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EFFECTS OF SURFACE WATER WITHDRAWAL ON FISHES IN RIVERS OF NORTHEAST LOUISIANA

K. Jack Killigure OTHE FILE COP.

Environmental Laborator,

DEPARTMENT OF THE ARMA Waterways Experiment Station, Corps of Erginners PO Box 631, Vicksburg, Mississipp 139180 (Fat

and

Neil H. Douglas

Northeast Louisiana University Department of Zoology Monroe, Louisiana 71209-0520





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19. ABSTRACT (Continued).

In addition to water volume, the influence of other physical and chemical variables on fish abundance was evaluated using stepwise multiple regression. Variability in fish abundance was best explained by water volume, dissolved oxygen, and conductivity (R⁻ = 0.77). Other variables (pH, water depth, percent cover, temperature, turbidity, water velocity, and discharge) had no significant influence on increasing the predictive capability of the regression equation. Although changes in dissolved oxygen and conductivity may accompany decreases in water volume, the ability to make these predictions as part of an impact assessment was beyond the scope of this study. However, these variables should be considered as potential limiting habitat factors in rivers of northeast Louisiana during the summer and early fall months and can provide a basis of monitoring habitat quality under actual surface water withdrawal conditions.

An index of biotic integrity (IBI) was used to compare the quality of the existing fish community structure between study sites. The IBI can be incorporated in a management plan to identify both pristine and degraded habitat conditions, but is not intended as a predictive technique to estimate changes in fish abundance for future conditions. The IBI integrates attributes of species richness and composition, trophic composition, and fish abundance to rate the fish community structure as excellent, fair, or poor. In northeast Louisiana, poor habitat conditions are usually associated with high numbers of juvenile shad, while excellent habitat is dominated by minnows, shiners, and darters. Bayou Bartholomew was the only major river in the study area that consistently exhibited high species richness, trophic composition representative of undisturbed habitat, and a relatively high number of total fishes compared with other rivers in northeast Louisiana. High quality fish habitats that do exist in other rivers, such as gravel substrate, continuous flowing water, and high amounts of instream cover, are intermittent and uncommon and are usually associated with small tributaries or are below water-control structures. Bayou Bartholomew is one of the few remaining water bodies in northeast Louisiana that is considered an ecologically significant stream because of its high diversity of streamdwelling fish species.

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PREFACE

This study provides an inventory of the existing fish community structure and evaluates the effects of surface water withdrawal on fishes in northeast Louisiana as part of a water supply study on the Boeuf River, Tensas River, and Bayou Bartholomew basins being conducted by the US Army Engineer District, Vicksburg (VXD). Funding for this project was provided by VXD; partial funding for the development of the Index of Biotic Integrity was provided by the Environmental Impact Research Program (Work Unit 32390). This report was prepared by Mr. K. Jack Killgore, Environmental Laboratory (EL), US Army Engineer Waterways Experiment Station (WES), and Dr. Neil H. Douglas, Northeast Louisiana University, Monroe, La. Assistance in the field was provided by Mr. Kenneth Conley, Mr. Frank Ferguson, and Ms. Teressa Naimo, Aquatic Habitat Group (AHG), WES, and Mr. Jan Hoover, University of Oklahoma. Technical Monitor from VXD was Mr. Marvin Cannon. Technical reviews of the report were provided by Dr. Barry S. Payne, Dr. Andrew C. Miller, Mr. Johnny Franklin, AHG, and Dr. John M. Nestler, Water Quality Modeling Group, WES, and Dr. James A. Gore, University of Tulsa. The report was edited by Ms. Lee T. Byrne of the WES Information Products Division, Information Technology Laboratory.

This study was conducted under the general supervision of Mr. Richard Coleman and Mr. Edwin A. Theriot, Acting Chiefs, AHG; Dr. Conrad J. Kirby, Chief, Environmental Resources Division; and Dr. John Harrison, Chief, EL, WES.

Commander and Director of WES was COL Dwayne G. Lee, CE. Technical Director was Dr. Robert W. Whalin.

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CONTENTS

	Page
PREFACE	1
LIST OF FIGURES	3
CONVERSION FACTORS, NON-SI to SI (METRIC) UNITS OF MEASUREMENTS	4
PART I: INTRODUCTION	5
Background and Purpose Description of the Study Area Approach and Assumptions	5 5 6
PART II: FIELD METHODS	10
PART III: IMPACT ASSESSMENT OF SURFACE WATER WITHDRAWAL	13
Data Analysis Description of Future Water Demand Scenarios Determination of Minimum Water Volume Fish Abundance and Water Volume Without Water Demands Fish Abundance and Water Volume With Water Demands	13 15 15 17 18
PART IV: MULTIVARIATE RELATIONSHIPS BETWEEN HABITAT AND FISH ABUNDANCE	29
Data Analysis Results and Discussion	29 29
PART V: INTEGRITY OF THE FISH COMMUNITY STRUCTURE	32
Description and Development of the IBIApplication of the IBI on Rivers in Northeast Louisiana	32 33
PART VI: CONCLUSIONS AND RECOMMENDATIONS	35
REFERENCES	37
TABLES 1-10	
APPENDIX A: PROCEDURE TO ESTIMATE EXISTING WATER LEVELS	. Al
APPENDIX B: LISTS OF FISH SPECIES THAT OCCUR IN THE STUDY AREA	. Bl

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LIST OF FIGURES

No.		Page
1	Location of study area and field collection sites	11
2	Location of stream gaging stations used to determine historic water	
	levels in the Boeuf River, Tensas River, Bayou Macon, Bayou	
	Bartholomew, and Big Creek	14
3	Number of fishes with and without water demands (MGD) and the MV	
	for rivers in Richland Parish	19
4	Number of fishes with and without water demands (MGD) and the MV	
	for rivers in Tensas Parish	20
5	Number of fishes with and without water demands (MGD) and the MV	
	for rivers in Madison Parish	21
6	Number of fishes with and without water demands (MGD) and the MV	
	for rivers in Catahoula Parish	22
7	Number of fishes with and without water demands (MGD) and the MV	
	for rivers in Franklin Parish	23
8	Number of fishes with and without water demands (MGD) and the MV	
	for rivers in Morehouse Parish	24
9	Number of fishes with and without water demands (MGD) and the MV	
	for rivers in East Carroll Parish	25
10	Number of fishes with and without water demands (MGD) and the MV	
	for rivers in West Carroll Parish	26



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CONVERSION FACTORS, NON-SI TO SI (METRIC) UNITS OF MEASUREMENTS

Non-SI units of measurement used in this report can be converted to SI (metric) units as follows:

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Multiply	Ву	To Obtain
acre-feet	1,233.489	cubic metres
cubic feet	0.02831685	cubic metres
feet	0.3048	metres
gallons	3.785412	cubic decimetres
square feet	0.09290304	square metres

EFFECTS OF SURFACE WATER WITHDRAWAL ON FISHES IN RIVERS OF NORTHEAST LOUISIANA

PART 1: INTRODUCTION

Background and Purpose

1. Surface water demands are rapidly increasing in northeast Louisiana for irrigation and commercial fish farming and, to a lesser extent, for industrial and municipal purposes (Henning 1985). Because a diverse fish community inhabits the numerous streams, bayous, and rivers in northeast Louisiana (Douglas 1974), an increase in surface water demand will result in competition for available water supplies between human consumption and aquatic habitat. Water is required for crop irrigation and commercial fish farming throughout the summer and early fall months (Henning 1985) when stream levels are low. Therefore, reduction in water levels from surface water withdrawal may affect the spatial requirements of fishes for foraging, spawning, and predator avoidance (Fraser 1972, Petts 1984), leading to a decrease in their condition and abundance.

2. The purpose of this study was to determine the effects of surface water withdrawal on fishes in northeast Louisiana during the summer and early fall months, as part of a water supply study of the Boeuf River, Tensas River, and Bayou Bartholomew basins. The objectives of this report were to evaluate changes in fish abundance resulting from various future water demand scenarios developed by Henning (1985), determine important physical and chemical variables that may limit fish abundance, and document the fish community structure that currently exists in the study area.

Description of Study Area

3. This study focused only on relatively small rivers in northeast Louisiana that were bordered by irrigated, agricultural land. These included the Boeuf River, Tensas River, Bayou Macon, Big Creek, and Bayou Bartholomew. These rivers have undergone extensive water resource development in the form of channelization, single-purpose dams, and various types of weirs, dikes, jettys, and outlet structures. Rivers in the study area are usually nonflowing during the summer and early fall as the result of low water or

backwater effects from dams, diversions, logjams, and larger rivers (Black and Ouachita Rivers). However, measurable discharge does occur immediately below some larger dams (e.g., Gumby Dam on the Tensas River) and other outlet structures. The substrate is composed of clay and sand. An exception is the Bayou Bartholomew, where gravel riffles still exist and flowing water occurs yearround. Trees commonly fall into the rivers and provide the only substantial instream cover available to the fishes.

4. Although removal of water from rivers should have no substantial effect on water levels in the immediate vicinity of the pumps because of inflow from upstream or downstream sources, the numerous low-water dams, logjams, and diversions that occur throughout the study area prevent a freeflowing exchange of water from the headwaters to the mouth during the summer and fall. For example, rock dams are commonly placed below a water intake to form a pool and ensure an adequate water supply during low-water periods. Consequently, surface water withdrawal should result in a decrease in water volume near the vicinity of the intake structure. The dams and diversions also make it difficult to establish a reliable stage-discharge relationship for gaging stations in the study area in order to apply hydraulic models for predicting changes in discharge over time.* Although stage height fluctuates throughout the summer and fall, it generally results from backwater effects after rainfall and continuous surface water withdrawal for irrigation.* This is particularly pronounced in the upper reaches of the Boeuf River, which becomes virtually dewatered as a result of surface water withdrawal.

