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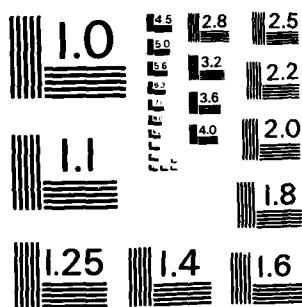
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**AQUATIC PLANT CONTROL
RESEARCH PROGRAM**

TECHNICAL REPORT A-78-2

**LARGE-SCALE OPERATIONS MANAGEMENT
TEST OF USE OF THE WHITE AMUR FOR
CONTROL OF PROBLEM AQUATIC PLANTS**

Reports 2 and 3

FIRST AND SECOND YEAR POSTSTOCKING RESULTS

Volume V

The Herpetofauna of
Lake Conway, Florida: Community Analysis

by

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LARGE-SCALE OPERATIONS MANAGEMENT TEST OF
USE OF THE WHITE AMUR FOR CONTROL OF
PROBLEM AQUATIC PLANTS

Report 1: Baseline Studies

Volume I: The Aquatic Macrophytes of Lake Conway, Florida

Volume II: The Fish, Mammals, and Waterfowl of Lake Conway, Florida

Volume III: The Plankton and Benthos of Lake Conway, Florida

Volume IV: Interim Report on the Nitrogen and Phosphorus Loading Characteristics
of the Lake Conway, Florida, Ecosystem

Volume V: The Herpetofauna of Lake Conway, Florida

Volume VI: The Water and Sediment Quality of Lake Conway, Florida

Volume VII: A Model for Evaluation of the Response of the Lake Conway, Florida,
Ecosystem to Introduction of the White Amur

Volume VIII: Summary of Baseline Studies and Data

Report 2: First Year Poststocking Results

Report 3: Second Year Poststocking Results

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20. ABSTRACT (Continue on reverse side if necessary and identify by block number) This report summarizes the first and second year poststocking results of the Large-Scale Operations Management Test (LSOMT) on the amphibian and reptile populations of Lake Conway, Florida. Data for the 15-month baseline study from June 1977 through September 1978 are compared with those from the two-year poststocking study from October 1978 through September 1980. During the study 11,928 individuals of 12 species of amphibians and 17 species of reptiles were recorded from Lake Conway. South Pool had the (Continued)		

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greatest species diversity (23 species) followed by Middle and East Pools and Lake Gatlin with 20 species, and West Pool with 17 species. Approximately 90 percent of the species known from Lake Conway were recorded with the first 2,900 specimens (24 percent) sampled, and this occurred during the first nine months of the three-year study. The turtle Sternotherus odoratus was the most common species on Lake Conway and accounted for 29.5 percent of the total sample. Of the 29 species recorded in Lake Conway, 14 (48 percent) were recorded in all pools. These 14 species included the most common amphibians and reptiles in the lake and accounted for 95.3 percent of all records.

Thirteen species were identified as functionally important species in the community dynamics of the Lake Conway herpetofauna. Two large salamanders decreased significantly during the study. In contrast to the salamanders, frogs, as documented by the number of calling males, showed an overall increase in density during the study. The density of alligators decreased and that of most turtles declined significantly through the study. The density of snakes also declined dramatically on the lake. Human disturbance of several types was identified as the major causative factor associated with the population declines of many species on Lake Conway. However, reductions in density or changes in feeding activity and habitat use of one salamander (Siren lacertina) and three turtles (Pseudemys floridana, Pseudemys nelsoni, and Sternotherus odoratus) were directly the result of or were affected by the feeding activity of the white amur.

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PREFACE

The work described in this volume was performed under Contract No. DACW39-76-C-0047 between the U. S. Army Engineer Waterways Experiment Station (WES), Vicksburg, Miss., and the University of South Florida, Tampa, Fla. The work was sponsored by the U. S. Army Engineer District, Jacksonville, and by the Office, Chief of Engineers, U. S. Army, Washington, D. C. The Museum Section, U. S. Fish and Wildlife Service, Washington, D. C., provided technical assistance and personnel support for part of the radiotelemetry study of amphibians and reptiles.

This is the fifth of eight volumes that constitute a series of reports documenting a Large-Scale Operations Management Test of use of the white amur for control of problem aquatic plants in Lake Conway, Florida. Report 1 of the series presented the results of the baseline studies. This report combines the results of the first and second year poststocking studies in a community analysis. A third report presents the species accounts from the three-year study. Data from the radiotelemetry study will be published elsewhere.

This volume was written by Dr. Roy W. McDiarmid, U. S. Fish and Wildlife Service, National Museum of Natural History, Washington, D. C., and Mr. G. Thomas Bancroft and Mr. J. Steve Godley, Department of Biology, University of South Florida, Tampa, Fla. The authors acknowledge with thanks Messrs. W. E. Ackerman and M. Lopez and Mdms. D. T. Gross, N. N. Rojas, and D. A. Sutphen for help in the field, Barbara and Tom Davis for hospitality and use of a cottage as a field laboratory, and Marianna B. Scott for help in preparing this report.

The work was monitored at WES in the Environmental Laboratory (EL), Dr. John Harrison, Chief. The study was under the general supervision of Mr. B. O. Benn, Chief, Environmental Systems Division (ESD), EL. Mr. J. L. Decell was Manager, Aquatic Plant Control Research Program, EL. Principal investigators were: Messrs. R. F. Theriot, J. D. Lunz, and E. G. Buglewicz and Dr. A. C. Miller, all of ESD, EL.

Commanders and Directors of WES during the contract period and report preparation were COL J. L. Cannon, CE, COL Nelson P. Conover, CE, and COL Tilford C. Creel, CE. Technical Director was Mr. F. R. Brown.

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CONTENTS

	Page
PREFACE.....	1
PART I: INTRODUCTION.....	4
PART II: METHODS AND MATERIALS.....	7
Funnel Trapping.....	7
Herp-Patrol.....	8
Extended Herp-Patrol.....	9
Selective Herp-Patrol.....	10
Alligator Census.....	10
Other Methods.....	10
PART III: THE LAKE CONWAY SYSTEM.....	11
South Pool.....	12
Middle Pool.....	15
East Pool.....	16
West Pool.....	17
Gatlin Canal.....	19
PART IV: THE HERPETOFAUNA OF LAKE CONWAY.....	20
Species Distribution and Abundance.....	24
Permanent Shoreline Sites.....	26
Yearly Comparisons - Lake Conway.....	55
Deepwater Trapping Stations.....	57
PART V: DISCUSSION AND CONCLUSIONS.....	58
REFERENCES.....	63
TABLES 1-9	
FIGURES 1-62	
APPENDIX A: SUMMARY OF PLANT SPECIES AND SUBSTRATUM TYPE FOR ALL PERMANENT SHORELINE HERPETOFAUNAL TRAPPING STATIONS ON LAKE CONWAY DURING THE FIRST POSTSTOCKING STUDY PERIOD.....	A1
APPENDIX B: SUMMARY OF PLANT SPECIES AND SUBSTRATUM TYPE FOR ALL PERMANENT SHORELINE HERPETOFAUNAL TRAPPING STATIONS ON LAKE CONWAY DURING THE SECOND POSTSTOCKING STUDY PERIOD.....	B1

LARGE-SCALE OPERATIONS MANAGEMENT TEST OF USE OF THE
WHITE AMUR FOR CONTROL OF PROBLEM AQUATIC PLANTS

FIRST AND SECOND YEAR POSTSTOCKING RESULTS

The Herpetofauna of Lake Conway, Florida:
Community Analysis

PART I: INTRODUCTION

1. Beginning in 1975, the U. S. Army Engineer Waterways Experiment Station began planning for a Large-Scale Operations Management Test (LSOMT) to investigate the suitability of the white amur (Ctenopharyngodon idella) as a potential biological control for aquatic plants, especially hydrilla (Hydrilla verticillata). Because the State of Florida had an aquatic plant problem and was receptive to using the fish under the concept of an LSOMT, Lake Conway near Orlando, Florida, was selected as the study site (Addor and Theriot 1977). Lake Conway is an urban lake consisting of five interconnected pools with a combined surface area of approximately 7.4 km². The project was envisioned as consisting of a one-year study for compilation of baseline data (prestocking period) and approximately three years of poststocking study. By September 1976, the following contracts had been awarded and work started: water chemistry and sediment quality--Orange County Pollution Control Authority; aquatic macrophyte populations--Florida Department of Natural Resources; phytoplankton, zooplankton, and benthic invertebrate populations, and an ecosystem model--University of Florida; and fish, aquatic bird, and mammal inventories--Florida Game and Fresh Water Fish Commission. In June of 1977, a contract was awarded to the University of South Florida to characterize and monitor the amphibian and reptile populations.

2. In September of 1977, a monosex population of the white amur was introduced into Lake Conway at an average stocking density of one fish per 0.1 hectare. This value, previously determined from a stocking rate model (Ewel and Fontaine 1979), was designed to control aquatic plant populations, particularly those of hydrilla, in Lake Conway. From this point on (poststocking period), populations were monitored in Lake Conway to evaluate the impact, if any, of the white amur on the component systems of the lake community through September 1980. In September 1979, radiotracking was added

to the project's methodologies to gain insight on movements of the white amur and select species of amphibians and reptiles. Radiotelemetry continued into 1982 for some species of turtles.

3. The herpetofaunal part of the Lake Conway study was designed to accomplish the following objectives:

- a. Determine the species of amphibians and reptiles inhabiting Lake Conway.
- b. Ascertain the habitat requirements, distribution, ecology, and seasonal activity of each species in the lake system.
- c. Establish quantitative baseline population data for the more common or otherwise important species in each of the Lake Conway pools including density by habitat, relative age (size) structure, movements, diel and seasonal activity patterns, growth, reproduction, food habits, and related parameters as deemed feasible.
- d. Monitor quantitatively following introduction of the white amur any changes in the species composition of each pool, movements within and between pools, density changes, or other population parameters of the amphibians and reptiles.
- e. Determine whether any observed changes are the result, directly or indirectly, of the white amur weed control program.

4. The Baseline Studies Report (Godley et al. 1981) summarized the herpetofaunal work for the 15-month period from June 1977 through September 1978. Although the white amur was introduced into Lake Conway in September 1977, only three months after the amphibian and reptile work began, the data presented in the first report were by necessity considered "baseline" and referred to herein as Study Year 1 (SY1). We are convinced that these first year's data were suitable as a baseline to which subsequent years could be compared because the white amur had little detectable impact on the lake system in the first poststocking year (Schardt et al. 1981; Land 1980). The baseline report included detailed descriptions of the herpetofaunal study sites, sampling methods and techniques, and data collection procedures. This baseline report also provided a list of the amphibian and reptile species encountered on Lake Conway during SY1, analyses of their temporal and spatial densities and distributions, and a composite of those parameters deemed important to understanding community dynamics within the system.

5. This report summarizes our findings for Study Years 2 and 3 (SY2 and SY3). Because of our late start on the LSOMT, our SY2 is the first poststocking period and SY3 is the second poststocking period for the herpetofaunal work. In contrast, our periods of SY2 and SY3 are equivalent to poststocking years two and three of other studies. In this report we update our methodologies, describe new and expanded sampling procedures, and provide a brief history of the permanent study sites during the three study years. The community analyses that examine distributions and densities within and among pools essentially follow those discussed in the Baseline Report. Detailed accounts of each species emphasizing natural history parameters and changes therein through the three-year study are presented separately (Pancroft et al. 1983). Results of the radiotracking study will be available in the doctoral dissertation of J. S. Godley to be filed at the University of South Florida.

PART II: METHODS AND MATERIALS

6. Details of the Lake Conway herpetofaunal sampling programs are provided elsewhere (Godley et al. 1981), and only a brief update and summary of methodologies are presented here. In each pool of the Lake Conway complex one permanent shoreline site (Figure 1) was used for mark and recapture population studies; destructive samples for analyses of stomach content and reproductive condition of selected species of amphibians and reptiles were taken from other areas of similar habitat within the lake system. A description of the permanent shoreline sites and any habitat changes that occurred at the sites during the study is provided in Part III, The Lake Conway System. The deepwater transects shown in Figure 1 were discontinued after the baseline studies because of extremely low trap success per unit effort (see paragraph 7 of Godley et al. 1981).

7. In general, the herpetofaunal sampling program involved spending three days and two nights every other week on the lake so that each permanent shoreline site was censused by herp-patrols and sampled with funnel traps twice a month (see below). Alternate weeks were spent in the laboratory processing data and gathering reproductive and ecological information from the destructive samples. Brief descriptions of the major sampling methods are discussed below.

Funnel Trapping

8. Funnel traps, 60x30x30 cm, were designed specifically to sample aquatic salamanders, tadpoles, small carnivorous turtles, and several species of aquatic snakes. Most funnel traps were constructed of 3-mm black plastic Vexar netting (DuPont De Nemours & Co., Model No. 5-59-V-360-BABK) stretched over welded metal frames with funnel entrances at each end. Some wire mesh (6.35-mm hardware cloth) funnel traps of the same dimensions were used in areas where rice rats (*Oryzomys palustris*) were common and often gnawed holes in the Vexar cover of traps.

9. During the baseline study period funnel traps were set at each permanent shoreline site twice a month for a 24-hr period. Traps were baited with fresh, cut fish. Because we had only half as many traps as sampling stations (N=163), the traps had to be moved to new sites each sampling trip;

each trap move required a considerable expenditure of field time. Beginning in December 1979, funnel traps were set at each site only once a month but for a 48-hr period. The traps were checked and freshly baited each day; animals collected on the first day were processed and usually released the same day. Thus, the total trapping effort remained constant (48 hr/month/trap station) but more field time was available for other activities. Relative density comparisons within and between sites and study years are based on total trap success per total number of trap-days (sum of all traps set per 24-hr periods).

Herp-Patrol

10. All permanent shoreline sites were censused twice a month at night from a 5.33-m johnboat. This censusing technique, termed "herp-patrol," involved the use of an electric trolling motor and two 12-volt, 120,000-candlepower spotlights. The permanent shoreline sites were censused with herp-patrols by motoring slowly along the edge of the littoral zone. One spotlight from the rear of the boat was directed towards the emergent vegetation on the shore side, while the other in the front of the boat was shined on the opposite side in adjacent, open water. A third individual collected animals with a dip net or by hand. The identification, time, location, water depth, vegetation type, substratum, activity, and behavior of all specimens observed or heard calling were recorded on standardized data sheets. All captured individuals were brought to the laboratory and sexed, measured, weighed, marked, and released at their capture point the following day.

11. In the baseline study period herp-patrols on permanent sites were replicated each sampling night with the electric trolling motor and assigned a run number of I or II. The same collecting path and direction were used on each run. The number of males of each frog species calling per 10-m increment of shoreline was recorded on both runs; other species of amphibians and reptiles were captured and processed as reported above. First year results indicated no significant differences in the mean number of calling frogs recorded on run I compared to run II (paired t-tests, Steel and Torrie 1960). However, many turtles eluded capture because of the slow speed and poor mobility of the boat when powered by the trolling motor. To improve capture success but still monitor frog populations, this herp-patrol procedure was

modified slightly in subsequent years: two runs were performed each sampling night but frogs were recorded only on the first run which was done with the electric motor, the second run was accomplished with the 25-hp (18.6 kw) outboard motor at slow speeds. Between-year comparisons of relative densities on permanent sites are based on the number of calling males heard only on run I (frogs only) or the total number of individuals observed per total search time for both runs (all other taxa).

Extended Herp-Patrol

12. As work progressed, three potentially confounding trends in populations of several turtle species on Lake Conway appeared: (1) the number of individuals observed on permanent shoreline sites was decreasing through time, (2) several species apparently had home ranges much larger than the size of the permanent shoreline sites, and (3) many marked turtles had moved off the permanent sites and into nearby areas perhaps in response to repeated capturing. To provide a better understanding of the demography and movements of turtles, the amount of shoreline censused on herp-patrols gradually was expanded in two pools (=extended herp-patrol).

13. In August 1978, the herp-patrol in South Pool was extended from the southern end of the permanent site to the Detwilder outflow canal (Figure 1). In November 1978, a second South Pool extension was initiated from the bridge to the northern end of the site (Figure 1). Thus, beginning in November 1978, herp-patrols in South Pool covered the permanent site and two extensions, or approximately half of the South Pool shoreline. Also, in November 1978, a herp-patrol extension was begun in West Pool, covering the shoreline from the southern end of the West Pool permanent site to the mouth of Gatlin Canal.

14. Extended herp-patrols involved one run at night at slow speeds with the outboard on each of the bimonthly sampling trips to Lake Conway for the remainder of the study. Extended herp-patrols were conducted on the same nights as permanent site herp-patrols but with three differences in methodology: (1) only one run per night was done on each extended herp-patrol while the two-run procedure was continued on permanent sites, (2) calling frogs were not recorded on the extensions, and (3) stinkpots (*Sternotherus odoratus*) captured on the extensions were processed in the boat and immediately released

at the capture site; all other species were brought back to the lab for processing, as was done on all permanent sites.

Selective Herp-Patrol

15. We were concerned that the floy tags used to tag Sternotherus odoratus may have become a significant mortality factor and may have contributed to declines in this species on permanent sites. To control for this possibility, "selective herp-patrols" were established in November 1978 in South Pool and September 1979 in West Pool (Figure 1). Selective herp-patrols were similar in all respects to extended herp-patrols except that on selective herp-patrols individuals of S. odoratus were not captured or marked, only visually censused and recorded. On selective herp-patrols, all other turtle species were captured and processed in the usual manner (see Godley et al. 1981).

Alligator Census

16. The alligator population of the entire Lake Conway complex was estimated every other month using nocturnal censusing procedures previously described in detail (Godley et al. 1981). These procedures provided estimates of the number, location, and approximate sizes of all alligators on Lake Conway. In addition, nesting was monitored by searching all stretches of suitable shoreline for alligator nests and determining the relative success of each nest during the summer nesting season.

Other Methods

17. In addition to the above sampling methods, several other techniques (gill netting, shoreline census, hyacinth sieving, electrofishing) were used for selected species as time and man-power permitted. These techniques and the procedures used for measuring and marking all captured amphibians and reptiles are described elsewhere (Godley et al. 1981). Late in the study we successfully used "muddling" to capture large turtles beneath dense waterhyacinth mats. Specimens were located by hand in a random search beneath mats of waterhyacinth often anchored by emergent cattails. Most success was achieved when water depth was less than 1 m and mats exceeded 20 m².

PART III: THE LAKE CONWAY SYSTEM

18. Lake Conway is a 743.0-ha urban lake located in South Orlando, Orange County, Florida (Figure 1). The lake consists of five interconnecting pools which include Lake Catlin, Little Lake Conway (East and West Pool), and Lake Conway (Middle and South Pool). The lake system is mesotrophic with gradually increasing eutrophic conditions as one proceeds north through the five pools (Corley et al. 1979). The substratum is primarily sand, except in areas of thick vegetation near shore or in dredged canals where organic detritus or silt has accumulated. The bottom contours are rather steep when compared with most central Florida lakes. Greater than 40% of the total lake bottom is deeper than 6.0 m (Schardt et al. 1981).

19. During the baseline study period, Illinois pondweed (Potamogeton illinoensis) and eelgrass (Vallisneria spiralis) were the dominant shallow-water (<2.0 m) submergent macrophytes in most pools; stonewort (Najas) and hydrilla (Hydrilla verticillata) predominated in water from 2-6 m deep (Schardt et al. 1981). During the first and second poststocking years, most macrophyte species decreased in distribution and abundance. An exception was eelgrass, which is not a preferred food of the white amur (Schardt et al. 1981).

20. Most of the emergent vegetation on Lake Conway had been removed for beach development prior to the initiation of the LSOMT. This trend of habitat modification and destruction has been documented for the five years prior to our work (Williams et al. 1982) and continued during the study. However, a narrow fringe of maidencane (Panicum hemitomon), lake rush (Fuirena scirpoides), pickerelweed (Pontederia lanceolata), or cattail (primarily Typha latifolia) persisted in some areas.

21. Given below are brief descriptions of each permanent shoreline sampling site and a chronology of important changes in the habitat at these sites. Appendices A and B of this report summarize the vegetation and substratum characteristics of each permanent trapping station for all sites during the first and second poststocking periods, respectively. These appendices can be compared directly with baseline conditions (Appendix A of Godley et al. 1981).

23. The inclusion of primrose willow (Ludwigia peruviana) and arrowhead (Sagittaria lanceifolia) in Appendix B but not in Appendix A of this report should not be interpreted as the sudden appearance of these plant species on the study sites during SY1. Rather, their appearances reflected changes in trap location as a result of fluctuations in water level. During SY3, high summer water level necessitated the placement of traps in previously drier habitats in which these two species occurred.

24. Hurricane David swept through the Orlando area on 7 September 1979 and uprooted the littoral zone vegetation at several sites. The effects of this natural disturbance are considered in Appendix B (second poststocking period) and not Appendix A because (1) this hurricane occurred with only 3.5 weeks remaining in the first poststocking period, and (2) the biannual sampling of the littoral zone vegetation and substratum conditions on the permanent sites was completed prior to the hurricane.

South Pool

24. The South Pool permanent shoreline site originally was 530 m in length (Figure 1) and included the only major section of undeveloped shoreline in the pool with large expanses of Panicum hemitomon, Puirena scirpoides, and Pontederia lanceolata (Godley et al. 1981). Development of this site for housing occurred during the baseline study period and continued through both poststocking periods (Tables A1 and B1). During the baseline year, littoral zone vegetation was removed and the area converted to white sand beach between markers 240 and 310, at marker 350, and at markers 380 and 490 (Godley et al. 1981, paragraphs 5 and 55, Table A1). By the end of SY1, all upland vegetation from markers 240 to 300 and 350 to 460 had been removed. In the first poststocking period (SY2), drastic changes to the littoral zone vegetation occurred from markers 0 through 90, where all vegetation was removed and the habitat converted to white sand beaches (Table A1). In addition, all remaining upland and transitional zone habitats (except between markers 190-240 and 300-345) along the South Pool site were bulldozed in site preparation for houses. In SY3, shoreline development continued with the loss of all native vegetation from markers 140 through 230 and at 450 and 460 (Table B1). Thus,

by September 1980, the only remaining littoral zone vegetation on the South Pool site occurred in scattered patches between markers 100 and 130 and 320 and 440.

25. Although not quantified, several other changes in the littoral zone at this site were noted. Localized increases in water turbidity on the South Pool site were pronounced, especially following heavy rains. Higher turbidity levels probably resulted from the increased runoff from the newly developed shoreline. During the same period, a decrease in the vigor of the remaining beds of Panicum hemitomon, Pontederia lanceolata, and Fuirena scirpoides was observed even though these beds were not physically disturbed. This decrease in stem density and general condition probably also was related to the development of the adjacent shoreline. Immediately offshore from the South Pool site, several changes in the species composition of submergent plants were recorded. In several areas, especially between markers 150 and 340, large beds of Potamogeton illinoensis either were replaced by Vallisneria americana or showed no growth following the winter decline.

26. The South Pool area sampled with extended herp-patrol I was 840 m long and covered the shoreline from the Nela Avenue Bridge south to the north end of the South Pool permanent shoreline site (Figure 1). A 450-m section nearest the bridge bordered a deep hole (to 7 m depth) while most of the remaining 390 m was less than 2 m deep up to 50 m from shore. Illinois pondweed (Potamogeton illinoensis) was the only submerged macrophyte in the shallows of this extension. Pondweed was sparse in coverage except for the last 300 m where moderately dense beds existed. The only emergent vegetation along the entire length of this extension was several isolated patches of torpedo grass (Panicum repens); the remainder of the shoreline was white sand beach.

27. The extended herp-patrol II on South Pool stretched from the south end of the permanent site to the Detwilder outflow canal (Figure 1). This extension covered 1420 m of shoreline and had a uniformly shallow bottom with depths of less than 2 m extending 30 to 60 m offshore. Illinois pondweed was the dominant shallow-water submerged macrophyte along the entire length of this extension but grew with stonewort (Nitella megacarpa) in water depths of 1 to 2 m. Pondweed was more abundant on the South Pool extension II than on

extension I but noticeably declined in coverage during the study. An isolated patch of eelgrass (Vallisneria americana) occurred near shore between x-coordinates 380 and 390. All of the shoreline was white sand beach except for two small (<50 m) isolated patches of panic grass (Panicum hemitomon) and two patches (40 and 50 m in length) of cattail (Typha latifolia). This extension ended at the Letwilder outflow canal in an area known as 7-11 Cove. This cove (60x100 m) was unique in South Pool in that it contained the only quiet backwaters. The banks were lined with shrubs and bordered by several stands of Panicum hemitomon in the shallows. Patches of spatterdock (Najas luteum) and waterlily (Nymphaea odorata) grew in several areas of the cove. The bottom of 7-11 Cove was heavily silted with no submerged macrophytes; the water depth was less than 2 m.

28. The South Pool selective herp-patrol covered 2540 m of shoreline, extending around the remaining shoreline from 7-11 Cove north and west to the Nela Avenue bridge (Figure 1). The easternmost shore of South Pool from 7-11 Cove to coordinates 400, 180 had steep contours (to 4 m) with sparse beds of pondweed and stonewort offshore. Along the shore of this section were three Panicum hemitomon beds of 140, 110, and 60 m in length; the remainder was beach habitat. North and west of this section to a point near coordinates 275, 208 the habitat was relatively uniform: sand beaches with gradually sloping contours and sparse pondweed interspersed with stonewort in deeper areas. From a point near coordinates 275, 208 west to the Nela Avenue bridge the bottom contours were rather steep with sparse pondweed near shore and stonewort offshore. White sand beaches lined most of this shore except for a small (40 m) patch of cattail near the bridge.

29. In summary, the herpetofaunal habitat in South Pool underwent significant changes during the study. Most of the littoral zone of the permanent shoreline site was developed during SY1 and SY2. Offshore from this site the dense beds of pondweed were reduced in size and coverage by the white smur during SY2 and remained sparse and well cropped for the remainder of the study. By the time the two extended herp-patrols (August and November 1978) and the selective herp-patrol (November 1978) were established, shallow-water macrophyte abundance had declined noticeably in South Pool. General observations in these areas during SY1 suggested that pondweed in particular

was as common here as on the permanent site in that year. Schardt et al. (1981; Tables B1 and B16, Figure J9) documented a 98.9% reduction in Potamogeton illinoensis biomass and an 85.7% reduction in percent coverage from baseline conditions to SY3 for the entire South Pool.

Middle Pool

30. The Middle Pool permanent shoreline site (Figure 1) was 200 m in length and averaged 60 m in width. It was located at the northern end of a large cattail (Typha latifolia) marsh which had a dense, inner zone of herbaceous aquatics. The site was sampled with three transects: an inner herbaceous zone where traps were set (1000 series transect), and two herp-patrol transects (2000 series = center of cattails, 3000 series = outer edge of cattails). During April 1978 of the baseline study, all shoreline and upland vegetation between markers 1000 and 1120 was cleared with bulldozers and draglines (Godley et al. 1981) effectively removing all vegetation from trapping stations 1000 to 1120 and all cattail from markers 2000 to 2120. A 10-m-wide outer fringe of cattail (markers 3000-3120) and all marsh habitat to the west of the site (including 80 m of the 1000, 2000, and 3000 series transects) were left intact.

31. No additional development occurred on the Middle Pool site during the study. On the previously disturbed section where traps were set (markers 1000-1120), lake rush (Fuirena scirpoides) and a few cattail (Typha latifolia) gradually invaded the bare, sandy shoreline (Tables A2 and B2). However, the 10-m outer fringe of cattail between markers 3000 and 3060 showed no regrowth following the winter of 1978-79, possibly because of previous disturbance. Several small, low growing patches of Potamogeton illinoensis and Nitella sp. invaded the open water between markers 2000 and 2120 and 3000 and 3120 during the summer of 1979, and persisted through the remainder of the study.

32. On the undisturbed sections, waterhyacinth (Eichhornia crassipes) became the dominant plant at several trapping stations and replaced Pontederia lanceolata and Typha latifolia (Tables A2 and B2). No changes in vegetation were detected on the undisturbed sections of the 2000 (2120-2200) and 3000 (3121-3200) series transects in Middle Pool.

East Pool

33. The permanent sampling site in East Pool was located at the northwest end of an uninhabited island (Figure 1). This site initially was 200 m in length and consisted of a 10- to 15-m outer fringe of cattail (Typha latifolia) and an inner zone of herbaceous aquatics, especially waterhyacinth (Eichhornia crassipes). No habitat destruction occurred on the East Pool permanent shoreline site during the study. However, several changes in plant species dominance and substratum conditions in the poststocking periods were recorded (Table A3 and B3). Waterhyacinth (Eichhornia crassipes) invaded or increased in dominance at several trap stations with concomitant decreases in the abundances of Panicum hemitomon and Pontederia lanceolata. At the East Pool trapping stations, the mean depth of the detrital layer decreased from a mean coded value of 4.00 (= 11.7 cm) in the baseline period (see Table A3 of Godfrey et al. 1981) to 3.00 (= 8.0 cm) in the first poststocking period (Table A3, this report) to 2.10 (= 5.3 cm) in the second poststocking period (Table B3). The decreases in mean detrital depth apparently were caused by two factors: (1) compaction of the substratum along the trapline as a result of continued human foot traffic; and (2) declining water levels, which required that traps be set slightly farther from shore in more sandy habitats.

34. Funnel trap results from the baseline and first poststocking periods showed that (1) the East Pool island site had the highest trap success of any permanent site on Lake Conway, but that (2) capture success was declining on the sampled section of the island. A preliminary survey of the entire island in September 1979 indicated that major portions of the island were similar to the permanent site in plant species composition and substratum conditions. To determine if declining herpetofaunal populations on the permanent site were the result of repeated trapping and human disturbance or were characteristic of the entire island, the trapline on the island was extended from 1210 to 1400 in October 1979 and again in December 1979 to encompass the entire island (1010-1470). Vegetation and substratum conditions for the East Pool extension trap stations are provided in Table B3.

35. Immediately offshore from this site, no change from baseline

conditions was noticed in the abundance of Vallisneria americana, but the percent cover and density of Potamogeton illinoensis apparently decreased, probably as a result of the white amur feeding activities (Schardt et al. 1981). The only noticeable change in vegetation as a result of Hurricane David was the uprooting of some cattail (Typha latifolia) that fringed the littoral zone of this site.

West Pool

36. The West Pool site was 570 m in total length, included the only large, continuous section of emergent vegetation in the pool, and was bordered by beach habitat at each end. The littoral zone vegetation included a mixture of Panicum hemitomon, Pontederia lanceolata, Typha latifolia, and Eichhornia crassipes (Godley et al. 1981). Several changes in the habitat occurred at this site during the first poststocking period (Table A4). Trap stations 0 through 70, which previously were open sand and beach habitats with sparse vegetation, overgrew considerably with Eriocaulon scirpoides and Panicum repens. Waterhyacinth (Eichhornia crassipes) continued its expansion and was the dominant vegetation at 17 trap stations by the end of the first poststocking period. The waterhyacinth primarily replaced beds of Pontederia lanceolata but Panicum hemitomon also was reduced in coverage. Typha latifolia beds on the West Pool site also increased in size.

37. The effect of Hurricane David on the littoral zone vegetation was more severe in West Pool than at any other permanent sampling site. Many extensive beds of Pontederia lanceolata were uprooted and several waterhyacinth mats either were set afloat or displaced from their original position (Table B4).

38. Offshore from this site a reduction in the coverage of Potamogeton illinoensis was noted, presumably as the result of white amur feeding activities. The Vallisneria americana beds, which covered most of the bottom offshore, remained the same in distribution and coverage but showed a reduction in stem height through both poststocking periods.

39. The West Pool herp-patrol was initiated in November 1978 and covered the shoreline from the southwestern end of the permanent site to the

entrance of Gatlin Canal (Figure 1), a distance of about 1450 m. Along this shoreline, water depth generally was less than 2.0 m, and the bottom was sandy.

40. At the time this extension was initiated, several dense stands of shallow-water macrophytes existed. At coordinates 137, 403 a luxuriant bed (40 m by 100 m) of eelgrass (Vallisneria americana) mixed with pondweed (Potamogeton illinoensis) occurred. About 300 m to the north (144, 434) a particularly dense stand of pondweed also existed. Another dense stand of eelgrass was present from this area north to the entrance to Gatlin Canal. In the intervening shallow-water areas sampled by the extended herb-patrol, the bottom was either bare sand or sparsely vegetated by Potamogeton illinoensis. Offshore from this site, the bottom was bare or supported stonewort and hydrilla. Most of the shoreline was bare sandy beach but cattail (Typha latifolia) beds of 90 m and 80 m, respectively, existed at two sites (128, 415 and 147, 457) and small (less than 30 m), scattered beds of Panicum repens, or P. hemitomon, occurred in several other areas (156, 469; 146, 427; 155, 445).

41. The most important change that occurred on the West Pool extension was the reduction or elimination of pondweed from most areas. Eelgrass was not eliminated but showed an apparent reduction in stem height in the two areas (noted above) where it was most abundant. No major changes occurred in the littoral zone of this site. Deepwater vegetation was not monitored by us but Schardt et al. (1981, Figure 3b) found a 61.5% reduction in total percent coverage in West Pool from August 1977 to August 1980.

42. The West Pool selective herb-patrol was 1450 m in length and extended from near the East Pool entrance into West Pool south to the permanent shoreline site (Figure 1). By the time this herb-patrol was initiated (September 1974) most of the shallow-water macrophytes had been eliminated. Most of the bottom was bare sand with sparse Potamogeton illinoensis near shore and bare bottom or stonewort offshore. Most of the shoreline was beach habitat but two cattail beds (55 and 60 m at 236, 403 and 224, 395, respectively) and three stands of Panicum hemitomon (200, 60, and 40 m centered at 285, 501; 271, 465; and 247, 415, respectively) were present. No major changes in littoral zone vegetation occurred along this site for the remainder of the study.

Gatlin Canal

43. The permanent shoreline site for Lake Gatlin extended the entire length of the canal (470 m) from Lake Gatlin to West Pool (Figure 1). During the baseline study period, funnel traps were set on both the west (0-40) and east (1050-1190) sides of Gatlin Canal (Godley et al. 1981). One homeowner on the east side of Gatlin Canal objected to traps being placed along his shoreline. As a result, all trap stations were relocated to the west side beginning in February 1979. Because few changes in vegetational composition and substratum occurred at the sites on the east shoreline from October to February in SY2, data for the east side can be found in Godley et al. (1981: Table A5). A summary of the plant species and substratum types for the west shoreline only is presented in Table A5 of this report.

44. In addition to modifying the location of the trapping stations, several other changes occurred in Gatlin Canal during the first poststocking period. Cabomba caroliniana, which initially was restricted to the west offshoot canal between markers 1120 and 1150, spread to occupy most of the open water in the canal along its entire length. However, constant motorboat traffic kept the center of Gatlin Canal relatively free of vegetation. The Typha latifolia bed located between markers 260 and 280 was sprayed with herbicide in July 1979 and showed no regrowth until the following spring. Also, waterhyacinth (Eichhornia crassipes) became established in the Nuphar luteum bed between markers 10 and 40 during the first poststocking period.

45. Several important events occurred in Gatlin Canal during the second poststocking period. On 6 May 1980 heavy rains washed out large mats of waterhyacinth from the western off-shoot canal near the bridge (Figure 1) into Gatlin Canal. These hyacinths were sprayed with herbicide in June 1980 and sank within six weeks. Early in September, heavy rains again washed hyacinths from this canal into Gatlin Canal completely blocking boat traffic. Herp-patrols beyond marker 70 were discontinued in the last month of the study.

PART IV: THE HERPETOFAUNA OF LAKE CONWAY

4b. Beginning in June 1977 and continuing through September 1980 a total of 11,428 individuals of 12 species of amphibians and 17 species of reptiles was recorded on Lake Conway. All 29 species were taken in the lake or along its shore and are dependent in some way (e.g., aquatic or semiaquatic as adults, terrestrial as adults with aquatic larval stages, feed on aquatic or semiaquatic species, etc.) on the lake environment. During the 15-month baseline study (Gosley et al. 1981), 8,700 individuals representing 11 species of amphibians and 16 species of reptiles were recorded. From October 1978 through September 1980 (1Y2), an additional 2,728 individuals and two additional species were added to the study. One specimen of the dwarf siren, Pseudobranchius striatus, was collected in East Pool and two specimens of the introduced red-eared turtle, Pseudemys scripta, were taken in North Pool. No additional species were among the 2,728 individuals taken in 1Y2.

4c. Prior to initiation of the inventory we projected that the herpetofauna of Lake Conway might include 49 species. As we became familiar with the lake system and its shoreline vegetation, it soon became obvious that several species of frogs that commonly breed in temporary ponds in central Florida did not occur on the lake. Notable among these was Bufo quercicus, Bufo lentiginos, Limnodynastes ocularis, Pseudacris nigrita, and Scaphiopus holbrookii. Two species of snakes, Nerodia taxispilota and Akistrodon piscivorus, often found in and around cypress swamps and other permanent bodies of water were absent, probably because of the lack of suitable riparian forest or swamp habitat. Three other aquatic species, the peninsula newt (Notopthalmus viridescens), the bullfrog (Bufo carolinensis), and the black swamp snake (Lampropeltis pyrocea) were expected but to date have not been recorded. In our experience, the peninsula newt, when present, is common. The bullfrog is easily recognized by its large size and distinctive call. We are unable to explain the absence of these two species. We suspect the presence of Lampropeltis pyrocea in the lake system but, because of the marginal nature of the habitat, it probably is rare and as yet undetected. In this respect, two other species (Pseudobranchius striatus and Hoplosternon littorale) that are rare in Lake

Conway are relatively common in habitats frequented by the black swamp snake, S. pygmaea. Coluber constrictor, a terrestrial snake, was recorded twice in South Pool, once in a drift fence trap in water and once on a shoreline census. An introduced species, the red-eared turtle (Pseudemys scripta) was taken on three occasions in South Pool. Originally we considered these three individuals to represent escaped or released pets but subsequently we established that a breeding population occurs in Lake Conway and thus included the species in the fauna. Based on the three years of sampling, the herpetofauna of Lake Conway includes four species of salamanders, eight anurans, one crocodilian, nine turtles, and seven snakes (Table 1).

