.0 %, and 0.6%, respectively. Reference values of $\delta^{18}O$, $\Delta^{17}O$, and $\delta^{37}Cl$ for USGS38 are 52.5 %, 73.3 %, and -87.2 %, respectively.

See Equation 1.4 for definition of Δ^{17} O.

APPENDIX B Comparison of Cl and O Stable Isotope Values in ClO₄- by SIMS and IRMS

A set of sample splits (aliquots of purified KClO₄) were sent to Professor John Eiler of Caltech for isotope ratio measurements using the Cameca ims-7f GEO secondary ion mass spectrometer (SIMS), also known as ion microprobe, at the Caltech Microanalysis Center. The primary strength of SIMS stable isotope analysis is that very small sample quantities are required (µg quantitites). However, the method has yet to be fully validated for analysis of Cl and O isotopes in ClO₄, so the data and method should be considered experimental for this application. This instrument uses a ~1 nA primary Cs+ ion beam to sputter the sample surface, which generates secondary ions that are accelerated through a magnetic field to a set of detectors for isotope ratio measurement. The KClO₄ samples were provided as fine powders and a small amount of each sample was pressed into an indium sample mount for analysis. A thin layer of gold was deposited on the sample surface to prevent sample charging, then the sample mount was placed in the high-vacuum sample chamber of the ion microprobe for isotopic analysis. Analyses were performed by rastering the primary ion beam over a 50 µm x 50 µm area to average out the effects of rough sample surface topography. The measurements of ³⁷Cl/³⁵Cl were performed separately from the measurements of ¹⁸O/¹⁷O/¹⁶O under slightly different sets of analytical conditions (i.e., beam tuning and counting time) to optimize results. The results of the SIMS analyses are compared with IRMS results in Table B-1. In a memo accompanying the report of these results by Professor Eiler, he concludes that the SIMS method remains experimental and it should be considered less reliable than the IRMS method.

As previously noted, one of the strengths of SIMS is that it can produce isotope ratio measurements from tiny amounts of sample, much smaller than the amounts used for IRMS measurements. However, this can also be a weakness if the sample being measured is not isotopically homogeneous. With respect to measurements of ClO₄⁻ isotope ratios, the sample is measured in the form of solid KClO₄, which is precipitated from an aqueous solution after purification. Because there are small (approximately per mil level) mass-dependent isotopic fractionations of O and Cl between solid KClO₄ and aqueous ClO₄⁻, isotopic zoning will inevitably occur during precipitation of the solid phase. This can cause the last fraction of KClO₄ precipitate to be significantly depleted in the heavier isotopes ¹⁸O and ³⁷Cl relative to the cumulative solid KClO₄. For example,

if the isotopic fractionation factor is such that the solid KClO₄ is one per mil enriched in ¹⁸O and ³⁷Cl relative to the aqueous ClO₄⁻, then when 99% of the aqueous ClO₄⁻ has been removed from solution, the solid being precipitated at that instant is about 5 ‰ depleted in both ¹⁸O and ³⁷Cl relative to the cumulative solid KClO₄. Also, at the initial stage of precipitation, the solid phase is one per mil enriched in ¹⁸O and ³⁷Cl relative to the aqueous ClO₄⁻. In addition, trace contaminants that would have relatively little effect on bulk analysis could be heterogeneously distributed in a sample such that a highly focused ion beam could analyze a highly contaminated small portion of the sample. In contrast, during the IRMS measurements, a sample aliquot is transferred to a glass tube in liquid or bulk powder form and the entire aliquot is converted to O₂ and CH₃Cl for isotopic analysis, thus minimizing effects of trace contaminants or isotopic fractionation during precipitation of the KClO₄.

Table B-1. Comparison of IRMS and SIMS results for split samples collected from the Study Area.

	Local							
	well	Analytical	USGS	$\delta^{37}\text{CI}$	$\delta^{18}O$	$\delta^{17}O$	$\Delta^{17}O$	
Well	name	method	Sample number ¹	(‰)	(‰)	(‰)	(‰)	
1S/5W-02K1	WVWD 11	IRMS	N17910	-0.4	-15.7	-7.3	1.0	
1S/5W-02K1	WVWD11	SIMS (11/2013)	N17910	-3.4	-20.9	-8.4	2.7	
1S/6W-23D2	F-17B	IRMS	N17880	-13.5	-6.6	4.5	8.0	
1S/6W-23D2	F-17B	SIMS (7/2013)	N17880	-13.3	-5.3	4.5	7.3	
1N/5W-32N1	F-26A	IRMS	N19245	-9.7	-8.3	1.7	6.1	
1N/5W-32N1	F-26A	SIMS (7/2013)	NA ²	-11.9	-9.1	-0.8	4.0	
1N/5W-32N1	F-26A	SIMS (11/2013)	N19245	-8.7	-11.3	1.3	7.3	
1S/5W-04N1	F4A	IRMS	N17914	-11.6	-8.0	3.4	7.6	
1S/5W-04N1	F4A	SIMS (7/2013)	N17914	-13.9	1.6	7.1	6.2	
1S/5W-04N1	F-4A (J613; 2007 sample)	IRMS	N19688	-11.9	-8.0	3.3	7.6	
_	F-4A (J613; 2007 sample)	SIMS (11/2013)	N19688	-10.7/-8.6	-15.1	-1.5	6.5	
1N/5W-21N2	PW-2	IRMS	N17906	1.4	-17.2	-9.2	-0.1	
1N/5W-21N2		SIMS (11/2013)	N17906	-2.2	-15.1	-9.2	-1.2	
1S/5W-02G1	Rialto 06	IRMS	N17916	1.3	-16.9	-8.8	0.2	
15/5W-02G1	Rialto 06	SIMS (11/2013)	N17916 N17916	-7.6	-10.9	-3.4	3.2	
	LICOCO T	IRMS ³	and a		47.0			
	USGS37	SIMS ³	standard	0.6	-17.0	-9.0	0.0	
	USGS37	SIIVIS	standard	0.6	-17.0	-9.0	0.0	
	USGS38	IRMS ³	standard	-87.2	52.4	102.4	73.3	
	USGS38	SIMS (7/2013) ³	standard	-91.0	53.3	98.8	69.3	
	USGS38	SIMS (11/2013) ³	standard	-87.9	53.7	102.7	72.8	

¹ IRMS data are for specific samples (column + purification aliquot) that were split for SIMS analysis with the exception of F-17B, for which the δ^{37} Cl value is from the same resin column but different purification aliquot (see Table 3.6).

 $^{^{\}rm 2}\,$ It is unclear from the data report whether the F26A sample analyzed on 7/2013 was N19245.

³ IRMS measurements were calibrated to both USGS37 and USGS38. Values for the calibration standards are provided.

SIMS measurements were calibrated to USGS37, whereas USGS38 was analyzed as an unknown.

 $[\]Delta^{17}$ O values for USGS38 measured by SIMS were recalculated using Equation 1.4 for consistency with other data in this report

⁽ Δ^{17} O values for other samples do not differ significantly with different definitions of Δ^{17} O).