Approach and Assumptions

5. A threefold approach was used in this study and incorporated both abiotic and biotic variables to predict impacts of surface water withdrawal on fishes, identify potential limiting factors on fish abundance, and classify streams according to the quality of the fish community structure. First, impacts of future water demands on fish abundance were determined according to a relationship between water volume and number of fish. Second, the

^{*} Personal communication, July 1987, Robert Walsworth, US Geological Service, Ruston, La., and Tommy Reynolds, US Army Engineer District, Vicksburg, Vicksburg Miss.

importance of other physical and water quality variables on fish abundance was evaluated using multiple regression techniques. Third, species and trophic compositions of the fish community structure were compared between rivers in the study area to identify ecologically significant river reaches.

Relationship between water volume and fish abundance

6. The lack of reliable stage-discharge relationships at the various gaging stations, the influence of the numerous dams, diversions, and water intakes on the flow regime, and the size of the study area (approximately 700 river miles) prevented the utilization of established water quality and hydraulic models in the impact assessment process. Consequently, the mean number of fish per unit volume of water was calculated based upon field data collected throughout the study area and multiplied by future changes in water volume according to Henning's (1985) water demand report. It was assumed that the magnitude of declines in fish abundance from future surface water withdrawal could be estimated from an existing mean number of fishes per unit volume of water. Validation of this assumption could be made by monitoring fish populations under actual withdrawal conditions. In addition, it was assumed that water removed from specific river reaches would not be replaced from other instream sources because of the influence of dams, diversions, and other structures and that long-term reductions in water levels would occur. Shortterm fluctuations in water levels might only displace fishes, followed by recovery of stream volume and habitat quality. Conversely, long-term declines in water levels over several months could decrease fish abundance due to an overall reduction of usable stream habitat. Since water demands for irrigation and commercial fish farming occur from May through October (Henning 1985), long-term reductions in water levels will likely occur if surface water ic used.

7. The primary objective of this study was to estimate changes in fish abundance resulting from Henning's (1985) future water demand predictions. However, Henning's report was not written for use in biological impact analysis, and the application of his results to this study presented several major problems. First, Henning determined changes in water volume demands without consideration to other habitat variables such as water quality. Second, Henning determined total water demands without indicating the amount required from surface water. Walter (1982) provides the only data (for 1980) on water

usage that distinguishes surface from ground-water withdrawal (Table 1). It was assumed that the rate of future ground-water withdrawal would equal that of 1980. Only ground water was used in Tensas and Madison Parishes according to Walter (1982); therefore, it was assumed that there would be no surface water demands for these two parishes. For parishes requiring some surface water in 1980, it was assumed that all increased demands in the future would be met solely by increased use of surface water. Thus, the percent surface water used in each parish according to Walter's (1982) estimates was multiplied by total water demands determined by Henning (1985) to obtain future demands of surface water only. Third, Henning did not indicate the source of water (river, lake, or pond) but simply expressed demands by parish. As a worst-case scenario, it was assumed that all surface water demands would be met by the five rivers in the study area. Although there are numerous oxbow lakes along the Mississippi and other rivers in this area, their inclusion in this analysis was beyond the scope of the study.

Multivariate habitat analysis

8. The distribution and abundance of fishes are influenced by a variety of habitat factors other than just water volume (Whiteside and McNatt 1972; Platts 1979; Fausch, Karr, and Yant 1984; Miranda, Shelton, and Bryce 1984; Schlosser 1985). The multivariate approach to impact analysis on fish populations has become an established technique to determine fish abundance due to changes in physical, chemical, and biotic variables. For example, using multiple regression, Binns and Eiserman (1979) developed a habitat quality index (HQI) that related standing crop of coldwater fishes to nine habitat attributes. Oswood and Barber (1982) used a similar approach for salmonids. Several multivariate habitat models for warmwater streams have also been developed to predict fish standing crop (Paragamian 1981, Layher and Maughan 1985), although some are subjective and not well verified (McClendon and Rabeni 1987). Other habitat assessment methods that use indices to describe the quality of the environment to fishes include the Instream Flow Incremental Methodology (IFIM) (Bovee 1982) and the Habitat Evaluation Procedures (HEP) (US Fish and Wildlife Service (USFWS) 1980). These and other habitat-based models are still evolving in an attempt to develop an acceptable fishery habitat classification system for resource planning and management (Platts 1980).

9. A multivariate analysis was employed in this study to identify selected physical and water quality variables that may limit fish abundance. The results were not used to predict changes in fish abundance according to Henning's (1985) water demands, because, as previously mentioned, variables other than water volume were not provided for future conditions and the ability to predict changes in other habitat variables resulting from surface water withdrawal was beyond the scope of this study.

Species and trophic compositions of fishes

10. The final approach used in this study was to describe and compare the integrity of the fish community structure in northeast Louisiana using the index of biotic integrity (IBI) proposed by Karr (1981). This index evaluates an aquatic resource based on the attributes (species composition, trophic composition, and health and abundance of fish) of the indigenous fish community (Leonard and Orth 1986). The fish community can be classified as excellent, fair, or poor according to the final IBI score. This index can ultimately be used to identify ecologically significant areas in northeast Louisiana that are sensitive to water resource development and to monitor changes in the quality of the fish community structure.

PART II: FIELD METHODS

11. Relationships between fish abundance (number of fishes) and physiochemical variables were examined at 15 different river locations in the study area during low-water periods in September and October (Figure 1). Numerous studies have shown that fish assemblages can be quite different due to the quality of fish habitat (Hynes 1970; Krumholz 1980; Ross, Matthews, and Echelle 1985); therefore, sites were selected to account for variability in channel shape and dimensions, amounts of instream cover, water velocity, and substrate type (gravel or clay). Each individual site, however, was relatively uniform in habitat features. The length of the sites ranged from 33 to 380 ft* in order to assess the relationship between water volume and fish abundance. However, the mean (+SD) length of all sites was 148 +92.

12. Each site was isolated with upstream and downstream blocknets (0.5-mm mesh). Three consecutive passes of equal effort were made through the site using a boat-mounted electroshocker (output was 350 to 400 v at 4 to 7 amp). All fishes collected were identified to species and measured (standard length). The number of fishes per unit volume of water was determined using the Zippin Depletion Method (Zippin 1958), and the standard error of these estimates was determined as described by Platts, Megahan, and Minshall (1983). Based upon frequency of occurrence (individuals per class divided by total individuals), fishes were separated into the following groups: harvestable sport and commercial species, juvenile sport and commercial species, minnows, darters, madtoms, and rough fishes. Fishes of harvestable size were identified using length criteria furnished by the US Army Engineer District, Vicksburg (VXD).

13. After fishes were collected, depth, velocity, cover (presence or absence), and water quality were measured at 2- to 5-ft intervals across a transect and used to describe the morphology, water volume, and average water quality conditions for the study site (Table 2). Because each individual site was relatively uniform with respect to channel shape and location of instream cover, a single transect was considered adequate to represent the habitat variability within the site. Water depth was measured to the nearest 0.1 ft

A table of factors for converting non-SI units of measurement to SI (metric) units is presented on page 4.



Figure 1. Location of study area and field collection sites

using a metered rod. Water velocity was measured to the nearest 0.1 ft/sec using a Marsh-McBirney Model 201 current meter. If total depth (TD) was less than or equal to 3.0 ft, velocity was measured at 0.6 TD. If TD exceeded 3.0 ft, velocity was measured at both 0.2 and 0.8 TD. Percent cover was determined by dividing the intervals with cover by the total number of intervals across the transect. Water quality parameters were measured with a Martek (Mark XV) water quality analyzer and included temperature, dissolved oxygen, pH, conductivity, and turbidity.

PART III: IMPACT ASSESSMENT OF SURFACE WATER WITHDRAWAL

Data Analysis

14. Based on the fish population and water volume measurements collected at the study sites (Table 2), the mean number of fish (\pm 95-percent confidence interval) per l-million gal of water was 1,064 \pm 291. One site was deleted (Bayou Macon 3.5 miles above Hwy 2 near Oak Grove at riffle RM 132) because the fish abundance value (19,489 fish/l-million gal) was disproportionately high and not considered representative of the study area. Based upon 20 years of fish collecting in the study area, riffle habitat in Bayou Macon is rare and is usually created from silt deposition or logjams. Therefore, use of this value would artificially increase the variability of the mean estimate and would not accurately depict fish-volume relationships for a basin-wide study. However, this site shows the importance of riffletype habitat on fish abundance and should be considered in a site-specific assessment.

15. The mean estimate was held constant and multiplied by water volumes currently existing in each river under wet to drought conditions and as a result of future water demands. Water volumes were determined from historic stage height data collected at 13 gaging stations (Figure 2) spread throughout northeast Louisiana. Each gaging station represented hydrological conditions at a specific reach of river in the study area. The procedure used to calculate water volumes that exist for wet to drought conditions and associated number of fishes is outlined in Appendix A.