48. The cumulative number of species is plotted against the cumulative number of individuals for each of the five pools and for the total lake system in Figure 2. In each graph, the circle represents the last individual recorded for that pool. Approximately 90% of the species known from the Lake Conway herpetofauna had been recorded with the first 2400 (241) specimens sampled and this occurred in the first nine months of the three-year study. South Pool has the highest number of species (23 species) followed by Middle, East, and Gatlin (20) and West (17). The fastest rate of species accumulation as a function of the number of individuals sampled was that of South Pool, followed closely by West, Middle, East, and Gatlin. Because the total samples from each pool are unequal and the result of differential sampling effort, we compared the species accumulations for the first 1863 individuals recorded from each pool. This value was equal to the smallest available sample, that from Lake Gatlin. The first 1863 individuals from the total sample recorded 20 species. In this comparison, South Pool had the highest species diversity (22 species) followed by Middle and Gatlin (20), East (18), and West (17). The number of species recorded for Middle and West Pools using the first 1000 individuals was equal to the total recorded diversity. Using this initial sample, the species diversity in South Pool was 95.6% of the known diversity at the end of year three and that in East Pool was 90.0%. However, with a sample of four more individuals (1867), the 18th species (95.6% of the known fauna) was added to East Pool (Figure 2). The first 10% of the total individuals observed in each pool gave a high of 70% of the recorded species in South Pool, followed by 71% for West Pool, 67% for Middle Pool, 60% for East Pool, and a low of 45% of the

recorded species in Lake Gatlin. More than 80% of the species in each pool had been sampled when half the total individuals from that pool had been recorded. This high percent was obtained with between 932 captures (Gatlin) and 160 captures (South). Although sampling effort for the total fauna varied between pools, the species curves are generally similar and convince us that we have adequately sampled the amphibian and reptile fauna of each pool in the Lake Conway system.

40. Sternotherus odoratus, the most common species encountered during our study, made up 24.5% of the total sample. The next most abundant species were Hyla cinerea (22.0%), Acris gryllus (11.4%), Pseudemys floridana (7.5%), and Rana auricularia (4.1%). The other 24 species were uncommon, each representing less than 4.5% of the total sample. The most common salamander on Lake Conway was Amphiuma means (N=681), followed closely by Siren lacertina (234). Adults of the frogs Hyla cinerea and Acris gryllus were recorded 5.4 and 2.8 times more frequently than the next most commonly observed frog species, Rana auricularia (N=491), which was slightly more abundant than Bufo terrestris (473). Alligators were not abundant on Lake Conway, but because of their size and easy detection they accounted for 283 observations during the three-year study. Among turtles, Sternotherus odoratus was nearly four times more frequently recorded than Pseudemys floridana (N=401) which was recorded four times more often than Pseudemys nelsoni (226), the next most frequently recorded species of turtle. Nerodia cyclopion was the most common snake (N=133) on the lake and accounted for 77.2% of the total observations for snakes.

41. The frequency distribution of species by sampling method for the three-year study is presented in Table A1 of Bancroft et al. (1984). Specimens collected on the extended and selective herp-patrols during FY3 are included with the regular herp-patrol data, as are specimens taken for reproductive and stomach analyses. Specimens taken in gill nets, by mudling, or while radiotracking are included in "other methods." In this and all subsequent table citations, data for the different life stages (eggs, larvae, adults), when available, are treated separately. Herp-patrol was the most successful sampling technique and accounted for 10,346 (86.7%) of the individuals sampled and 23 (79.3%) of the species. Funnel traps took 781

(6.5%) of the individuals and 20 (68.9%) of the species. Shoreline censuses recorded 24 (79.3%) of the species with only 245 (2.0%) of the individuals. Shoreline censuses were very effective in determining the species present in each pool but did not contribute substantially to the quantitative data. The other sampling methods each took nine or fewer species and less than 1.0% of the individuals.

51. Of 29 species observed during the study, five were recorded only with one sampling technique: all were relatively rare species (Table A1, Bancroft et al. 1983). The salamander Eurycea quadridigitata and the snake Thamnophis sirtalis were sampled only during the shoreline censuses and the frogs Hyla femoralis and H. squirella and the turtle Leiochelys reticularia were recorded only on herp-patrols. All other species were sampled with at least two methods. Herp-patrol, the most productive sampling method, accounted for very high percentages of the observations of adult frogs and sizable percentages of Siren lacertina (44%), all species of turtles, and the two species of water snakes, Nerodia fasciata (59%) and N. cyclopion (50%). Funnel trapping resulted in a high percentage of our observations of Amphiuma means (92%), and sizable percentages of the records of the salamanders Siren lacertina (47%), the turtles Chelydra serpentina (57%), Kinosternon baurii (29%), and K. subrubrum (33%), the more common snake Nerodia cyclopion (33%) and the rarer Regina alleni (57%) and Farancia abacura (33%), and most frog larvae.

52. Any methodological comparisons derived from Table A1 of Bancroft et al. (1983) clearly indicate the diversity of species of amphibians and reptiles in Lake Conway and point to the importance of a multifaceted sampling program for a study of this magnitude. Comparisons of the relative abundance of species by sampling method also indicate the differential susceptibility of some species to different techniques. For example, vocalizing frogs make detection from a distance very easy as compared to most turtles that must be located visually from a distance of a metre or two. Also, species that are susceptible to trapping or have food (bait) preferences are more easily sampled than large individuals with different food habits (e.g., larvae versus adult Rana utricularia; small, carnivorous versus large, herbivorous turtles). In addition, the susceptibility of a species to capture by a specific method is confounded by its absolute abundance and the sampling effort devoted to a

method, i.e., rarer species are likely to be represented by only a single, commonly employed method. Thus, the biological attributes of a species and the effort put into a sampling technique introduced a certain bias in the data and made cross-method comparisons difficult and less informative.

Species Distribution and Abundance

53. The distributions of amphibians and reptiles observed or captured from the five pools of the Lake Conway system are summarized for the three years in Table A2 of Bancroft et al. (1983). This summary includes the total number of individuals recorded by all sampling methods on the permanent sites and throughout each pool. Because sampling effort and catch varied between pools and between years, the data presented in the table provide only an estimate of the species diversity in each pool and the species distributions among pools.

54. The surface area of each pool and its shoreline together with the number of species and number of individuals recorded during the study are summarized in Table 2. Gatlin Canal is included with Lake Gatlin in this comparison. As expected, the highest diversity (23 species) was recorded from South Pool where the largest sample (3364 individuals) was taken. The lowest diversity was in West Pool (17 species) with an intermediate sample of 2146 individuals. Lake Gatlin had the smallest sample (1603 individuals) but with a species diversity (20 species) equivalent to those of Middle and East Pools. A comparison of species diversity to surface area of each pool did not follow the expected; that is, the largest pool did not have the highest diversity and smallest did not have the lowest. The higher than expected diversity in East Pool may be, in part, the result of its more extensive shoreline. East Pool with 42.7% of the surface area but 91.1% of the shoreline of Middle Pool has the same number of species (20). In some instances (e.g., South Pool) the higher diversity is a reflection of greater effort and the presence of a few rare species (e.g., Hyla squirella, Pseudemys scripta, Coluber constrictor, and Thamnophis sirtalis). In other instances (e.g., West Pool) the low recorded diversity appears to be a true reflection of the herpetofauna in the pool and is not as easily explained as are the values found for other pools. Nearly 50%

of the total West Pool sample consisted of adult Hyla cinerea (1064 individuals), a sample much larger than values recorded for any other pool (Table A2, Bancroft et al. 1983). In contrast, the numbers of individual captures for the two native species of Pseudemys were considerably below those for these two species in the other four pools. Thus, the low diversity recorded for West Pool apparently is reflective of the West Pool environment. In certain instances (e.g., West Pool), the recorded differences accurately reflect the pool's species composition (species unevenness) but, in others (Middle Pool and Lake Gatlin), the differences are clearly the result of inadequately sampling some of the rarer species. These differences will be treated in more depth in the permanent site data presented below, and in the individual species accounts in a separate report.

5b. The known occurrence of species of amphibians and reptiles in Lake Conway varied among pools (Table A2, Bancroft et al. 1983). Of the 29 species recorded from the lake system, 14 (48%) are known from all pools. These 14 included the most common species in the lake and accounted for 95.26% of all records. A few of these 14 species (Kinosternon subrubrum, Trionyx ferox, and Nerodia fasciata) were relatively rare in the lake and each represented less than 1% of the total sample. Of the 15 remaining species with distributions in one to four pools, none represents more than 2.5% of the total capture. The most common of these 15 species was Rana grylio (1.96% adults and 0.49% larvae) which was absent from South Pool. We suspect that it may occur in South Pool transiently but it has not been seen or heard there during the three-year study. Judging from the habitats occurred in other pools, we attribute the apparent absence of R. grylio in South Pool to a lack of suitable habitat and to the drastic reduction and loss to shoreline development during the study of nearly all natural habitat, marginal as it was for R. grylio. We suspect that some of the rarer species (e.g., Pseudobranchius striatus, Kinosternon baurii, Deirochelys reticularia, Farancia bacura, Regina alleni) have wider distributions among pools than our data indicate and that additional sampling effort would support this view. In fact, some of these have been collected in the fourth year from pools in which they were unknown through the three-year study. Other rare species (including Eurycea quadridigitata and Hyla squirella) probably occurred in suitable habitat throughout the lake but have

been eliminated because of the extensive habitat modification and loss.

Permanent Shoreline Sites

56. The distributions and total numbers of individuals of amphibians and reptiles recorded from the permanent sites in each pool are presented in Tables 2 of this report and A3 of Bancroft et al. (1983). A total of 7754 observations (65.0% of the total sample) was made on these five sites during the three-year study. The bulk of the quantitative data (86.34%) from these permanent sites was obtained with herp-patrols. Funnel trapping added 9.03% and shoreline census another 2.41%. The other five sampling techniques added between 0.86% (drift fence) and 0.04% (alligator census) to the total sample.

57. The permanent sites covered 1770 m of shoreline in Lake Conway (Table 2). The linear distance sampled and time spent on herp-patrols and with funnel traps varied among pools (Table 2). Of the 29 species of amphibians and reptiles known from Lake Conway, 28 were recorded on the permanent sites. Only the treefrog Hyla femoralis was not taken on a permanent site. One salamander (Eurycea quadridigitata), one frog (Hyla squirella), one turtle (Deirochelys reticularia), and four snakes (Coluber constrictor, Regina alleni, Thamnophis sauritus, and Thamnophis sirtalis) were recorded only from permanent sites. In addition, more than 90% of the samples of the salamander Amphiuma means, the frogs Acris gryllus, Gastrophryne carolinensis, Hyla cinerea, Rana grylio, and Rana utricularia, and the turtles Chelydra serpentina and Kinosternon baurii were taken on the permanent sites.

58. The mean relative densities of each species on the five permanent shoreline sites as determined by the two major sampling methods are shown for funnel traps (mean number/100 trap days) in Table A4 of Bancroft et al. (1983), and for herp-patrol (mean number/hr) in Table A5 of Bancroft et al. (1983). Between-pool differences in the relative abundance of a species were determined by using the chi-square approximation of the nonparametric Kruskal-Wallis extension of the Mann-Whitney U-test (Farr et al. 1974) for herp-patrol trips, and the difference among proportions chi-square test (Freund 1973) for funnel trapping. The mean tested on herp-patrols was the mean number of individuals of a species observed per hour for all trips with run numbers (see paragraph

11) on each permanent site. The proportion tested was the total number of a species captured at a site divided by the total number of trap days at that site during each year of study. If significant ($P < 0.05$) between-pool differences were found, pair-wise comparisons of pools were made using the Mann-Whitney U-test (herp-patrols) or the difference among proportions test (funnel traps). Because the same data were analyzed for this second test, the alpha level of significance was increased to $P < 0.025$. In general, the sum ranks test used for herp-patrol comparisons is less robust than the chi-square test used for funnel trap analysis. This means that trends seen in a number of species on herp-patrols probably are real, but more difficult to demonstrate statistically than trends observed in funnel-trapped species.

59. The total mean relative densities (mean number/hr) of seven species recorded from the permanent sites on herp-patrols varied significantly among pools during the three-year study (Table A5, Bancroft et al. 1983). Adult frogs of the species Acris gryllus, Hyla cinerea, Rana grylio, and Rana utricularia varied significantly among pools. Of these four, only Acris gryllus had significantly different means during each study year and for the combined years. The other three species (Pseudemys floridana, Pseudemys nelsoni, and Sternotherus odoratus) each had a significant difference for one year and a total mean difference but no significant differences for all three years. Because seasonality is an important component of frog reproduction and reproductive activity varied among species, we will consider the frog data separately. The turtles Sternotherus odoratus and Pseudemys floridana showed significant between-pool differences in mean relative densities for each year as well as for the total combined values. Pseudemys nelsoni had significant between-pool differences in SY1, SY2, and total but not in SY3. The small sample (19 individuals) of P. nelsoni from the permanent sites in SY3 was not adequate to demonstrate statistically any between-pool differences although between-pool trends in SY3 generally followed those established in SY1 and SY2 (Table A5, Bancroft et al. 1983). Only Siren lacertina of the remaining 12 species recorded on herp-patrols from the permanent sites showed a significant between-pool difference and that was recorded in SY3 in East Pool.

60. Eight species of amphibians and reptiles captured in funnel traps had significant differences in total mean relative densities among pools (Table A4,

Bancroft et al. 1983). Three of these eight species (Amphiuma means, Rana gryllis adults and larvae, and Sternotherus odoratus) showed significant differences among pools in all three years. The salamander Siren lacertina and larvae of the frog Rana utricularia showed significant between-pool differences in all years except SY3 and the snake Nerodia cyclopion showed significant differences in all years except SY2. Hyla cinerea larvae and Kinosternon subrubrum had significant between-pool differences in SY1 but not in SY2 or SY3. The high SY1 values contributed to the totals which were significant for both species. Only the turtle Chelydra serpentina had a significant between-pool difference in one year (SY2) but no total difference. The fact that no specimens were taken in Gatlin Canal in SY3 and the overall small sample sizes affected the comparisons for C. serpentina.

61. The relative densities of the two salamanders as recorded with funnel traps decreased on the permanent sites through the three-year study (Table A4, Bancroft et al. 1983). Siren lacertina densities declined in Middle and East Pools and to a lesser extent in West Pool and Gatlin Canal and those of Amphiuma means declined in all pools but most noticeably in East Pool where relative density went from 30.41 in SY1 to 1.43 in SY3. Declines in A. means densities to apparent extirpation on the permanent sites in South and Middle Pools corresponded with the massive habitat destruction that occurred there in SY1 and SY2 (Table A1, Godley et al. 1981; Appendix A, Tables A1 and A2, this report). For both species of salamander, East Pool had the highest densities and South Pool the lowest. Densities in Gatlin Canal were low but showed the least fluctuations among years.

62. Interestingly, the funnel traps proved effective in sampling the relative density of larvae of three of the eight species of frogs recorded from the lake system. All of the Hyla cinerea larvae were taken in three pools during SY1 when water levels for the lake were high. Both species of Rana call from and deposit their eggs in the littoral zone usually over deeper water than the other frog species. Their tadpoles are larger and spend more time in deeper water than those of Acris gryllus, Bufo terrestris, and Gastrophryne carolinensis which are found primarily in very shallow water (<10 cm). Because little is known about movements of anuran larvae, we assume that local densities reflect the reproductive success of frogs at the same site. Habitat

disruption and seasonality as it affects frog breeding will influence larval densities. Little can be said about between-pool comparisons for frogs taken in funnel traps except that Rana utricularia larvae were considerably more abundant in SY2 in all pools than in either of the other years.

63. Funnel traps contributed 60.0% of all captures on the permanent sites of the turtles Chelydra serpentina, 52.6% of those for Kinosternon subrubrum, but only 2.3% of those for Sternotherus odoratus. Yet S. odoratus was the only species in the funnel trap data that showed significant mean density differences among pools in all years. The relatively frequent captures in Gatlin Canal in SY2 resulted in the significant between-pool differences for C. serpentina in that year (Table A4, Bancroft et al. 1983). A similar pattern existed for Kinosternon subrubrum in South Pool in SY1. The decline in K. subrubrum on the permanent sites in South Pool was the result of habitat disruption. Although no specimens were trapped in South Pool in SY3, individuals were recorded on herp-patrols (Table A5, Bancroft et al. 1983) but at lower relative densities.

64. The snake Nerodia cyclopion declined on all permanent sites and most strikingly on South, Middle, and East Pools. We attribute the decline in South and Middle Pools in part to high mortality during shoreline development. We suspect the reduction in East Pool may have resulted from high predation by river otter (Lutra canadensis) in SY1 and SY2.

65. The mean relative densities of calling males of the more common of species of frogs recorded from the permanent sites are compared in Table A6, Bancroft et al. (1983). Only data from the breeding seasons are compared. In contrast to most other species of amphibian and reptile samples from the permanent sites, densities of most frogs increased through the three years. Significant differences among pools were recorded in all three years and for total years only for Acris pygmaea. Three other species with mean differences among pools for total years and one or two study years included Gastrophryne carolinensis in SY2, and Hyla cinerea and Rana pygmaea in SY1 and SY3. While we are unable to account completely for the increase in density of calling male frogs on our permanent sites, several explanations exist. The increase may represent a high in normal fluctuations of density that are correlated with weather conditions highly favorable to breeding during the study. Another

possibility is that the increase may represent an increase in survivorship as the number of predators, especially snakes that feed on adult frogs, decreased. Also, larval survivorship and growth may have increased as a secondary effect of the shifts in vegetation.

66. Included below are detailed community analyses of the five permanent shoreline sites on Lake Conway. For each site a "point analysis" and a "trip analysis" are presented. Point analyses show the numerical distributions of amphibians and reptiles observed or captured along 10-m increments of the shoreline sites during each study year. Trip analyses show the numerical distributions of species through time on the bimonthly sampling trips to Lake Conway. The species codes used in all point and trip analyses figures are listed in Table 1.

67. At least nine figures (one for funnel trapping [total captures] and two for herp-patrols [anurans only, and salamanders and reptiles only] in each of three years) are presented for the point and for the trip analyses of each site. Each funnel trap trip analysis figure includes the total number of funnel traps set at a site (per trip or per trap station) and, for herp-patrols, the total time (minutes) spent on each herp-patrol trip at a site. Thus, each figure is scaled by sampling effort. In some cases the total number of individuals recorded on the point analysis for a site will be less than the number of individuals recorded on the trip analysis for that site; this means that some individuals on a trip were not given a sample point and thus do not appear on the point analysis.

68. Between-year comparisons of the relative abundances of species as determined by herp-patrol and funnel trapping for each permanent shoreline site on Lake Conway were summarized and significant between-year differences (not in Tables A4-A6, Bancroft et al. 1983) have been incorporated into the text of this report.

South Pool

69. The permanent shoreline site in South Pool had the highest recorded diversity (22 species) of any site but also received the most effort (Tables A4 and A6, Bancroft et al. 1983). Alligator mississippiensis was the only species known from South Pool that was not recorded on the permanent site. Four species taken from the South Pool sites are unknown elsewhere on Lake

Conway and include Hyla squirella, Pseudemys scripta, Coluber constrictor, and Thamnophis sirtalis. The relative density of Kinosternon subrubrum was significantly greater (78.9% of the total sample from all permanent sites) on the South Pool site than on any other permanent site in Lake Conway (Table A³, Bancroft et al. 1983). High densities for Acris gryllus adults (62.1% of the total sample from all permanent sites) and Nerodia cyclopion (51.8%) also were recorded on the South Pool permanent site. The highest total number of observations for four other species was recorded from the South Pool site and included the turtles Kinosternon baurii, Sternotherus odoratus, and Trionyx ferox and the snake Farancia abacura. Thus, many components of the Lake Conway community of amphibians and reptiles are known best from South Pool. A conspicuous absence from the South Pool site is the frog Rana grylio. Interestingly, relative densities of Rana utricularia also are lower on the South Pool sites than on the Middle, West, and Gatlin permanent sites.

70. Unfortunately, the shoreline of the South Pool permanent site was developed gradually for houses during the three-year study until nearly all the littoral vegetation was removed or the habitat strongly modified. At the beginning of the study (July 1977), 460 m of the 530 m permanent site was a continuous, undeveloped section of shoreline habitat (Godley et al. 1981). The northern 70 m of the site was a sandy, developed beach. By the end of SY1, 78% of the upland and transitional zone habitat bordering the permanent site had been cleared of understory vegetation to the waterline and about 25% of the emergent littoral zone vegetation was removed. In SY2 most of the remaining upland habitat was cleared and an additional 25% of the emergent littoral zone vegetation was eliminated. The same trend continued through SY3 so that less than 20% of the upland and shoreline habitat and only 30% of the original littoral zone habitat remained on the permanent site. The impact of this habitat modification and loss had a profound effect on several species in South Pool, notably Amphiuma means, Kinosternon subrubrum, and, especially, Nerodia cyclopion. Densities of Nerodia cyclopion decreased 97% (from 101 records in SY1 to 4 records in SY2 and 3 records in SY3). Details of this change are presented in Bancroft et al. (1983).

71. The number of species sampled each year in South Pool and on the permanent site decreased between SY2 and SY3 (Table 3). In order to evaluate

this apparent decrease in diversity, the individual samples from SY2 and SY3 were compared to the cumulative plots of number of species and number of individuals collected in SY1 (Table 3 and Figure 2). We assumed that the curves for SY1 were representative of baseline conditions and that the permanent sites adequately represented the pools. The number of species based on sample sizes from SY2 and SY3 were read from the curve for SY1 and used as the expected species diversity for the sample. This expected value was compared to the actual species number recorded in SY2 and SY3 (Table 3). For the total of South Pool, the SY2 sample of 1019 individuals of 21 species was the same number as predicted if the sample had been taken in SY1. However, the SY3 sample size predicted 21 species but only 17 were taken. For the permanent site a similar pattern occurred. The SY1 sample of 1221 would have resulted in 21 species for the total pool plot (Figure 2) but only recorded 20. The much smaller SY2 sample predicted only 18 species but 20 were recorded. By SY3, the expected diversity should have held at 18 but only 14 species were recorded. Thus, a reduction in diversity of four species was noted in the South Pool and on the permanent sites for SY3.

72. Point analysis. The distributions of all amphibians and reptiles captured in funnel traps along the South Pool permanent shoreline site for each year are shown in Figures 3, 4, and 5. Most (91.8%) of the 66 captures in South Pool in SY1 were concentrated between markers 0 and 100 and between markers 360 and 460 where 47.9% of the traps were set. In general, these two productive areas at each end of the transect were characterized by a diverse, emergent littoral zone flora and mud substratum, while the depauperate, central area was dominated by Euirena scirpoides and Panicum hemitomon and had a sand substratum (Godley et al. 1981, Table A1).

73. Even though trapping effort increased (Table A4, Bancroft et al. 1983) in SY2, only 17 captures were recorded. Three of these captures were of relatively rare species including the snakes Farancia abacura and Pegana alleni and the turtle Pseudemys scripta. Larvae of Rana utricularia were taken at four sites and three Sternotherus odoratus were trapped at marker 0. Siren lacertina occurred at markers 10, 150, and 440. The major concentrations of the fauna recorded during SY1 (Figure 3) at each end of the site was not evident in SY2 (Figure 4). Most of this faunal reduction was attributed to the

loss of upland and transitional habitat in late SY1 and SY2 and the removal of emergent littoral zone vegetation between markers 6 and 40 in SY2. Most of the reduction at the ends of the site was due to the disappearance of individuals of Amphiuma means, Kinosternon subrubrum, and Nerodia cyclopion (Figures 1 and 4).

74. Trapping effort in SY1 was increased in SY2 (Table A4, Bancroft et al. 1987) in SY3, yet only one capture was made. A single specimen of the snake Nerodia cyclopion was taken near marker 40 (Figure 1). Interestingly, the upland and transitional habitat in the intermediate area was still intact and some littoral zone vegetation remained (Appendix B, Table A1).

75. Comparisons among years were made for densities of turtles as determined by funnel trapping. The salamander Amphiuma means ($\chi^2 = 1.4$), the turtles Kinosternon subrubrum ($\chi^2 = 88.4$) and Chrysemys picta ($\chi^2 = 1.4$), and the snake Nerodia cyclopion ($\chi^2 = 11.4$) had highly significant differences among years. In all four species the mean relative densities were highest in SY1 (they in the other two years; no differences between SY2 and SY3 were detected).

76. The spatial distribution of turtles and salamanders observed or collected on herp-patrols for the three years (Figures 1, 2, and 3) on the same section of shoreline was similar to that observed for the funnel-trapped animals. These comparisons apply only to the area between markers 6 and 40; traps were not set between markers 40 and 100 and observations from this section are not comparable. Most observations (91.4%) of turtles on herp-patrols between markers 6 and 40 in SY1 were located between markers 6 and 100 and 100 and 400. In SY2 (81.8%) and SY3 (86.1%) similar trends were observed.

77. Much of the high densities recorded at the ends of the transect was the result of turtles. In SY1, 96.5% of the sample in these areas was made up of turtles of five species; in SY2 the values were 90% and four species and in SY3 they were 100% and four species. One possible explanation for the greater observed densities at each end is that we spent more time and covered a greater area with the turnarounds at each end of our transect than at other stations. Examination of distributions at the ends of transects in other pools does not support this interpretation. These high concentrations, especially of

Sternotherus odoratus and Pseudemys floridana, at each end of the transect may be the result of offshore habitat preferences rather than a preference for the littoral zone with its emergent vegetation. At the ends of the transect broad areas of shallow water (<1.5 m) extended out some distance from shore and most turtles were captured in these shallower areas (Figures 6, 7, and 8). Along the central section of the South Pool site, deep water (>2 m) occurred immediately offshore from the edge of the emergent vegetation and turtle observations were fewer. The area from marker 460 to 530 had the greatest concentration of turtles on the site (Figures 6-8) with 34.7% of the total sample recorded in SY1, 41.3% in SY2, and 45.4% in SY3. Importantly, this section of developed shoreline had extensive, heavily vegetated (Potamogeton illinoensis) shallows but no emergent vegetation. We conclude that these high turtle densities are a reflection of offshore habitat preferences.

20. Some trends are obvious in comparing the distribution of species along the permanent site through the three years (Figures 6, 7, and 8). More gaps in the local distribution of species are obvious in SY2 and SY3 than in SY1. Eleven species are represented in SY1, five in SY2, and only four in SY3. Individuals of Sternotherus odoratus were recorded at 40.7% of the total points on the site and 90.4% of the points with animals in SY1, on 14.8% of the total site and 94.6% of the sites with animals in SY2, and on 64.8% of the total site and 90% of the sites with animals in SY3. The relative abundances of most species decreased during the three years but that of Sternotherus odoratus did not. S. odoratus made up 69.2% of the herp-patrol observations of reptiles and salamanders in SY1, 87.2% of the observations in SY2, and 93.2% of the observations in SY3. This suggests that even though the absolute density of stickpoles decreased through the study, S. odoratus responded differently than most other species.

21. Only a single species recorded on herp-patrols showed a significant decrease among years. The turtle Pseudemys floridana had highly significant ($\chi^2 = 2.54$) differences among years with the mean relative density in SY1 higher than that in SY2 and SY3, which were not significantly different from each other. During SY1, 57.5% of the herp-patrol sample of P. floridana (N=47) was recorded between markers 0 and 100 and 26.4% between markers 430 and 530. Three of the seven turtles taken in SY2 and the entire sample (N=3) in SY3 were

from the same area. We associate this dramatic reduction in relative density to the equally dramatic reduction of vegetation at the ends of the transect. At the beginning of the study (August 1977) extensive beds of Potamogeton illinoensis covered the bottom along the site but especially in the shallower northern areas. Hydrilla verticillata was common along the southern section of the site. By August 1980, P. illinoensis and H. verticillata had been eliminated from the site by grazing white amur (Schardt et al. 1981, Figures J1 and J2). Although sample sizes were much smaller, a similar pattern of decline existed for the other large, herbivorous turtle (Pseudemys reisoni) on the site.

80. The distribution of calling frogs on the South Pool site is given in Figures 9, 10, and 11. The cricket frog, Acris gryllus, was the most common frog on the site (Tables A1 and A6, Pancroft et al. 1981). During SY1 males of A. gryllus were recorded calling along most of the permanent site except from the area between markers 400 and 500 which was a beach (Figure 9). The same pattern existed in SY2 (Figure 10) but by SY3 (Figure 11), the impact of habitat disruption and shoreline development was obvious on the dispersion of calling males. In SY1 and SY2, A. gryllus called from 77.8% and 75.0% of the points on the transect. In SY3, males were recorded from only 42.0% of the points. Noticeable absences in areas where males were common in previous years occurred between markers 0 and 100, an area developed in April and May of 1979 (Appendix A, Table A1) and between markers 140 and 200, which was developed in May of 1980 (Appendix B, Table B1). Other species were more patchily dispersed and relatively uncommon on the South Pool site. Even so, Hyla cinerea was often recorded in areas where Eichhornia crassipes was the dominant plant and Rana utricularia often occurred in areas where Typha latifolia occurred (Appendices A, Table A1, and B, Table B1).

81. Trap analysis. The temporal distribution of amphibians and reptiles collected in funnel traps on the South Pool permanent shoreline site is shown in Figure 12. Trap success decreased with time even though the number of traps increased within SY1 and among years. The highest success rate occurred early in the study, between 21 July and 24 September 1977, when traps were set only between markers 0 and 120 (Godley et al. 1981). These traps accounted for 74 (66.6%) of 51 total captures from this section of shoreline during SY1. When

the trapline was expanded on 31 March 1978 to include the entire 400-m transect, most specimens were taken between markers 0 and 120 and between markers 40 and 400, although the capture rate for the section between 0 and 100 was lower than originally found. Because sample sizes in SY2 and SY4 were small, clear seasonal patterns in funnel-trapped animals were not obvious. In general, more animals were trapped in the summer months (June through September) than at other times of the year. Activity patterns of single-species are difficult to evaluate for this site alone because of the relatively low trapping success achieved, especially in SY4 and SY6. More striking patterns may become more obvious when data are combined for all trapping techniques and across all roads; these are discussed in a separate report (Chandler et al. 1987). Suggestive patterns of activity were found for Nerodia cyclopion which was absent in traps from November through mid-May. During this time, individuals were observed leaving the permanent sites and moving into upland overwintering sites. Site preparation of individuals for housing during the winter months effectively estimated the North Fork population of Nerodia cyclopion. Amphiuma means and Pleurodonta subtruncata seemed to be active throughout the year.

83. The temporal distributions of salamanders and reptiles on North Fork herp-patrol trips are given in Figure 10. Most individuals were observed during the months of July through November of the first year and decreased in abundance thereafter. A general pattern of observational peaks in activity in the warmer months (May through November) and valleys in the colder months (December through February) exists (Figure 10). The large number of observations on 17 November 1977 represented a high, local density of animals with observations of 26.74 individuals per hour obtained (Godley et al. 1981).

84. The majority of observations on nearly all herp-patrols was of the star-nosed, Stenrotherus odoratus. The abundance of this turtle accounted for the prominent seasonal patterns of the first two years (high densities in summer and fall and low densities in winter and spring). In SY4, S. odoratus made up 93.2% of the total sample.

85. During SY1, 19 specimens of Nerodia cyclopion had been recorded from the site. Only one specimen was observed on herp-patrol after 4 May 1978 and that was in November 1978, the same month the previous year when 10 individuals

were observed. No specimens were seen on herp-patrol after 10 November 1976. Reductions in density of this snake also were seen in the funnel trap results (Figure 12) and reflect their general decline as a result of habitat destruction.

iii. Frog calling activity on the South Pool site decreased through the study, was highly seasonal, and varied by species and year (Figure 14). The sample consisted of six species and was dominated by Acris gryllus whose calling activity established the general seasonal pattern. Calling males of five species including A. gryllus were noted from late February through September with the peak of activity occurring in June through September depending on the year. Rana utricularia was heard calling only from November through early May with most activity between February and April.

Middle Pool

iv. Nineteen species of amphibians and reptiles were recorded from the Middle Pool permanent site. The only species known from Middle Pool but not recorded on the permanent site was Hyla femoralis. The only Acris gryllus larva collected during the study was recorded from Middle Pool in SY1. In addition, the highest densities for permanent sites (Table A4, Bancroft et al. 1983) of the following species were recorded in Middle Pool: Rana grylio adults and larvae, Rana utricularia adults and larvae, Alligator mississippiensis, Regina alleni, and Thamnophis sauritus. Middle Pool was the only site where the relative densities of the salamanders Amphiuma means and Siren lacertina were nearly equal (Table A4, Bancroft et al. 1983). (One of the two female alligators known to nest on Lake Conway successfully hatched 12 to 16 young in the marsh of the Middle Pool site in August 1977. Unfortunately, the nest site (marker 1110) of this female Alligator mississippiensis was destroyed in April 1978 and she apparently did not nest again on Lake Conway during our study.

v. The Middle Pool permanent site underwent dramatic habitat change during SY1. On 26 April 1978 all upland, marsh, and shoreline vegetation between markers 1040 and 1120 and between markers 2000 and 2120 was cleared with bulldozers and draglines in preparation for housing (Godley et al. 1981). This activity removed more than half of the natural habitat along the inner transect that was used for funnel trapping and along the inner herp-patrol transect (2700 series) through the center of the Cyrba latifolia marsh. A 10-m

outer fringe of cattail (markers 1000-1120) and the habitat to the west of the cleared area (50 m of the 1000, 2000, and 3000 series transects) were all that remained natural on the site. No other major disturbance occurred on the site through the remainder of the study.

2b. A comparison of the number of species within Middle Pool and on the permanent site among years (Table 4) indicated a general reduction in diversity on the permanent site but not in the total pool through the study. For the total pool, samples in SY2 predicted 16 species (based on SY1 samples, Table 4 and Figure 2) and 16 were observed. In SY3, 16 species were found but the sample only predicted 15. On the permanent site, 16 species were predicted and observed in SY1. The smaller sample (16-40) of SY2 recorded 14 species and 14 were predicted. However, the SY3 sample predicted 14 but only 11 species were recorded. Noteworthy among the species not collected were Anthonysia sp., Pseudomys nelsoni, and Nerodia cyclorion.

2c. Point analysis. The distribution of three-transect samples on the permanent site is shown for each year in Figures 15, 16, and 17. Capture decreased in SY3 (Figure 17) even though the trapping effort was doubled. The species captured decreased from 8 in SY1 and SY2 to 5 in SY3. In SY1 and SY2 (Figures 15 and 16) animals were concentrated in the undisturbed portion of the habitat (markers 1150 to 1200). Sample sizes of individuals and species at SY3 were too small ($N=5$) to determine a meaningful pattern but some concentration of the area cleared in SY1 occurred (Figure 17), probably from the adjacent, undisturbed areas where Regina alleni and Fiber locustina were captured the previous year. A comparison of species distributions on the permanent site in SY1 before and after the habitat change (Godley et al. 1981, Figures 1 and 2) indicated that the removal of emergent vegetation severely reduced the diversity and diversity of the herpetofauna in the disturbed area and that surviving individuals apparently moved into the adjacent, undisturbed habitat. With the exception of the single larva of Pana strimmaria taken at marker 1150, all captures of animals from markers 1150 to 1160 (Figure 15) were made before the habitat disruption occurred. Also, the numbers of individuals and species in the undisturbed area increased following the adjacent habitat disruption. Most captures on the permanent site in SY1 were in areas where organic detritus was deep and covered the bottom and where waterhyacinth (Eichhornia crassipes)

often was the dominant plant (Godley et al. 1981, Table A.1). In that part of the transect where sand or shallow mud predominated, captures were few and occurred in the mole (markers 100-1120) vegetated by Peridium hemisphaerum and Salicornia virginica. In the wet mud area (markers 1120-1200) dominated by Typha latifolia or Pontederia lanceolata produced no captures. Similar trends in vegetational association and capture rate were evident in SY2 and SY3.

4. Comparisons among years (Table A2, Benaroff et al. 1981) showed that four species including Ambystoma macrodactylum ($\chi^2 = 16.6$), Ambystoma macrodactylum ($\chi^2 = 16.6$), Bombina orientalis larvae ($\chi^2 = 16.6$), and Larophis cyclonotus ($\chi^2 = 16.6$) had highly significant differences among years as determined by funnel trapping. In all of these except Bombina orientalis, the means were higher in SY1 than in the other two years but were not different between SY2 and SY3. In Bombina orientalis larvae, each year was distinct with the highest density recorded in SY2.

5. The spatial distributions of salamanders and reptiles as recorded from berry-patrols on the Middle Pool permanent site for the three years are presented in Figures 18, 19, and 20. For purposes of presentation, data from the 2000 and 100 series transects are figured together. The peak of abundance seen (Figure 18) near the center of the 2000 series transect (markers 2000 to 2090) was the result of predisturbance observations in SY1. No specimens were recorded from this area after 25 April 1978 (Godley et al. 1981, Figures 11 and 12). The pattern shown in this area in Figure 19 was a disturbance picture of SY1. By SY2, nine captures of turtles between markers 2060 and 2090 resulted in a pattern similar to that of SY1. As previously discussed (Godley et al. 1981, paragraph 75) a marked decrease in mean observations from 1.21 individuals/trip to 0.40 individuals/trip occurred on the disturbed portion of the transect after clearing. No differences between the predisturbance and postdisturbance abundances ($\bar{X} = 1.20$ individuals/trip) on the undisturbed portion of the 2000 series transect were detected.