16. Future water demands were developed for all parishes in Louisiana by Henning (1985). However, only the parishes where the five rivers occurred were considered in this study and included Catahoula, East Carroll, Franklin, Madison, Morehouse, Richland, Tensas, and West Carroll. Henning predicted the amount of water required for crop irrigation and aquaculture from May to October over a 40-year period (1990 to 2030) according to four scenarios: 50-percent chance of water need without conservation measures, 50-percent chance of water need with conservation measures, 90-percent chance of water need without conservation measures, and 90-percent chance of water need with conservation measures. For each scenario, a mean water demand was calculated by month from the values of the 5 target years (1990, 2000, 2010, 2020, 2030)



Figure 2. Location of stream gaging stations used to determine historic water levels in the Boeuf River, Tensas River, Bayou Macon, Bayou Bartholomew, and Big Creek to simplify data presentation. Water demand was expressed as million gallons of water per day (MGD) and was calculated by dividing the number of days in each month into the total monthly water volume. Therefore, this value indicates the amount of water available for withdrawal on a daily basis using constant pumping rates and does not refer to a stream discharge rate.

Description of Future Water Demand Scenarios

17. The four water demand scenarios are described below according to Henning (1985). Scenario 1 (50-percent chance of water need without conservation measures) projected supplemental irrigation water requirements for rice, soybeans, cotton, wheat, corn, grain sorghum, and the necessary amount of water to maintain catfish ponds under conditions of average (50-percent chance of water need) rainfall in northeast Louisiana. Scenario 2 (50-percent chance of water need with conservation measures) estimated water use for average rainfall conditions with conservation employed to irrigation and aquaculture practices. On-farm conservation measures included land leveling, flow measurement devices, recycling of water, and matching irrigation systems to soil and crop conditions. Off-farm conservation measures, although not normally practiced, included lining conveyance canals and laterals, weed control along conveyance channels, and improved scheduling allocation. Scenario 3 (90-percent chance of water need without conservation measures) estimated water use when rainfall was below normal. Below normal rainfall conditions were considered to be a drought situation where historic average rainfall was expected to exceed estimated rainfall 9 years in 10. Scenario 4 (90-percent chance of water need with conservation measures) estimated water use for drought conditions with conservation measures previously described.

Determination of Minimum Water Volume

18. A threshold value (the minimum volume necessary to maintain a viable fishery) was determined for the rivers in each parish using a modification of the Tennant Method (Tennant 1976) and referred to as "minimum volume" (MV). The Tennant Method uses a predetermined percentage of the historic average water discharge (volume in this study) to indicate the quality of fish habitat that ranges from "flushing or maximum" to "severe

degradation." This method was chosen because it is relatively unbiased in that it does not incorporate subjective reasoning into the recommended water volume and it can be developed quickly and inexpensively (Annear and Conder 1984); however, it does not incorporate habitat preferences of fishes or seasonal variability in water levels in the decision-making process.

19. In this study, 40 percent (defined as "good" by Tennant) of the median monthly water volume (50-percent exceedance value) was used to determine the MV. The median monthly water volume was determined by parish according to the procedure in Appendix A. If two or more rivers existed in one parish, their median water volumes were summed, and the MV was calculated from this value. However, the MV can be determined for each individual river in a parish using the data provided in Table 4. Although other studies commonly used 30 percent of the mean annual water volume (Annear and Conder 1984), this value was considered too low, since in most cases it would result in water Lovels typical of severe drought situations. Tennant (1976) concluded that 30 percent of the mean annual water volume would provide adequate water levels to cover most substrates, provide some instream cover for fishes, allow most side channels to carry some water, and provide adequate water temperatures that would not become limiting to the fishes. Therefore, 40 percent of the median water volume by month is considered a conservative value and incorporates monthly variability in water levels.

20. No single technique is available to objectively define a minimum water volume necessary to maintain a viable fishery in rivers where flow occurs intermittently because of dams and other water restrictions. The Tennant Method was originally applied to salmonid riverine habitat and was based on percentages of mean annual discharge. The use of the median water volume, rather than the mean annual value, provided a reasonable minimum volume based upon local hydrological conditions, but should be used with caution because of its lack of empirical verification. Other "minimum flow" techniques previously used in impact analysis include the wetted perimeter method, the habitat retention method, use of physical habitat simulation models, and subjectively identified inflection points on hydrographs of habitat-discharge relationships (Annear and Conder 1984). However, these methods have been designed for flowing water conditions and are not considered appropriate for rivers in northeast Louisiana.

Fish Abundance and Water Volume Without Water Demands

21. Based upon a review of historic fish collections at the Museum of Zoology at Northeast Louisiana University, a total of 116 species have been identified in the study area (excluding Ouachita and Black Rivers), with 51 common and 65 uncommon species (Appendix B). Bayou Bartholomew has the highest number of species (97), followed by Boeuf River (69), Big Creek (50), Bayou Macon (43), and Tensas River (38) (see Appendix B). In this study, 59 species were collected by electroshocking. Juvenile shad (36.6 percent) was the most abundant species (Table 3), particularly in rivers other than Bayou Bartholomew. Shad have broad niche requirements and are often abundant in sluggish rivers, impoundments, and areas where habitats have been disturbed by water resource development (Carlander 1969, Pflieger 1975, Becker 1983). Minnows, darters, and madtoms were the second most abundant group of fishes (34.6 percent) throughout the study area and were usually the dominant group of fish at Bayou Bartholomew. These species usually have narrow niche requirements, are sensitive to environmental degradation, and dominate the fis' assemblage in warmwater rivers that sustain flows year-round (Pflieger 1975, Pennington et al. 1981, Becker 1983, Page 1983). Juvenile sunfishes were the only other group of fishes common in all rivers. Harvestable sport and commercial fishes made up less than 10 percent of the total number of fishes collected.

22. The available water volume (MGD) that currently exists in the study area and associated numbers of fish are shown in Table 4 by river and parish. The water volume is provided for a range of high-water (river stage is exceeded 30 percent of the time) to extreme low-water (river stage is exceeded 90 percent of the time) conditions. In addition, the stage height that corresponds to the volume of water is given for a representative gaging station (Figure 2). Rivers occasionally formed partial boundaries between parishes and were assigned to the parish where the highest number of river miles occurred.

23. The water volume was highest in May at all parishes and steadily decreased throughout the summer months because of lower amounts of rainfall. West Carroll and East Carroll Parishes had the least amount of surface water, whereas Tensas and Franklin Parishes had the highest amount of surface water. Monthly rish abundance values ranged from 6,000,000 (199,700 fish/MGD) in the

lower reaches of Bayou Bartholomew in Morehouse Parish during high-water periods to less than 5,000 (100 fish/MGD)in the headwaters of the Boeuf River in West Carroll Parish during drought conditions. The abundance of harvestable sport and commercial fishes can be determined for a given river reach by multiplying the percentage of a particular group shown in Table 3 by the fish abundance value.

Fish Abundance and Water Volume With Water Demands

24. The effects of surface water withdrawal on fish abundance are presented by parish in Figures 3 through 10 for each of the four scenarios previously described. Each figure indicates water volume (MGD) and associated number of fish with and without demands. In addition, the MV for normal and drought conditions is provided to indicate the amount of water available to partially meet the demands and still provide adequate habitat to maintain the existing fish community structure. If two or more rivers occurred in one parish, their water volume and fish abundance values were summed by month for the "without demand" variable (see Table 4). The "without demand" variable for Scenario I (50-percent chance of water need without conservation measures) and Scenario 2 (50-percent chance of water need with conservation measures) reflect the 50-percent exceedance values (median water volume) shown in Table 4, while Scenario 3 (90-percent chance of water need without conservation measures) and Scenario 4 (90-percent chance of water need with conservation measures) reflect the 90-percent exceedance values (drought conditions). A "loss" of fishes referred to in subsequent paragraphs can be caused by fishes leaving the area for an extended period of time, high mortality rates for those fishes stranded because of extremely low-water conditions, or a reduction in recruitment of future year classes because of degraded habitat conditions.

25. Water demands exceeded total water supply for most parishes. Except for Richland, Tensas, and Madison Parishes, a 100-percent loss of fishes would occur in 1 or more months as the result of complete dewatering of the river (Table 5). Fish losses were similar with and without conservation measures. In Richland Parish, where water demands were relatively low, Scenarios 1 and 2 resulted in only a slight decrease in water volume with a maximum fish loss of 2 percent (Figure 3). During drought conditions







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Figure 7. Number of fishes with and without water demands (MGD) and the MV for rivers in Franklin Parish

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has the highest water supply, followed by Tensas, Catahoula, Richland, Morehouse, East Carroll, Madison, and West Carroll Parishes.

30. Juvenile shad would be the primary fish lost because of surface water withdrawal. This species is an important forage fish for sport and predator fishes (Becker 1983) and can form a major part of the diet for at least 17 sport fishes (Miller 1960). Therefore, the cumulative effects of losing high numbers of shad may have significant effects on sport fish production. In addition to a loss of forage fishes, a relatively high number of harvestable sport and commercial fishes will be lost as the result of surface water withdrawal. For example, out of a total of 1-million fishes lost in typical river reaches, 2,000 would be harvestable bass, 3,000 would be harvestable sunfishes, 4,000 would be harvestable crappie, 5,000 would be harvestable catfish, and 17,000 would be harvestable buffalo. Recruitment would also be affected, especially for sunfishes, since 18 percent of the fishes collected were juvenile sunfishes. A high number of minnows, darters, and madtoms (34 percent) would also be lost because of dewatering effects, particularly in Bayou Bartholomew. Although these fishes do not directly contribute to the sport and commercial fishery, they are important forage fishes and ecologically significant. They also indicate a high diversity fish community, and studies have suggested that these types of aquatic systems are more affected by perturbations than those of low diversity (Petts 1984).

31. This study shows that the water demands predicted by Henning (1985) will create major impacts to the fish community structure in northeast Louisiana under the assumptions previously stated. Alternative water supplies will have to be identified in order to meet future water demands, or many of the rivers will be dewatered as a result of surface water withdrawal. The assumptions made in this analysis should be critically reviewed if new data become available. For example, assuming that future ground-water withdrawal will remain at the 1980 rate may be erroneous, but no new information is available to modify these predictions. If new water demand data do become available, the fish abundance estimates can be used to modify the impacts of surface water withdrawal to the fish community. The validity of the fish abundance-water volume relationship can also be tested in any long-term monitoring efforts that may occur in the future under actual water withdrawal conditions.

PART IV: MULTIVARIATE RELATIONSHIPS BETWEEN HABITAT AND FISH ABUNDANCE

32. Fish abundance can be influenced by a variety of habitat variables other than water volume. The interaction of both physical and chemical properties of the aquatic environment can regulate the size and distribution of fish populations. This section provides an analysis of the physical habitat and water quality variables measured in association with fish population estimates to determine which and how many variables are most important in predicting fish abundance; this section also identifies potential limiting factors on fish abundance in the rivers of northeast bouisiana.

Data Analysis

33. The 10 physical and water quality variables measured in the field (Table 2) were separately correlated to the fish population estimates. Variables with correlation coefficients near or greater than 0.30 were entered into a stepwise multiple regression using the maximum R^2 improvement technique (SAS Institute Inc. 1985). A predictive equation was developed to explain the majority of variations in fish abundance according to these physical and water quality variables. The precision of the equation was examined by regressing predicted fish abundance values against actual fish abundance values measured in the field and the examining correlation coefficients (McClendon and Rabeni 1987).

Results and Discussion

34. Of the 10 habitat variables measured in the field, only 5 had correlation coefficients (R) near or greater than 0.30 (Table 7). These included water volume (0.75), conductivity (-0.46), pH (0.46), water depth (0.32), and dissolved oxygen (0.29). However, water volume was the only variable significantly (P < 0.05) correlated to fish abundance. Because of the narrow range of pH values (7.0 to 7.8) throughout the study area, this variable was eliminated. Therefore, only water volume, conductivity (Cond), water depth, and dissolved oxygen (DO) were subjected to stepwise multiple regression.

35. Fish abundance was best determined from water volume, conductivity, and dissolved oxygen. The predictive equation is expressed as follows:

Number of fish = 9.946 + 570.461 (Vol) + 58.505 (D0) - 0.837 (Cond) (1)

This equation explained 77 percent of the variation in fish abundance (R⁺ = 0.77) and was significant at P < 0.01. Water depth increased the R^{2} to only 0.78 and was therefore not used in the predictive equation. Correlation of the predicted and observed fish abundance values shows a relatively high level of precision (R^{2} = 0.88) and was significant P < 0.01). However, the accuracy of the equation can be determined only from an independent data set collected in future years.

36. As water volume decreases, the amount of usable fish habitat is reduced, and inter- and intraspecific competition for food, predator avoidance, and suitable spawning areas becomes more likely. Therefore, water volume should be highly correlated to fish abundance, unless other habitat variables become limiting. The results of this analysis has identified dissolved oxygen and conductivity as two potential limiting factors on fish abundance. Dissolved oxygen is important to the physiological, biochemical, and behavioral processes in fishes (Davis 1975). Low dissolved oxygen usually results in low species richness and abundance of fishes. For example, fish abundance and species richness were relatively low at the Boeuf River 1 mile above Hwy 15 near Alto, where the lowest dissolved oxygen value (4.7 mg/ ℓ) was measured during the study (Table 2). Conductivity was important to fish abundance due to the influence of Bavou Bartholomew. The water in Bavou Bartholomew has a lower conductivity than the other four rivers sampled as well as a higher number of fish per unit volume (see Table 2). Therefore, a negative correlation existed between conductivity and fish abundance (Table 7). The degree of land utilization practices and nutrient loading can be indicated by conductivity. Most rivers in agricultural environments, such as northeast Louisiana, are subjected to high rates of sedimentation, usually composed of fine clav materials high in colloidal material or organic matter "Schmidt 1972), causing the water to be highly conductive. Thus, an increase in conductivity may coincide with degrading habitat conditions.

37. The predictive equation presented previously can be used in determining changes in the number of fish resulting from decreasing water volumes

with associated changes in dissolved oxygen and conductivity. However, the difficulty of predicting water quality changes resulting from altered hydrology of the rivers in northeast Louisiana may limit the usefulness of the equation. Furthermore, the predictive equation does not necessarily imply a cause and effect relationship, since fish populations can be regulated by other unmeasured variables such as competition, predation, and extreme climatic conditions.

PART V: INTEGRITY OF THE FISH COMMUNITY STRUCTURE

38. Another approach to impact analysis of water resource projects is to determine changes in the biotic component of aquatic environment. The 1b: proposed by Karr (1981) provides a means of evaluating the status of the fish community structure according to biotic variables that can be measured in the field. The IBI can assess the biological integrity of the stream resource and, along with information on physical and chemical conditions, should provide a sound basis for management decisions (Angermeier and Karr 1986). An IBI was developed for rivers in northeast Louisiana to compare the quality of the fish community structure between study sites. In addition, the IBI can be used to monitor impacts of surface water withdrawal, as well as other water resource projects in northeast Louisiana, to the fish community structure if fish population data are collected under future impact conditions.

Description and Development of the IBI

39. The IBI consists of three biotic categories, each composed of different attributes (metrics). The categories include species richness and composition, trophic composition, and health and abundance of fish (Table 8). The value of each metric within the three categories reflects a level of stream degradation. The basic premise is that low habitat quality is associated with relatively low species richness, fewer numbers of total fishes, and a high number of omnivores. Further explanation of each metric is explained in Karr (1981); Fausch, Karr, and Yant (1984); Angermeier and Karr (1986); and Leonard and Orth (1986).

40. The observed value of each metric was determined from the fish population estimates taken at each study site (see Table 3). Several study sites were deleted from this analysis (both sites at Bayou Macon near Oak Grove and both sites on Lake LaFourche) because they were considered too small to accurately represent the fish community structure. Prior to calculating the IBI score, all species were placed into the trophic categories of omnivores, insectivorous cyprinids, and top carnivores according to literature based information (Table 9). The observed metric values were then assigned a score from 1 (worst) to 5 (best) based on their relationship to pristine or relatively undisturbed habitat conditions. After all metric criteria were

set, the individual metric scores were added to obtain a total IBI score that was used to classify the fish community as excellent, fair, or poor Angermeier and Karr 1986). Input from biologists familiar with the study area is necessary to develop a defensible IBI. Furthermore, an IBI may be specific to a drainage basin because metric values will vary with stream size and zoogeographic region (Karr 1981; Fausch, Karr, and Yant 1984).

41. Modifications of IBI metrics proposed by other researchers were necessary to account for different fish assemblages and habitat conditions that exist in northeast Louisiana. In the category of species richness and composition, high proportions of green sunfish (Karr 1981) or creek chubs (Leonard and Orth 1986) usually indicate degraded habitat conditions. However, in rivers of northeast Louisiana, high numbers of shad are more appropriate indicators of degraded habitat conditions. The presence of intolerant species (Karr 1981) was deleted from this category because their selection was considered a subjective process (Leonard and Orth 1986). Two additional metrics were deleted from the category of fish abundance and health proportion of individuals as hybrids, and proportion of fish with disease or anomalies. Identification of hybrids is difficult, even for an experienced ichthyologist, and fishes with disease or anomalies were not observed during field collections. However, the existence of hybrids and fish with disease or anomalies over 1 percent indicates highly degraded habitat conditions (Karr 1981; Fausch, Karr, and Yant 1984; Leonard and Orth 1986). It should be noted that hybrids in the genus Lepomis and Notropis have been collected in all rivers in northeast Louisiana over the past 20 years.

Application of the IBI on Rivers in Northeast Louisiana

42. The values of the fish community metrics at Bayou Bartholomew were equal to or higher than the metrics measured at other rivers in northeast Louisiana (Table 10). Species richness, as well as fish abundance, was generally higher at Bayou Bartholomew because of relatively high numbers of minnows, shiners, and darters. Shad were relatively lower in number cless than 36 percent of the total number of fishes) at Bayou Bartholomew than at other rivers and were usually confined to the lower reaches of Bayou Bartholomew near its confluence with the Ouachita River. Conversely, shad were distributed throughout all reaches of the other four rivers, except for the extreme headwaters. The proportion of individuals as insectivorous cyprinids was highest (indicating high quality habitat) at Bayou Bartholomew, except for Bayou Macon near Delhi, where a high number of insectivorous red shiners were collected. The proportion of individuals as omnivores was higher in rivers of northeast Louisiana, including Bayou Bartholomew, than in rivers where IBT metrics have been developed (Fausch, Karr, and Yant 1984; Angermeier and Karr 1986), primarily because of the high numbers of omnivorous shad collected at each study site. However, this is to be expected because of regional variability in fish assemblages and habitat quality. With rew exceptions, the number of sunfish species, the number of sucker species (buffalo were the only suckers collected during this study), and the proportion of individuals as top carnivores were consistent across study sites.