6. The numbers of species and individuals recorded from the 2000 series transects decreased from 7 species and 119 individuals in SY1 (Figure 18) to 4 species and 12 individuals in SY2 (Figure 19) to 2 species and 19 individuals in SY3 (Figure 20). As expected, individuals of Sternotherus odoratus made up most of the SY2 and SY3 samples.

93. Even though the 100 series transect was not affected directly by tree lines used to clear part of the permanent site in April 1978, distributions clearly changed along the transect (Godley et al. 1981, Figures 13 and 14). Before clearing, the mean number of individuals was 11.4/trip and was reduced to 1.7/trip afterwards. Little change was noted for the undisturbed portions of the 100 series transect where mean abundance went from 2.10/trip before disturbance to 1.7/trip after disturbance. Even though the number of species changed during the three study years from 4 to 1 to 3 along the 100 series transect and the number of individuals from 417 to 1 to 11 from SY1 to SY3, the pattern of distribution was very similar (Figures 13, 14, and 20). Peaks in abundance occurred in each year near marker 400 and to a lesser extent between markers 4140 and 4150. A very dense patch of Potamogeton illinoensis occurred near marker 400 and persisted there through time even though Schardt et al. (1981) reported an overall reduction in percent coverage of this species in Middle Pool from 28% to 8% from 1977 to 1980. The lake bottom was generally bare of vegetation from marker 400 to 4120. Near the end of the Typha latifolia stand (marker 4150) another dense patch of P. illinoensis occurred. Apparently, Potamogeton illinoensis provides excellent habitat for Sternotherus odoratus perhaps through its support of specific prey for the stinkpot.

94. The distributions of calling frogs changed as a result of the habitat modification in SY1. Godley et al. (1981, Figures 15 and 16) reported four species from markers 1000 to 1120 before disturbance and only two following disturbance. Both species of Rana, Hyla cinerea, and Acris gryllus were recorded in this area prior to disturbance but only Acris gryllus and Bufo terrestris, two species that inhabit and successfully call from grassy shore or bare beach, used the area following disturbance in SY1. All species were recorded in the undisturbed area (Figure 21). During SY2 Acris gryllus was the primary species calling from the disturbed area but a few individuals of Hyla cinerea (7), Rana utricularia (7), and Rana grylio (1) also were recorded (Figure 22). By SY3, these three species were abundant and dispersed throughout the area (Figure 23). Hyla cinerea, Rana grylio, and Rana utricularia call from areas of dense, emergent vegetation; this habitat had been removed completely in April of 1978 and only gradually reappeared by SY3.

With the redevelopment of the emergent, littoral zone vegetation, the four species moved back into the area.

95. Comparisons among years (Table A4, Hancock et al. 1983) showed significant differences in two species as determined by Kerr-patrols. Adult Rana utricularia were significantly different ($\chi^2 = 11.290$) among years with the SY1 sample being higher than those in the other two years which statistically were the same. Pseudemys floridana had significant differences among years ($\chi^2 = 10.55$) with SY1 being higher than SY2 and SY3 which were statistically the same.

96. Trip analysis. The temporal distributions of amphibians and reptiles taken in funnel traps are shown in Figure 24. An overall decrease in captures occurred through the three-year study. Adult and larval specimens of Rana sylvia and Rana utricularia made up the bulk of the sample especially in the winter months and in SY1 and SY3. Nerodia cyclopion was active from March through August in SY1 but recorded only once in April of SY2. No specimens were captured during the balance of SY2 or in SY3. The activity of Amphiuma means was restricted to the period from March through October. Small samples of the other species prevented meaningful analysis. During SY1, trap success was significantly greater on a disturbed section prior to habitat destruction than afterwards ($\chi^2 = 4.12$, $P < 0.05$). More animals were collected on the undisturbed section after habitat disruption than on the adjacent part of the transect before, but the difference was not significant (Godley et al. 1983).

97. In Figure 25, the distributions of salamanders and reptiles recorded on Kerr-patrols on the Middle Pool permanent sites combines the 200 and 400 series data for the three years. An overall decrease in abundance is obvious following the April 1978 habitat disruption. Sternotherus odoratus dominated the three-year sample, especially in SY2 and SY3 when it made up 94.0% of the sample. The high number (1500) of S. odoratus recorded on 7 December 1977 (Figure 25) was the result of an incredible concentration of 17 turtles in a 10-m² section of habitat between markers 60 and 70 of the 200 and 400 series transects. No other concentrations of this density were observed in Middle Pool during the remainder of the study.

98. Few frogs called on the Middle Pool site from October through February of SY1 and SY2 (Figure 26). However, 66 Rana utricularia were

recorded during the same period in SY3. Occasional individuals were calling in every month in SY3. Rana grylio usually began calling after the peak in R. utricularia and was more common during the later spring and summer months. Hyla cinerea called from May through September as did Acris gryllus but occasional individuals of the latter species sometimes began calling as early as March. The frog densities on the permanent site increased in SY4 to near SY1 densities. The low densities of SY2 probably were the result of the habitat destruction in April of 1978. Rana grylio was the only frog that showed highly significant increases ($\chi^2 = 17.47$) in relative densities (Table A6, Bancroft et al. 1983) in SY3; values for SY1 and SY2 were statistically similar.

East Pool

99. Only 17 of the 20 species known to occur in East Pool were recorded on the permanent site (Table A6, Bancroft et al. 1983). The toad Bufo terrestris, the turtle Kinosternon subtratum, and the snake Ferania abacura are in East Pool but were not recorded from the permanent site. The salamander Eurycea quadridigitata was taken only once in the lake system and those two specimens were obtained in SY1 from the East Pool permanent site. The dwarf siren Pseudobranchius striatus was known only from East Pool and one of the two specimens was taken on the permanent site in SY1. The relative densities of three species (Amphiuma means, Siren lacertina, and Hyla cinerea larvae) were higher on East Pool than any other permanent site. All of these species were commonly associated with waterhyacinth (Eichhornia crassipes) communities, which dominated much of the site (see Tables A1 and B1). The lowest densities of adult Rana utricularia and Sternotherus odoratus recorded on any permanent site occurred on the East Pool site.

100. In contrast to the conditions described for South and Middle Pools, no habitat destruction occurred on the East Pool permanent site during the study. However, some changes in the depth of the detrital layer or the transect and replacement of some plant species by Eichhornia crassipes were noted. Also, the percent cover and standing crop of Potamogeton illinoensis decreased dramatically as a result of the white smut (a reduction of 81.1% was noted from August 1977 to August 1980; Schardt et al. 1981).

101. The species diversity for the total of East Pool and the East Pool

permanent site is presented in Table 5. Applying the same logic as described for the South Pool comparisons, little change occurred in the East Pool community through the three years. The same number of species recorded in SY1 were also there in SY3 with a smaller sample size. For the total pool, the SY2 sample predicted (see first year's data in Table 5 and Figure 2) only 14 species, but 17 were recorded. The permanent site samples recorded 15 species (16 predicted) in SY1, 15 (11 predicted) in SY2, and 15 (14 predicted) in SY3.

10%. Certain characteristics of the East Pool permanent site and its inclusive samples are important to understanding the observed diversity. First, the East Pool site is insular and in this regard different from the four other sites. As a result, a few species that were common elsewhere in the pool or on the other permanent sites are unknown from the East Pool site and include Bufo terrestris and Gastrophryne carolinensis. Also, the salamander Eurycea quadridigitata probably was common on most sites prior to upland development but now is known only from the undeveloped island in East Pool. The fact that there has been no development on or adjacent to the East Pool permanent site makes it different from the other permanent sites and may have resulted in its relatively higher diversity in SY3. Finally, trapping area and effort in East Pool (Table A4, Bancroft et al. 1985) was expanded considerably in SY3 and this may have contributed to the maintenance of a species diversity equivalent to that recorded in SY1. The species diversity on all other permanent sites was lower in SY3 than in SY1. The extended trapping effort is discussed separately but included in the figures for the point and trip analyses.

10%. Point analysis. The distribution of captures in funnel traps was more uniform on the East Pool site than on other permanent sites in SY1 (Figure 27) and SY2 (Figure 28). This probably reflects the relatively high habitat homogeneity in East Pool compared with other permanent sites. Because captures on the original site in SY1 were only 5.0% of those in SY1 and 12.9% of those in SY2, no meaningful comparisons can be made for the SY3 data. Traps set from markers 1090 to 1200 captured 66.5% of the sample in SY1 and 71.4% of the sample in SY2. This area was dominated by waterhyacinth in SY1 (Codley et al. 1981, Table A1) which spread through most of the permanent site in SY2 and SY3 (Appendices A, Table A1, and B, Table B1). Amphiuma means was captured at every marker in SY1 (Figure 27) and every marker except 1040, 1050, and 1160 in

the Figure 291). In SY3 only two specimens were recorded on the original transect, both from marker 1120 (Figure 291). In SY1 Aren. laevis was taken at all stations except between 1050 and 1100. In SY2 A. laevis was recorded from only eight of the 14 stations (from 1000 to 1100). Only two individuals were taken on the original site in SY2, one at marker 1000 and the other at marker 1100. Heredia cyclorion was more common (14 of 14 individuals recorded in SY2) at stations with greater plant species diversity (Killey et al. 1981). Only three individuals were taken in funnel traps in SY2 and four from the original site in SY1. All were from markers with greater recorded plant diversity than adjacent stations where waterhyacinth was the dominant plant.

3d. The extended trapping in SY2 covered the area from marker 1000 to marker 1400. As mentioned in paragraph 3a, this extended trapping site was set up to determine if declining populations of amphibians and reptiles on the permanent site were the result of repeated trapping and human disturbance or were characteristic of the entire island. Large portions of this extension (especially between markers 1100 and 1400) were comparable in habitat characteristics and presumed habitat quality to the original permanent site (see Table A-1 and text). Of the 16 animals taken on this extended site (Figure 30), 11 were Amphiuma means. If the relative density in SY2 of Amphiuma means (1.42) and 41, for SY1 at 0.1 (1981), the commonest funnel-trapped species in East Pool, is partitioned into a density from the original permanent site (proportionally to the area) and one from the extended site, the following values are obtained: original site relative density (0.48), extended site (1.72). Even though the density on the new site was 4.5 times higher than that on the original site in SY1, the density of A. means was 79.3% lower than the SY2 density and 94.1% lower than the SY1 density on the permanent site, and was considerably below that expected for an undisturbed and untrapped section of high quality habitat. These comparisons clearly indicate a general reduction in the density of Amphiuma means (and to a lesser extent for the other species sampled by funnel trapping) on the island and suggest that our disturbance and repeated trapping contributed little to the dramatic decline of the A. means population on the East Pool permanent site. To some extent Amphiuma means may become trap-shy with repeated captures.

105. Predation by river otter (Lutra canadensis) may have contributed to the decline in capture rate and resulted in a loss of a substantial proportion of the populations of Amphiuma means, Siren lacertina, and Nerodia cyclopion. From the middle of SY2 through the remainder of the study, some traps on the East Pool site were raided on 40-50% of the trips. Bait was removed and eaten as were captured animals. The otters typically visited most traps in the line. However, even on trips when otters did not raid the traps, relatively few captures were recorded on the permanent sites or on the extended trapline. We conclude that human disturbance, otter predation, and trap shyness contributed to the apparent population declines, especially of A. means on the East Pool permanent site. These factors, however, cannot explain the observed low density on the extended trapline. We assume that these population fluctuations are largely the result of unknown natural causes.

106. Comparisons of relative densities among years for the East Pool permanent site showed significant declines in several funnel-trapped species. Differences in Amphiuma means densities among years (Table A4, Bancroft et al. 1983) were highly significant ($\chi^2 = 14.56$). The same pattern existed for Siren lacertina ($\chi^2 = 64.36$). Larvae of Hyla cinerea ($\chi^2 = 32.83$) and Nerodia cyclopion ($\chi^2 = 22.06$) were highly significant with SY1 different from SY2 and SY3 which were not different from each other. The larvae of Rana utricularia were highly significantly different among years ($\chi^2 = 28.53$) with no significant difference between SY1 and SY2 densities but both were higher than SY3. The larvae of Rana grylio also were highly significantly different among years ($\chi^2 = 10.86$) with a significant decline only between SY1 and SY3. Adults of Rana grylio showed a significant decline among years ($\chi^2 = 6.38$) with the same pattern as described for R. grylio larvae.

107. Comparisons of the distributions of salamanders and reptiles recorded on herp-patrol on the East Pool site are shown in Figures 30, 31, and 32. The decrease in sample size from SY1 (N=60) to SY2 (N=22) and subsequent increase in SY3 (N=46) does not correspond to the total hours spent on herp-patrol, i.e., the mean total number of animals per hour was 5.21 in SY1, 2.44 in SY2, and 4.90 in SY3. The species composition of the sample also changed through the three years. The turtle Emydoidea blandingii made up 68.3% of the sample in SY1 (Figure 30), 20.7% in SY2 (Figure 31), and 1.5% in

SY3 (Figure 32). During the same period, the percent of Siren lacertina increased from 1.67% in SY1 to 22.7% in SY2 to 28.3% in SY3 and that of Pseudemys floridana from 20.0% in SY1 to 56.4% in SY2 to 45.6% in SY3. A majority of animals recorded on herp-patrol were from the second half of the transect (markers 2110 to 2200) and included 71.7% of the sample in SY1, 72.7% in SY2, and 73.9% in SY3. The second half of the transect was dominated by a dense cover of waterhyacinth near the shore. Extensive beds of Potamogeton illinoensis also covered the lake bottom in SY1 especially along the second half of the transect. By August of SY3, the percent cover of this plant in East Pool had been reduced from 57% to 7% and it was essentially gone from the northern side of the island (Schardt et al. 1981). At the same time, beds of Vallisneria spiralis were constant or expanded in coverage and density on the transect, or increased from 15% to 40% in the pool (Schardt et al. 1981, Figure 31b). Particularly noticeable were dense beds of Vallisneria offshore from markers 2140 to 2200. Declines in Sternotherus odoratus densities have been noted previously in South Pool and correlated with the decrease in coverage and density of Potamogeton. The cover provided by Vallisneria beds apparently is important to maintenance of populations of Siren lacertina; Pseudemys floridana is known to feed on Vallisneria extensively (Bancroft et al. 1983). In addition, the increase in waterhyacinth during SY2 and SY3 provided ideal cover for large P. floridana and a suitable basking area during the winter months.

10c. The distributions of calling frogs on the permanent site in East Pool are shown in Figures 33, 34, and 35. Hyla cinerea was especially common at this site and individuals called from every marker in each year. Distributions varied from year to year but densities were usually lowest between markers 194 and 1100. Coincidentally, this area had relatively high densities of the snake Nerodia cyclopion. As was the case in other pools, frog density generally increased in SY3 in East Pool.

10d. Trip analysis. Figure 36 shows the distribution of tunnel-trapped amphibians and reptiles on the East Pool permanent site through the three-year study. Most of the pattern in Figure 36 is defined by the distribution of Scaphiopus trous. The low activity characteristic of the winter months of 1979-80 followed by high activity during the remainder of SY3 was not repeated in SY2 or SY3. Although sample sizes were smaller, the highest peak in SY2

occurred in January of 1972 and tris during June, August, and September of 1972 recorded no animal activity. Larvae of the frogs Hyla cinerea, Rana gryllus, and Rana utricularia were collected from July through September in SY1 in April, May, July, and August of SY2 and February and September of SY3. The February record of R. gryllus suggests that larvae of that species may take six months or more before metamorphosing. Results of trapping the extended area in SY3 are shown as an insert in Figure 4 (see paragraphs 104-105).

110. The distributions of salamanders and reptiles recorded on the East Pool permanent site on terr-patrols through the three years is shown in Figure 37. The trend towards a decline in overall activity is the result of the large numbers of Desmognathus pleurolineatus that were taken early in SY1. This species decreased in abundance on the East Pool site through the three years. In contrast, Pseudemys floridana increased in abundance. Peaks of activity in P. floridana were concentrated in December and January of SY1 and SY3. Observations in SY2 were low and amounted to 29.3% of those in SY1 and 49.0% of those in SY3. From January through June of 1972 only seven animals of four species were recorded (Figure 37).

111. Calling frogs showed distinct peaks from April through September of each year on the East Pool permanent site (Figure 38). Most of the summer activity was due to calling Hyla cinerea and Acris gryllus with a few individuals of Rana gryllus. During the study, the number of calling Acris gryllus increased on the permanent site while that of Hyla cinerea decreased. Acris gryllus showed a significant increase in relative density ($\chi^2 = 6.57$, $P = 0.04$) among years with SY3 densities distinctly higher than those of SY1 or SY2 which were statistically the same (Table A6, Bancroft et al. 1983). The fourth frog species on the site, Rana utricularia, called primarily from November through March with a few individuals recorded in April or May (Figure 38).

West Pool

112. The West Pool permanent site had the lowest number of species (N=15) of any permanent site and in this sense paralleled the low diversity (N=17) of the pool (Table 1). The low number of species in West Pool was in part attributable to the absence of several species of snakes (only two of seven species in the lake are known from West Pool: Nerodia cyclopion and N.

associates) and the widely distributed turtle Kinosternon flavum. Also absent from the site but known from the pool are Trionyx ferox and Alligator mississippiensis. Several species had their lowest densities of any permanent site on West Pool, including Acris gryllus, Kinosternon subrubrum, Desmognathus floridana, and B. nelsoni. Sternotherus odoratus, the most abundant turtle on Lake Lanaway, also had a low density in West Pool. Interestingly, the highest number of total observations for Gastrophryne carolinensis and Hyla cinerea was recorded from this site.

113. The West Pool site did not suffer the habitat destruction characteristic of South and Middle Pools. However, some natural shifts in plant coverage and dominance were noted. Appendices A, Table A4, and B, Table B4. Sagittaria arifolia and Potamogeton amplifolius filled in most of the open space between markers 1 and 20. Antennary growth increased considerably in coverage as did Cypha latifolia betw. between August 1971 and August 1972. Hydrilla verticillata was eliminated from the pool and Botanocystis elliptica was reduced from 14% to 2% coverage (Robert et al. 1972). During the same period, most Vallisneria spiralis beds were reduced in size (Robert et al. 1972, Figure 33b), but plants showed evidence of stem height reduction, the apparent result of cropping by the white egret.

114. The number of species and sample sizes for the total West Pool and the permanent site are shown in Table 4. All species known to occur in the pool were recorded in 1972. This number (17) was three species higher than predicted from the SY1 curve (Table 2 and Figure 1) and included Bana gryllus and Nerodia fasciata, species known only from SY1 in West Pool. The SY1 sample (N=44) had one less species than was predicted from the SY1 data. Trends of the permanent site were similar to those in the pool with more species known from 1972 than the other years. For the permanent site, SY1 diversity was two species below the pool diversity and that of 1972 was one species less than predicted. The 1972 sample recorded only nine species from the permanent site, three less than expected. Species not recorded in 1972 included Eleutherodactylus, Acris gryllus, Nerodia cyclopion, and N. fasciata.

115. Point analysis. The spatial distributions of amphibians and reptiles sampled with funnel traps on the West Pool permanent site are shown in Figures 34, 40, and 41. As with results from the other permanent sites,

Densities as measured by funnel traps decreased through the three years. Amphiuma means made up the bulk of each year's sample (71.1% in SY, 66.4% in FY, 68.8% in FY2) and was widely distributed through the site except for markers 1 through 10. No animals were captured in this part of the site during the three-year study even though the area showed considerable change in plant cover and composition (compare data from Godley et al. 1961, Table A4, to those in Tables A4 and B4 of this report). The trend seen in FY (Godley et al. 1961), where nonbeach trapping sites dominated by waterhyacinth produced the greatest proportion of capture (64.7%) with only 26.1% of the traps, apparently continued into FY2 even though sample size was smaller. Those markers dominated by waterhyacinth accounted for 74.4% of the FY2 captures with 66.6% of the trapping effort. All four captures in FY2 (Figure 41) were from sites dominated by waterhyacinth (Table B4).

11b. Highly significant differences among years were found for the two salamanders (Siren lacertina ($\chi^2 = 12.46$) and Amphiuma means ($\chi^2 = 61.01$). Siren lacertina densities from funnel traps were lower in SY than in FY and FY2 which were statistically the same. Mean densities for Amphiuma means were different each year. No other species showed significant differences among years on the West Pool site.

11c. The spatial distributions of species recorded on terrapins are shown in Figures 42, 43, and 44. A reduction in annual density occurred from FY1 to FY2 on the West Pool site. Patterns among years were not obvious. In contrast to the funnel-trapped sample, FY1 animals were widely distributed even in the area from marker 1 to 50. Sternotherus odoratus made up 61.4% of the FY1 sample and shows concentrations near marker 10 in an area where a dense stand of Potamogeton illinoensis occurred in shallow water (Godley et al. 1961); this stand disappeared from the area in later years as did stinkpots. A high density of Siren lacertina was noted near marker 10. Rare individuals of Ambystoma cyclopion were spaced across most of the undisturbed shoreline (markers 10 to 50). In FY2 Sternotherus odoratus made up 41.6% of the sample and individuals were concentrated between markers 210 and 250. None was recorded within 40 m of marker 50. No Siren lacertina and only one Ambystoma cyclopion (at marker 50) were seen. By FY2 (Figure 44) 65.1% of the sample was Sternotherus odoratus and individuals were concentrated from markers 110

the 1st 1/2 and from 2/2 through 2/2. An individual of three other turtle species made up the remainder of the 2/2 sample.

The size distributions of Chelonia mydas from the West Pool site were slightly more clumped in 2/2 and 2/2 than in 2/2 (Figures 4b, 4c, and 4d). In 2/2 Chelonia mydas made up 2/2 of the sample and called from 1/2 of the sites. However, 2/2 of the Ch. mydas called from seven markers (1b, 1c, 1d, 1e, 1f, 1g, 1h). In 2/2, Ch. mydas made up 2/2 of the sample and called from 2/2 of the available sites. No distinct clusters of males were distinguishable. In 2/2 (Figure 4f) Ch. mydas made up 2/2 of the sample and called from 2/2 of the sites. Clusters existed around markers 1d and 1e, 1c to 1h, and 1c to 1h but did not contribute as major a part to the sample as in the other years in 2/2. The density and distribution of Desmarestia desmarestii at marker 1b in 2/2 and 2/2 was 2/2 of the sample calling at 1b sites in 2/2, to 2/2 individuals calling at 1b sites in 2/2, to 2/2 individuals calling at 1b sites in 2/2 (Figures 4b to 4d). Also in 2/2 Desmarestia was recorded calling from 1b sites close to the beach area of the transect. Desmarestia was recorded only from marker 1b in 2/2. Alternatively, the transect at the West Pool transect site was not suitable for this species.

The size distributions of Chelonia mydas at marker 1b and 2/2 sites on the West Pool site are shown in Figure 4e. The size distribution of Chelonia mydas in February and early lower samples than the 2/2 samples in 2/2 through August. Because of small sample sizes in 2/2 and 2/2, the size distribution of 2/2 calling at 1b, 1c, 1d, 1e, 1f, 1g, 1h was not evident in 2/2 and 2/2. Most captures were of Amphibia peens and occurred in every month except January and February. Forty-four individuals were recorded in 2/2 between 14 May and 1 September 1969 with 12 of them trapped on 1 July.

The size distributions of Chelonia mydas and Desmarestia on the West Pool site are given in Figure 4f. Peaks of abundance occurred in September of 1969, in March of 1970, in March of 1971, and February of 1972, and consisted primarily of individuals of Desmarestia desmarestii. No specimens of Desmarestia desmarestii were taken from 2/2 until March of the first two years. Observations declined dramatically in late 2/2 and 2/2. From May of 1969 until September of 1970, 2/2 calls were only 2/2 of the sample of their species and those 2/2 were recorded on 2/2 sites.

121. Three major peaks in frog activity are identified in Figure 10 and vary temporally from year to year. In 1971, the peak in summer breeding occurred in July and August, consisted of three species, and was preceded by some activity in May. In 1972, the summer breeding was spread from late June through late August and included five species. In 1973, the major calling activity was from May through July with a secondary peak in September and included four species. Hyla squirei reached its greatest densities on West Lake and accounted for most of the pattern seen in Figure 10. Rana striatularia was the only species calling from October to November April of 1972 and from November of 1973 to April, 1974. Even though most individuals restricted their calling to the colder months, a few H. striatularia called on warm nights in late May, June, and July on the West Pool site. Bufo terrestris was noted calling only in May of 1973. Desmognathus carolinensis called in all months from May through September.

Gatlin Canal

122. All of the 21 species reported from Lake Gatlin (Figure 10) have been taken on the permanent site in Gatlin Canal (Table A6, Pancroft et al. 1974). Gatlin Canal, typical of the very shallow, grassed canals in the system, has a surprisingly high number of species for a disturbed, eutrophic, man-made habitat. The highest density of Bufo terrestris of any permanent site and the only records of Bufo terrestris larvae were from Gatlin Canal. Rana striatularia densities were also very high compared to other sites. The Gatlin Canal site was the only place where all species of native aquatic turtles were recorded and the only permanent site where Emydoidea reticularia was taken. The introduced turtle Pseudemys scripta has yet to be taken in Gatlin Canal. In addition to the high turtle diversity, the highest density of Pseudemys nelsoni was recorded in Gatlin Canal; high densities also characterized the populations of Pseudemys floridana and Stemnoterus odoratus at this site. Considering the sampling effort (especially with herp-patrols, Table A6, Pancroft et al. 1974), the diversity and density of snakes was relatively low, even though the highest density of Lerodia fasciata for any site was recorded in Gatlin Canal.

123. The numbers of species of amphibians and reptiles in the total Lake Gatlin sample and the permanent site sample for each year are shown in Table

11. The total lake contained 11 species, the last of which was recorded in FY at a sample size of 100 individuals. In general, the species diversity in each year was equal to or higher than that extrapolated from the FY curve (Table 1 and Figure 10). For the total lake, the sample in FY₁ predicted 10 and 11 species were recorded; in FY₂ the sample predicted 10 and 11 species were recorded. In each instance the addition of three or four more individuals to the sample could have raised the predicted species number to equal the recorded number. For the permanent site, the FY₁ sample predicted the number of species was 10; in FY₂ the sample predicted 10 and 11 species were recorded; in FY₃ the sample predicted 10 and 11 were recorded. In FY₃, nine more individuals in the permanent site sample would have raised the extrapolated diversity to that which was recorded (Table).

12a. The Gatlin Canal site experienced several changes in plant structure and composition during the study. The density of Salix capillaris increased in FY₁ and Eleocharis acicularis increased dramatically during FY₂ and FY₃. Herbicide was used to control the waterhyacinth in FY₁ and to control the growth of Lythra latifolia during the summer of FY₂. As a result of these modifications, the terrestrial diversity in later years was high even though the total abundance decreased. In part, this high diversity is the result of the diverse array of microhabitats within the canal (Table A1, p. 1). The location of Gatlin Canal between two other major, alternate habitats, West Lake and Lake Gatlin, Figure 11 may have contributed to the maintenance of the relatively high diversity. The high diversity is even more surprising considering that most of the shoreline bordering Gatlin Canal consisted of water rowed almost to the waterline. The lack of good terrestrial habitat and the level of development and proximity of humans, most of whom have a tendency to kill snakes, certainly lowered the snake diversity.

12b. Point analysis. The spatial distributions of amphibians and reptiles at funnel trap stations for the three years along the Gatlin Canal site are shown in Figures 51, 52, and 53. Because part of the trapline had to be moved from the east side to the west side of the canal in February 1973 (see paragraph 4), Figure 53 contains data from both sides of the canal. Because habitats on each side of the canal were different, comparisons of funnel-trapped animals among years along the canal were difficult. Trapping

effort was more consistent among years in Gatlin Canal than any other site (Table A4, Bancroft et al. 1986). Even so, capture rates were low for all three years with the largest sample (N=16) having been taken in SY2 and the smallest (N=17) in SY3. Trends seen in other pools were evident in Gatlin Canal. Densities of Amphiuma means went from seven in SY1 to one in SY2 and those of Nerodia cyclopion from four in SY1 to one in SY3. Sternotherus odoratus also decreased in the funnel trap sample from ten in SY1 to three in SY2. Helicoverpa serpentina was the only species to show highly significant differences ($\chi^2 = 9.441$) among years when five captures were recorded in SY2 and none in SY1 or SY3.

126. The distributions of salamanders and reptiles recorded on herp-patrols during the study are shown for the Gatlin Canal permanent sites (west side - W series; east side - E series) in Figures 14, 15, and 16. Herp-patrols in September 1981 only included the area from the mouth of Gatlin Canal to marker 11-12. The remaining 1.0 m of the canal was blocked with waterhyacinth and impassable to foot. Catches decreased by 77.6% from SY1 to SY2 and 88.3% from SY2 to SY3. Sternotherus odoratus was the most common species in Gatlin Canal in SY1 and SY2 but was replaced numerically by Pseudemys floridana in SY3 (Figure 14). The largest concentration of stinkpots occurred at the mouth of Gatlin Canal (markers 1-40, 100-104.5), particularly on the west side of the canal where a large patch of Najas lutea was established. Other areas in Gatlin Canal that produced large numbers of stinkpots were marker 41 in SY1 and SY2 and markers 109 to 141 in SY1. Individuals of Pseudemys floridana and B. texoni were often concentrated around Najas lutea near markers 10 and 11 and 101 and 102, 41 to 45 and 141 to 145, 1980. Because of the diversity of habitats and the relatively narrow nature of the canal, any concentration of turtles with specific habitats was not obvious from herp-patrol data. Only Sternotherus odoratus showed a highly significant difference ($\chi^2 = 9.441$) among years in Gatlin Canal. The relative density of stinkpots decreased each year.

127. Relatively few frogs were recorded on the Gatlin Canal shoreline sites in SY1 as compared to other sites and years. Frogs were more common in Gatlin Canal in SY2 and SY3. The distributions of frogs during the three-year study are shown in Figures 17, 18, and 19. Distributions were noticeably patchy

with concentrations of Hyla cinerea in areas dominated by Eichhornia crassipes usually with some Typha latifolia (e.g., markers 160-180). Bufo terrestris called from open, beach habitats or in areas where mowed lawns came to the water's edge (e.g., markers 1000, 1400-1450). Adults of Rana grylio showed a highly significant difference ($\chi^2 = 14.80$) among years in Gatlin Canal as did those of Hyla cinerea ($\chi^2 = 6.74$). A distinct increase in relative densities, especially in SY3, made each year statistically different for R. grylio but only made SY1 different from SY3 for H. cinerea.

128. Trip analysis. The temporal distributions and abundances of amphibians and reptiles captured in funnel traps in Gatlin Canal are shown in Figure 60. Most captures occurred between May and November of SY1 and SY2, with peaks in June, July, and August of both years and in November of 1978. Very few animals were taken in SY3 but a peak occurred in September 1979. For an unknown reason no animals were captured from November of 1977 through February of 1978. A few were captured during these months in SY2 and SY3. Amphiuma means was active throughout the year and Nerodia cyclopion was recorded from June through November.

129. The distributions of salamanders and reptiles observed on strip-patrols in Gatlin Canal during the study are shown in Figure 61. The density of turtles, especially Sternotherus odoratus, was high early in the study but gradually decreased through September of 1977. After the first year, peaks of turtle activity occurred in February and August 1978 and June of each year. No animals were observed in Gatlin Canal on three trips in February of 1979.

130. The temporal distributions of calling frogs in Gatlin Canal are shown in Figure 62. A clear trend towards increasing activity and density from SY1 to SY3 is evident with the frogs. The bulk of the activity was concentrated during July and August of SY1 and SY2 but began earlier (May) and extended through August in SY3. Most of the summer activity was made up of Hyla cinerea, Bufo terrestris, and Rana grylio. Bufo terrestris showed distinct patterns in each year. Calls were recorded calling in April and May of 1978 and 1979 and again in August of 1978, and in July of 1979. Rana tricolor called primarily from November through March but a few individuals were calling during the warmer months from June through September. The other frog species

from Catlin Canal generally called during the summer months.

Yearly Comparisons - Lake Conway

131. A comparison (Table 8) of the numbers of individuals and species recorded in all of Lake Conway with the permanent sites only (all pools combined) shows an overall decrease of six species (shared each year) from FY1 to FY3 in both the total lake and the permanent sites. The permanent site sample for the three years combined ($N=7,754$) was only 63.0% of the total lake sample. Yet, only one species recorded from the total lake was not also found on the permanent sites. Four adults of that species Hyla femoralis were calling on one night in FY1 in Middle Pool. Thus, the permanent sites and their samples adequately represented the Lake Conway environment and its herpetofauna. The decrease in individuals between FY1 and FY2 was 41.3% but the decrease in species was only 11.1%. In FY3, the individuals sampled were 21.7% less than in FY2 while the species number decreased by 12.5%. Similar trends occurred on the permanent sites (Table 8). From FY1 to FY2, the sample of individuals decreased 55.1% but species number went down only 11.1%. In FY3, the sample was only down 9.0% while the number of species sampled decreased 14.0% from the FY2 value. The sampling effort was generally comparable among the three years (see Methods section) or showed slight increases (funnel traps, Table A4, Bandcroft et al. 1980). Yet, densities of most species decreased and diversity went down during the three-year study. Most of the decrease in densities occurred between FY1 and FY2. For the total lake sample, the FY2 sample predicted (based on FY1 curve in Figure 2 and data in Table 8) 26 species but only 24 were recorded. In FY3, the predicted species number was 26 but only 21 were recorded. A similar pattern for the permanent sites existed. The FY1 sample predicted 26 and 26 species were recorded. In FY2, the predicted number of 22 was one less than the recorded number of 23 species. In FY3, however, the predicted species number was 22 but only 20 were recorded. Species numbers in the total lake were below the expected in FY2 and FY3, but on the permanent sites, only below the expected in FY3.

132. Ten species showed significant yearly changes in relative densities

in the lake system: Amphiuma means, Siren lacertina, Hyla cinerea, Bombina gryllus, Bombina atricolor, Polydora serpentina, Ambystoma subdubium, Desmognathus alternatus, Desmognathus floridana, and Nerodia cyclopion. Two salamanders, larvae of two frogs, three turtles, and one snake showed significant changes as detected by funnel trap analysis. The turtle showed significant differences in relative densities with the herp-control and funnel trap analyses. The frog and one turtle showed significant differences on herp-control.

1971. The two larger aquatic salamanders were taken primarily in funnel traps and had highly significant differences among years (Amphiuma means, $\chi^2 = 20.14$; Siren lacertina, $\chi^2 = 24.04$) with a significant decrease in relative density in most pools through the study. Larvae of Hyla cinerea and those of Bombina atricolor showed highly significant changes in relative densities through the study. The densities of Hyla cinerea larvae decreased and were highly significant among years ($\chi^2 = 21.02$) with densities in SY1 statistically different from those in SY2 and SY3 which were the same. Differences in relative densities of Bombina atricolor larvae were highly significant among years ($\chi^2 = 21.02$) and showed a peak in SY2 and a large decrease in SY3. The relative densities of the turtle Chelydra serpentina were significant among years ($\chi^2 = 21.02$) with a slight increase in funnel trap captures in SY2. The population of Ambystoma subdubium increased in relative density with SY1 values statistically different from SY2 and SY3 values which were not distinct from each other. For this turtle, a highly significant difference among years ($\chi^2 = 21.02$) was shown by the funnel trap analysis. Populations of Nerodia cyclopion decreased sharply between SY1 and SY2 but with relative densities of SY2 and SY3 showing no significant difference. Overall relative densities were highly significant ($\chi^2 = 21.02$) among years for this species. Desmognathus alternatus was the most common species on the lake and showed highly significant differences among years for the funnel trap analysis ($\chi^2 = 21.02$) and highly significant differences among years for the herp-control analysis ($\chi^2 = 21.02$) with the relative densities in SY1 statistically different from those of SY2 and SY3 which were very similar to each other. The populations of Bombina gryllus increased in relative densities with those of SY2 significantly higher than those of SY1 or SY3 which were the same. A highly significant difference among

years ($\chi^2 = 11.53$) was shown by the breeding season analysis of horn-patrol data. The populations of Pseudemys floridana decreased in density with values for SY1 being statistically higher than those for SY2 and SY3 which were the same. The relative densities of this turtle from horn-patrol data were significantly different among years ($\chi^2 = 8.06$). No other species showed significant differences among years.

Deepwater Trapping Stations

134. A single salamander, Siren lacertina, was collected in 244 trap-days at deepwater sampling sites during SY1 in 1.2 m of water on the East Pool site during the July 1978 sampling period (Godley et al. 1981). The mean trap success of amphibians and reptiles at deepwater sites was 0.41 individuals/100 trap-days. This value was 40.92 times lower than the mean for all amphibians and reptiles ($\bar{X} = 16.77$ individuals/100 trap-days) and 5.72 times lower than the mean for S. lacertina ($\bar{X} = 2.34$ individuals/100 trap-days) trapped at permanent shoreline sites during the same period. Deepwater trapping was discontinued in SY2 and SY3.

PART V: DISCUSSION AND CONCLUSIONS

135. The herpetofauna of Lake Conway is a diverse and complex assemblage of amphibians and reptiles. Some species by virtue of their size or abundance are important, integral components of the ecosystem and play primary roles in the food webs and community dynamics of the lake. Other species, by virtue of their low densities or marginal occurrences, probably are less important components of the systems. The functional roles of these species are not well understood and remain to be defined. Included in the category of primary species are the salamanders Amphiuma means and Pleurodeles latissima; the frogs Acris gryllus, Hyla cinerea, Rana gryllus, and Rana picicollaris; the alligator Alligator mississippiensis; the turtles Chelydra serpentina, Pseudemys subtridens, Pseudemys floridana, Pseudemys nelsoni, and Strognomys carolinensis; and the snake Nerodia cyclopion.