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43. Based upon the values of each individual metric across study sites, 20 years of collecting fish in the study area, and considering Bayou Bartholomew as an indicator of relatively pristine habitat conditions, scoring criteria were developed for each metric (Table 8). Total IBI scores ranged from a high of 35 at Bayou Bartholomew near Bastrop to a low of 15 at the Boeuf River 1 mile above Alto (Table 10). The IBI scores for the three sites at Bayou Bartholomew were equal to or greater than the scores determined for the other rivers, indicating the ecological importance of Bayou Bartholomew in northeast Louisiana. In conclusion, river reaches with IBI scores greater than 29 should be considered excellent habitat, whereas IBI scores between 20 and 29 represent fair habitat and scores below 20 represent poor habitat. Excellent fish habitat do occur in isolated areas in northeast Louisiana other than Bavou Bartholomew (e.g., Big Roaring Bayou on the Tensas River, see Table 2), but are usually confined to small tributaries that have not been subjected to severe bank erosion and that have high amounts of instream cover (greater than 30 percent), and to short reaches below water control structures with measurable discharge during the summer and fall months.

PART VI: CONCLUSIONS AND RECOMMENDATIONS

44. The existing water volume and associated number of fishes, as well as the maximum amount of water that can be withdrawn on a monthly basis for each parish while still maintaining a viable fishery, are provided over a range of wet to dry conditions for each river and parish. These data can be used to estimate the effects of any water withdrawal scenario on the fish community structure.

45. Future water demands determined by Henning (1985) will result in the rivers being dewatered for 1 or more months with fish losses ranging from 5,000 to 4,000,000 for the following parishes: Catahoula, East Carroll, Franklin, Morehouse, and West Carroll. Relatively low water demands occur in Richland Parish and have little effect on fish abundance. No surface water demands occur for Tensas and Madison Parishes according to published water usage reports. The assumptions made in this analysis should be critically reviewed if new water demand data become available. However, given the current data base, alternative water supplies should be considered, or substantial effects will occur to the fish community structure as the result of dewatering from surface water withdrawal.

46. Water volume, dissolved oxygen, and conductivity were identified as three important habitat variables that can potentially limit fish abundance in rivers of northeast Louisiana, and when incorporated into a multiple regression equation, provide a relatively high predictive capability to estimate number of fishes. Other physical and chemical habitat variables (water depth, water velocity, discharge, percent cover, water temperature, pH, and turbidity) had no significant influence on fish abundance.

47. A diverse fish community exists in northeast Louisiana with over 100 species of fishes residing in the numerous rivers, streams, and bayous. Bayou Bartholomew has the most diverse and abundant fish community because it has flowing water year-round with scattered amounts of gravel substrate. The dominant group of fishes collected at Bayou Bartholomew consisted of minnows, shiners, darters, and madtoms. In contrast, other major rivers in northeast Louisiana have clay or sand substrate and are usually nonflowing during the summer and early fall because of various types of water restrictions. In addition, the fish assemblage is composed primarily of juvenile shad. An IBI identified Bayou Bartholomew as an ecologically significant area in northeast Louisiana with excellent habitat conditions, whereas other major rivers in the study area were classified as fair to poor habitat.

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Parish	Percent Ground Water	Percent Surface Water
Catahoula	25.1	74.9
East Carroll	92.2	7.9
Franklin	14.1	85.9
Madison	100.0	0.0
Morehouse	41.4	58.9
Richland	98.6	1.4
Tensas	100.0	0.0
West Carroll	72.6	27.4

Percent Ground Water and Surface Water Used in 1980 To Meet Total Agriculture and Aquaculture Water Use Demands*

* Based on Walter (1982).

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Table l

Description of theh Abundance. Morphometry, and Mater Quality for Each study she to wortheest 5 wheleve

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511.	Number of Fish r SE	Number of Fish Species	Million Gallone	Fish per L million gellon	• • • • • • •	Mean Depth ft	Meen Velucity frise	latu - harge - to	Percent Lover	1 2 4 8 1 1 2 1 7 4 2 1 2 1 1	Diamolued Daygen 1871	ž	tondur - tivity umbos/	Turbiates MIT
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Group	Percent Occurrence	Length Criteria
Harvestable sunfishes	0.3	>127
Harvestable crappie	0.4	>191
Harvestable black bass	0.2	>241
Harvestable bullheads	0.1	× _ 03
Harvestable shad	0.3	>279
Harvestable catfish	0.5	> 305
Harvestable gar	0.9	> - 05
Harvestable buffalo	1.7	> 305
Harvestable carp	1.1	>356
Juvenile sunfishes	18.4	≤127
Juvenile shad	36.6	s279
Juvenile crappie, bass, bullheads, catfish, gar,	. 2	
burralo, and carp	4.5	-
Minnows, darters, and madtoms	34.6	all sizes
Drum, bowfin, and herrings	0.6	all sizes

Table 3

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Percent Occurrence of Harvestable Fishes for Rivers in Northeast Louisiana

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Total Exercation Event state	Mail Laun	Tenses	Hwy 80 at	71 miles trom the	Mary	10.14	28.4	30.2	(8./	15.7	16.7	64	1	1.1	15.5	1.1	
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$ \begin{array}{c c c c c c c c c c c c c c c c c c c $				Madisun Parish	Jul	6.89	10.5		6.16	6.)	÷ 4	5.81	4 1	4 . 4	\$ 14	т ~	1.1
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				Madiaun Tensas	Sep	5.75	4.4	1.4	5.60	3. 5	8 1	5 T T	1 2	1.9	12.5	1.1	-
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	1 - 1 - 3 - 8	Tensaa	Hev 4 near	47 miles trom the	¥∎¥	8 47	263.4	280.1	1.14	1 677	4.447	. 8	2.01	184	7 2	111.6	
$ \begin{array}{cccccccccccccccccccccccccccccccccccc$			New I IKhr	Madison-lenses	lun	B. 74	1.44.	265.7	17.5	1 191	1/1.4	15.1	143.0	152.1	1	10.7.01	
				Partsh line to the	Jul	15.1	1.16.2	144.4	7 7	1.16.4	134 5	8.42	18 6	4 F	4 [7		. :
$ \begin{array}{cccccccccccccccccccccccccccccccccccc$				Tensas fatahouia	Aug	34.3	121 5	1.21	14.1	0.251	1 101	1 12	(67	4 14		r 4	. /
$ \begin{array}{cccccccccccccccccccccccccccccccccccc$				Puttin line	Sep	14 2	1 341 6	0.911	14.1	5 671	11/ 8	23 e	1 2 1	1. 1	6 77	-	,
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$ \begin{array}{cccccccccccccccccccccccccccccccccccc$	- at strong a	Tensa .	Heve 15 Lin	29 miles from the	Hav	4 47	H.861	254.1	6.1.3	110.7	124 2	, ,;	160 2	1/1: 5	14.1		140
$ \begin{array}{cccccccccccccccccccccccccccccccccccc$			(layton	Tensas-Catahoula	hun	14.1	2.10.0	144.1	11.2	164.1	174 6	14 4	5 65 1		 -:	4 81	
$ \begin{array}{cccccccccccccccccccccccccccccccccccc$				Parish line to the	Int	34.2	137.5	146.3	939	130.7	1 961	11 6	1.28	- 4(1		-	7
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$ \begin{array}{c ccccccccccccccccccccccccccccccccccc$				*1.<=-	Sep.	1.0	107.7	6 071	11.5	4 11 1	0.021	20.4	10 4	, 		11	-
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$ \begin{array}{cccccccccccccccccccccccccccccccccccc$				Trunsylvania (**	171	64.0	1.7	1.6	61.1		1.5	- 14	-	-	5.7 S	-	÷
$ \begin{array}{cccccccccccccccccccccccccccccccccccc$				the base carroll	Aug	4.5.0	2.4	2.5	63.0	(e. 1	-	R 12	r 2	1. 		2	
$ \begin{array}{cccccccccccccccccccccccccccccccccccc$				Pladison Failth Hae	đe,	1.1.1	11.4	0.1	6.2.H	0.4	n. I	1.1.4	÷	-			
$ \begin{array}{rcccccccccccccccccccccccccccccccccccc$					1 10	6.2.7	i i	0.7	4.14		1.1		:	4	-	÷	-
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Description Main Diff. Bits				Fast Varrell	171	1.51	8.7		1.52	¢.5	6.4		;		- 4.		-
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					1 10		5.4	1.0	81.1		-	7,			•		

Table 5 Percant Luan of Flahae Resulting from Surface Mater Atthdramal According to Scenariou 1-4

	:	Yak				Iun				1414				Auduk								October		,
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Pacinh -			-	4	 -	~	 -	-	 -	~	 -	+	 	~	 -	4	 -	~		4	-	~	-] 4
s at above a	c	-	-	^	::	91	100	001	89	75	100	100	001	66	100	100	80	61	001	100	7	-	3	001
Prankijo	-	14	6	14	4	7	100	001	001	001	001	100	100	100	100	001	100	100	100	001	94	01	÷	6
R. rehouse	÷	7	001	100	5 7	**	001	166	57	55	001	001	44	44	001	001	17	20	100	001	17	41	25	;
kaat vattoil	2	+	-	,	001	100	100	100	100	001	001	100	001	100	100	100	001	100	001	1001	5		001	HR
Vest tarril	ŕ	9	Ŧ	ĩ	001	100	×	ŝ	100	106	61	62	001	001	f 4	74	001	001	81	81	100	100	19	Ţ
Richtand	-	Ţ	~	-	~,	7	11	7	-	1	11	17	-	~	41	1	-	7	77	41	÷	÷	-	

Table b

Comparison Between Mater Demanda for Mormal* and Dry ** Conditions With (C) and Without (MC) conservation and the Maximum Available Mater supply That Can Be Withdrawn and Still Maintain a Viable Hishery

		2			10	au au		Jul.	1		Aug	LINE		iepr	ember		1 10	ober
	Den d	pu	Maximum Availability	De De	Pu	Maxiaum Availability	Den U	pu -	Maximum Availability	Ben HGI		Maximum Availability Availability		pu	Maximum Availability	Den Den		Mastaua Availability
Tation Auf theory of wares need	-1	el			zl	1111	-			-	E		-	Ĕ	A.F	-	z!	8- E
(atahoula	01	0	126	н	04	14	89	68	18	114	871	11	8	501	6/	5	:1	2
Franklin	;	(5	254	61 1	181	202	101	190	153	411	177	151	269	058	851	74	÷	451
Marehouse	101	501	105	770	151	70	1/2	15.2	÷.	157	151	1	10	9	01	~.	-	4
Fast (arroll	~	~	77	7	71	~	Ξ	7	~	2	î	~	÷	\$	~	-		3
West Larroll	\$	~	\$	61	74	~	۶. ۶	8,	~	9 9	85	7	44	65	۰.	÷	x	2
R1. hland	-	-	111	~	-	56	-	4	61	-	-	11	-	-	16	÷	÷	4 P
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003 . hanse of water need (afahoula	71	51	E J	÷#		c	917	087	c	H-, 7	\$45	c	451	465	c	1.61	T.	2
8 r duh 1 fu	; ;	\$\$	124	234	80	89 v	\$15	184	17	648	244	99	404	141	-	113	1,1	
Murehuuse	104	56.1	0	£ #.7	141	÷	111	7117	16.5	161	H 577		ī	÷	Ξ	2	÷	
tant Introll	•	*	0	<u>۴</u>	R	÷	5	1,	÷	0	.:	۰.	с. Т	5	÷.,		÷	÷.,
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Murmaul = "M-perient chance of water need,
 Use = "M percent chance of water need,

Habitat Variables	Symbol .	<u> </u>	<u> </u>	F	<u>P</u>
Volume of Water, ft ³	x,	0.75	0.56	14.29	0.003
Conductivity, umhos/cm	x ₂	-0.46	0.21	3.01	0.11
рH	x ₃	0.46	0.22	3.04	0.11
Depth, ft	x ₄	0.32	0.10	1.29	0.28
Dissolved oxygen, mg/ ϵ	× ₅	0.29	0.08	1.01	0.34
Percent cover	x ₆	0.18	0.03	0.35	0.56
Turbidity, NTU	x ₇	-0.12	0.01	0.16	0.70
Velocity, ft/sec	x ₈	0.06	<0.01	0.04	0.84
Temperature, °C	x ₉	0.02	<0.01	0.01	0.93
Discharge, ft ³ /sec	x	0.02	<0.01	<0.01	0.96

Table 7Correlations Between Numbers of Fish and Habitat Variables

for Rivers in Northeast Louisiana

Table 8

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Scoring Criteria for Metrics Used in the Index of Blotic Integrity for Rivers in Northeast Louisiana*

		Sci	oring Crite	rla
(ategory	Metric	5 (Best)	3 (Fair)	1 (Worst)
Species richness and composition	Total number of fish species	> 20	15-20	<15
	Number of darter species	>3	1-3	0
	Number of sunfish species	۲<	5-7	<5
	Number of sucker species	>2	1-2	0
	Proportion of individuals as shad	201>	10-502	> 50%
Trophic composition	Proportion of individuals as omnivores	× 507	50-802	> 807
	Proportion of individuals as insectivorous cyprinids	~ 20%	5-202	< 5.7
	Proportion of individuals as top carnivores	%9	3-62	< 37
F1sh abundance	Number of individuals/0.5 acre	>600	300-600	< 300

Modiffed from Fausch, Karr, and Yant (1984) and Leonard and Orth (1986). *

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Trophic Classification of Fishes Collected in Rivers of Northeast

		Insectivorous		
Species	Omnivores	Cyprinids	Piscivores	<u>Unknown/Other</u>
Chestnut lamprey			х	
Spotted gar			Х	
Longnose gar			Х	
Shortnose gar			Х	
Bowfin			Х	
Skipjack herring			Х	
Threadfin shad	Х			
Gizzard shad	Х			
Common carp	Х			
Silver chub		Х		
Emerald shiner		Х		
Bullhead minnow		Х		
Redfin shiner		Х		
Spotfin shiner		Х		
Blacktail shiner	Х			
Weed shiner	Х			
Bluntnose minnow		Х		
Silvery minnow	Х			
Ribbon shiner				Х
Red shiner	Х			
Smallmouth buffalo	Х			
Bigmouth buffalo	Х			
Black buffalo	Х			
Channel catfish	Х			
Blue catfish	Х			
Flathead catfish			Х	
Yellow bullhead	Х			
Freckled madtom				Х
Freshwater eel			Х	
Blackspotted topminnow	Х			
Mosquito fish	Х			
Brook silverside				Х
Spotted bass			Х	
Largemouth bass			Х	
Green sunfish			Х	
Warmouth			Х	
Orangespotted sunfish	Х			
Bluegill	Х			
Dollar sunfish	Х			
Longear sunfish	Х			
Redear				Х
Spotted sunfish	Х			

Louisiana from August-October 1986

(Continued)

		Insectivorous		
Species	Omnivores	Cyprinids	Piscivores	Unknown/Other
Black crappie			X	
White crappie			Х	
River darter				Х
Dusky darter				Х
Speckled darter				Х
Logperch				Х
Bluntnose darter				Х
Cypress darter				Х
Freshwater drum	Х			

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APPENDIX A: PROCEDURE TO ESTIMATE HISTORIC WATER LEVELS IN THE STUDY AREA

The purpose of this appendix is to describe the procedure used to determine the volume of surface water (million gallons) in five rivers over a range of high-water to drought conditions by parish.

Step 1. Obtain stage height readings of historic water levels conditions for each river. Stage duration tables were obtained for 13 stream gaging locations in northeast Louisiana (see Figure 2 in main text). Stage duration tables are cumulative frequency distributions of daily or monthly river stage heights measured over a 10- to 20-year period of record by the US Army Engineer District (VXD), Vicksburg, or the US Geological Survey (USGS) and are expressed as percent exceedance values. The 90-percent exceedance value indicates extreme dry conditions when stream levels are lowest, whereas the 10-percent exceedance value coincides with relatively high-water levels, usually resulting from an above normal amount of rainfall. Gage locations were chosen to represent morphological and hydrological conditions of specific reaches of the Boeuf River, Tensas River, Bayou Macon, Big Creek, and Bayou Bartholomew. Table 4 (see main text) shows the stage duration values for each river reach in the study area.

Step 2. Develop correlations between river stage and water volume for each gaging station and summarize existing water volume by parish. Each gage location represented a defined reach of a particular river. However, Henning (1985)* presented future water demand scenarios by parish and did not indicate the source of the water. Therefore, to relate existing water volume to future demands, the monthly volume of water that occurred at each representative gage location during wet to dry water conditions (30- to 90-percent exceedance values) was extrapolated to the entire length of one or more rivers lying in each parish and was expressed as million gallons of water per day (MGD). This was accomplished by first obtaining stream width and depth measurements collected at various stage heights by USGS and VXD survey crews and converting them into regression equations to predict stream width and average depth at stage heights representing the 30- to 90-percent exceedance values. The number

• See References at the end of main text.

of river miles in each parish was then determined from USGS 1:24,000 topographic maps using a cartometer. For each stage height, water volume (million gallons) was calculated by multiplying surface area (stream width < stream length: by the appropriate water depth and dividing by 1 million. This value was then divided by the number of days that occurred in each month to obtain MGD. If more than one river existed in a parish, the sum of their water volumes were taken to represent total water volume by parish. The mean fish abundance value (number of fishes per 1-million gal of water) determined from the field-measured population estimates was multiplied by the volume of water for wet to dry water conditions (see Table 4). As discussed in the main text, this procedure assumed that water removed from a specific reach of the river will not be replaced from other upstream or downstream sources because of the numerous dams and other water restrictions that exist throughout the study area and that changes in fish abundance can be explained from an existing relationship between numbers of fish and water volume. Although there will certainly be exceptions, it was concluded that these assumptions will hold true for the majority of river reaches in northeast Louisiana.

Step 3. Synthesize stage heights for ungaged stream reaches. The lower reach of Bayou Bartholomew did not have an established gaging station, and the water levels were substantially different from those of upstream locations where gaging stations were located. Therefore, historic water levels had to be synthesized using field-measured discharges at the lower reaches of Bayou Bartholomew and correlated to the upstream gaging location at Jones. A transect was established across several downstream locations (Hwv 139 and near the mouth), and discharge was measured. The stage-discharge readings measured by VXD were obtained for the gage near Jones on the same day the discharge was measured at the downstream locations. The percent change in discharge was determined between the downstream segments and the gage at Jones. It was assumed that discharge was steady from the gage at Jones to the lower reaches of Bayou Bartholomew. The water volume (MGD) that occurred at Jones over the range of historic stage heights was increased by this percentage to represent the water volume at downstream reaches of Bayou Bartholomew.