136. An examination and understanding of the population dynamics and ecology of these 13 species during the three-year study adequately illustrate the major trends that characterize the amphibian and reptile populations on Lake Conway. In addition, the data gathered during the study provide a much better understanding of the functional role that each species have in the overall lake system. With these new insights we have been able to interpret the observed changes in density and composition of certain species with some degree of reliability and, in specific instances, to attribute these changes to the complex of environmental factors acting on the system including the impact of the white shrimp as a plant control agent.

137. The selection of Lake Conway for the LSW was unfortunate from the viewpoint of the amphibian and reptile populations because of confounding influences of human disturbance on the system. The destruction and loss of both terrestrial and aquatic littoral zone habitats for housing and beach development have had a significant negative effect on all species of amphibians and reptiles on Lake Conway. In certain instances, significant portions of populations were eliminated by land clearing and site preparation practices. In other instances, species were affected indirectly by loss of suitable foraging habitat, calling sites, nesting areas, and food resources. High mortality in snake populations as a result of direct human activity and in

populations of larger turtles from excessive boat traffic has been a significant factor influencing the population dynamics of these species. The impact of various kinds of human disturbance, both during and prior to our work, makes it difficult to isolate specific causes responsible for the population declines documented for many species on the lake. However, reductions in the density of certain open-water species clearly are correlated with decreases in aquatic plant biomass as a result of the feeding activity of the white amur. These declines apparently are not confounded by human disturbance factors. Table 4 summarizes density responses of the major groups of the 14 primary species of amphibians and reptiles and attempts to identify and rank the major causative agents affecting their densities. Unfortunately, comparative data from other detailed integrative studies of a community of amphibians and reptiles in an aquatic environment comparable to Lake Dorway are unavailable.

138. Populations of the two large salamanders decreased significantly through the study. Habitat loss through human disturbance resulted in major decreases in two pools but does not explain the concomitant declines in the other three pools. A significant change in the density of open-water populations of Siren lacertina occurred (data presented in Bancroft et al. 1983) as extensive areas of Potamogeton illinoensis were reduced or eliminated by white amur feeding. Mortality from boat traffic also was documented for Siren lacertina. Both of these factors were considered minor compared with the impact of habitat loss on Amphiuma means and S. lacertina. The facts that individuals of Siren were never recaptured in funnel traps and that recaptures of Amphiuma decreased through the study suggest that both species may have become trap-shy and, therefore, less subject to detection. Also, the documented decrease in water level of the lake since 1977 may have resulted in primary littoral zone habitat going dry and eventually resulting in a decline of salamander populations through reduction and loss of nesting sites. Finally, Hurricane David may have contributed to the decline of salamander densities in West Pool where considerable habitat disruption occurred. Relatively few animals compared to the previous year were recorded in West Pool for several months following the hurricane.

139. Frogs showed an overall increase in density through the study with

documented decreases in those pools where considerable habitat destruction occurred (e.g., South and Middle Pools). In addition, water levels recorded by the U. S. Geological Survey were lower during the 1971-72 season at any time during the previous 15 years. During our study, water levels were lowest in 1971, intermediate in 1972, and highest in 1973. The low water level in 1971 left some primary habitats used by Spizella monticola dry and may have resulted in decreased activity and, hence, a lower perceived density. We also suspect that the amount of emergent lower or littoral zone vegetation, especially Hydrilla verticillata, at West Pool may have indirectly affected the calling activity in 1972 when water levels were lower than expected in 1971. The only other explanation was a population change activity and density the result of habitat loss through destruction. The general decrease in turtle activity during the study may be the result of decreased availability of food resources. If water levels during the period the activity was low were low, then the decrease in activity may be the result of decreased food availability. During the study, food resources were low and resulted in turtles being observed in areas where they normally did not feed. For example, very little movement was observed in the littoral zone by turtles and activity was restricted to the emergent zone. The decrease in activity may have been the result of decreased food availability and their inability to move to areas where food was abundant and, therefore, food was not available. The decrease in activity may have been the result of decreased food availability and their inability to move to areas where food was abundant and, therefore, food was not available. The decrease in activity may have been the result of decreased food availability and their inability to move to areas where food was abundant and, therefore, food was not available.

14. The decrease in collection of Lake Hovwy in the period of study disturbance either through direct predation or by loss of nest sites through habitat destruction. Removal of a large, large female from the population may have a dramatic effect on the population in a pool for several years. The same is true when nesting areas, which were in very limited supply, were destroyed. Examples of both were recorded during the study. Mortality from boat traffic was documented but probably was of minor importance compared to the other factors.

15. Losses of turtles on Lake Conaway were dramatic and confounded by several factors. Destruction of littoral zone habitat clearly affected turtle

densities in some areas of Lake Conway. More difficult to document, but certainly important to turtle populations, was the loss of suitable nesting sites around the lake. As the natural habitat was converted to residential development and much of the area planted in lawns, suitable nesting sites decreased. Whether conversion of some areas to white sand beach has countered the loss of natural nesting sites is unknown. It is known that turtles have a difficult time digging a nest hole in lawns. Our data clearly show that the major mortality factor for adult turtles on Lake Conway is the heavy boat traffic. Most adults of Pseudemys floridana have extensive scars from being hit by boat propellers. As boat traffic increases, so will this unnatural cause of mortality. We suspect that our use of Floy tags to mark Sternotherus odoratus also may have contributed to their decline. In some instances the Floy tags on individuals of S. odoratus became entangled in filamentous algae and resulted in the turtle drowning. At this time, we do not know the extent of the mortality implied by this observation but anticipate a thorough analysis of recapture data to evaluate its impact. Apparently, filamentous algae increased in Lake Conway during the study.

142. Turtles, probably more than any other group of amphibians or reptiles, are dependent on aquatic macrophyte productivity and most susceptible to its removal. Two common herbivorous species (Pseudemys floridana and P. nelsoni) are direct competitors with the white amur for aquatic plants and another species (Sternotherus odoratus) feeds primarily on snails that use the macrophytes as a substratum. We have documented pronounced shifts and reductions in density of these species in areas where the white amur has reduced or eliminated cover of particular plant species. Analyses indicate that P. floridana and S. odoratus have shifted their habitat use primarily from the littoral zone and Potamogeton beds to beds of Nitella, Vallisneria, and over bare bottoms as the composition of the aquatic macrophyte community changed from white amur feeding activity. These results are discussed in greater detail in Bamcroft et al. (1987).

143. Increased mortality had a dramatic effect on snake densities, particularly of Lerista cyclopion. Fragmentation of terrestrial sites for housing, particularly during the winter months, decimated the South Carol populations of L. cyclopion that were hibernating the colder season in burrows of

terrestrial mammals. This loss, combined with human predation in all pools and other predation of trapped individuals in East and West Pools, accounted for the overall decrease in snake densities during the study.

144. In those situations where the white amur was shown to have contributed to the decline in population densities or changes in activity and habitat use by certain species, especially turtles, the low stocking rate, continued presence of aquatic macrophytes, and their subsequent increase in SY4 has ameliorated the long-term effects on the turtle and salamander populations in the lake. If additional stocking or an higher initial stocking rate had occurred, the long-term negative impact on some species of amphibians and reptiles would have been more severe. Under these conditions some species probably would have disappeared from the system.

[illegible]

- 62

Table 1

Checklist of Amphibians and Reptiles Known from the Lake Conway System

<u>Scientific Name</u>	<u>Common Name</u>	<u>Species Code</u> *
AMPHIBIA		
CAUDATA		
AMPHIUMIDAE		
<u>Amphiuma means</u>	Two-toed amphiuma	A, 1
PLETHODONTIDAE		
<u>Eurycea quadridigitata</u>	Dwarf salamander	
SIRENIDAE		
<u>Pseudobranchius striatus</u>	Dwarf siren	S
<u>Siren lacertina</u>	Greater siren	L
ANURA		
BUFONIDAE		
<u>Bufo terrestris</u>	Southern toad	F, 7
HYLIDAE		
<u>Acris gryllus</u>	Florida cricket frog	Y, ★
<u>Hyla cinerea</u>	Green treefrog	H, \$
<u>Hyla femoralis</u>	Pinewoods treefrog	M
<u>Hyla squirella</u>	Squirrel treefrog	P
MICROHYLIDAE		
<u>Gastrophryne carolinensis</u>	Eastern narrow-mouthed toad	G, 2
RANIDAE		
<u>Rana grylio</u>	Pig frog	R, +
<u>Rana utricularia</u>	Southern leopard frog	U, #, &

(Continued)

* If applicable, the code for adults of each species is followed by that for egg and larval stages. These codes are used in Figures 3 through 62.

Table 1 (Concluded)

REPTILIA

CROCODILIA

CROCODILIDAE

<u>Alligator mississippiensis</u>	American alligator	E
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TESTUDINATA

CHELYDRIDAE

<u>Chelydra serpentina</u>	Florida snapping turtle	O
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EMYDIDAE

<u>Deirochelys reticularia</u>	Chicken turtle	D
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<u>Pseudemys floridana</u>	Peninsular cooter	F
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<u>Pseudemys nelsoni</u>	Florida red-bellied turtle	C
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<u>Pseudemys scripta</u> **	Red-eared turtle	S
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KINOSTERNIDAE

<u>Kinosternon baurii</u>	Striped mud turtle	I
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<u>Kinosternon subrubrum</u>	Eastern mud turtle	K
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<u>Sternotherus odoratus</u>	Stinkpot	S, @
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TRIONYCHIDAE

<u>Trionyx ferox</u>	Florida softshell	T, 4
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SQUAMATA

COLUBRIDAE

<u>Coluber constrictor</u>	Black racer	V
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<u>Farancia abacura</u>	Mud snake	X
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<u>Nerodia cyclopion</u>	Green water snake	N
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<u>Nerodia fasciata</u>	Florida water snake	W
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<u>Regina alleni</u>	Striped swamp snake	J
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<u>Thamnophis sauritus</u>	Peninsula ribbon snake	Z
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<u>Thamnophis sirtalis</u>	Eastern garter snake	Q
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Introduced species.

Table 2

Surface Area, Shoreline, Number of Species, and Number of Individuals Recorded from Each Pool and the Entire Lake Conway System During the Three-Year Study*

	<u>South</u> <u>Pool</u>	<u>Middle</u> <u>Pool</u>	<u>East</u> <u>Pool</u>	<u>West</u> <u>Pool</u>	<u>Lake</u> <u>Gatlin</u>	<u>Lake</u> <u>Conway</u>
Total pool						
Surface area, ha	142.6	301.0	128.4	144.4	26.6	743.0
Shoreline, km	5.58	10.73	9.78	5.25	3.33	34.67
Number of species	23	20	20	17	20	29
Number of individuals	3364	2084	2471	2146	1863	11928
Permanent sites						
Shoreline, m	530	200	200	370	470	1770
Transect length						
HP, m	1060	800	400	740	940	3940
FT, m	460	200	490	370	200	1720
Time						
HP, hr	64.0	41.5	29.6	36.9	61.8	233.8
FT, days	3213	1359	2010	2255	1317	10154
Number of species	22	19	17	15	20	28
Number of individuals	1901	1309	1536	1621	1387	7754

* The transect lengths and sampling time for herp-patrol (HP) and funnel traps (FT) at each permanent site are shown. See Part III for additional explanation.

Table 3

Number of Individuals and Number of Species Recorded from All of South Pool and
from the South Pool Permanent Site Only During Each Year of the Study

<u>Study Year</u>	<u>Total Pool</u>		<u>Permanent Site</u>	
	<u>Individuals</u>	<u>Species</u>	<u>Individuals</u>	<u>Species</u>
SY1	1396	21	1221	20
SY2	1019	21	381	20
SY3	949	17	299	14
Total	3364	20	1901	22

Table 4

Number of Individuals and Number of Species Recorded from All of Middle Pool
and from the Middle Pool Permanent Site Only During Each Year of the Study

<u>Study Year</u>	<u>Total Pool</u>		<u>Permanent Site</u>	
	<u>Individuals</u>	<u>Species</u>	<u>Individuals</u>	<u>Species</u>
SY1	1212	19	755	18
SY2	446	16	248	14
SY3	424	16	306	11
Total	2084	20	1309	19

Table 5

Number of Individuals and Number of Species Recorded from All of East Pool and
from the East Pool Permanent Site Only During Each Year of the Study

<u>Study Year</u>	<u>Total Pool</u>		<u>Permanent Site</u>	
	<u>Individuals</u>	<u>Species</u>	<u>Individuals</u>	<u>Species</u>
SY1	1352	17	800	15
SY2	480	14	304	13
SY3	639	17	432	15
Total	2471	20	1536	17

Table 6

Number of Individuals and Number of Species Recorded from All of West Pool and
from the West Pool Permanent Site Only During Each Year of the Study

<u>Study Year</u>	<u>Total Pool</u>		<u>Permanent Site</u>	
	<u>Individuals</u>	<u>Species</u>	<u>Individuals</u>	<u>Species</u>
SY1	887	14	822	12
SY2	813	17	556	13
SY3	446	12	243	9
Total	2146	17	1621	15

Table 7

Number of Individuals and Number of Species Recorded from All of Lake Gatlin
and from the Gatlin Canal Permanent Site Only During Each Year of the Study

<u>Study Year</u>	<u>Total Pool</u>		<u>Permanent Site</u>	
	<u>Individuals</u>	<u>Species</u>	<u>Individuals</u>	<u>Species</u>
SY1	948	19	585	18
SY2	497	19	389	19
SY3	418	16	413	16
Total	1863	20	1387	20

Table 8

Number of Individuals and Number of Species Recorded from All Pools and All
Permanent Sites on Lake Conway During Each Year of the Study

<u>Study Year</u>	<u>Total Pools</u>		<u>Permanent Sites</u>	
	<u>Individuals</u>	<u>Species</u>	<u>Individuals</u>	<u>Species</u>
SY1	5745	27	4183	26
SY2	3257	24	1878	23
SY3	2876	21	1693	20
Total	11,928	29	7,754	28

Table 9

Density Responses of the Primary Herpetofaunal Groups on Lake Conway and the
Major Causative Agents*

	<u>Salamander</u>	<u>Frog</u>	<u>Alligator</u>	<u>Turtle</u>	<u>Snake</u>
Density Response	+	++	+	+	+
Causative agent					
Habitat loss	++	++	++	++	++
White amur	++	S		++	
Human predation			++		++
Boat mortality	+		+	++	
Water fluctuations	S	+	+	S	
Otter predation	+			S	+
Tag mortality	S			S	
Hurricane	S	S			
Trap shyness	S				

* ++ = major factor; + = minor factor; S = suspected factor.

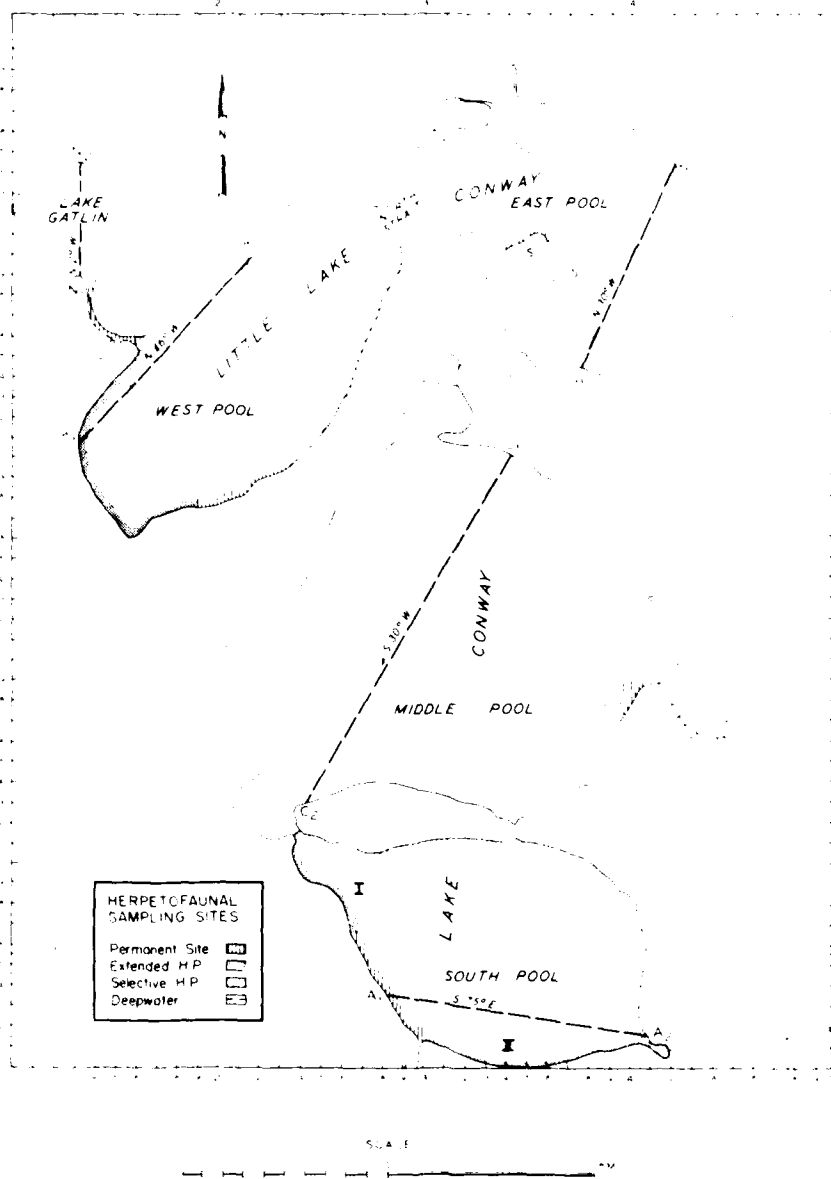


Figure 1. Map showing the five interconnected pools of Lake Conway and the permanent sampling sites for amphibians and reptiles.

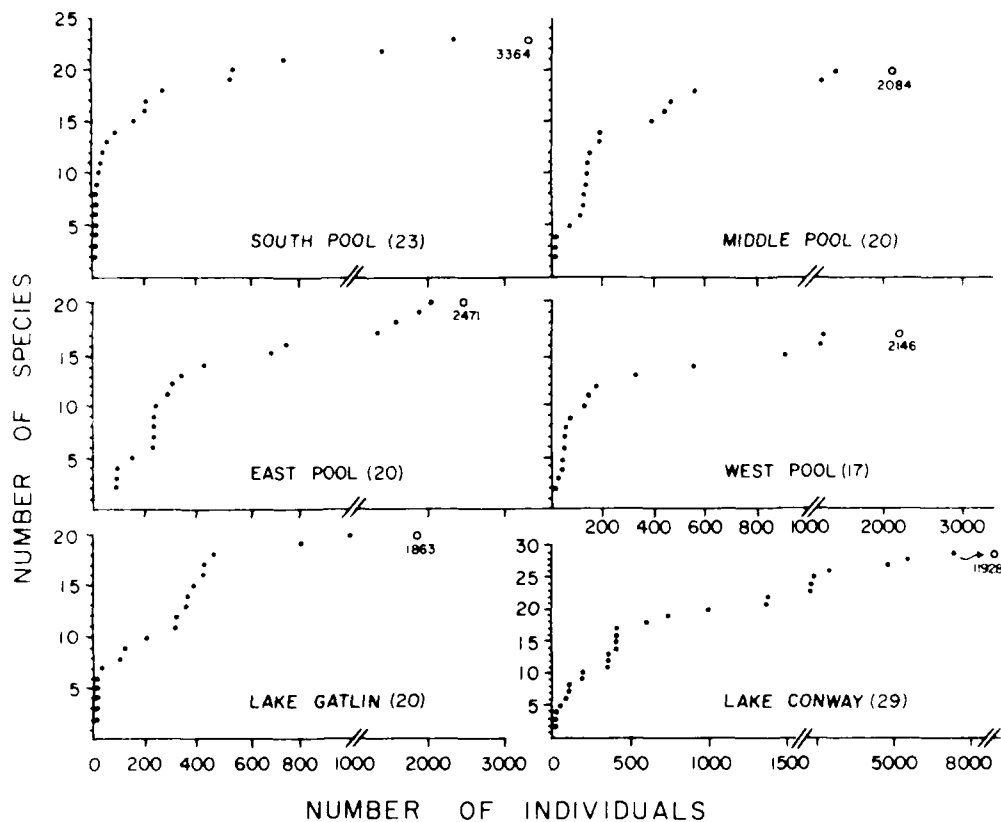


Figure 2. The cumulative number of amphibian and reptile species as a function of the cumulative number of individuals recorded for each pool and for the total of Lake Conway. Number in parentheses equals the number of species recorded. Circle indicates the last individual collected (and sample size) during the three-year study.

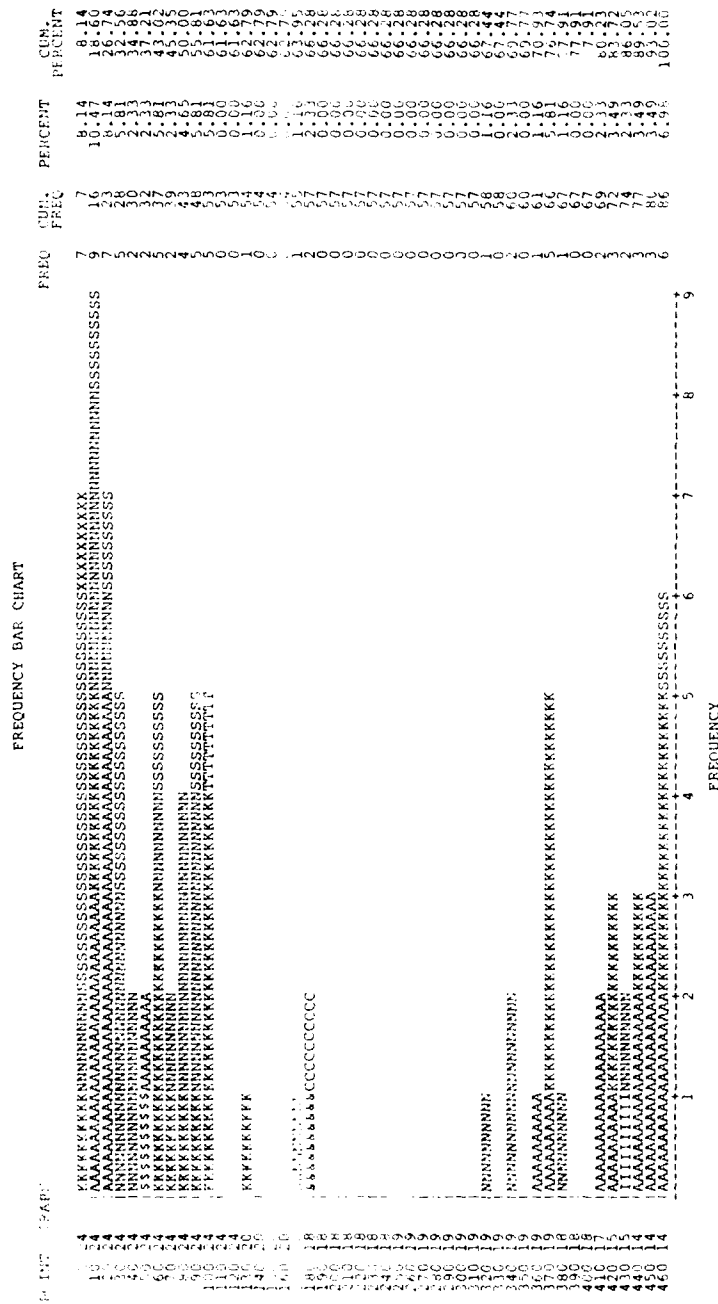
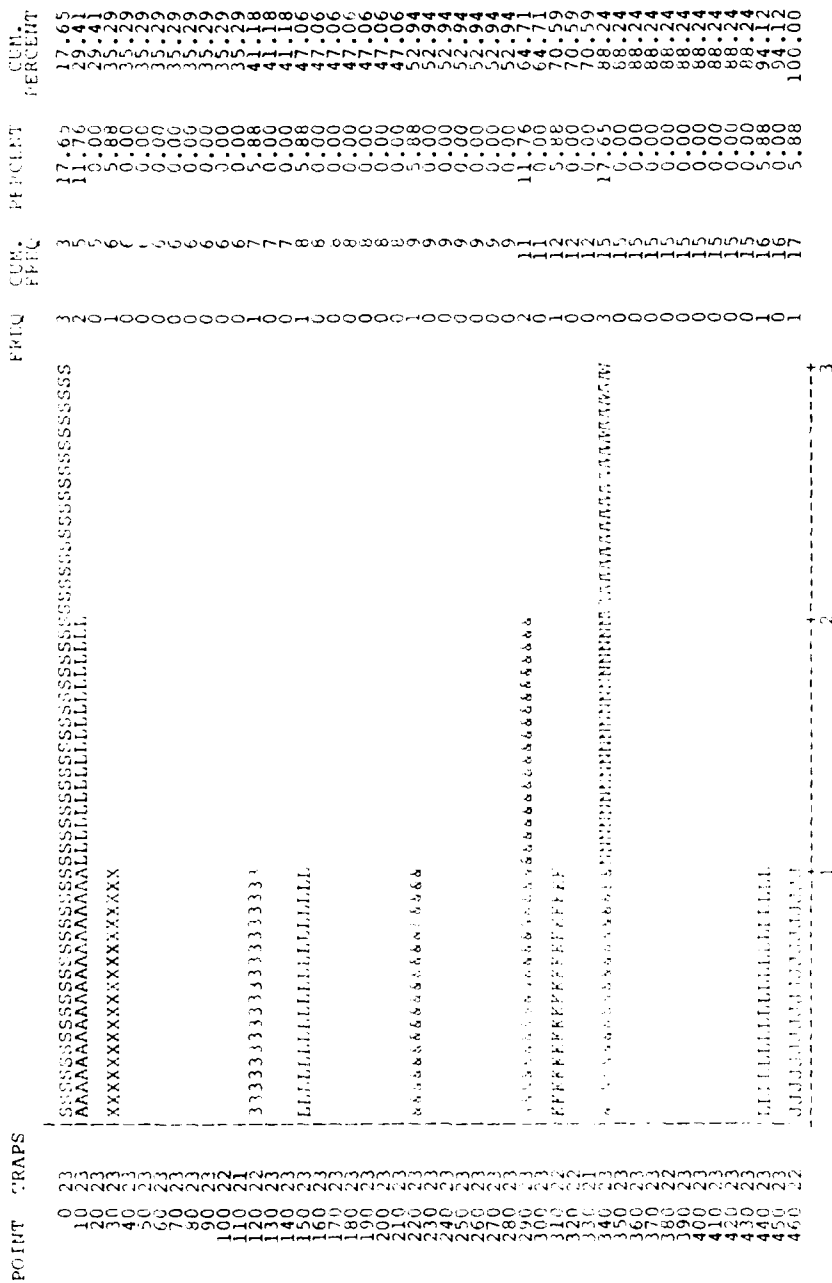


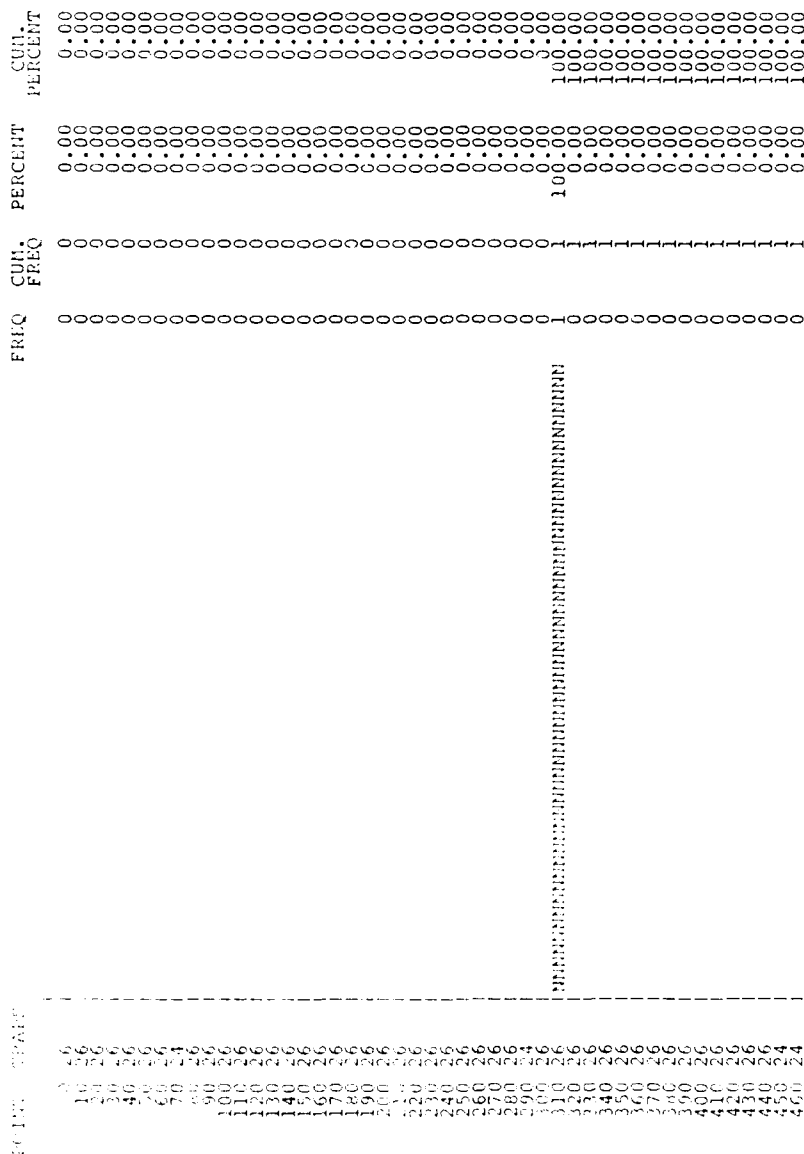
Figure 3. Funnel trap point analysis of amphibians and reptiles on the South Pool site during SY1. Point = location where traps were set; traps = total number of traps set at a sample point. See Table 1 for species codes for this and all subsequent figures.

FREQUENCY BAR CHART



Frequency bar chart showing the distribution of point traps across three categories (1, 2, 3). The y-axis represents the number of point traps (0 to 450), and the x-axis represents the frequency (1 to 3). The bars are labeled with their respective frequencies and cumulative percentages.

FREQUENCY BAR CHART



FREQUENCY

Figure 3. Panel trap point analysis of amphibians and reptiles on the sample road site during day. Point 1 location where traps were set; trap 1 total number of traps set at a single point.

FREQUENCY BAR CHART

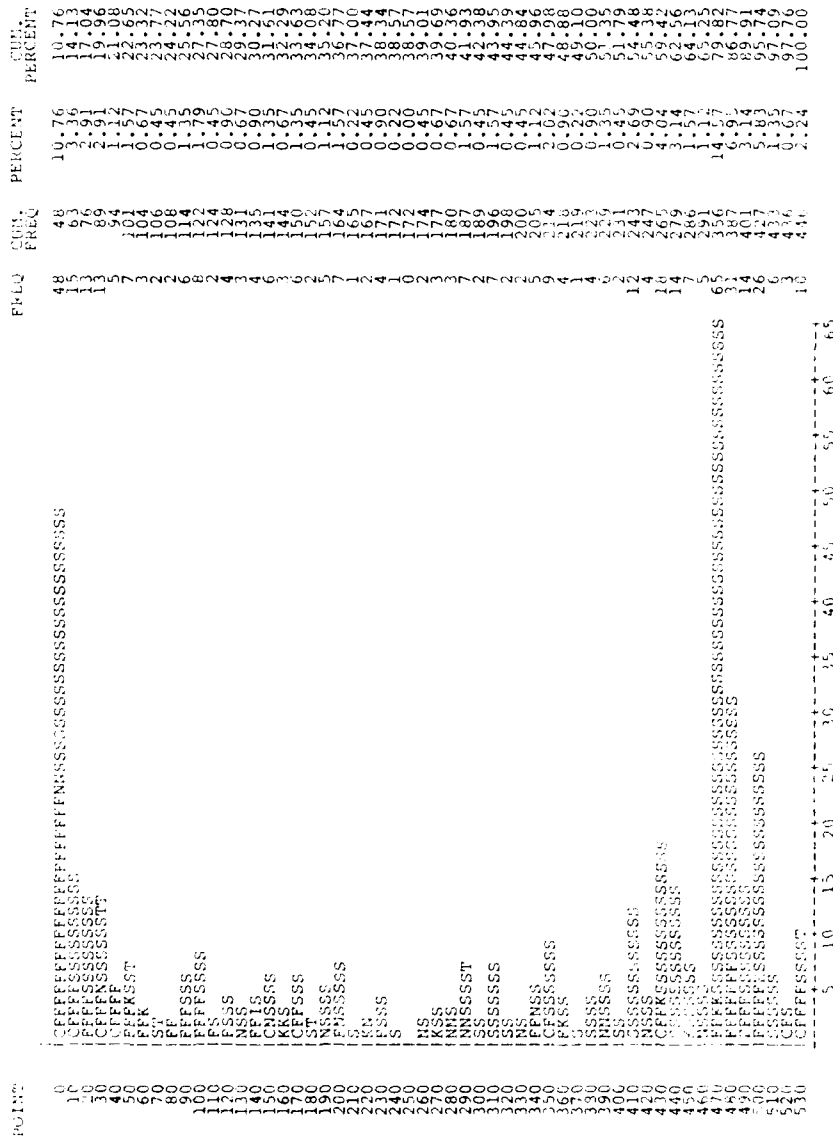


Figure 6. Histogram of point analysis results for the 100-point test. The distribution is approximately normal, with a mean of 50 and a standard deviation of 10. The data was generated using a random number generator.

FREQUENCY BAR CHART

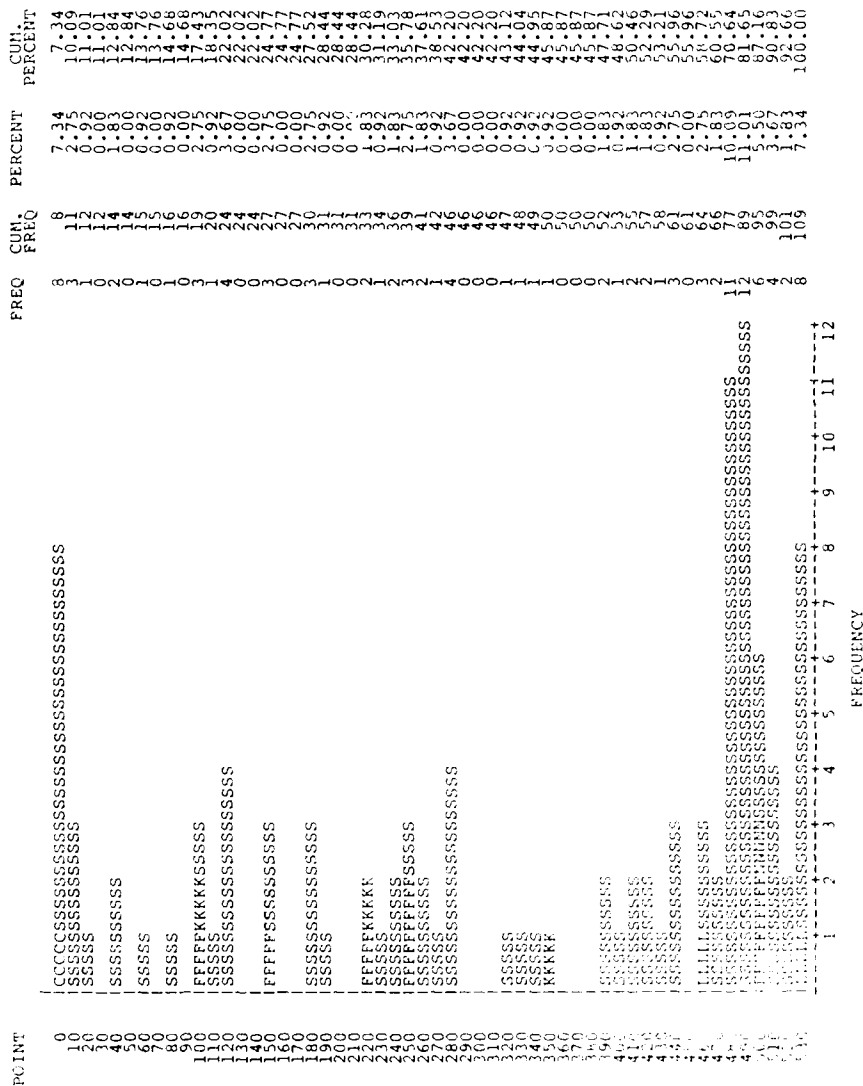


Figure 7. Horizontal point analysis of salamanders and reptiles on the
 central pool site during 1972. Point = midpoint of 10-m section
 of horizontal transect.

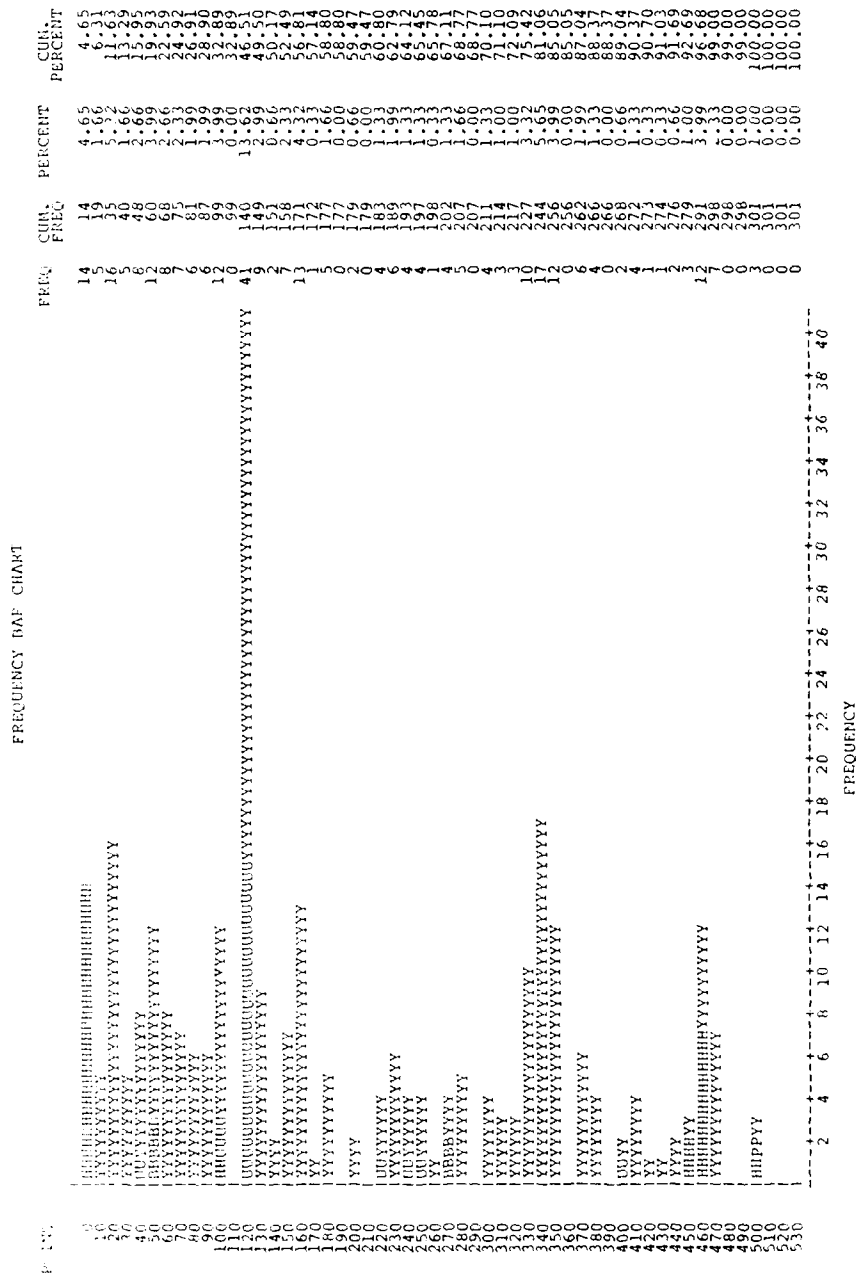


Figure 9. Herp-patrol point analysis of calling frogs on the South Pool site during SYL. Point = midpoint of 10-m section of herp-patrol transect.

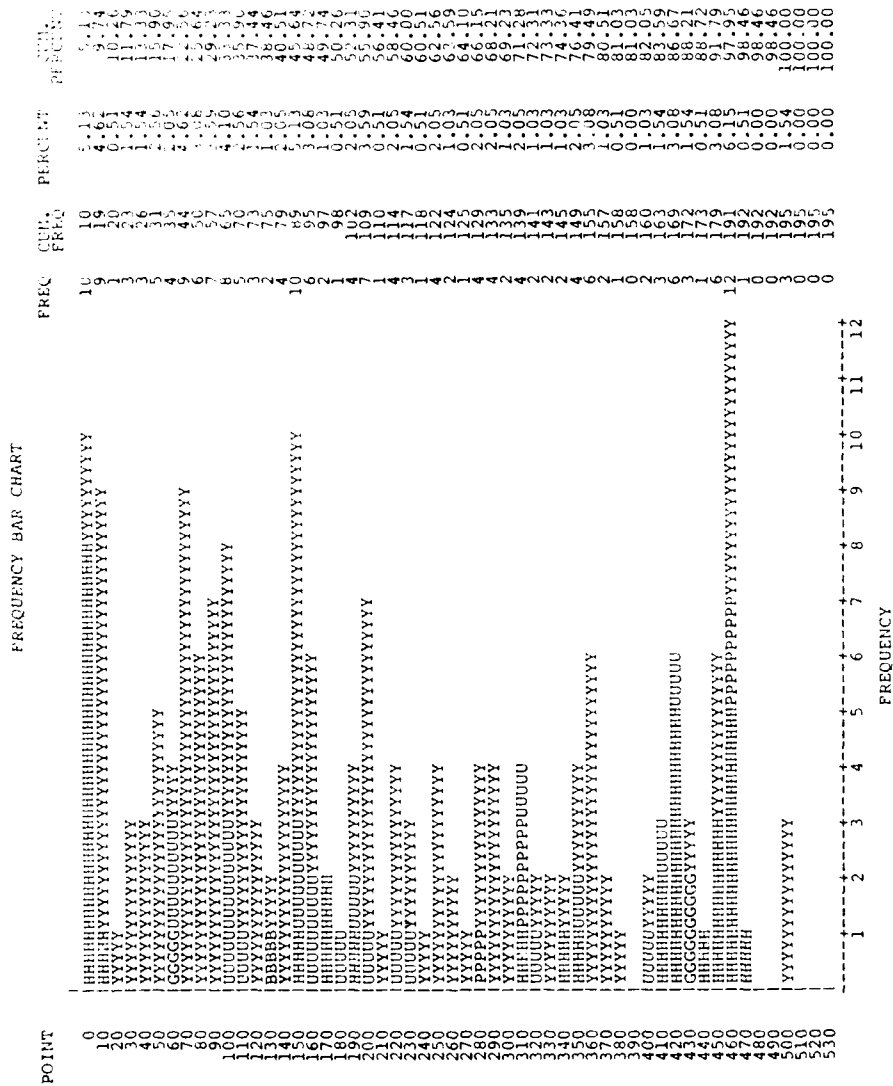


Figure 13. Herp-patrol joint analysis of calling frogs on the South Pond site during SY2. Point - midpoint of 10-m section of herp-patrol transect.

[illegible]

Figure 11. Herp-patrol point analysis of calling frogs on the South Pool site during SY3. Point = midpoint of 10-m section of herp-patrol transect.

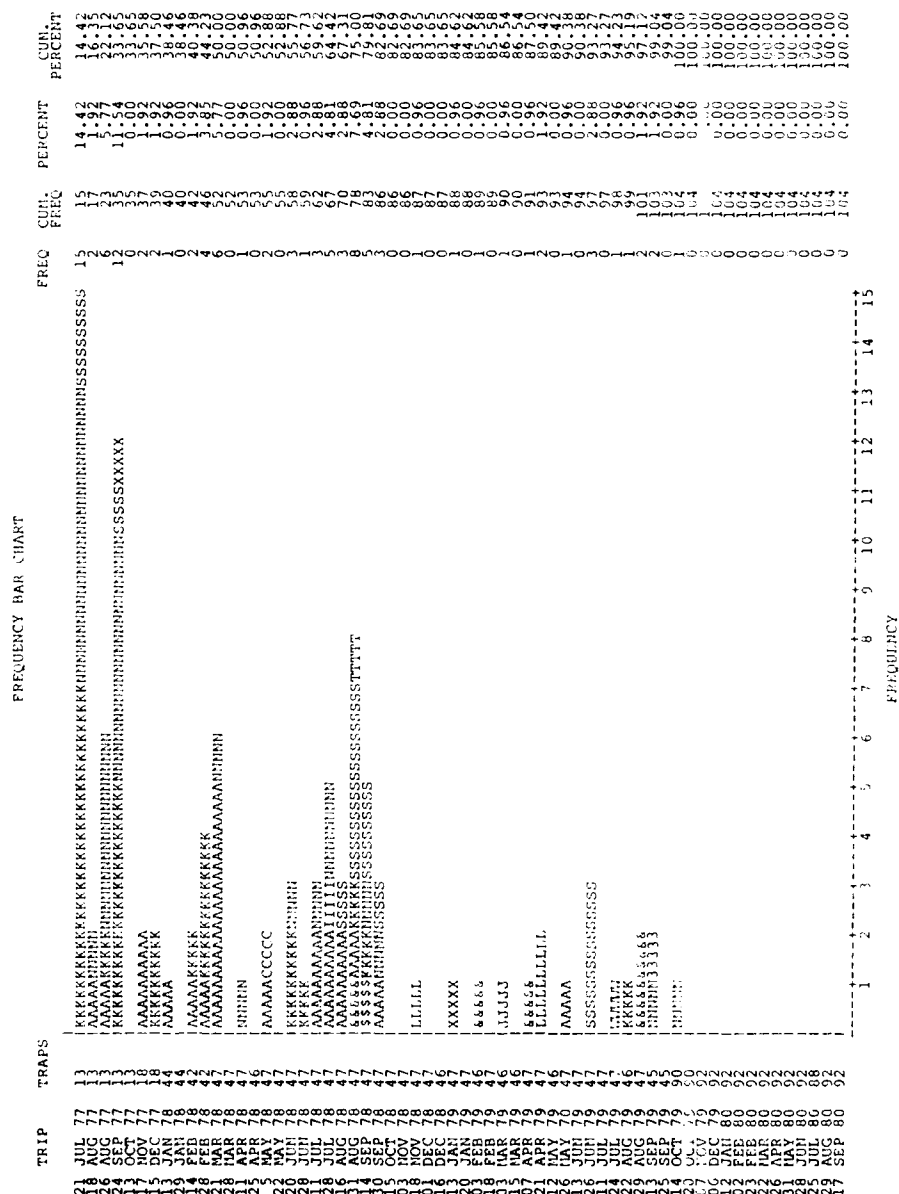


Figure 12. Funnel trap tri analysis of amphibians and reptiles on the South Pool site. Trip = date of trapping; traps = total number of traps set on a date.

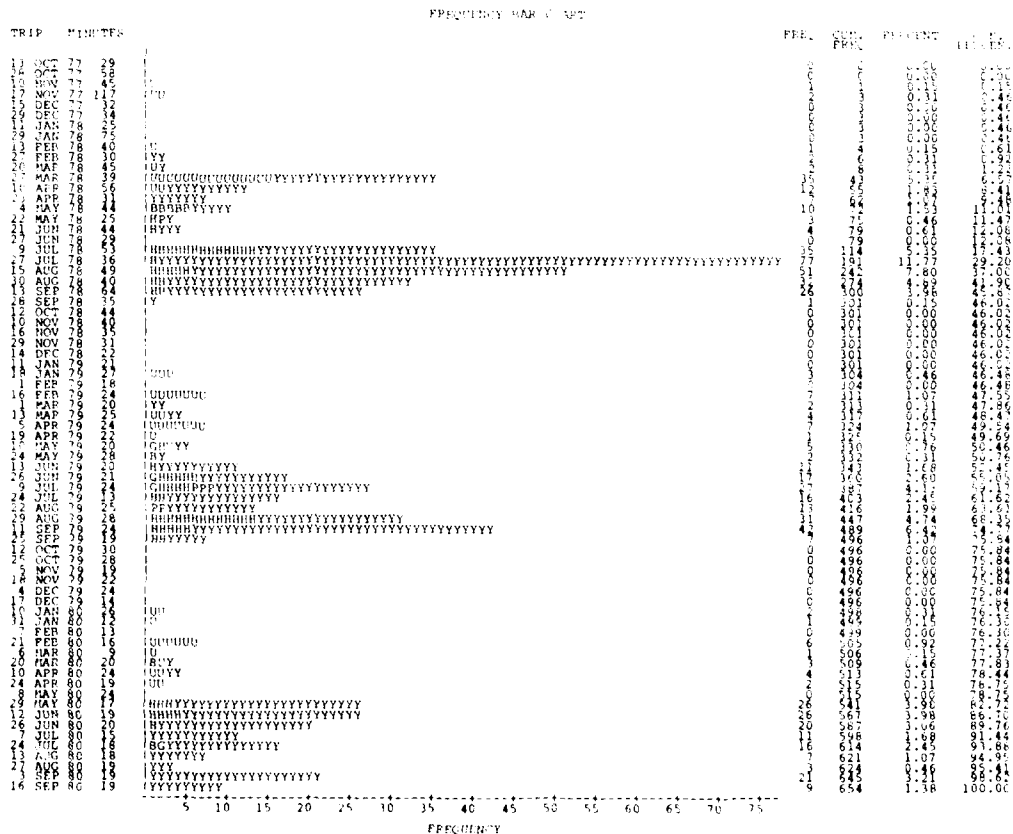


Figure 14. Herp-patrol trip analysis of calling frogs on the South Pool site. Trip = date of herp-patrol; minutes = total sampling time of a herp-patrol on a date (time not recorded prior to 13 October 1977).

FREQUENCY BAR CHART									
POINT	TRAPS					FREQ	CUM. FREQ	PERCENT	CUM. PERCENT
1000	21	KKKK				1	1	2.04	2.04
1010	21	NNNN	SSSS			2	3	4.08	6.12
1020	22	KKKK	SSSS			2	5	4.08	10.20
1030	22					0	5	0.00	10.20
1040	24					1	6	2.04	12.24
1050	24	SSSS				0	6	0.00	12.24
1060	24					0	6	0.00	12.24
1070	24					0	6	0.00	12.24
1080	24					0	6	0.00	12.24
1090	24	++++	LLLL			3	9	6.12	18.36
1100	24	AAAA	LLLL	NNNN		2	11	4.08	22.44
1110	24	AAAA	LLLL	NNNN		2	13	4.08	26.52
1120	24	SSSS	SSSS	NNNN		4	17	8.16	34.68
1130	24	++++	SSSS	NNNN		6	23	12.24	46.92
1140	16	++++	SSSS	NNNN		10	33	20.41	67.35
1150	16	LLLL	NNNN	NNNN		4	37	8.16	75.51
1160	16	AAAA	LLLL	NNNN		2	39	4.08	79.59
1170	16	AAAA	LLLL	NNNN		2	41	4.08	83.67
1180	16	****	AAAA	NNNN		8	49	16.33	100.00

Figure 15. Funnel trap point analysis of amphibians and reptiles on the Middle Island site during 1971. Point = location where traps were set; traps = total number of traps set at a sample point.

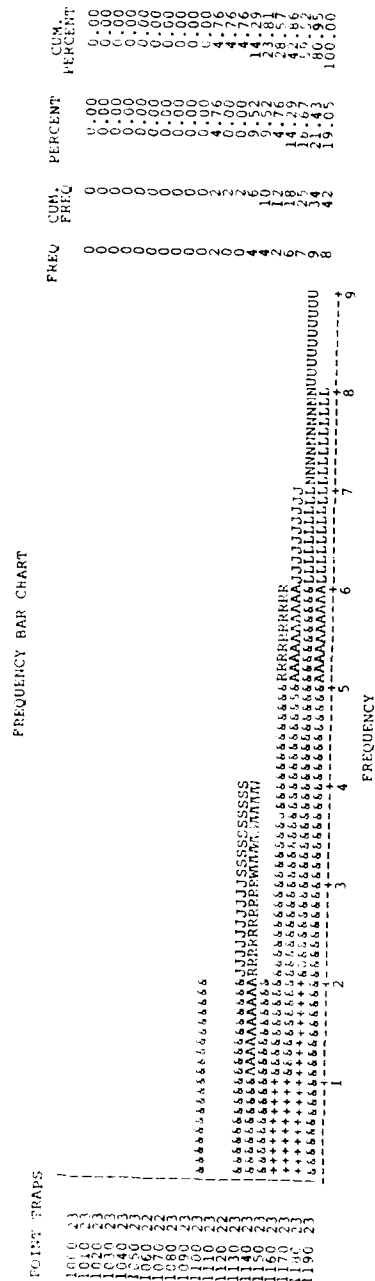


Figure 16. Funnel trap point analysis of amphibians and reptiles on the Middle Fork site during SY2. Point = location where traps were set; traps = total number of traps set at a sample point.

FREQUENCY BAR CHART

POINT	FREQ	CUM. FREQ	PERCENT	CUM. PERCENT
2000	3	3	4.35	4.35
2010	10	4	10.00	14.35
2020	0	4	0.00	14.35
2030	0	4	0.00	14.35
2040	0	4	0.00	14.35
2050	1	5	1.45	15.80
2060	0	5	0.00	15.80
2070	0	5	0.00	15.80
2080	0	5	0.00	15.80
2090	0	5	0.00	15.80
2100	0	5	0.00	15.80
2110	0	5	0.00	15.80
2120	0	5	0.00	15.80
2130	0	5	0.00	15.80
2140	2	7	2.90	18.70
2150	2	9	2.90	21.60
2160	2	11	2.90	24.50
2170	2	13	2.90	27.40
2180	0	13	0.00	27.40
2190	0	13	0.00	27.40
2200	0	13	0.00	27.40
2210	0	13	0.00	27.40
2220	0	13	0.00	27.40
2230	4	17	5.80	33.20
2240	1	18	1.45	34.65
2250	1	19	1.45	36.10
2260	1	20	1.45	37.55
2270	1	21	1.45	39.00
2280	1	22	1.45	40.45
2290	1	23	1.45	41.90
2300	1	24	1.45	43.35
2310	3	27	13.04	56.39
2320	7	34	30.23	86.62
2330	3	37	10.14	96.76
2340	3	40	14.49	100.00
2350	10	50	44.79	100.00
2360	0	50	0.00	100.00
2370	0	50	0.00	100.00
2380	1	51	1.45	101.45
2390	1	52	1.45	102.90
2400	1	53	1.45	104.35
2410	1	54	1.45	105.80
2420	7	61	10.14	115.94
2430	3	64	4.35	120.29
2440	1	65	1.45	121.74
2450	0	65	0.00	121.74
2460	0	65	0.00	121.74
2470	0	65	0.00	121.74
2480	0	65	0.00	121.74
2490	0	65	0.00	121.74
2500	1	66	1.45	123.19

1 2 3 4 5 6 7 8 9 10

FREQUENCY

Bar chart of point analysis of each machine and repetition on the Machine Pool
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and (2000 and 3000) and (2000 and 3000) and (2000 and 3000) and (2000 and 3000)

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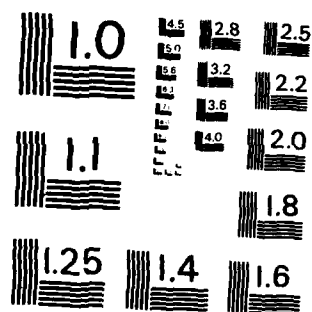
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FREQUENCY BAR CHART

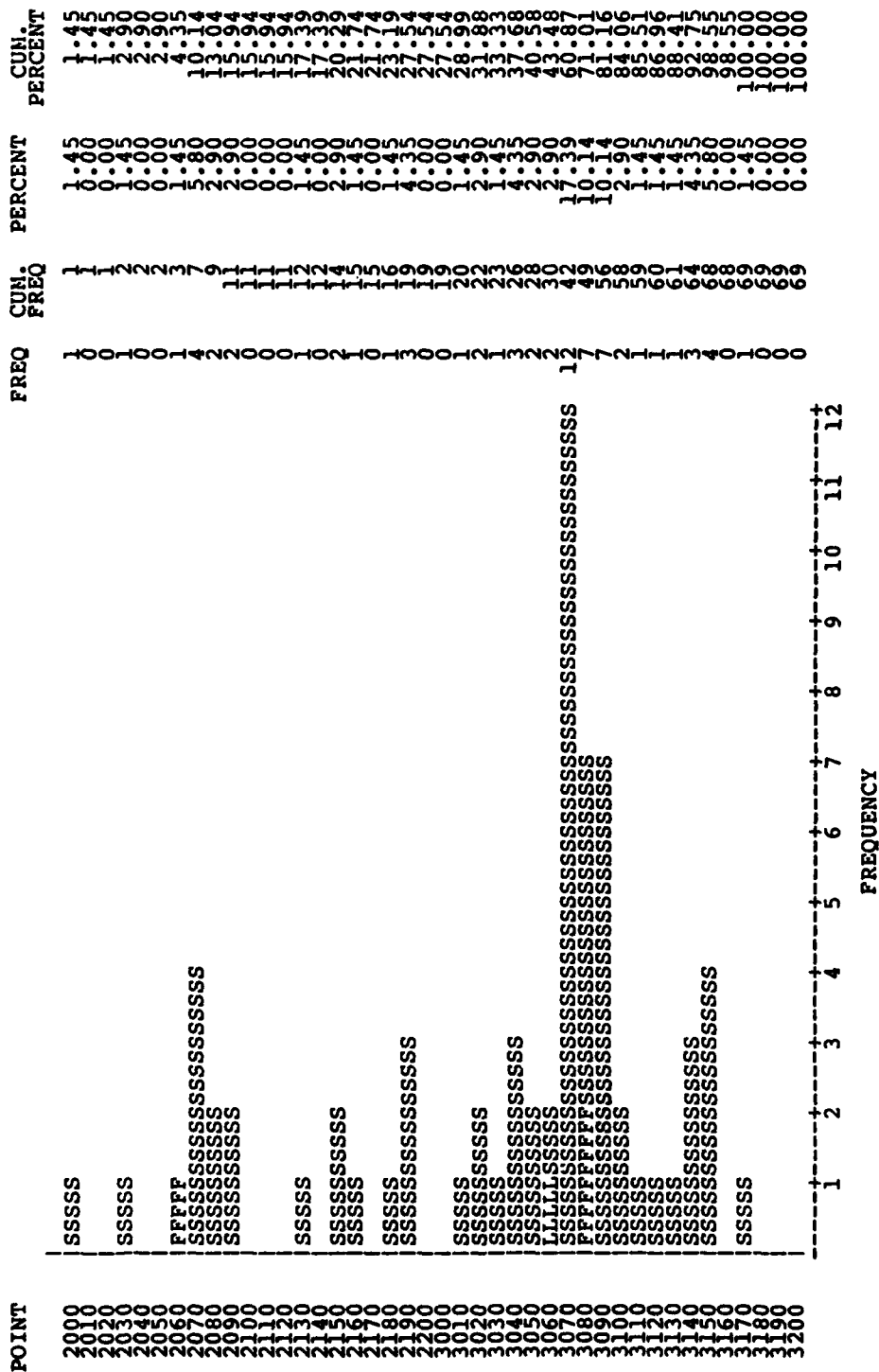


Figure 20. Herp-patrol point analysis of salamanders and reptiles on the Middle Pool site (2000 and 3000 series) during SY3. Point = midpoint of 10-m section of herp-patrol transect.

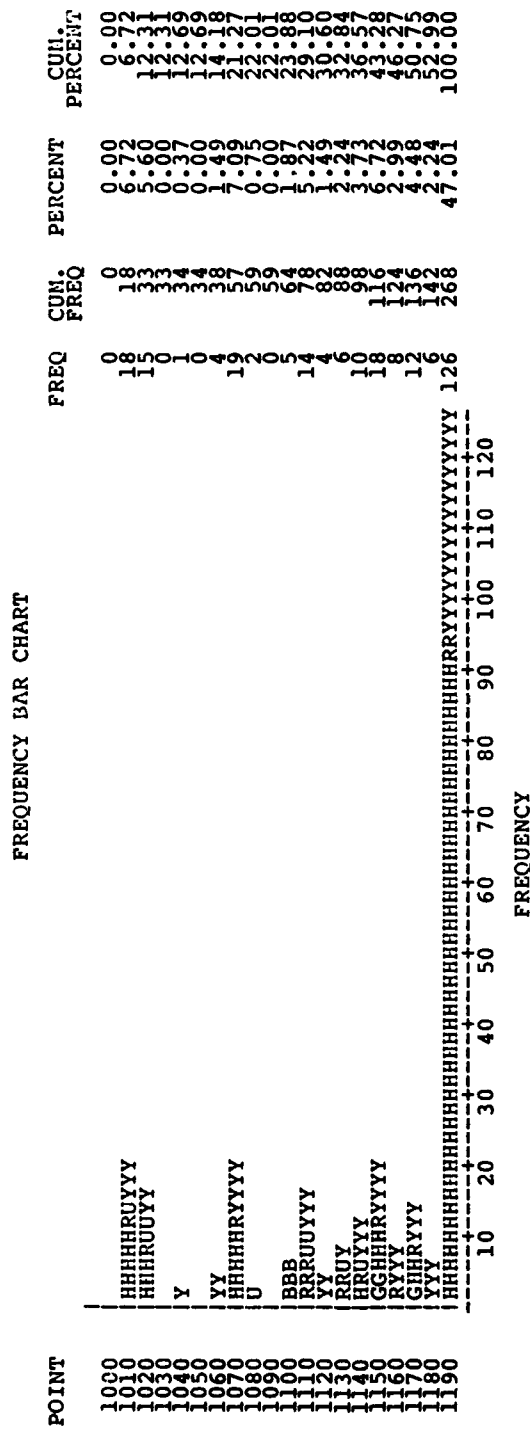


Figure 21. Herp-patrol point analysis of calling frogs on the Middle Pool site during SY1. Point = midpoint of 10-m section of herp-patrol transect.

FREQUENCY BAR CHART

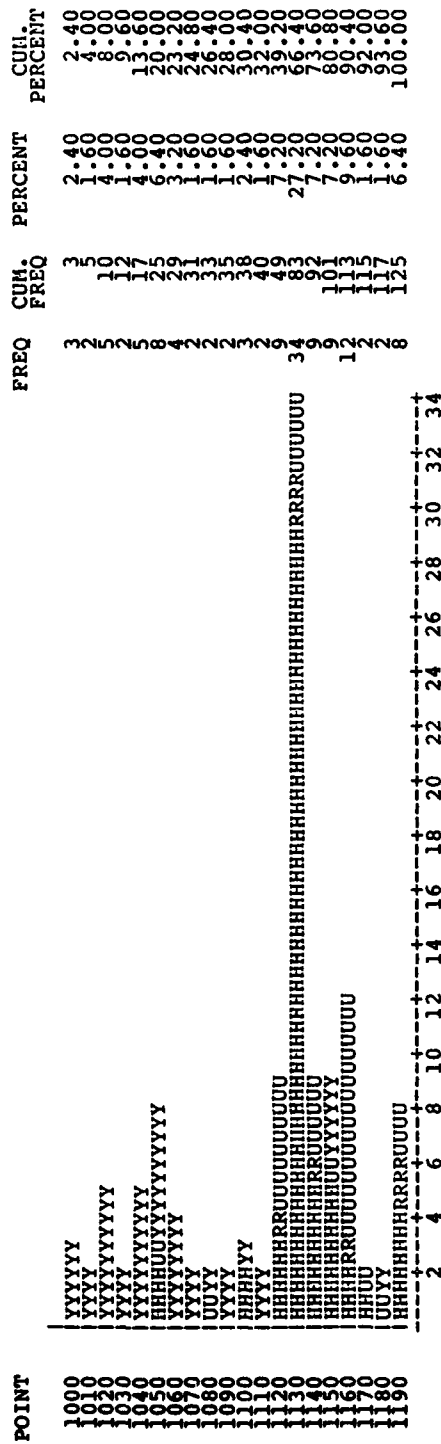


Figure 22. Herp-patrol point analysis of calling frogs on the Middle Pool site during SY2. Point = midpoint of 10-m section of herp-patrol transect.

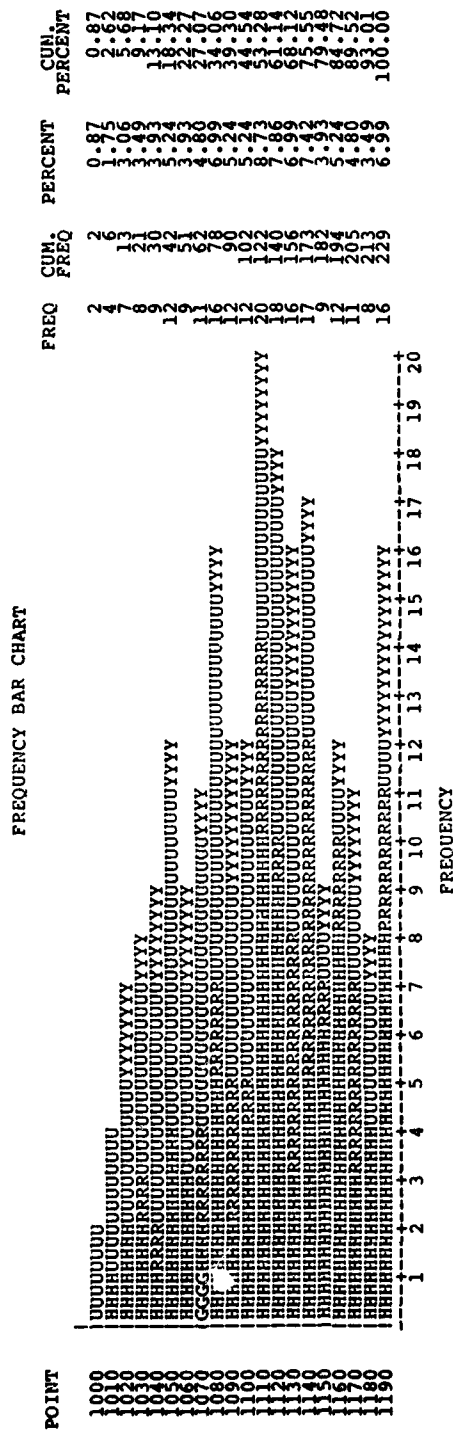


Figure 23. Herp-patrol point analysis of calling frogs on the Middle Pool site during SV3. Point = midpoint of 10-m section of herp-patrol transect.

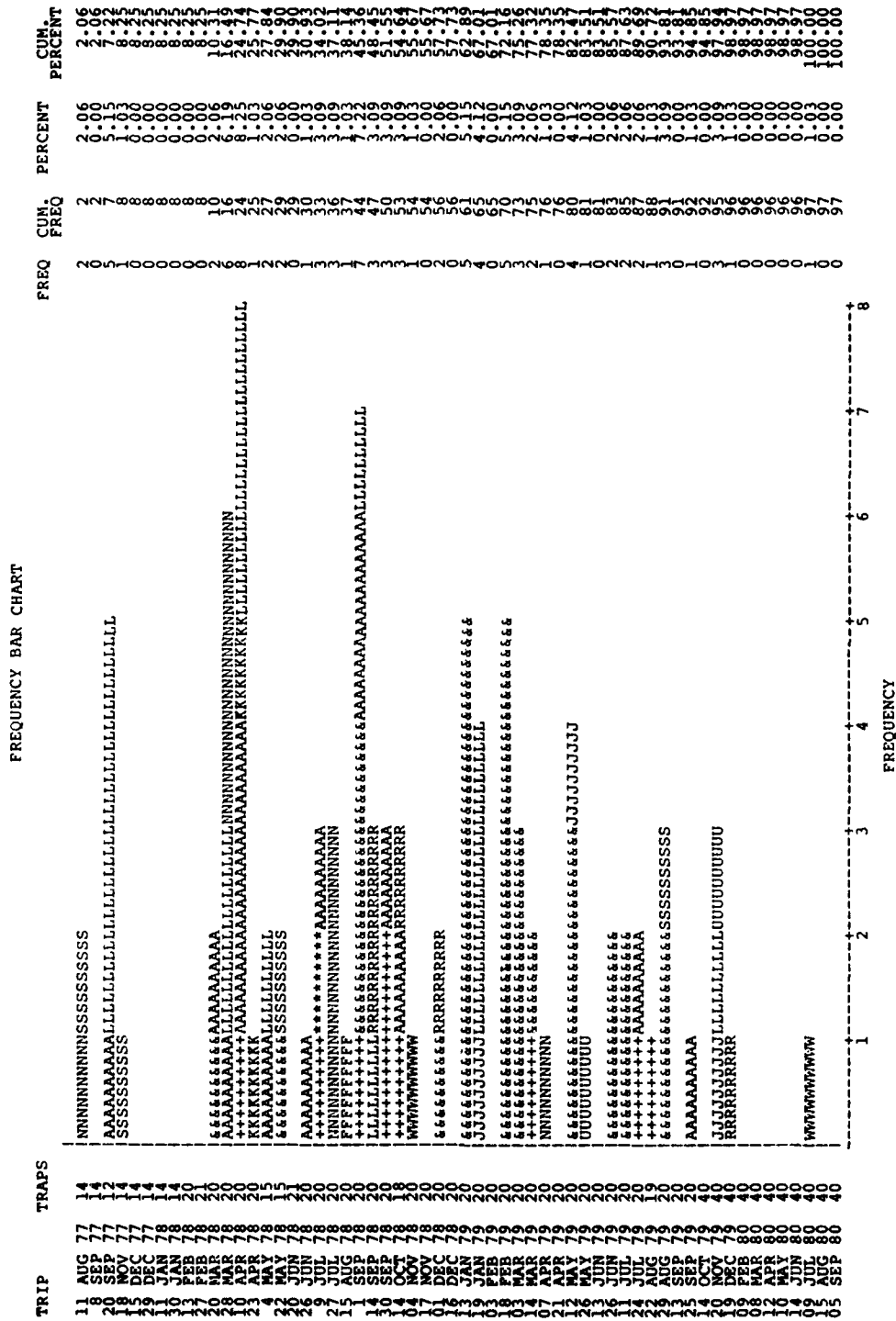


Figure 24. Funnel trap trip analysis of amphibians and reptiles on the Middle Pool site. Trip = date of trapping; traps = total number of traps set on a date.

[illegible]

Figure 25. Herp-patrol trip analysis of salamanders and reptiles on the Middle Pool site (2000 and 3000 series combined). Trip = date of herp-patrol; minutes = total sampling time of a herp-patrol on a date (time not recorded prior to 13 October 1977).

[illegible]

Figure 26. Herp-patrol trip analysis of calling frogs on the Middle Pool site. Trip = date of herp-patrol; minutes = total sampling time of a herp-patrol on a date (time not recorded prior to 13 October 1977).

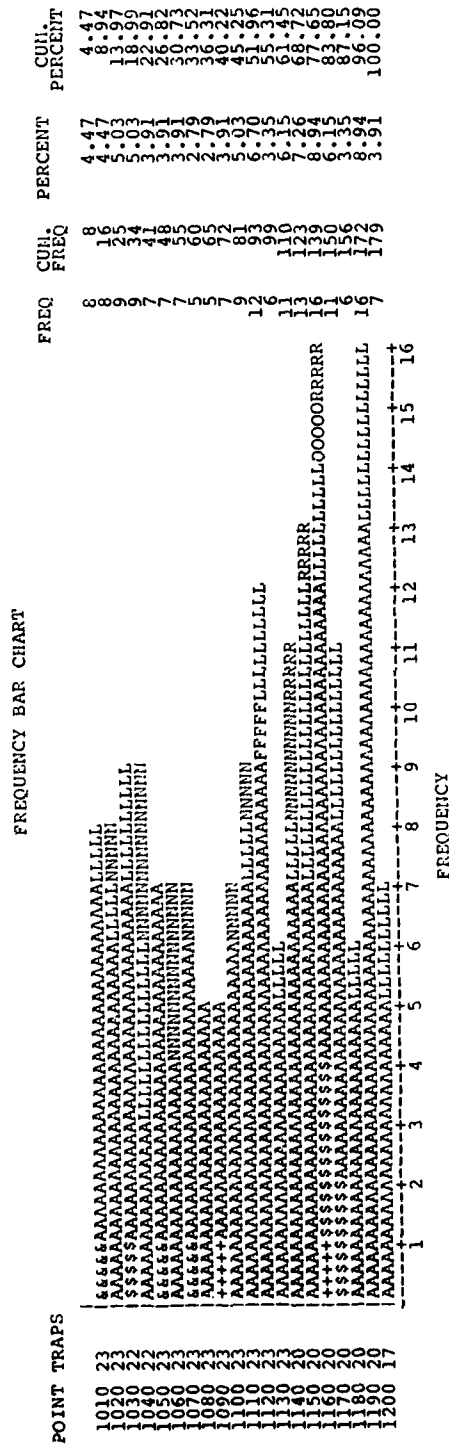


Figure 27. Funnel trap point analysis of amphibians and reptiles on the East Pool site during SYL. Point = location where traps were set; traps = total number of traps set at a sample point.

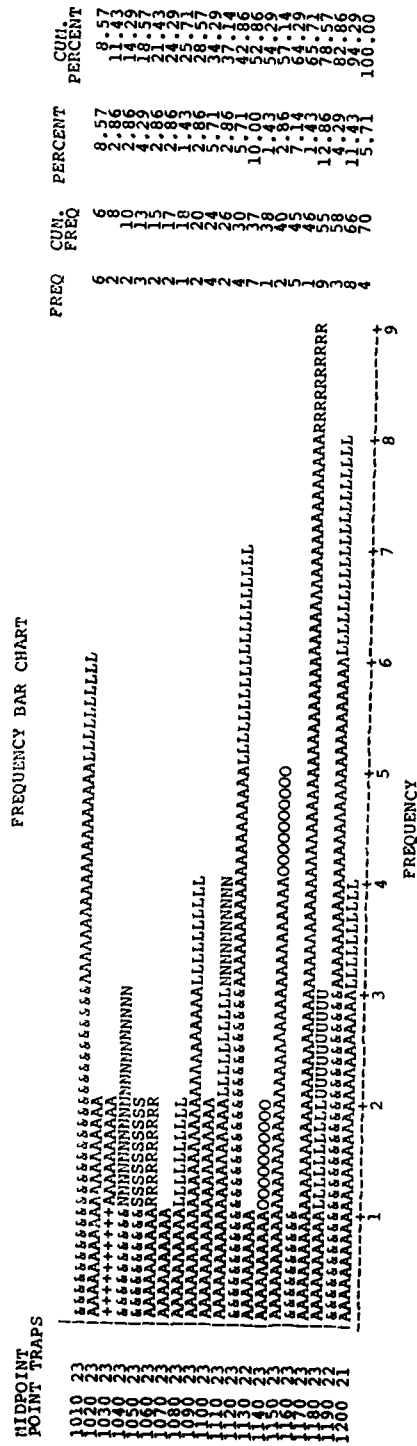


Figure 28. Funnel trap point analysis of amphibians and reptiles on the East Pool site during SY2. Point = location where traps were set; traps = total number of traps set at a sample point.