APPENDIX B: LISTS OF FISH SPECIES THAT OCCUR IN FIVE RIVERS IN NORTHEAST LOUISIANA

Table Bl

Distributional Status of Fish Species Throughout the Study Area

Common Species Uncommon Species Southern brook lamprey Spotted gar Goldstripe darter Longnose gar Chestnut lamprev Speckled darter Shortnose gar Shovelnose sturgeon Redfin darter Bowfin Paddlefish Logperch Gizzard shad Alligator gar Channel darter Threadfin shad Skipjack herring Blackside darter Grass pickerel Goldeve Saddleback darter Chain pickerel Mooneve Dusky darter Stoneroller River darter Carp Silverv minnow Goldfish Stargazing darter Golden shiner Grass carp Sauger Emerald shiner Cvpress minnow Walleve Ghost shiner Speckled chub Stiped mullet Pugnose minnow Silver chub Inland silversije Weed shiner Silver carp Redfin shiner Pallid shiner Blacktail shiner Bigeve shiner Mimic shiner Ironcolor shiner Bullhead minnow Southern striped shiner Lake chubsucker Ribbon shiner Smallmouth buffalo Bluehead shiner Bigmouth buffalo Longnose shiner Spotted sucker Red shiner Black bullhead Taillight shiner Yellow bullhead Sabine shiner Channel catfish Silverband shiner Flathead catfish Steelcolor shiner Tadpole madtom Bluntnose shiner Golden topminnow Creek chub Blackstripe topminnow River carpsucker Mosquito fish Quillback Pirate perch Blue sucker White bass Highfin carpsucker Flier Creek chubsucker Green sunfish Golden redhorse Blacktail redhorse Warmouth Orangespotted sunfish Blue catfish Brown bullhead Bluegill Brindled madtom Dollar sunfish Longear sunfish Freckled madtom Brown madtom Redear

(Continued)

Table B1 (Concluded)

Common Species

Uncommon Species

Spotted sunfish Bantam sunfish Spotted bass Largemouth bass White crappie Black crappie Bluntnose darter Cypress darter Drum Brook silverside American eel Yellow bass Crystal darter Western sand darter Scaly sand darter Mud darter Creole darter Swamp darter Slough darter Harlequin darter

Name	Scientific Name	Occurrence
Spotted gar	Lepisosteus oculatus	common
Longnose gar	Iepisosteus caseua	common
Shortnose gar	Lepiscoteus platootomus	common
Bowfin	Amia calva	common
Gizzard shad	Iorozoma cepedi rum	abundant
Threadfin shad	Iorozoma petenense	uncommon
Mooneye	Hiodon tengious	uncommon
Grass pickerel	Esox xmenioanus verminulatus	common
Chain pickerel	Esox niger	common
Carp	Syprinus carpic	uncommon
Cypress minnow	Hybognathus hayi	common
Silvery minnow	Hybognathus nuchalis	common
Speckled chub	Eybopsis aestivalis	abundant
Silver chub	Hybopsis storerizna	uncommon
Golden shiner	Notemigonus cryscleucas	abundant
Pallid shiner	Notropis armis	common
Emerald shiner	Sctropis atherinoides	abundant
Bigeye shiner	Notropis boops	rare
Ghost shiner	Notropis buchanani	common
Ironcolored shiner	Notropis chalybaeus	common
Southern striped shiner	Notropis chrysocephalus isolepis	rare
Pugnose minnow	Notropis emiliae	abundant
Pibbon shiner	Notropis fumeus	abundant
Bluehead shiner	Notropis hubbsi	uncommon
Red shiner	Notropis lutrensis	common
Taillight shiner	Nctropis maculatus	abundant
Weed shiner	Notropis texanus	abundant
Redfin shiner	Netropis umbratilis	abundant
Blacktail shiner	Notropis venustus	abundant
Mimic shiner	Notropis volucellus	abundant
Steelcolor shiner	Notropis whipplei	rare
Bullhead minnow	Pimephales vigilax	abundant
River carpsucker	Carpiodes carpio	common
Cuillback	Carpiodes cyprinus	rare
Highfin ca rpsucker	Carpiodes velifer	rare
Blue sucker	Cycleptus elongatus	rare
Creek chubsucker	Erimyzon oblongus	abundant
Lake chubsucker	Erimyzon sucetta	rare
Smallmouth buffalo	Ictiobus bubalus	common
digmouth buffalo	Istictus cyprine lus	common
Black buffalo	Istiobus niger	uncommon

Table B2

List of Fishes of Bayou Bartholomew Drainage*

(Continued)

* Species list compiled from an examination of 61,055 specimens taken from 189 collections at 41 locations.

(Sheet 1 of 🔅

Table B2 (Continued)

Name	Scientific Name	Occurrence
Spotted sucker	Minytrema melanops	abundant
Golden redhorse	Mewestema erythmamm	rare
Blacktail redhorse	Moxostoma poedilurur	rare
Blue catfish	Istalurus furcatus	common
Black bullhead	istalums melas	uncommon
Yellow bullhead	Istalume natalis	common
Channel catfish	Istalurus punctatus	abundant
Tadpole madtom	lotume gyrinus	common
Freckled madtom	Noturus nostumus	abundant
Flathead catfish	Eylodietis olivaris	uncommon
American eel	Anguélla rostrata	rare
Golden topminnow	Eundulus phryestus	common
Blackstripe topminnow	Fundulus notatus	uncommon
Starhead topminnow	Fundulus notti	uncommon
Blackspotted topminnow	Fundulus clivaceus	abundant
Mosquito fish	Ambusia affinis	abundant
Pirate perch	Aphredoderus suyanus	abundant
White bass	Norone shrysops	uncommon
Yellow bass	Morene mississippiensis	uncommon
Flier	Sentrarchus macropterus	uncommon
Green sunfish	Lepomis cyanellus	common
Warmouth	Lepomis gulosus	common
Orangespotted sunfish	Lepomis humilis	abundant
Bluegill	Iepomis macrochimus	abundant
Dollar sunfish	Lepomis marginatus	common
Long ear sunfish	Lepomis megalozis	common
Redear sunfish	Lepomis mierolopius	uncommen
Spotted sunfish	Lepomis punctatus	common
Bantam sunfish	Lepomie symmetricus	common
Spotted bass	Mieropterus punetulatus	uncommon
Largemouth bass	Micropterus salmoides	common
White crappie	Ecmoxis annuluris	uncommen
Black crappie	Eomoxis nigromaculitus	commen
Banded pigmy sunfish	Elassema zonatum	abundant
Crystal darter	Armoerypta asprella	raré
Western sand darter	Armoerypta elara	rare
Scaly sand darter	Armoerypta vivax	uncommen
Mud darter	Ethecstoma asprigere	common
Bluntnose darter	Etheostoma enloresemen	abundant
Swamp darter	Etheostoma fusiformo	uncommon
Slough darter	Ethecotoma gradily	commen
Harlequin darter	Ethesetoma Rietels	abundant
Goldstripe darter	Ethe stoma partirinne	rare
Cypress darter	stresstora propis pe	common
Speckled darter	Etheostoma indianaeur	uncommon

(Continued)

Sheet _ t is

Table B2	(Concluded)
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Name	Scientific Name	Occurrence
Redfin darter	Etheostoma whitplei	common
Logperch	Peroina caprodes	uncommon
Blackside darter	Eeroina maculata	uncommon
Saddleback darter	teroina cuachitae	uncommon
Dusky darter	Peroira eciera	common
River darter	Eensina shunandi	common
Stargazing darter	Eeroira uranidea	rare
Freshwater drum	Apicainotus grunniens	common
Brook silverside	Labidesthes sizeulus	abundant

Sheet 5 of 3)

Table B3

List of Fishes of the Bayou Macon Drainage*

Name	Scientific Name	Occurrence
Spotted gar	Lepisosteus coulatus	uncommon
Shortnose gar	Leviscoteus platostomus	common
Gizzard shad	Iorosoma cepediarum	abundant
Grass pickerel	Esca americanus vermiculatus	common
Carp	lyprinus carpic	common
Silvery minnow	Eybegnathus nuchalis	uncommon
Speckled chub	Eybopsis aestivalis	uncommon
Golden shiner	Notemizonus prysoleupus	common
Emerald shiner	Corropis atherinoides	uncommon
Ghost shiner	Notropis buchanani	common
Red shiner	Notropis Intrensis	common
Redfin shiner	Actropis umbratilis	common
Blacktail shiner	Notropis venustus	common
Mimic shiner	Notropis volucellue	uncommon
Bullhead minnow	Fimephales vigilar	abundant
Lake chubsucker	Erimyzon sucetta	uncommon
Smallmouth buffalo	Ictiobus bubalus	common
Bigmouth buffalo	Isticbus cyprinellus	common
Black buffalo	Ictiobus niger	uncommon
Yellow bullhead	Istalu r us natalis	common
Channel catfish	Istalu rus punctatus	common
Tadpole madtom	Noturus gyrinus	common
Freckled madtom	Noturus nocturnus	rare
American eel	Anguilla rostrata	uncommon
Golden topminnow	Fundulus chrysotus	uncommon
Starhead topminnow	Fundulus notti	uncommon
Blackspotted topminnow	Fundulus olivaseus	abundant
Mosquito fish	Gambusia affinis	abundant
Pirate perch	Aphredoderus sayanus	uncommon
Green sunfish	Iepomis cyanellus	common
Warmouth	Lepomis gulosus	common
Orangespotted sunfish	Iepomis humilis	cemmon
Bluegill	Lepomis macrochirus	abundant
Dollar sunfish	Lepomis marginatus	abundant
Longear sunfish	Lepomis megalotio	abundant
Spotted sunfish	Iepomis punctatus	uncommon
Bantam sunfish	Lepomis syrmetricus	uncommon
Largemouth bass	Micropterus salmoides	common
White crappie	Fomozis annularis	common
Black crappie	Eomoxis nigromaeulatus	common
Banded pigmy sunfish	Elassoma zonatum	common
Freshwater drum	Aplodinetus grunniene	common
Brook silverside	Labidesthes sicculus	common

* Species list compiled from an examination of 2,140 specimens taken from 14 collections at 11 locations.