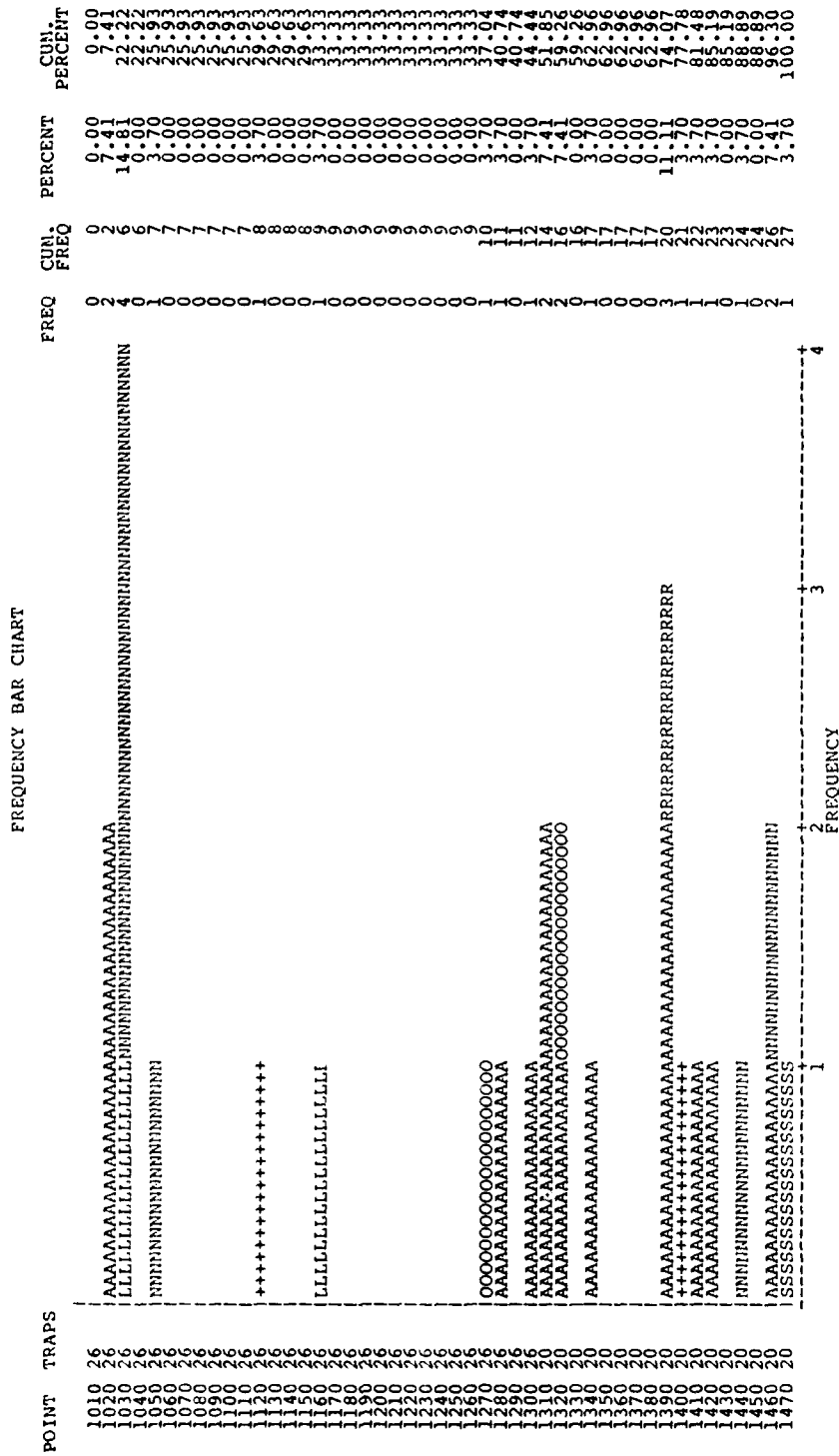


Figure 29. Funnel trap point analysis of amphibians and reptiles on the East Pool site during SY3, including the extended trapline. Point = location where traps were set; traps = total number of traps set at a sample point.

FREQUENCY BAR CHART

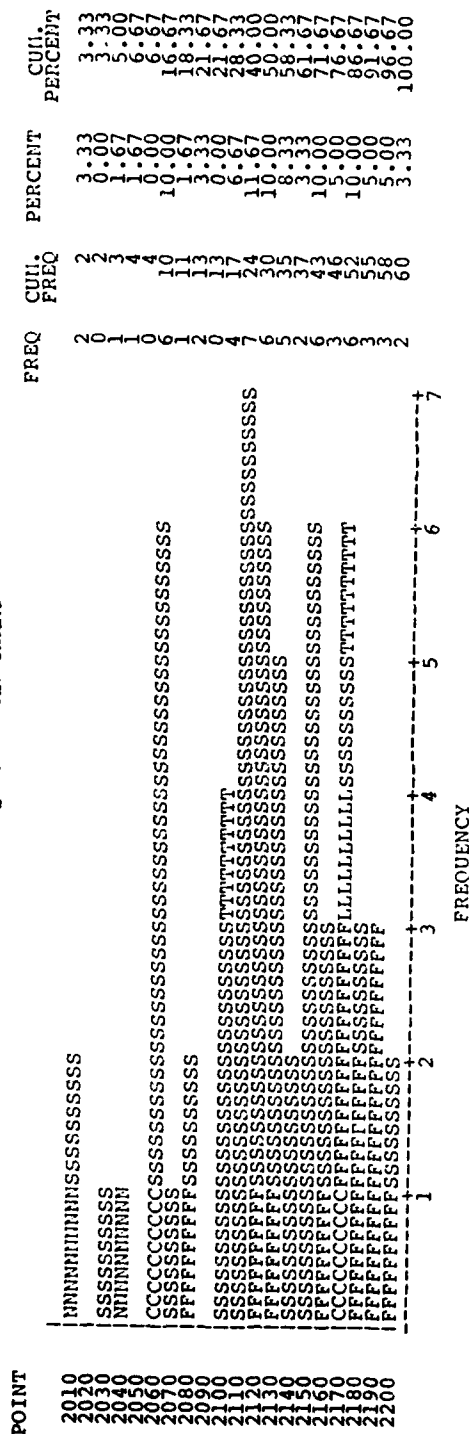


Figure 30. Herp-patrol point analysis of salamanders and reptiles on the East Pool site during SY1. Point = midpoint of 10-m section of herp-patrol transect.

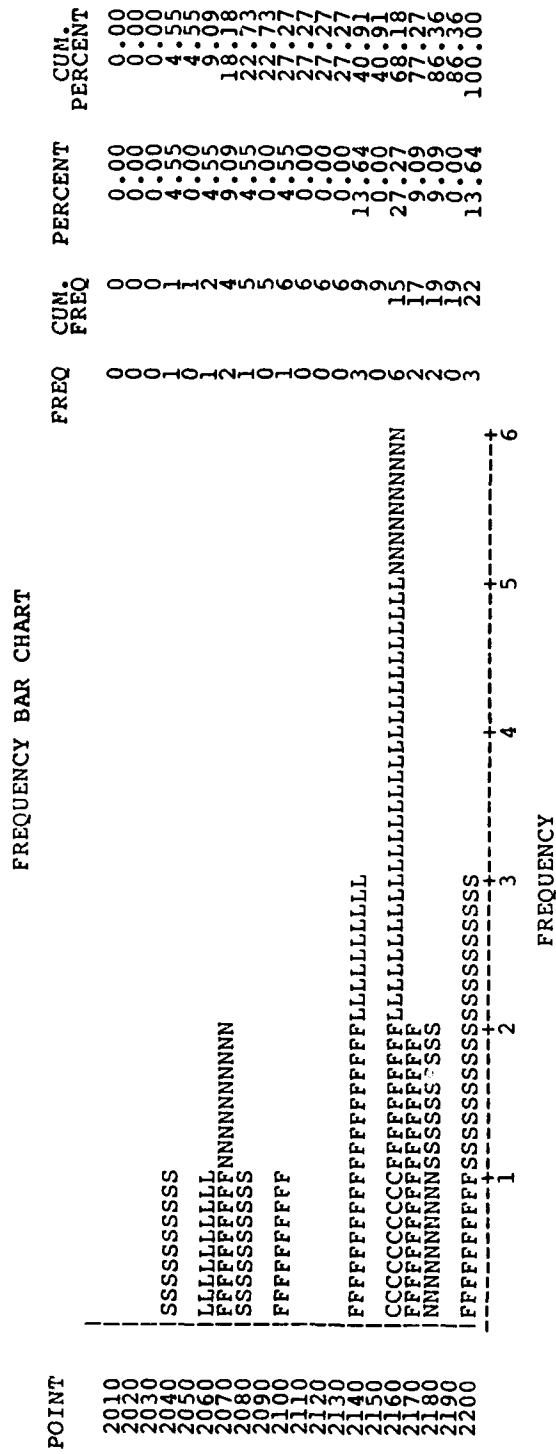


Figure 31. Herp-patrol point analysis of salamanders and reptiles on the East Pool site during SY2. Point = midpoint of 10-m section of herp-patrol transect.

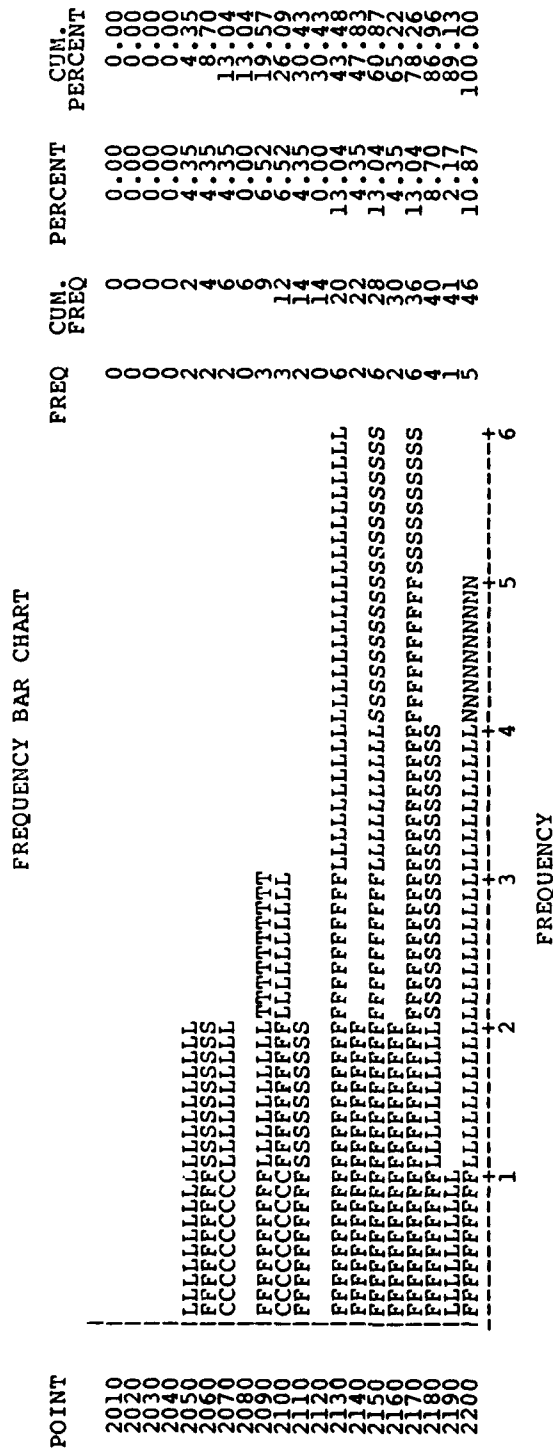


Figure 32. Herp-patrol point analysis of salamanders and reptiles on the East Pool site during SY3. Point = midpoint of 10-m section of herp-patrol transect.

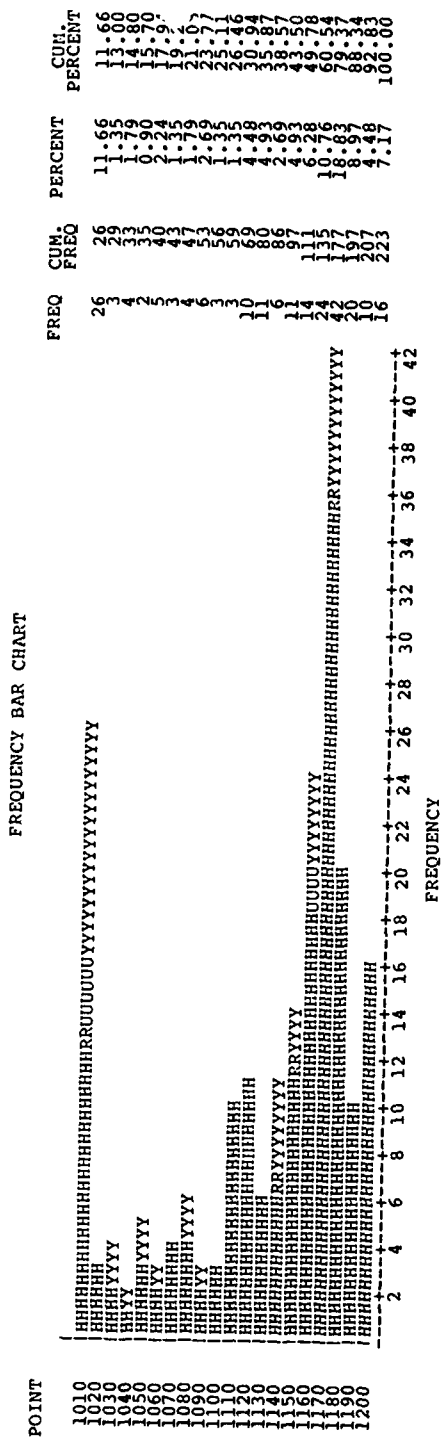


Figure 33. Herp-patrol point analysis of calling frogs on the East Pool site during Syl. Point = midpoint of 10-m section of herp-patrol transect.

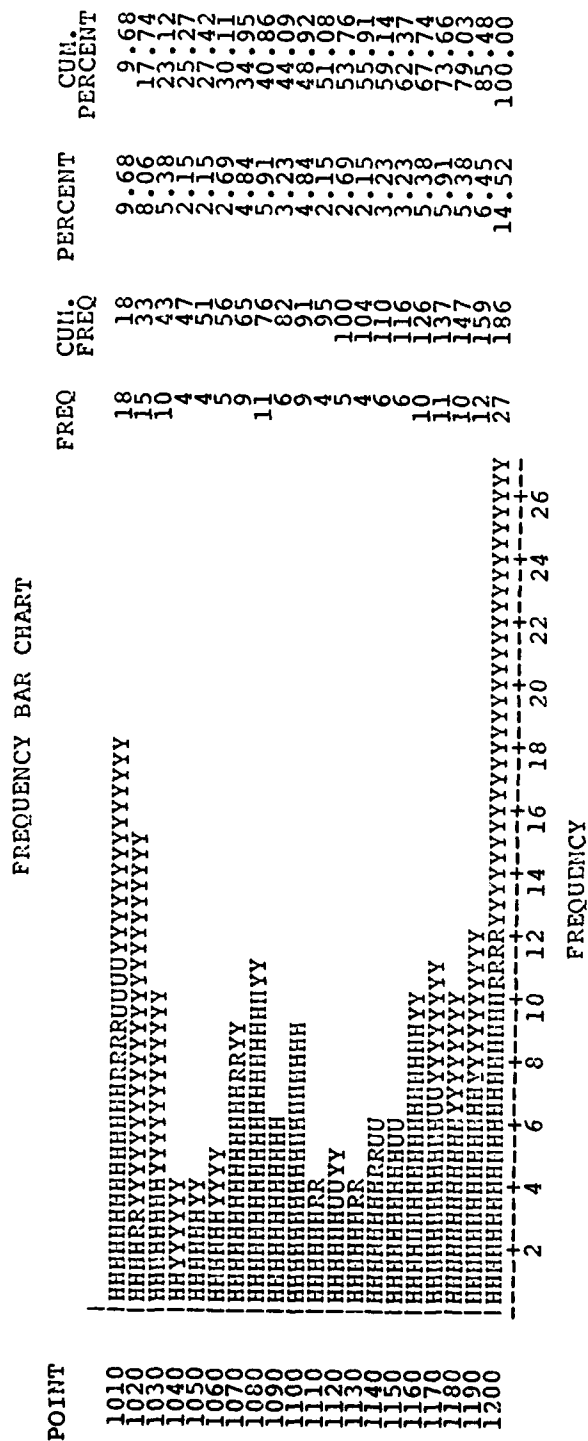


Figure 34. Herp-patrol point analysis of calling frogs on the East Pool site during SY2. Point = midpoint of 10-m section of herp-patrol transect.

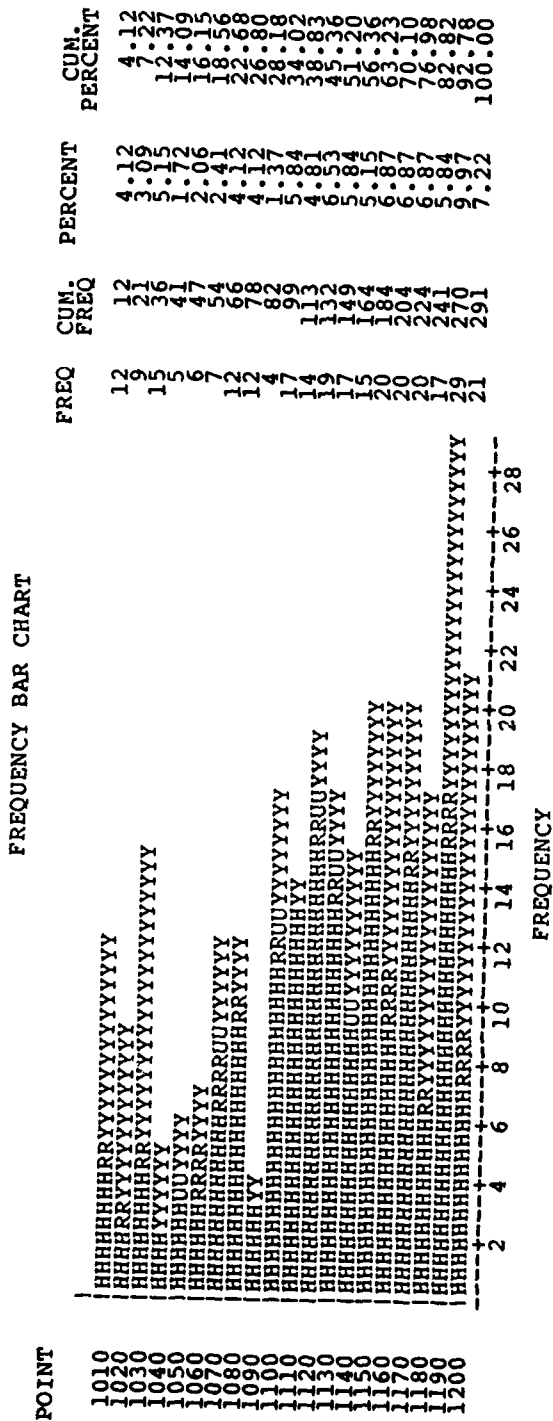
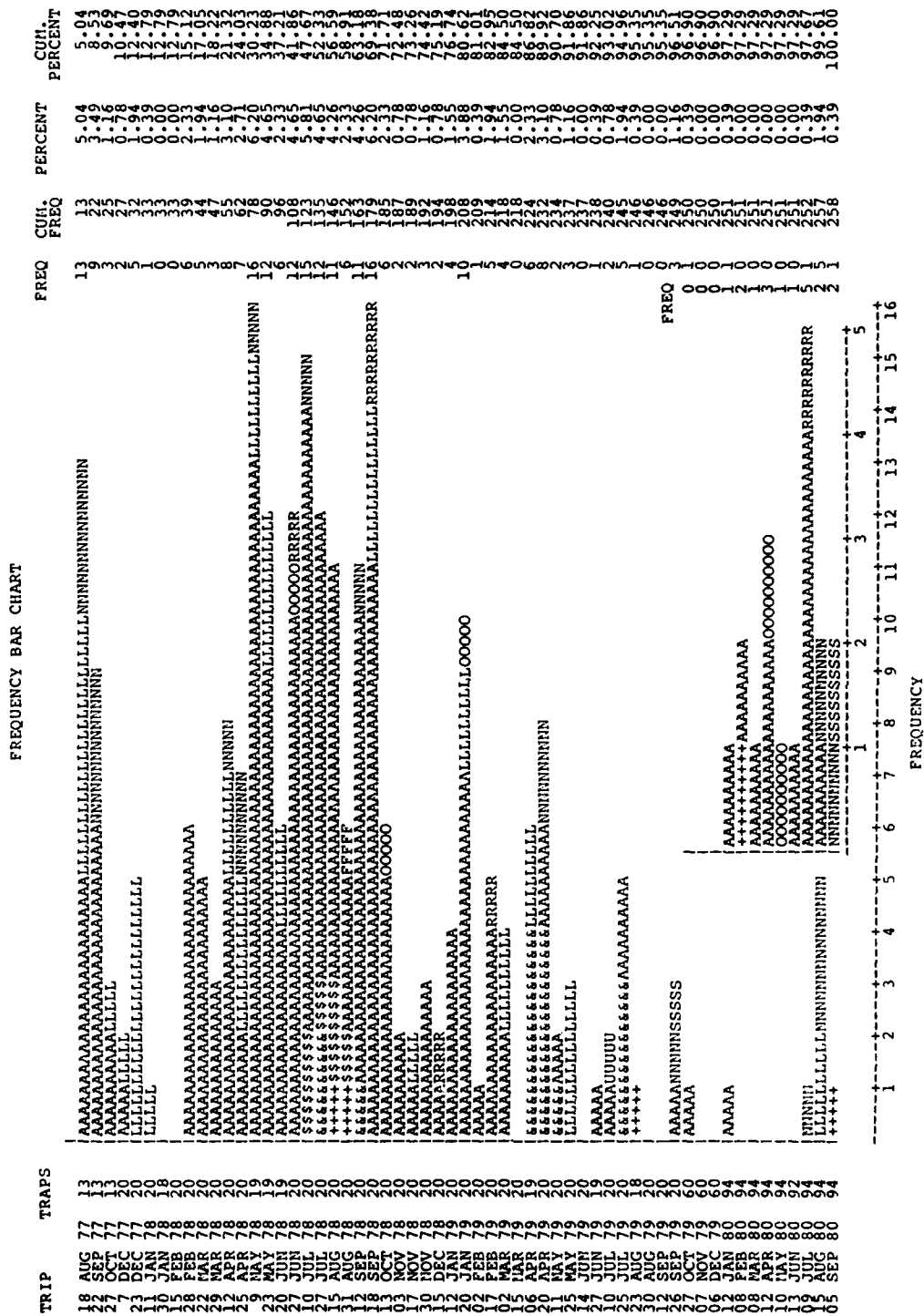


Figure 35. Herp-patrol point analysis of calling frogs on the East Pool site during SY3. Point = midpoint of 10-m section of herp-patrol transect.



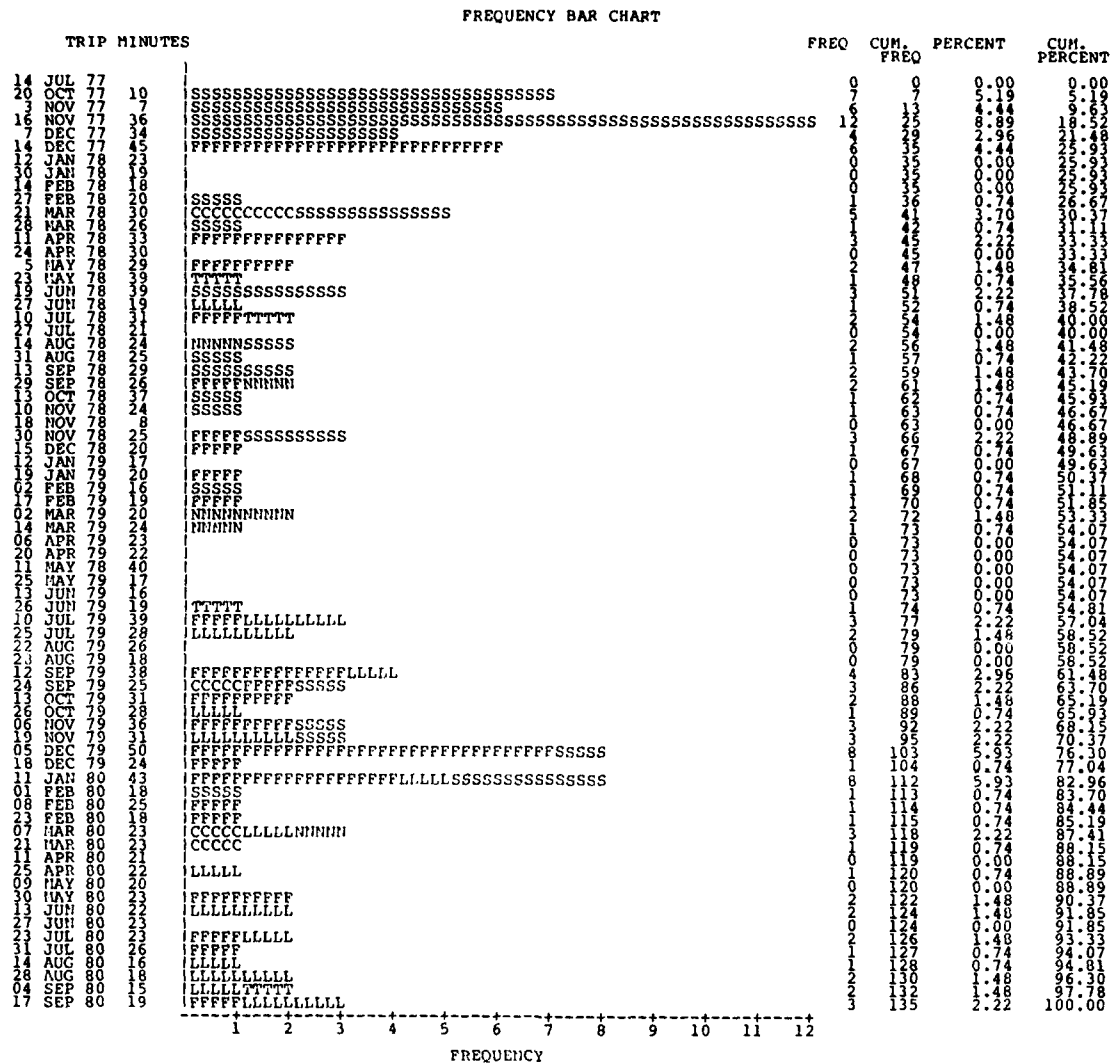


Figure 37. Herp-patrol trip analysis of salamanders and reptiles on the East Pool site. Trip = date of herp-patrol; minutes = total sampling time of a herp-patrol on a date (time not recorded prior to 13 October 1977).

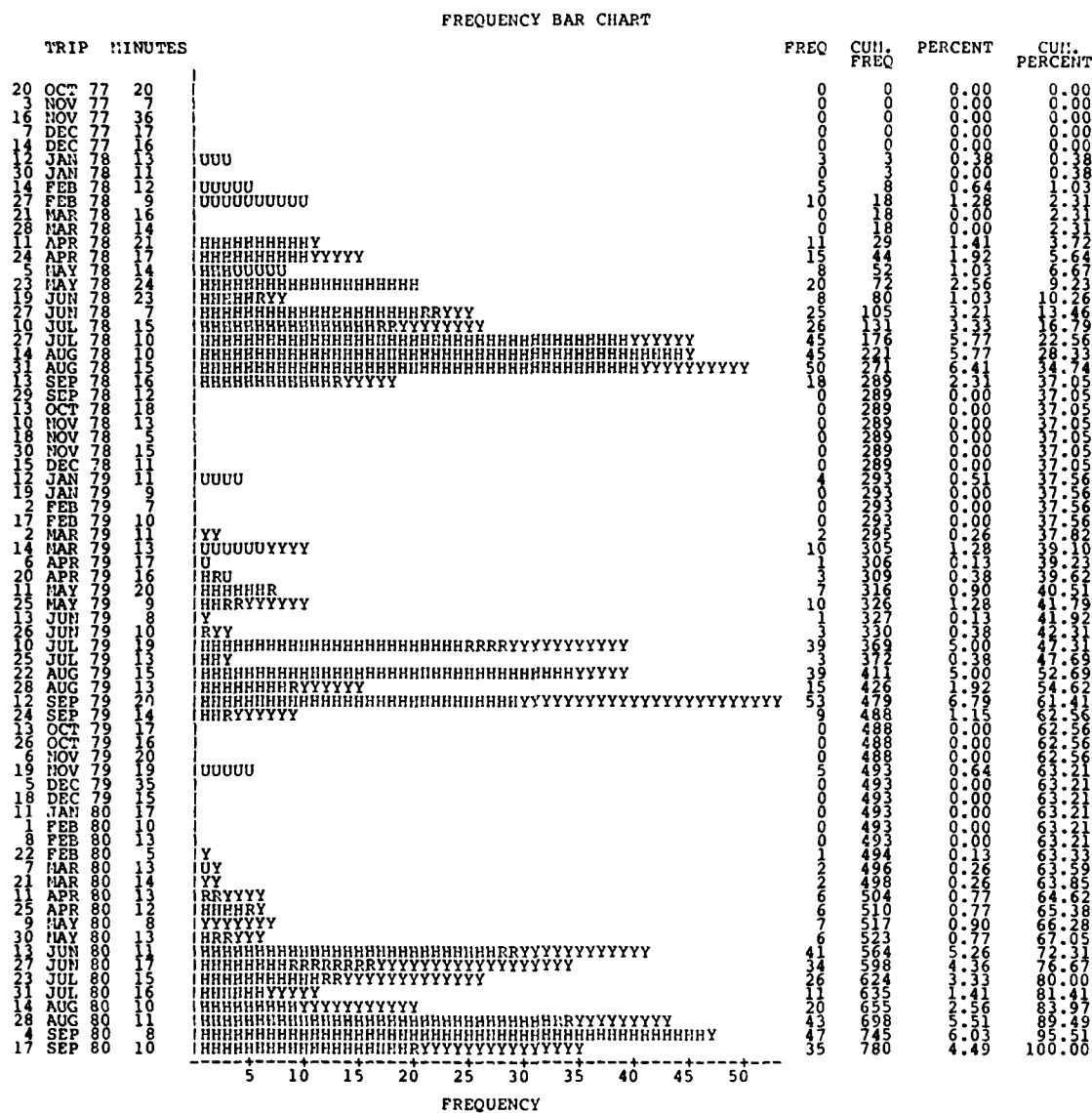


Figure 38. Herp-patrol trip analysis of calling frogs on the East Pool site. Trip = date of herp-patrol; minutes = total sampling time of a herp-patrol on a date (time not recorded prior to 13 October 1977).

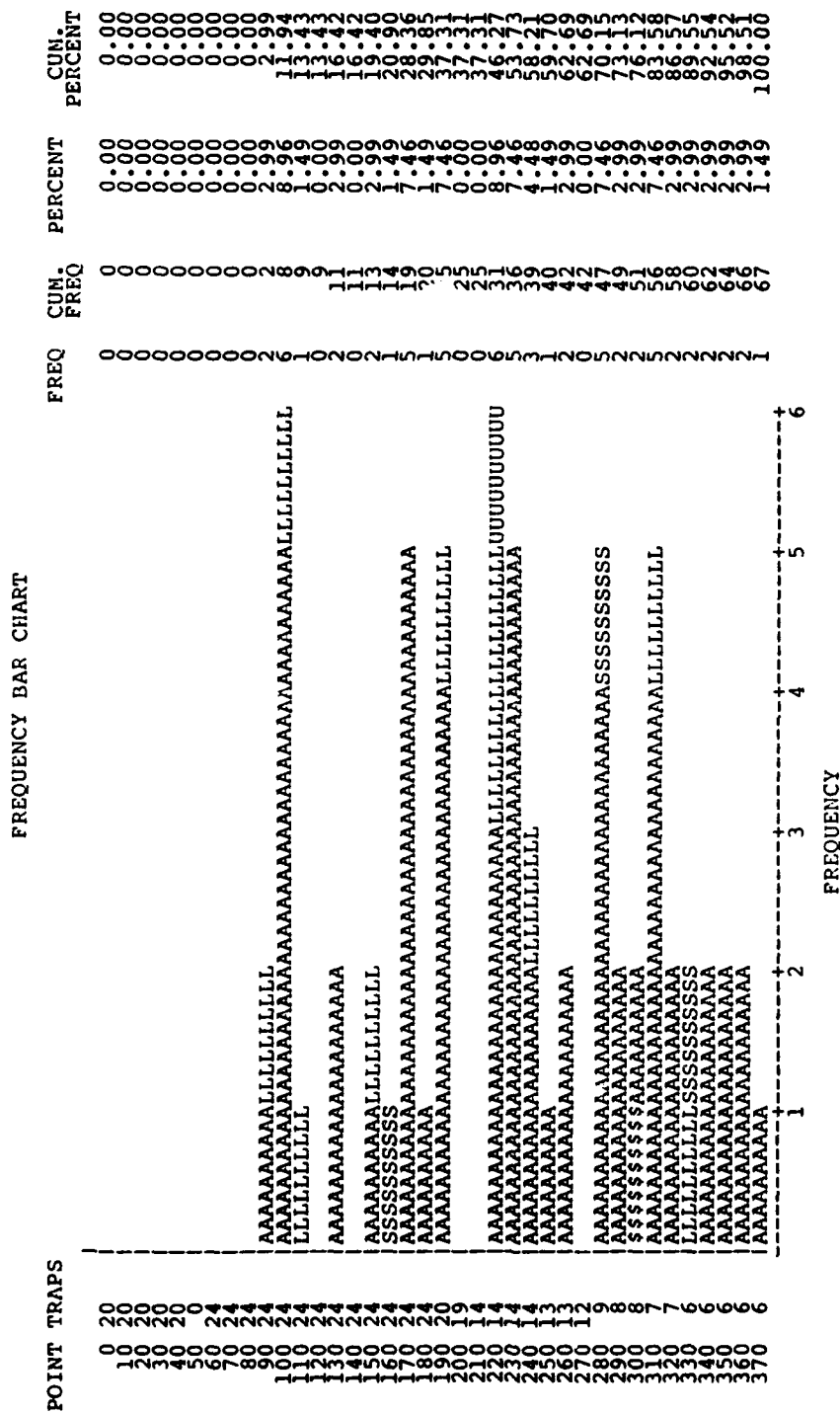


Figure 39. Funnel trap point analysis of amphibians and reptiles on the West Pool site during SY1. Point = location where traps were set; traps = total number of traps set at a sample point.

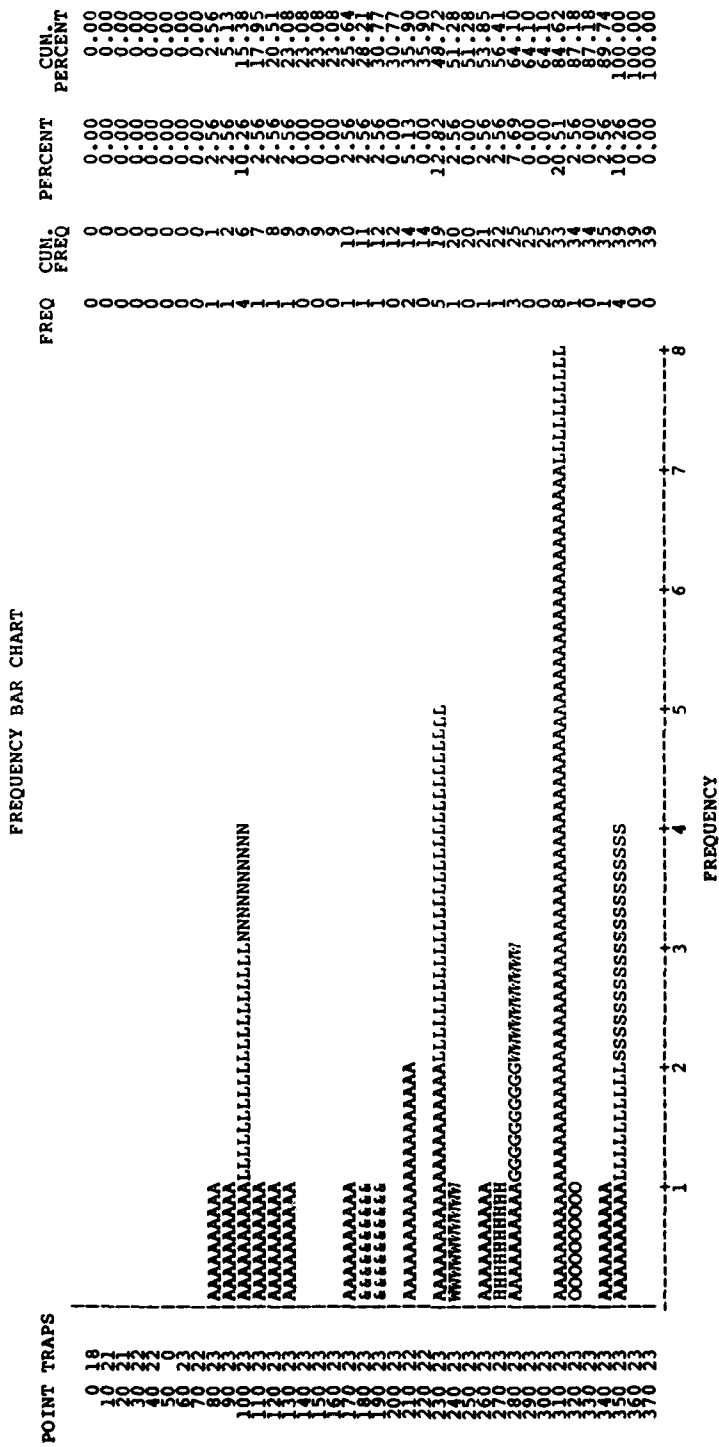


Figure 40. Funnel trap point analysis of amphibians and reptiles on the West Pool site during SY2. Point = location where traps were set; traps = total number of traps set at a sample point.