Table B4

List of Fishes of the Big Creek Drainage*

Name	Scientific Name	Occurrence
Spotted gar	Lepiscoteus coulatus	common
Longnose gar	Lepisosteus osseus	uncommon
Bowrin	Amia calva	uncommon
Gizzard shad	Icrosoma cepedianum	abundant
Threadfin shad	Corosoma peterense	common
Grass pickerel	Esox americanus vermiculatus	common
Chain pickerel	Esca riger	uncommon
Carp	Syprinus carpis	common
Silvery minnow	Eybognathus nuchalis	common
Speckled chub	Hybopsis aestivalis	common
Golden shiner	Notemigonus crysoleucas	common
Emerald shiner	Notropis atherincides	common
Ghost shiner	Notropis buchanani	common
Pugnose minnow	Notropis emiliae	common
Red shiner	Notropis lutrensis	abundant
Weed shiner	Notropis texanus	uncommon
Redfin shiner	Notropis umbratilis	common
Blacktail shiner	Notropis verustus	common
Mimic shiner	Notropis volucellus	common
Bullhead minnow	Pimephales vigilax	common
Lake chubsucker	Erimyzon sucetta	uncommon
Smallmouth buffalo	Ictiobus bubalus	common
Bigmouth buffalo	Ictiobus cyprinellus	common
Yellow bullhead	Ictalurus natalis	uncommon
Channel catfish	Ictalurus punctatus	common
Tadpole madtom	Noturus zyrinus	uncommon
Brown madtom	Scturus phaeus	rare
Golden topminnow	Eunaulus chrysotus	uncommon
blackspotted topminnow	Fundulus olivaceus	common
Mosquito fish	Gambusia affinis	abundant
Pirate perch	Aphredoderus sayanus	common
Flier	Centrarchus macropterus	uncommon
Green sunfish	Lepomis cyanellus	common
Warmouth	Leportis gulosus	common
Orangespo tted sunfish	lepomis humilis	common
Bluegill	lepomis macrochirus	abundant
Dollar sunfish	_eponis marginatus	common
lengear sunfish	Leporis megalotis	abundant
Redear sunfish	Leronis microlophus	uncommon
Spotted sunfish	Leporis punctatus	uncommon
Bantam sunfish	Lepomis symmetricus	uncommon
Largemouth bass	Micropterus calmoides	common

(Continued)

* Species list compiled from an examination of 3,069 specimens taken from _9 collections at 18 locations.

Table B4 (Concluded)

Name	Scientific Name	Occurrence
White crappie	Pomoxis annularis	common
Black crappie	Pomeris nigromaculatus	uncommon
Banded pigmy sunfish	Elassoma zonatum	uncommon
Scaly sand darter	Ammoerupta vivax	uncommon
Bluntnose darter	Etheostoma chlorosomum	abundant
Cypress darter	Etheostoma procliane	common
Freshwater drum	Aplodinotus americans	common
Brook silverside	Labidesthes sicculus	common

Table B5

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List of Fishes of Tensas River Drainage*

Name	Scientific Name	Occurrence
Spotted gar	Levisosteus oculatus	abundant
Shortnose gar	Lepisosteus platostomus	common
Gizzard shad	Iurosoma cepedianum	abundant
Grass pickerel	Esox americanus vermiculatus	common
Chain pickerel	Esox niger	common
Carp	Syprinus carpic	common
Silvery minnow	Hybognathus nuchalis	common
Golden shiner	Notemigonus orysoleucas	common
Emerald shiner	Notropis atherincides	common
Ghost shiner	Notropis buchanari	common
Weed shiner	Notropis texanus	common
Redfin shiner	Sotropis umbratilis	common
Blacktail shiner	Notropis venustus	abundant
Mimic shiner	Notropis volucellus	uncommon
Bullhead minnow	Fimephales vigilax	abundant
Lake chubsucker	Erimyzon sucetta	common
Smallmouth buffalo	Icticbus bubalus	common
Bigmouth buffalo	Ictiobus cyprinellus	common
Yellow bullhead	Ictalurus natalis	common
Channel catfish	Ictalurus punctatus	common
Tadpole madtom	Noturus gyrinus	uncommon
Golden topminnow	Fundulus chrysotus	common
Blackstripe topminnow	Fundulus notatus	uncommon
Starhead topminnow	Fundulus notti	uncommon
Blackspotted topminnow	Fundulus olivaceus	common
Mosquito fish	Gambusia affinis	abundant
Pirate perch	Aphredode rus sayanus	uncommon
Green sunfish	Iepomis cyanellus	common
Warmouth	Lepomis gulosus	common
Orangespotted sunfish	Lepomis numilis	common
Bluegill	Lepomis macrochirus	abundant
Spotted sunfish	Lepomis punctatus	uncommon
Bantam sunfish	Lepomis symmetricus	uncommon
Black crappie	Pomoxis nigromaculatus	common
Banded pigmy sunfish	Elassoma zonatum	common
Bluntnose darter	Etheostoma chloroscrum	common
Cypress darter	Etheostoma proeliare	common
Brook silverside	Labidesthes sicculus	abundant

* Species list compiled from an examination of 2,179 specimens taken from 23 collections at 16 locations.

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List of Fishes of the Boeuf River Drainage*

Name	Scientific Name	Occurrence
Spotted gar	Levisosteus oculatus	common
Shortnose gar	Lepisosteus platostomus	common
Bowfin	Amia ealva	common
Gizzard shad	Corosoma cepedianum	uncommon
Threadfin shad	lorosoma peterense	uncommon
Goldeye	Hiodon alcsvides	uncommon
Grass pickerel	Essæ americanus vermiculatus	common
Chain pickerel	Esca riger	common
Carp	Syprinus carrio	common
Cypress minnow	Eybognathus hayi	uncommon
Silvery minnow	Hybognathus nuchalis	abundant
Speckled chub	Hybopsis aestivalis	uncommon
Golden shiner	Notemigonus cryscleucas	abundant
Pallid shiner	Notropis amnis	rare
Emerald shiner	Notropis atherinoides	abundant
Ghost shiner	Notropis buchanani	common
Pugnose minnow	lotropis emiliae	common
Ribbon shiner	Notropis fumeus	uncommon
Red shiner	Notropis lutrensis	abundant
Taillight shiner	Notropis maculatus	uncommon
Weed shiner	Notropis texanus	common
Redfin shiner	Notropis umbratilis	common
Blacktail shiner	Notropis venustus	abundant
Mimic shiner	Notropis volucellus	common
Bullhead minnow	Fimephales vigilax	abundant
Creek chubsucker	Erimyzon oblongus	uncommon
Lake chubsucker	Erimyzon sucetta	uncommon
Smallmouth buffalo	Ictiobus bubalus	common
Bigmouth buffalo	Ictiobus cyprinellus	common
Black buffalo	Ictiobus niger	uncommon
Blue catfish	Ictalurus furcatus	rare
Black bullhead	Ictalurus melas	uncommon
Yellow bullhead	Ictalurus natalis	common
Brown bullhead	Ictalurus nebulosus	rare
Channel catfish	Ictalumus punctatus	abundant
Tadpole madtom	Noturus gyrinus	uncommon
Freckled madtom	Noturus nocturnus	rare
Flathead catfish	Pylodictis olivaris	rare
Golden topminnow	Fundulus chrysctus	common
Blackstripe topminnow	Fundulus notatus	common
Starhead topminnow	Fundulus notti	uncommon
Blackspotted topminnow	Fundulus olivaceus	common
Mosquito fish	Cambusia affinis	abundant

(Continued)

* Species list compiled from an examination of 15,844 specimens taken from 97 collections at 27 locations.

Table B6 (Concluded)

Scientific Name	Occurrence
innedeciente amoria	
Monone chnysone	
Vonere missipainniennie	uncommon
Contranable magneters	uncommon
isnomia a grafica	rare
	common
	common
	common
Lepomis macrochinus	abundant
Lepumis megalotis	abundant
Lepomis microlophus	common
_eponis punctatus	common
Lepomis symmetricus	rare
Micropterus punctulatus	common
Micropterus salmoides	common
Fomoxis annularis	uncommon
Pomoxis nigromaculatus	common
Elassoma zonatum	common
Armcerypta vivax	uncommon
Etheostoma asprigene	rare
Etheostoma chlorosomum	common
Etheostoma fusiforme	
Ethecstoma aracile	
Etheostoma proeliare	
Percina shumardi	rare
Aplodinotus amerijans	abundant
Tubidesthes signalus	abundant
	Scientific Name Aphredoderus sayanus Morone chrysops Morone mississippiensis Sentrarchus masropterus Lepomis cyanellus Lepomis gulosus Lepomis macrochirus Lepomis megalotis Lepomis megalotis Lepomis microlophus Lepomis punctatus Lepomis symmetricus Micropterus punctulatus Micropterus saimoides Fomoxis annularis Pomoxis nigromaculatus Elassoma zonatum Armocrypta vivax Etheostoma asprigene Etheostoma fusiforme Etheostoma gracile Etheostoma proeliare Percina shumardi Aplodinotus grunniens Labidesthes sicculus