FREQUENCY BAR CHART

POINT	TRAPS		FREQ	CUM. FREQ	PERCENT	CUM. PERCENT
0	22		0	0	0.00	0.00
10	22		0	0	0.00	0.00
20	22		0	0	0.00	0.00
30	22		0	0	0.00	0.00
40	22		0	0	0.00	0.00
50	0		0	0	0.00	0.00
60	22		0	0	0.00	0.00
70	22		0	0	0.00	0.00
80	22		0	0	0.00	0.00
90	22		0	0	0.00	0.00
100	21		0	0	0.00	0.00
110	22		0	0	0.00	0.00
120	20		0	0	0.00	0.00
130	22		0	0	0.00	0.00
140	22	OOO	1	1	25.00	25.00
150	22	AAA	1	2	25.00	50.00
160	22		0	2	0.00	50.00
170	22		0	2	0.00	50.00
180	22		0	2	0.00	50.00
190	21		0	2	0.00	50.00
200	22		0	2	0.00	50.00
210	22		0	2	0.00	50.00
220	20		0	2	0.00	50.00
230	20		0	2	0.00	50.00
240	22		0	2	0.00	50.00
250	22		0	2	0.00	50.00
260	22		0	2	0.00	50.00
270	20		0	2	0.00	50.00
280	22		0	2	0.00	50.00
290	22		0	2	0.00	50.00
300	22	AAA	1	3	25.00	75.00
310	22		0	3	0.00	75.00
320	22	AAA	1	4	25.00	100.00
330	22		0	4	0.00	100.00
340	22		0	4	0.00	100.00
350	22		0	4	0.00	100.00
360	22		0	4	0.00	100.00
370	22		0	4	0.00	100.00

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FREQUENCY

Figure 41. Funnel trap point analysis of amphibians and reptiles on the West Pool site during SY3. Point = location where traps were set; traps = total number of traps set at a sample point.

[illegible]

Abstract

[illegible]

Figure 43. Herp-patrol point analysis of salamanders and reptiles on the West Pool site during SY2. Point = midpoint of 10-m section of herp-patrol transect.

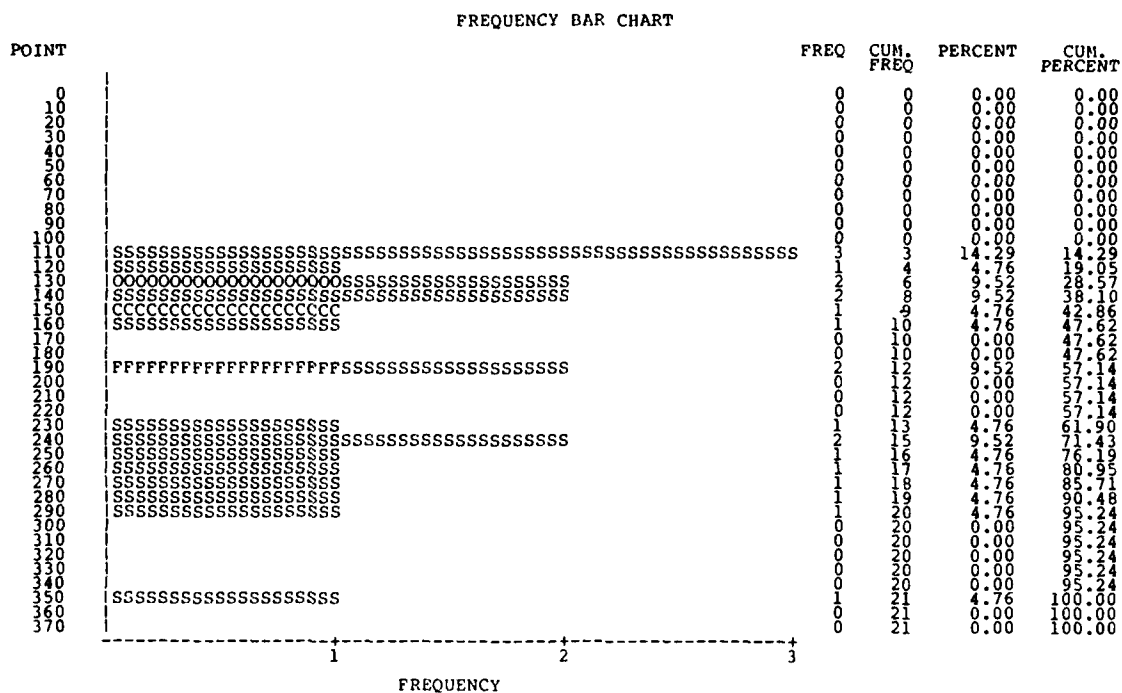


Figure 44. Herp-patrol point analysis of salamanders and reptiles on the West Pool site during SY3. Point = midpoint of 10-m section of herp-patrol transect.

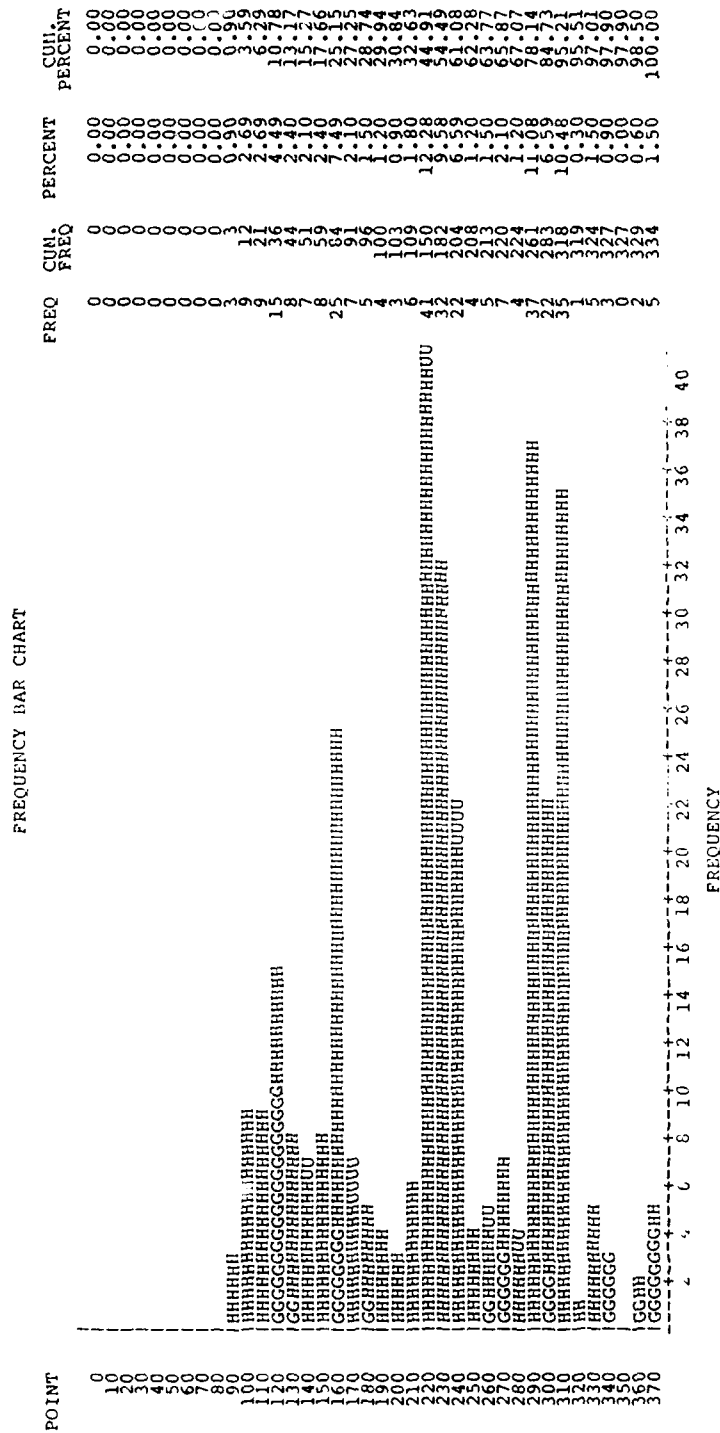


Figure 45. Herp-patrol point analysis of calling frogs on the West Pool site during SYL. Point = midpoint of 10-m section of herp-patrol transect.

FREQUENCY BAR CHART

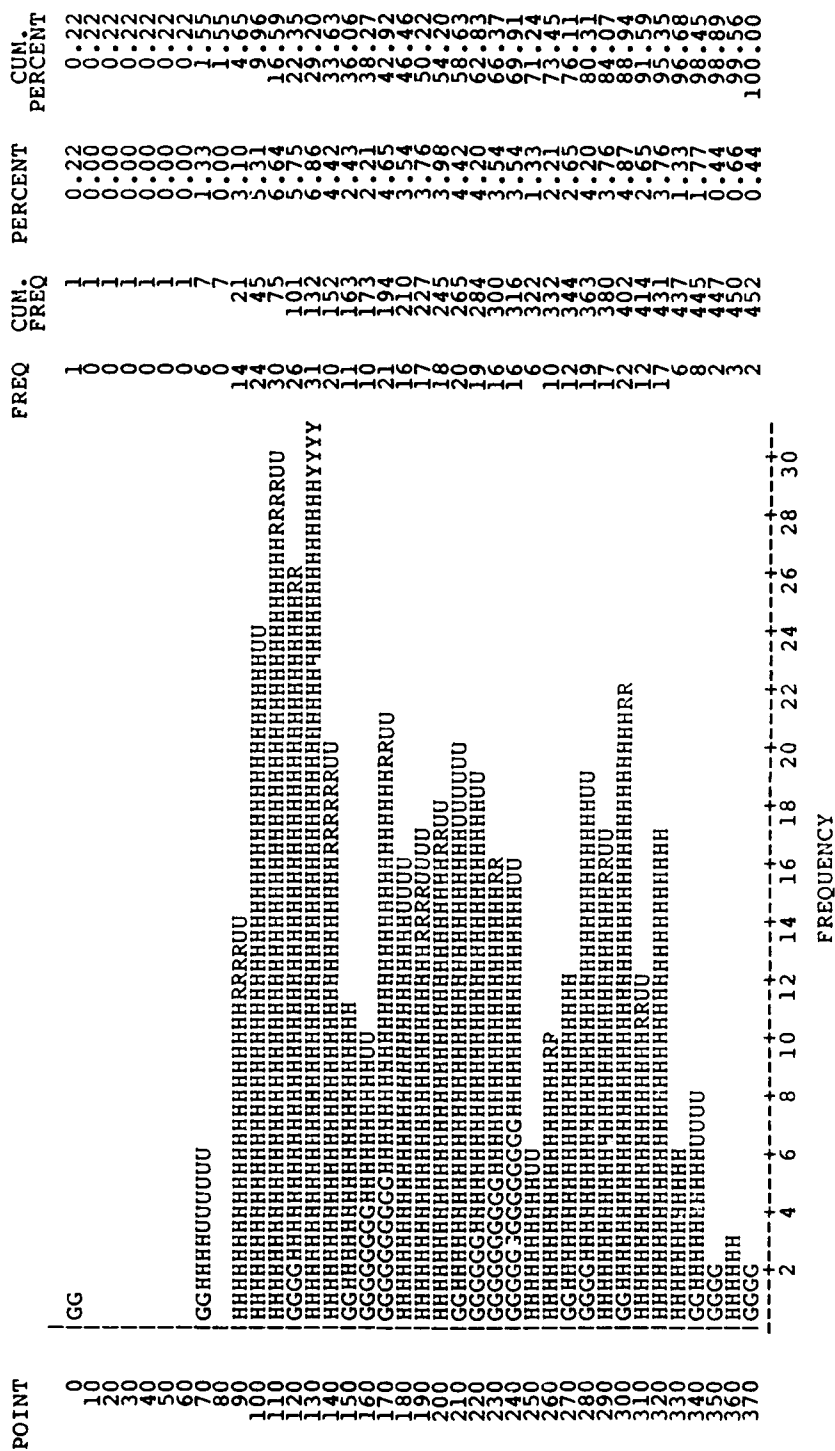


Figure 46. Herp-patrol point analysis of calling frogs on the West Pool site during SY2. Point = midpoint of 10-m section of herp-patrol transect.

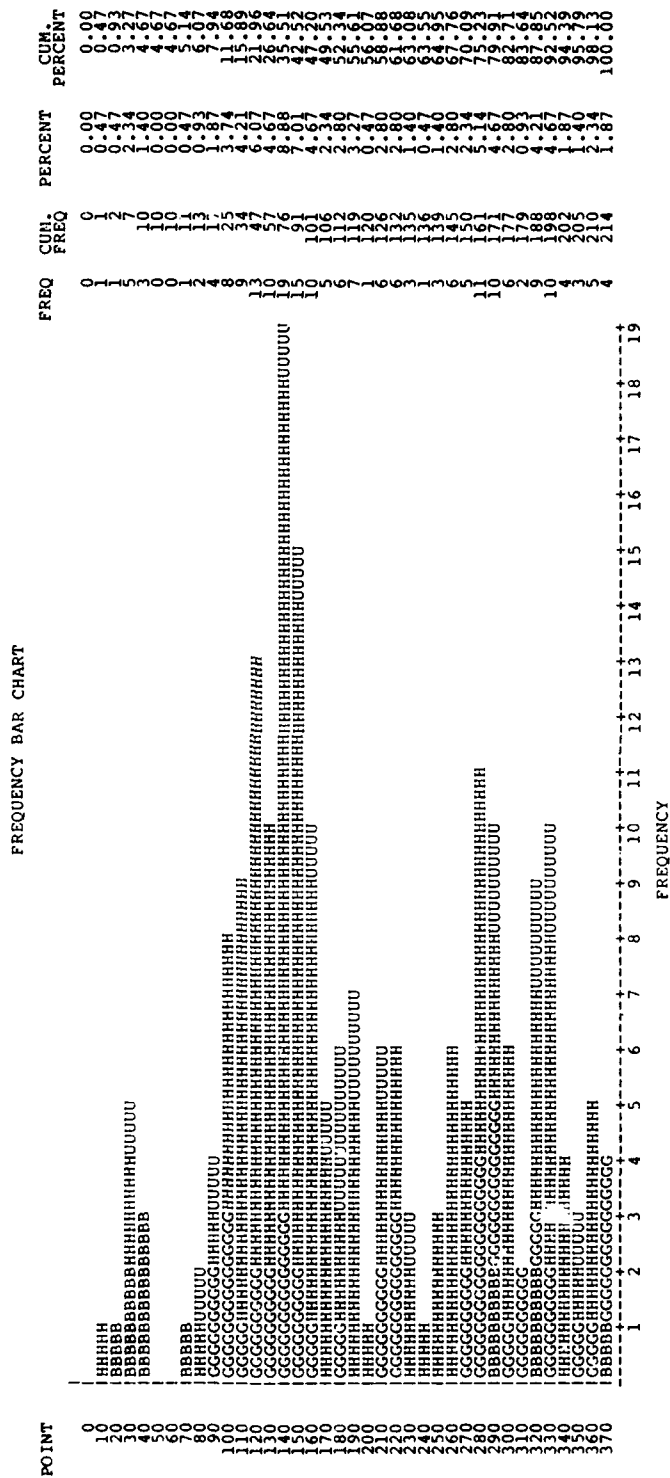


Figure 47. Herp-patrol point analysis of calling frogs on the West Pool site during SY3. Point = midpoint of 10-m section of herp-patrol transect.

FREQUENCY BAR CHART

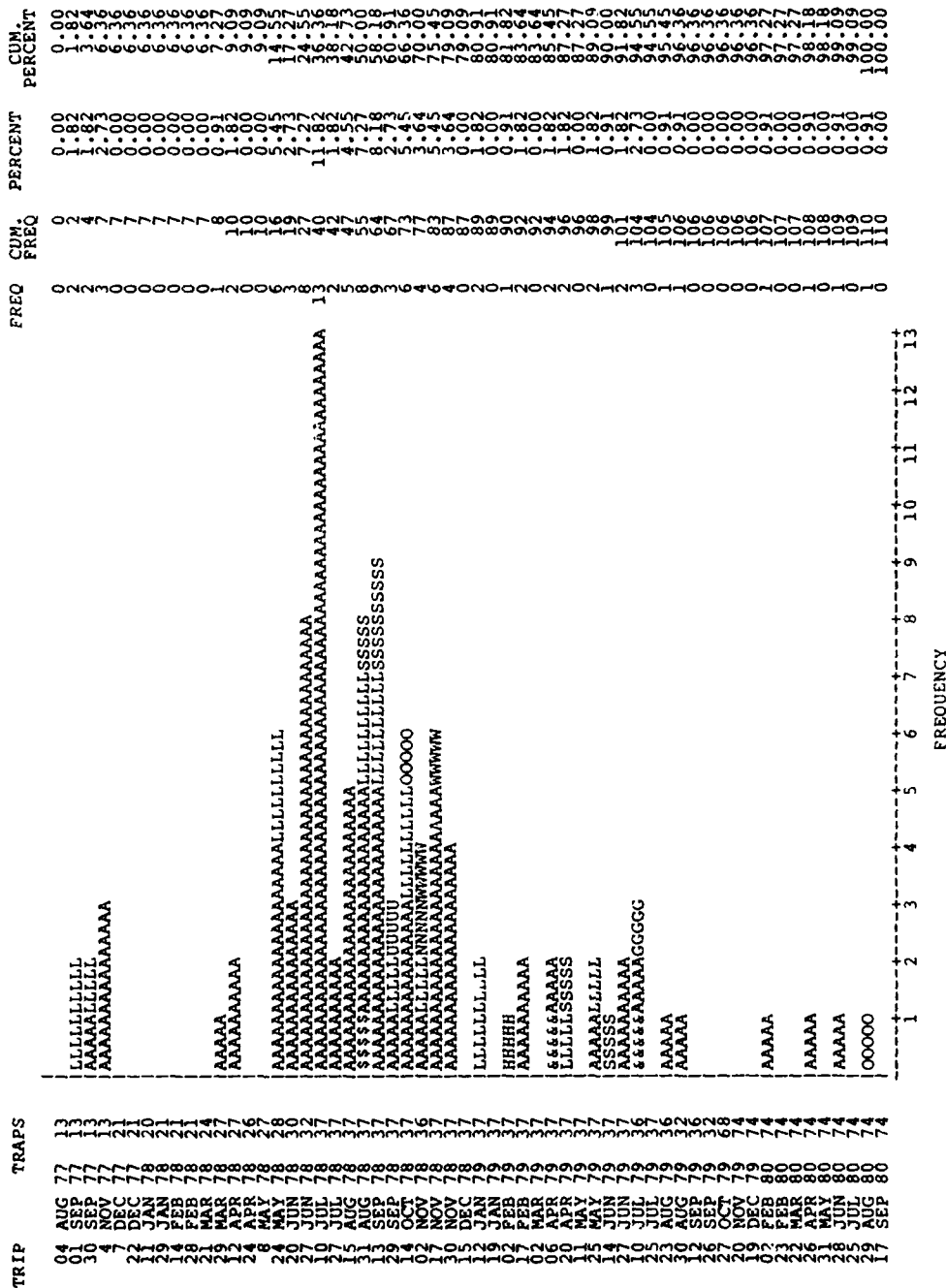


Figure 48. Funnel trap trap analysis of amphibians and reptiles on the West Pool site. Trip = date of trapping; traps = total number of traps set on a date.

[illegible]

Figure 49. Herp-patrol trip analysis of salamanders and reptiles on the West Pool site. Trip = date of herp-patrol; minutes = total sampling time of a herp-patrol on a date (time not recorded prior to 13 October 1977).

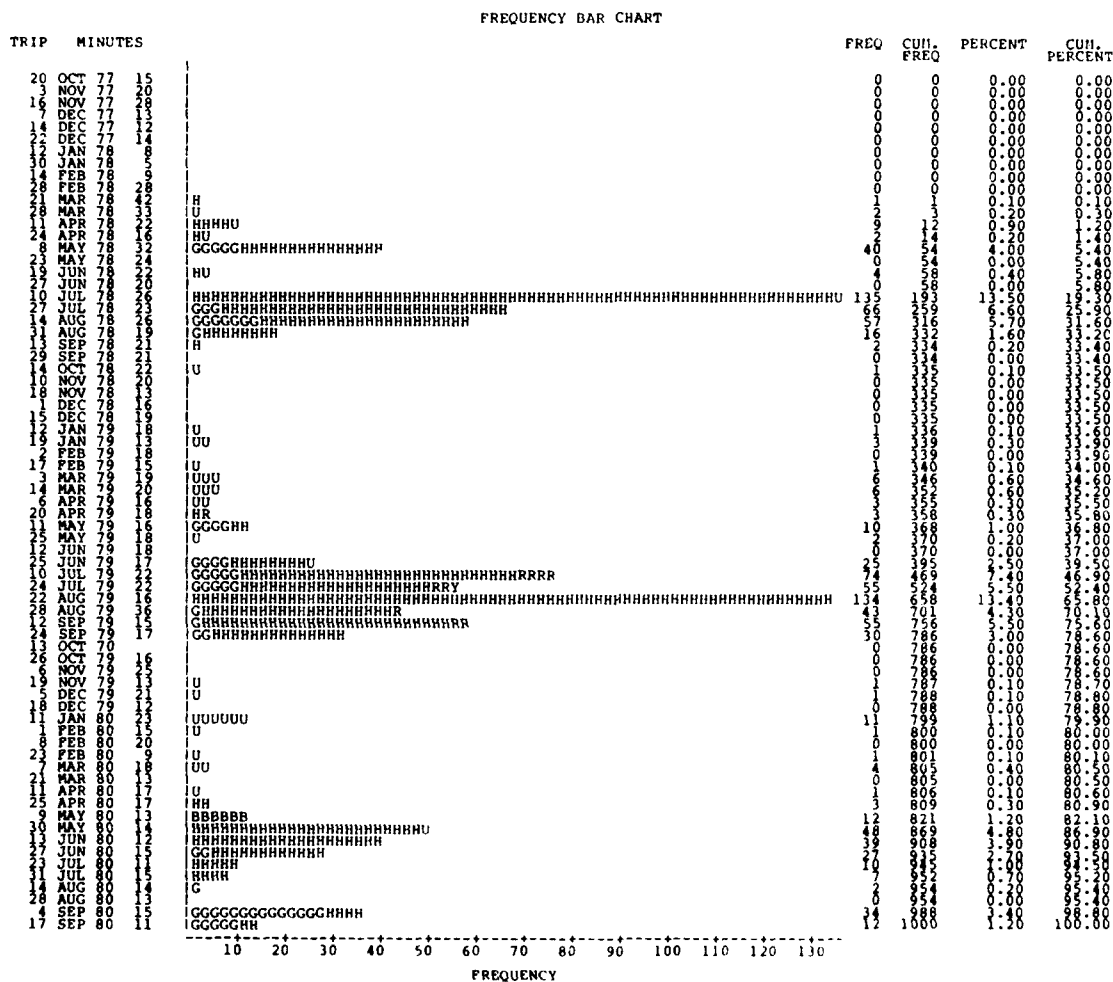


Figure 50. Herp-patrol trip analysis of calling frogs on the West Pool site. Trip = date of herp-patrol; minutes = total sampling time of a herp-patrol on a date (time not recorded prior to 13 October 1977).

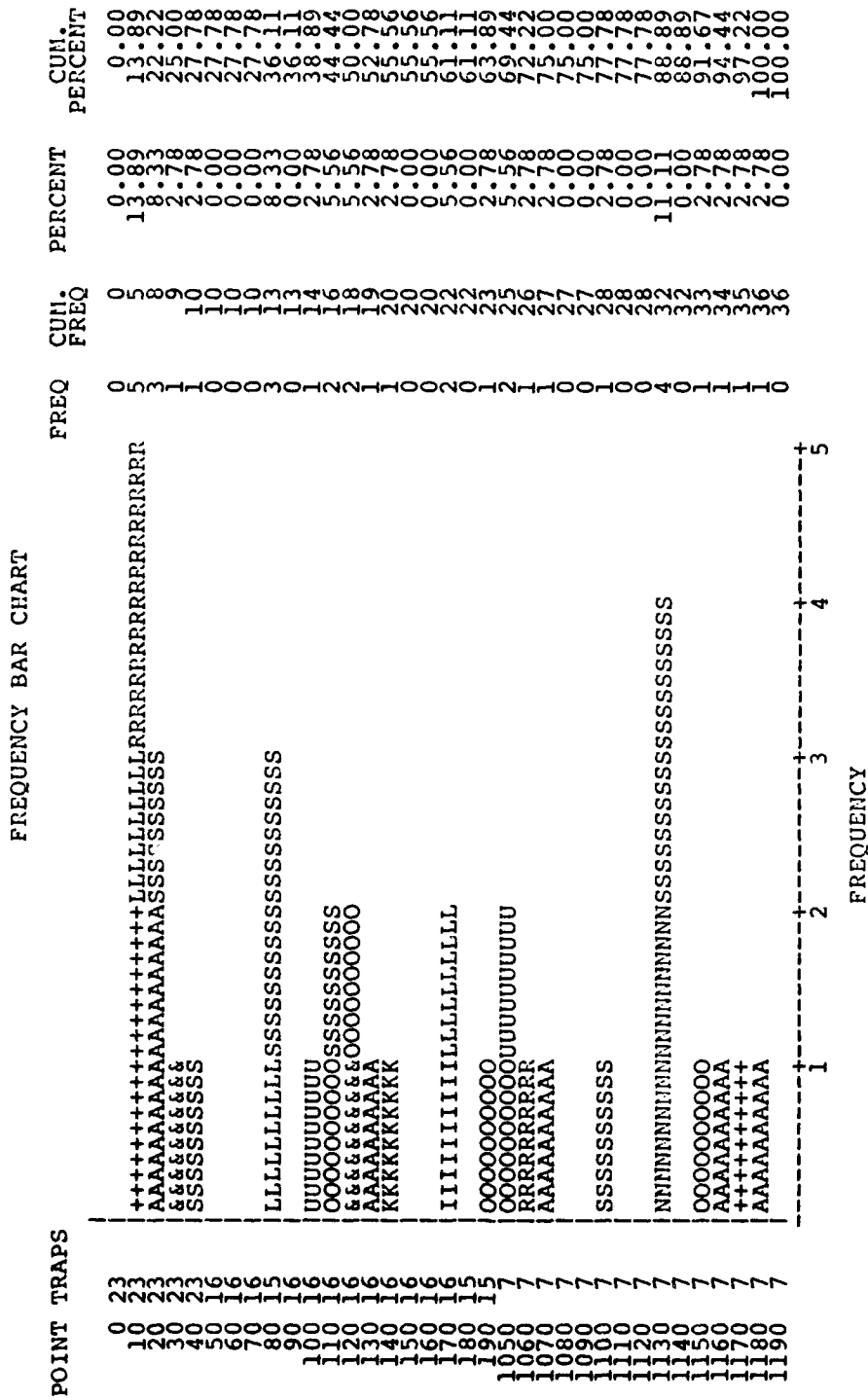


Figure 52. Funnel trap point analysis of amphibians and reptiles on the Gatlin Canal site during SY2. Point = location where traps were set; traps = total number of traps set at a sample point.

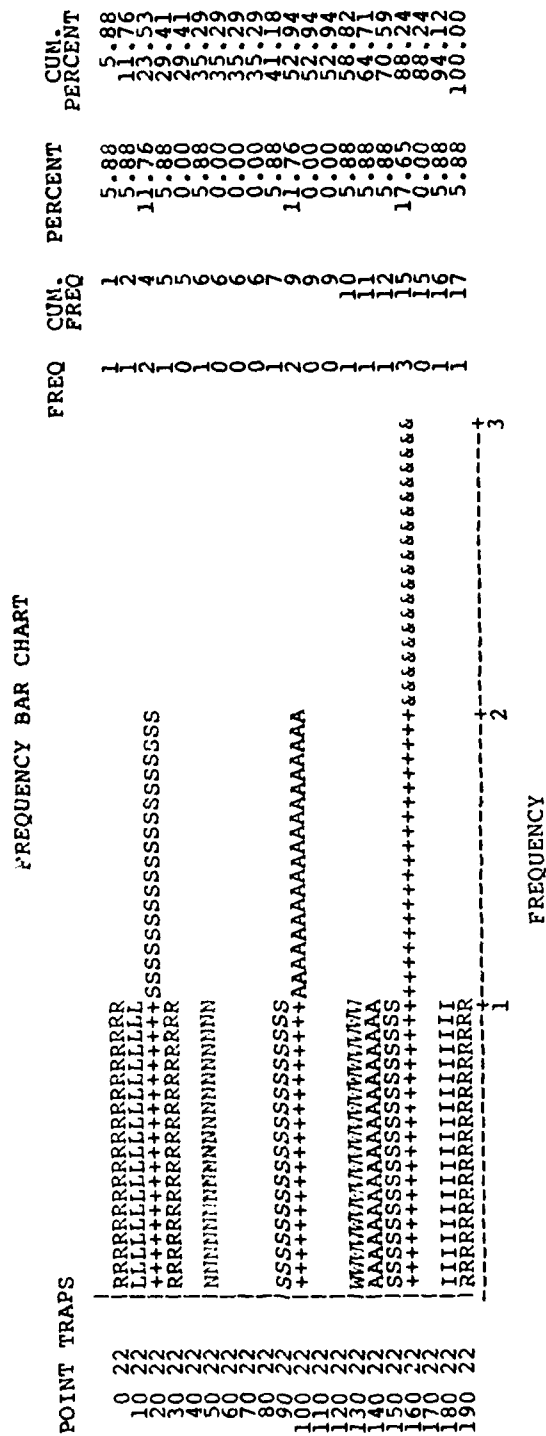


Figure 53. Funnel trap site analysis of amphibians and reptiles on the Gatlin Canal site during SY3. Point = location where traps were set; traps = total number of traps set at a sample point.

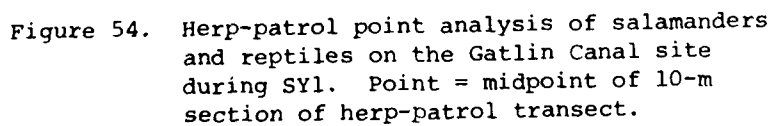


Figure 54. Herp-patrol point analysis of salamanders and reptiles on the Gatlin Canal site during SYL. Point = midpoint of 10-m section of herp-patrol transect.

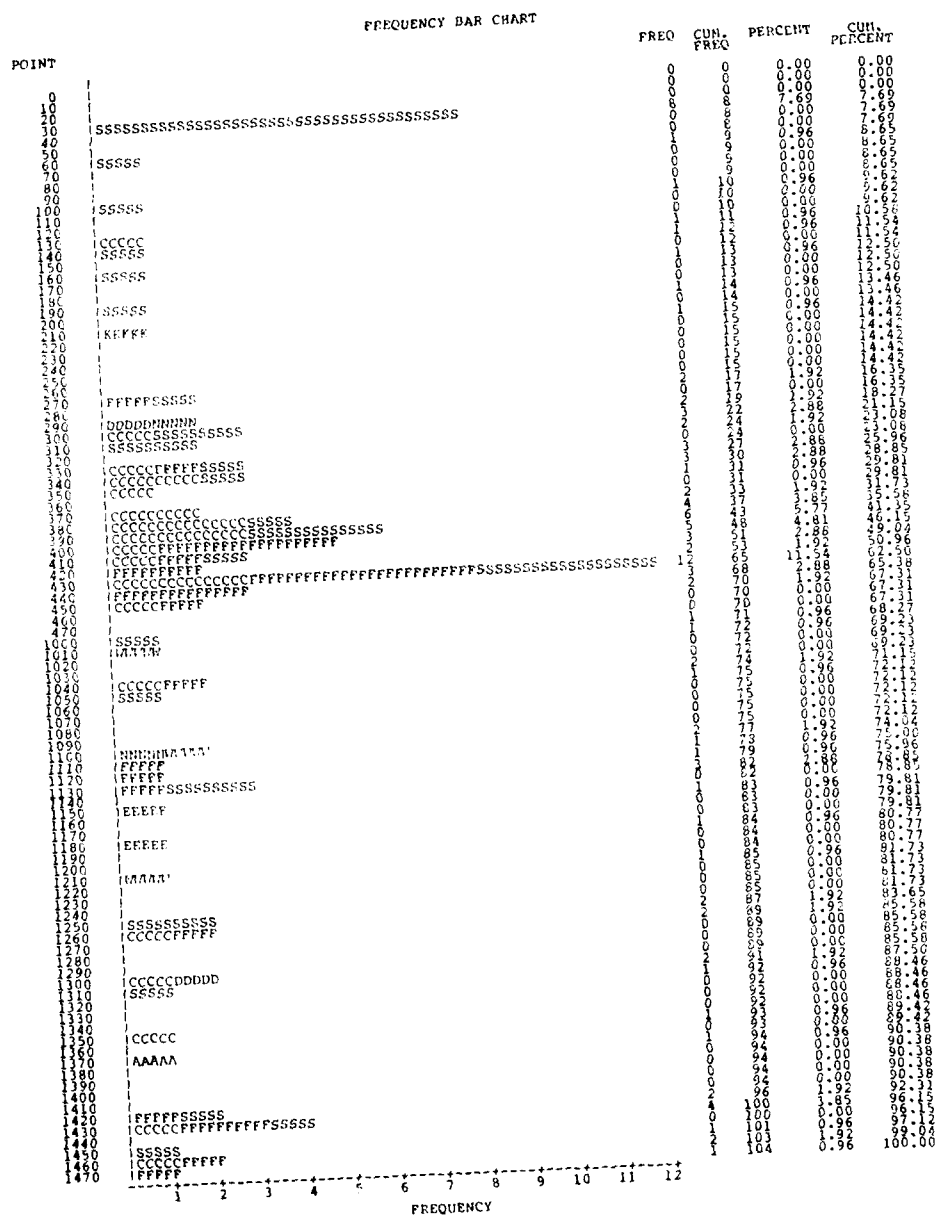
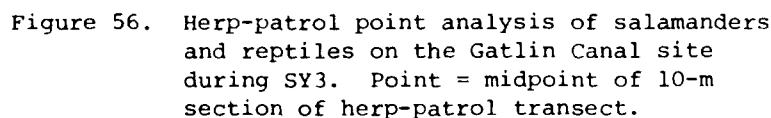


Figure 55. Herp-patrol point analysis of salamanders and reptiles on the Gatlin Canal site during SY2. Point = midpoint of 10-m section of herp-patrol transect.



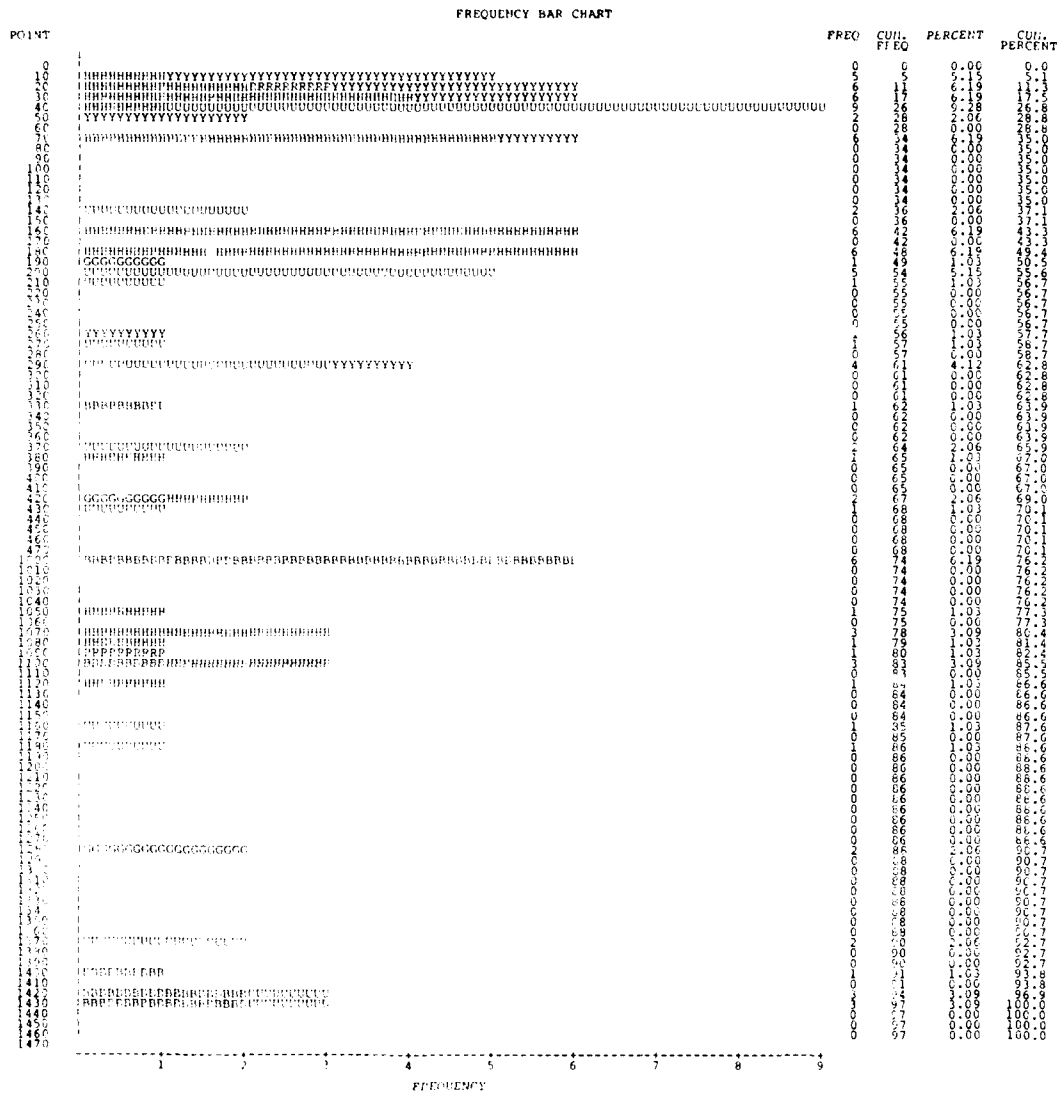


Figure 57. Herp-patrol point analysis of calling frogs on the Gatlin Canal site during SY1. Point = midpoint of 10-m section of herp-patrol transect.

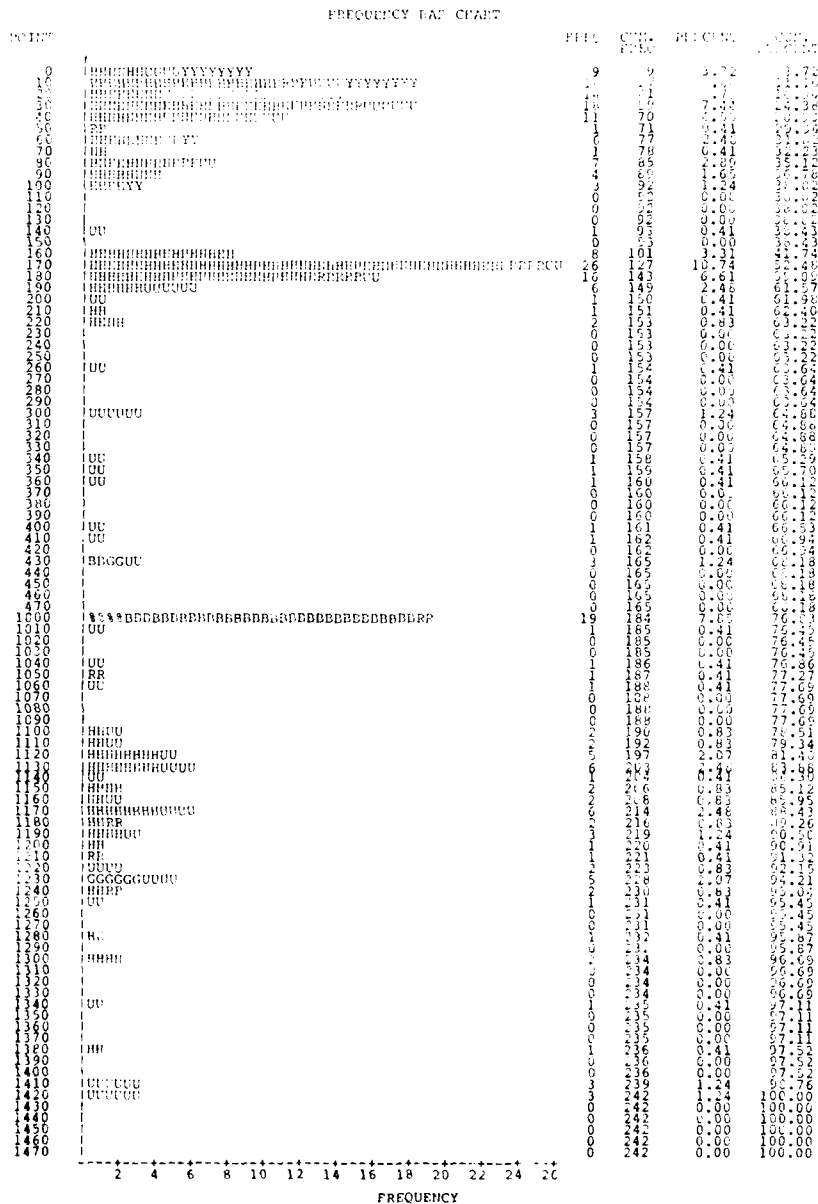


Figure 58. Herp-patrol point analysis of calling frogs on the Gatlin Canal site during SY2. Point = midpoint of 10-m section of herp-patrol transect.

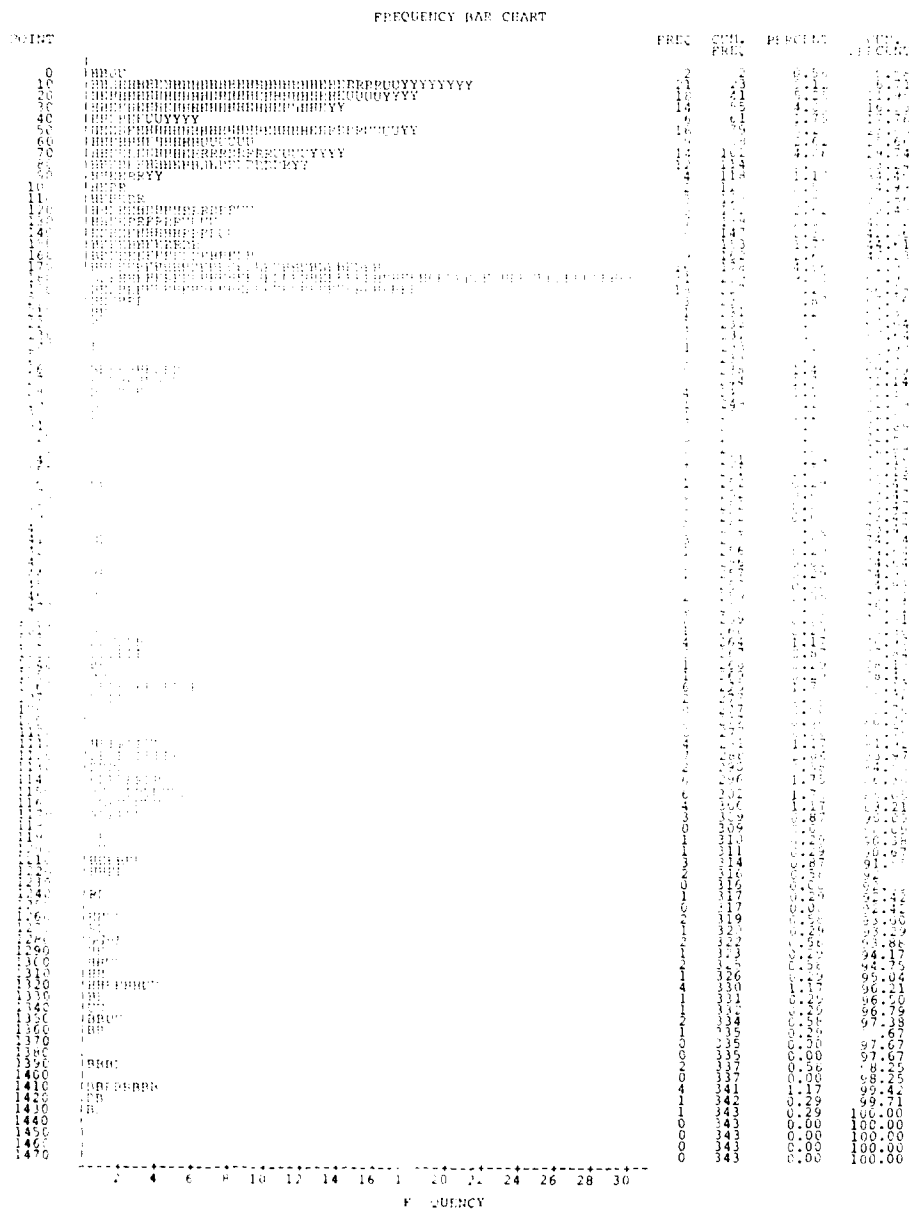


Figure 59. Herp-patrol point analysis of calling frogs on the Gatlin Canal site during SY3. Point = midpoint of 10-m section of herp-patrol transect.

FREQUENCY BAR CHART

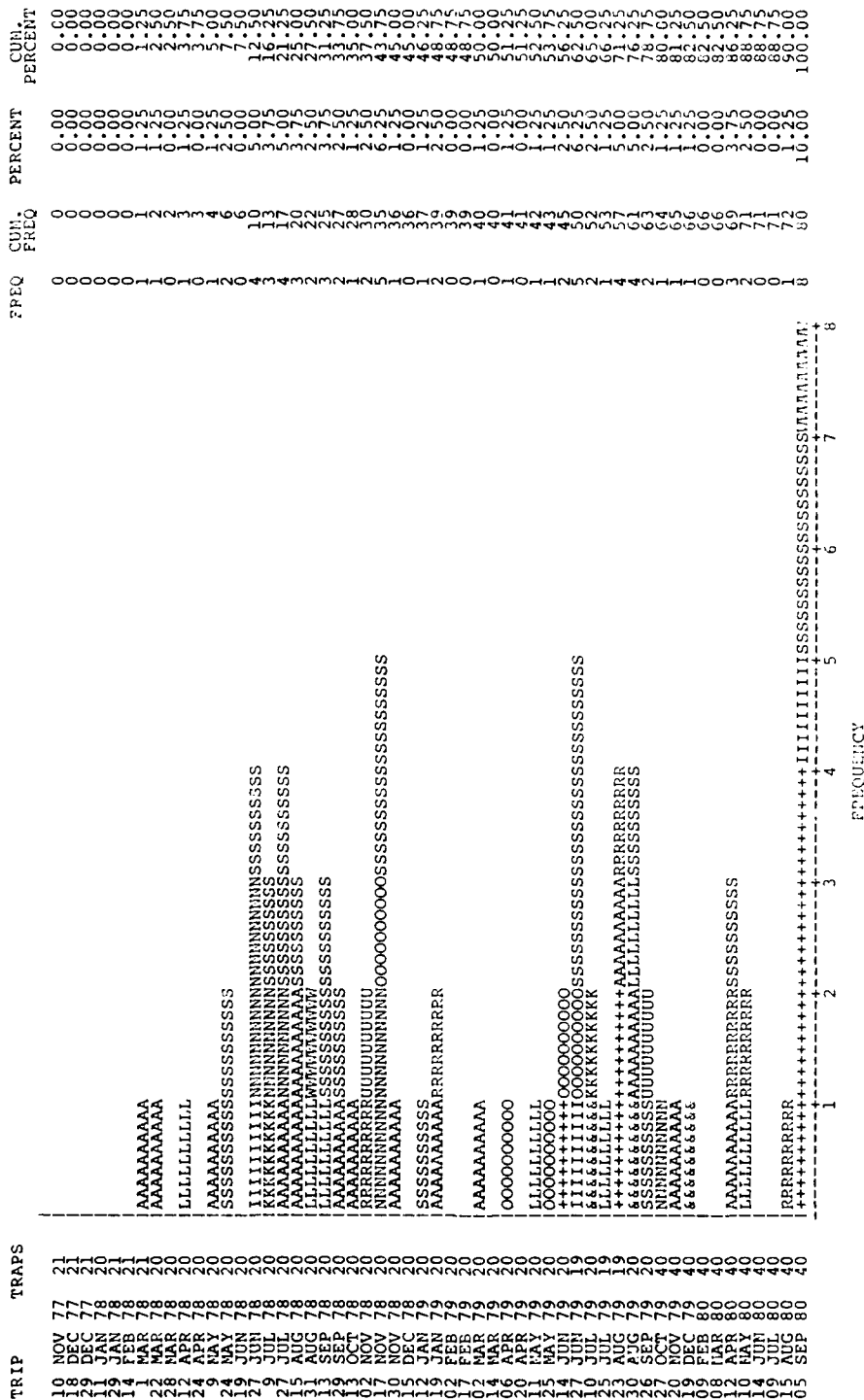


Figure 60. Funnel trap trip analysis of amphibians and reptiles on the Gatlin Canal site. Trip = date of trapping; traps = total number of traps set on a date.

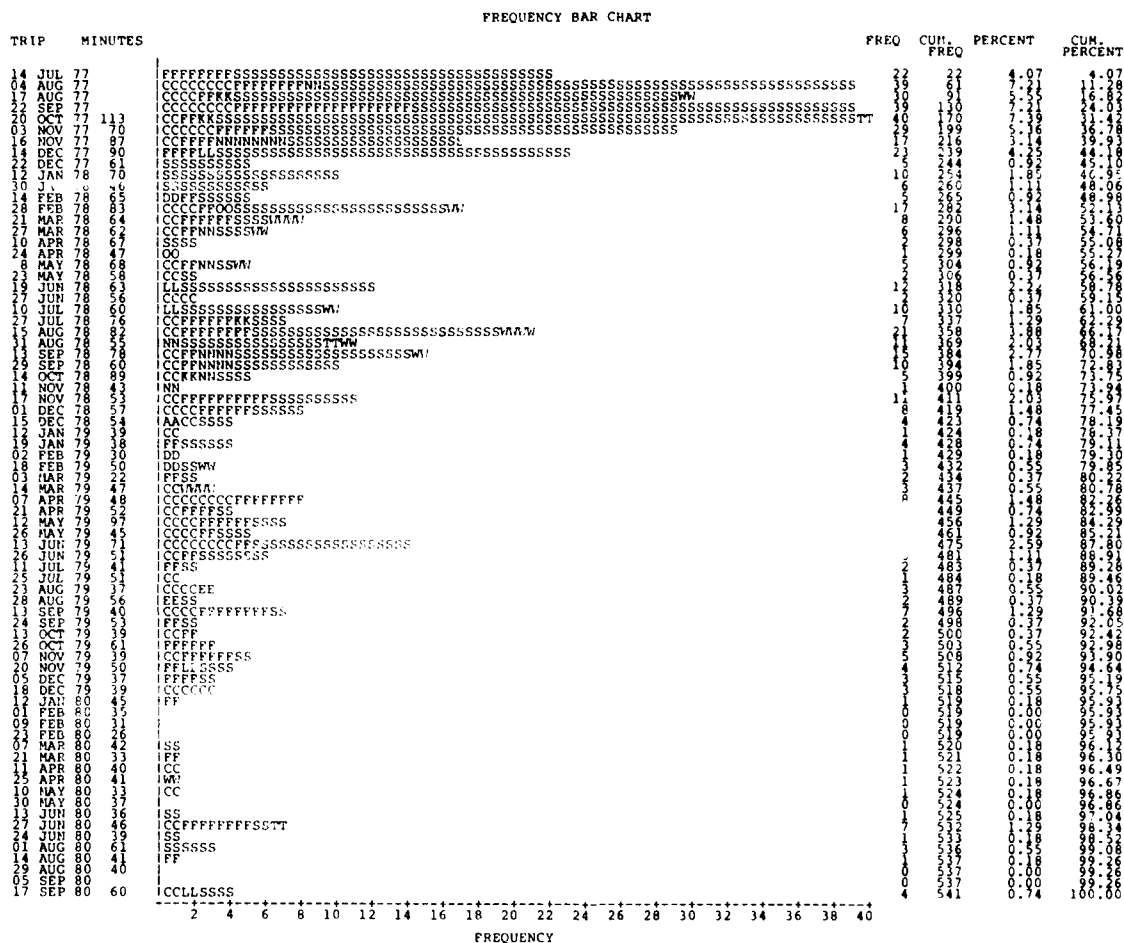


Figure 61. Herp-patrol trip analysis of salamanders and reptiles on the Gatlin Canal site. Trip = date of herp-patrol; minutes = total sampling time of a herp-patrol on a date (time not recorded prior to 13 October 1977).

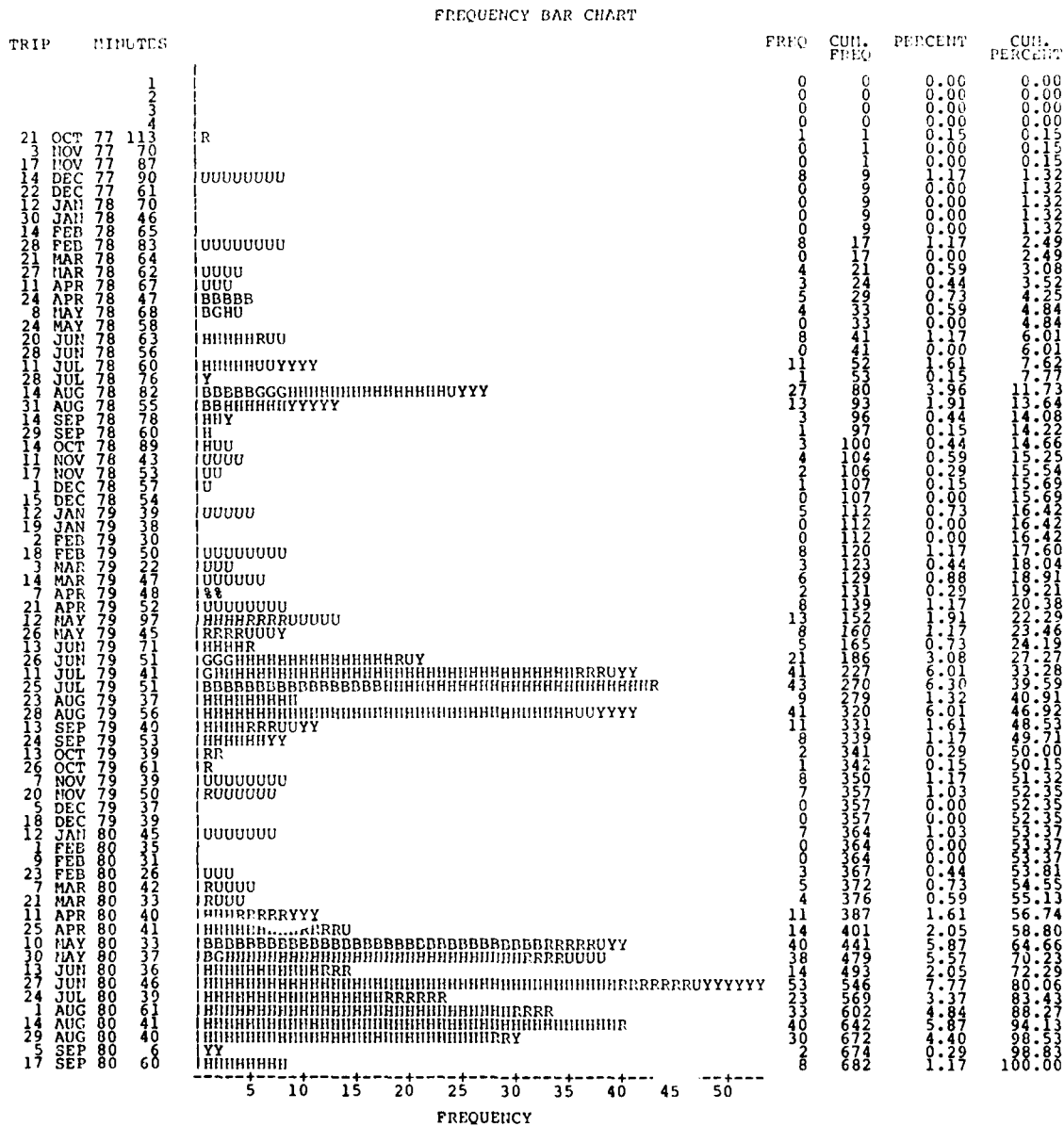


Figure 62. Herp-patrol trip analysis of calling frogs on the Gatlin Canal site. Trip = date of herp-patrol; minutes = total sampling time of a herp-patrol on a date (time not recorded prior to 13 October 1977).

APPENDIX A: SUMMARY OF PLANT SPECIES AND SUBSTRATUM TYPE FOR ALL PERMANENT SHORELINE HERPETOFAUNAL TRAPPING STATIONS ON LAKE CONWAY DURING THE FIRST POSTSTOCKING STUDY PERIOD. Given are the three most abundant plant species or habitat conditions coded in order of decreasing percent cover within a 2-sq-m area of each trapping station averaged over the January and August 1979 vegetation samples. If a significant proportion of the quadrat contained no vegetation but was in natural surroundings, it was coded as "Bare bottom"; likewise, "Beach" means man-made white sand beach, and "No other vegetation present" means that other plant species were monodominant or codominant in the quadrat. If plant cover changed as a result of man-made habitat modification during the first poststocking study period, the date of change and new conditions are given in parentheses. Substratum types are coded as follows: 1 = sand; 2 = 1-5 cm mud; 3 = 6-10 cm mud; 4 = 11-15 cm mud; 5 = 16-20 cm mud; 6 = > 20 cm mud.

Table A1

South Pool

Habitat Condition	Trap Station										
	0	10	20	30	40	50	60	70	80	90	100
<u>Eichhornia crassipes</u>	1	1	1	2							110
<u>Fuirena scirpoides</u>					2	2	2	2	2	2	3
<u>Nuphar luteum</u>											
<u>Panicum hemitomon</u>						3	3				
<u>Panicum repens</u>											
<u>Pontederia lanceolata</u>					3			3	3	3	3
<u>Typha latifolia</u>				3						2	1
No other vegetation present	2	2									
Bare bottom				1	1	1	1	1	1	1	2
Beach	(1)	(1)	(1)	(1)	(1)	(1)	(1)	(1)	(1)	(1)	
Substratum	3(1)	3(1)	3(1)	1	1	1	1	1	1	1	2
Date of habitat change	4/79	+	+	+	4/79	5/79	+	+	+	5/79	

(Continued)

(Sheet 1 of 4)

Table A1 (Continued)

Habitat Condition	Trap Station													
	120	130	140	150	160	170	180	190	200	210	220	230		
<u>Eichhornia crassipes</u>														
<u>Fuirena scirpoides</u>		1	1		1	1	3			1	1	1		
<u>Nuphar luteum</u>														
<u>Panicum hemitomom</u>			2	2	2									
<u>Panicum repens</u>														
<u>Pontederia lanceolata</u>	2			1		2	1	2	1	2	2	2		
<u>Typha latifolia</u>	1	2												
No other vegetation present								3						
Bare bottom	3	3	3	3	3	3	2	1	2	3	3	3		
Beach														
Substratum	1	1	1	1	1	1	1	1	1	1	1	1		
Date of habitat change														

(Continued)

(Sheet 2 of 4)

Table A1 (Continued)

Habitat Condition	Trap Station											
	240	250	260	270	280	290	300	310	320	330	340	350
<u>Eichhornia crassipes</u>												
<u>Fuirena scirpoides</u>									2	1	1	
<u>Nuphar luteum</u>												
<u>Panicum hemitomom</u>								2	1			
<u>Panicum repens</u>												
<u>Pontederia lanceolata</u>									3			
<u>Typha latifolia</u>												
No other vegetation present												
Bare bottom										2	2	
Beach	1	1	1	1	1		1	1				1
Substratum	1	1	1	1	1		1	1	1	1	1	1
Date of habitat change												

(Continued)

(Sheet 3 of 4)

Table A) (Concluded)

Habitat Condition	Trap Station									
	360	370	380	390	400	410	420	430	440	450 460
<u>Eichhornia crassipes</u>									1	1
<u>Fuirena scirpoides</u>	2	2		1						
<u>Nuphar luteum</u>										
<u>Panicum hemitomon</u>					2	2	2	2		
<u>Panicum repens</u>										
<u>Pontederia lanceolata</u>	1	1	2	3	1	1	1	1	1	
<u>Typha latifolia</u>										
No other vegetation present			3		3	3	3	3	2	2
Bare bottom	3	3								
Beach			1	2						
Substratum	1	2	2	2	2	2	2	2	2	3
Date of habitat change										

Table A2

Middle Pool

Habitat Condition	Trap Station									
	1000	1010	1020	1030	1040	1050	1060	1070	1080	1090
<u>Eichhornia crassipes</u>										
<u>Fuirena scirpoides</u>	2	2	2	2	2	2	2	2	2	2
<u>Nuphar luteum</u>										
<u>Panicum hemitomon</u>										
<u>Panicum repens</u>										
<u>Pontederia lanceolata</u>										
<u>Typha latifolia</u>									3	3
No other vegetation present										
Bare bottom										
Beach	1	1	1	1	1	1	1	1	1	1
Substratum	1	1	1	1	1	1	1	1	1	1
Date of habitat change										

(Continued)

(Sheet 1 of 2)

Table A2 (Concluded)

Habitat Condition	Trap Station									
	1100	1110	1120	1130	1140	1150	1160	1170	1180	1190
<u>Eichhornia crassipes</u>				1	1	1	1	1	1	1
<u>Fuirena scirpoides</u>	2	2	1							
<u>Nuphar luteum</u>										
<u>Panicum hemitomon</u>										
<u>Panicum repens</u>										
<u>Pontederia lanceolata</u>				2	2	2	2	2		
<u>Typha latifolia</u>				3	3	3	3	3		
No other vegetation present									2	2
Bare bottom										
Beach	1	1	1							
Substratum	1	1	1	3	3	3	3	3	4	4
Date of habitat change										

Table A3

East Pool

Habitat Condition	Trap Station										
	1010	1020	1030	1040	1050	1060	1070	1080	1090	1100	
<u>Eichhornia crassipes</u>	1	1			2	3	2	1	1	1	
<u>Fuirena scirpoides</u>											
<u>Nuphar luteum</u>											
<u>Panicum hemitomom</u>											
<u>Panicum repens</u>											
<u>Pontederia lanceolata</u>			1	1							
<u>Typha latifolia</u>	2	2		2	1	1	1	2	2	2	
<u>Ludwigia peruviana</u>						2	3	3	3	3	
No other vegetation present			2	3	3						
Bare bottom											
Beach											
Substratum	3	3	2	2	2	2	3	4	5	5	
Date of habitat change											

(Continued)

(Continued)

(Sheet 1 of 2)

Table A3 (Concluded)

Habitat Condition	Trap Station										
	1110	1120	1130	1140	1150	1160	1170	1180	1190	1200	
<u>Eriogonum crassipes</u>	1	1	1	1	1	1	1	1	1	1	
<u>Purpurea scirpoides</u>											
<u>Nyctar luteum</u>											
<u>Panicum hemitomon</u>											
<u>Panicum repens</u>											
<u>Fontinalis lanceolata</u>											
<u>Typha latifolia</u>	2	2		2							
<u>Ludwigia peruviana</u>	3						2	2	2	2	
No other vegetation present		3	3	3	2	2	3	3	3	3	
Bare bottom											
beach											
Substratum	5	2	2	3	2	3	3	4	5	5	
Date of habitat change											

Table A4

West Pool

Habitat Condition	Trap Station										
	0	10	20	30	40*	60	70	80	90	100	110
<u>Eichhornia crassipes</u>									1	1	1
<u>Fuirena scirpoides</u>	2	2	1	1	1	1	1				
<u>Nuphar luteum</u>											
<u>Panicum hemitomon</u>			2	2	2			1			
<u>Panicum repens</u>			3	3	3	2	2				
<u>Pontederia lanceolata</u>								2	2	2	
<u>Typha latifolia</u>											2
No other vegetation present						3	3	3	3	3	2
Bare bottom	1	1									
Beach											
Substratum	1	1	1	1	1	1	1	2	2	3	2
Date of habitat change											

(Continued)

*Station 50 was located under a boat dock and this trap was not set.

(Sheet 1 of 3)

Table A4 (Continued)

	Trap Station														
Habitat Condition	140	150	160	170	180	190	200	210	220	230	240	250			
<u>Echinochloa polystachya</u>	1	1	2	2	3	1	2		1	1	1	2			
<u>Eleusine indica</u>															
<u>Nyctaginia</u>															
<u>Panicum homitomon</u>			3				3	2							
<u>Panicum repens</u>		2													
<u>Pontederia lanceolata</u>	2				2	2	1	1	2		2	1			
<u>Typha latifolia</u>		1	1	1											
No other vegetation present	3	3		3		3		3							
Bar. bottom										2		3			
Beach															
Substratum	2	3	2	2	2	3	2	2	4	4	4	2			
Date of habitat change															

(Continued)

(Sheet 2 of 3)

Table A4 (Continued)

Habitat Condition	Trap Station											
	260	270	280	290	300	310	320	330	340	350	360	370
<u>Eichhornia crassipes</u>	1	3	2	3	1	1	1		2	1	1	1
<u>Eurkea scarpoides</u>												
<u>Nypa lutea</u>												
<u>Panicum hemitomon</u>								2	3		2	3
<u>Panicum repens</u>												
<u>Pontederia lanceolata</u>	2	2	1	1				1	1	2		
<u>Typha latifolia</u>	3	1	3	2	2	2	2				3	2
No other vegetation present					3	3	3	3		3		
Bar... bottom												
Beach												
Substratum	2	2	2	3	4	5	4	2	3	3	3	2
Date of habitat change												

Table A5

Gatlin Canal

Habitat Condition	Trap Station										
	0	10	20	30	40	50	60	70	80	90	100
<u>Eichhornia crassipes</u>											
<u>Fuirena scirpoides</u>											
<u>Nuphar luteum</u>											
<u>Panicum hemitomon</u>											
<u>Panicum repens</u>											
<u>Pontederia lanceolata</u>											
<u>Typha latifolia</u>											
No other vegetation present											
Bare bottom	1										
Beach											
Substratum	2	3	4	4	3	2	2	2	3	2	2
Date of habitat change											

(Continued)

Table A5 (Concluded)

Habitat Condition	Trap Station									
	120	130	140	150	160	170	180	190		
<u>Eichhornia crassipes</u>					3	1	1	2		
<u>Fuirena scirpoides</u>										
<u>Nuphar luteum</u>	3									
<u>Panicum hemitomon</u>										
<u>Panicum repens</u>	2	2	2	2	1			1		
<u>Pontederia lanceolata</u>	1	1	1	1						
<u>Typha latifolia</u>					2	2	2	3		
No other vegetation present		3	3	3						
Bare bottom										
Beach										
Substratum	3	3	3	3	2	3	3	3		
Date of habitat change										

APPENDIX B: SUMMARY OF PLANT SPECIES AND SUBSTRATUM TYPE FOR ALL PERMANENT SHORELINE HERPETOFAUNAL TRAPPING STATIONS ON LAKE CONWAY DURING THE SECOND POSTSTOCKING STUDY PERIOD. Given are the three most abundant plant species or habitat conditions coded in order of decreasing percent cover within a 2-sq-m area of each trapping station averaged over the January and August 1980 vegetation samples. If a significant proportion of the quadrat contained no vegetation but was in natural surroundings, it was coded as "Bare bottom"; likewise, "Beach" means man-made white sand beach; and "No other vegetation present" means that other plant species were monodominant or codominant in the quadrat. If plant cover changed as a result of man-made habitat modification during the second poststocking study period, the date of change and new conditions are given in parentheses. Substratum types are coded as follows:

1 = sand; 2 = 1-5 cm mud; 3 = 6-10 cm mud; 4 = 11-15 cm mud; 5 = 16-20 cm mud; 6 = > 20 cm mud.

Table B1

South Pool

<u>Habitat Condition</u>	<u>Trap Station</u>										
	0	10	20	30	40	50	60	70	80	90	100 110
<u>Eichhornia crassipes</u>											3 2
<u>Puirena scirpoides</u>											
<u>Ludwigia peruviana</u>											2 3
<u>Nuphar luteum</u>											
<u>Panicum hemitomom</u>											
<u>Panicum repens</u>											
<u>Pontederia lanceolata</u>							2				
<u>Typha latifolia</u>											1 1
No other vegetation present	2	2	2	2	2	2	3	2	2	2	
Bare bottom											
Beach	1	1	1	1	1	1	1	1	1	1	
Substratum	1	1	1	1	1	1	1	1	1	1	2 2
Date of habitat change											

(Continued)

(Sheet 1 of 5)

Table B1 (Continued)

Habitat Condition	Trap Station											
	120	130	140	150	160	170	180	190	200	210	220	
<u>Eichhornia crassipes</u>	1	1	1						2			
<u>Fuirena scirpoides</u>				1	1	1	3			1	1	
<u>Ludwigia peruviana</u>	1								3	3		
<u>Nuphar luteum</u>												
<u>Panicum hemitomom</u>			2	2	2							
<u>Panicum repens</u>												
<u>Pontederia lanceolata</u>				1		2	1	2	1	2	2	
<u>Typha latifolia</u>	3	3										
No other vegetation present			2	3	3	3		3			3	
Bare bottom							2	1				
Beach			(1)	(1)	(1)	(1)	(1)	(1)	(1)	(1)	(1)	
Substratum	2	1	1	1	1	1	1	1	2	2	2	
Date of habitat change			5/80	→	→	→	→	5/80	9/80	→	9/80	

(Continued)

(Sheet 2 of 5)

Table B1 (Continued)

Habitat Condition	Trap Station										
	230	240	250	260	270	280	290	300	310	320	330
<u>Eichhornia crassipes</u>											
<u>Eurena scirpoides</u>	1						2	3	1	1	1
<u>Ludwigia peruviana</u>	3								2	2	2
<u>Nuphar luteum</u>											
<u>Panicum hemitomom</u>								2			
<u>Panicum repens</u>											
<u>Pontederia lanceolata</u>	2										3
<u>Typha latifolia</u>											
No other vegetation present		2	2	2	2	2	2	3		3	
Bare bottom										1	1
Beach	(1)	1	1	1	1	1	1	1	1		
Substratum	2	2	2	2	2	2	2	1	1	1	1
Date of habitat change	9/80										

(Continued)

(Sheet 3 of 5)

Table B1 (Continued)

Habitat Condition	Trap Station										
	340	350	360	370	380	390	400	410	420	430	440
<u>Eichhornia crassipes</u>					1	1		1	1	1	1
<u>Fuirena scirpoides</u>	1		2	3		1	1				
<u>Ludwigia peruviana</u>	2						3				
<u>Nuphar luteum</u>											
<u>Panicum hemitomom</u>					2		2				
<u>Panicum repens</u>											
<u>Pontederia lanceolata</u>	3		3	2		2		2	2	2	3
<u>Typha latifolia</u>											2
No other vegetation present		2			3			3	3	3	
Bare bottom	1		1	1							
Beach		1			1	3					
Substratum	1	1	1	2	2	1	2	2	2	2	2
Date of habitat change											

(Continued)

(Sheet 4 of 5)

Table B1 (Concluded)

Trap Station

Habitat Condition	450	460
<u>Eichhornia crassipes</u>	1	1
<u>Fuirena scirpoides</u>		
<u>Ludwigia peruviana</u>		
<u>Nuphar luteum</u>		
<u>Panicum hemitomom</u>		
<u>Panicum repens</u>		
<u>Pontederia lanceolata</u>	2	
<u>Typha latifolia</u>		
No other vegetation present	3	2
Bare bottom	(1)	(1)
Beach		
Substratum	3 (2)	3 (2)
Date of habitat change	5/80	5/80

Table B2

Habitat Condition	Middle Pool											Trap Station
	1000	1010	1020	1030	1040	1050	1060	1070	1080	1090	1100	1110
<u>Eichhornia crassipes</u>												
<u>Fuirena scirpoides</u>	2	2	2	2	2	2	2	1	1	1	1	1
<u>Nuphar luteum</u>												
<u>Panicum hemitomon</u>												
<u>Panicum repens</u>												
<u>Pontederia lanceolata</u>								3	2	2	3	3
<u>Typha latifolia</u>												
No other vegetation present												
Bare bottom												
Beach	1	1	1	1	1	1	1	2	3	3	2	2
Substratum	1	1	1	1	1	1	1	1	1	1	1	1
Date of habitat change												

(Continued)

(Sheet 1 of 2)

Table B2 (Concluded)

Habitat Condition	Trap Station									
	1120	1130	1140	1150	1160	1170	1180	1190		
<u>Eichhornia crassipes</u>		1	1	1	1	1	1	1		
<u>Fuirena scirpoides</u>	1									
<u>Nuphar luteum</u>										
<u>Panicum hemitomom</u>										
<u>Panicum repens</u>										
<u>Pontederia lanceolata</u>		2	2	2	2	2				
<u>Typha latifolia</u>	3				3	3				
No other vegetation present		3	3	3			2	2		
Bare bottom										
Beach	2									
Substratum	1	3	3	3	3	3	3	3		
Date of habitat change										

Table B3

East Pool

Habitat Condition	Trap Station												
	1010	1020	1030	1040	1050	1060	1070	1080	1090	1100	1110	1120	1130
<u>Eichhornia crassipes</u>	1	1	1	2	3	1	1	1	1	1	1	1	1
<u>Fuirena scirpoides</u>													
<u>Ludwigia peruviana</u>						3	3	2	2	2	2	2	
<u>Nuphar luteum</u>													
<u>Panicum hemitomon</u>													
<u>Panicum repens</u>													
<u>Pontederia lanceolata</u>			2	3		2	2						
<u>Sagittaria lancifolia</u>													
<u>Typha latifolia</u>	2	2	3	1	1								2
No other vegetation present	3	3						3	3	3	3	3	3
Bare bottom													
Beach													
Substratum	3	3	2	2	2	2	2	2	2	2	2	2	2
Date of habitat change													

(Continued)

(Sheet 1 of 4)

Table B3 (Continued)

Habitat Condition	Trap Station													
	1140	1150	1160	1170	1180	1190	1200	1210	1220	1230	1240	1250	1260	
<u>Eichhornia crassipes</u>	1	1	1	1	1	1	1	1	1	1	1	1	1	
<u>Fuirena scirpoides</u>														
<u>Ludwigia peruviana</u>				2				2	2	2	2	2	2	
<u>Nuphar luteum</u>														
<u>Panicum hemitomon</u>														
<u>Panicum repens</u>														
<u>Pontederia lanceolata</u>														
<u>Sagittaria lancifolia</u>														
<u>Typha latifolia</u>														
No other vegetation present	2	2	2	3	2	2	2	3	3	3	3	3	3	
Bare bottom														
Beach														
Substratum	2	2	2	3	2	2	2	2	2	2	3	3	3	
Date of habitat change														

(Continued)

(Sheet 2 of 4)

(Continued)

(Sheet 2 of 4)

Table B3 (Continued)

Habitat Condition	1270	1280	1290	1300	1310	1320	1330	1340	1350	1360	1370	1380	1390
<u>Eichhornia crassipes</u>	1	1	1	1	1	1	1	1	1	1	1	1	1
<u>Fuirena scirpoides</u>													
<u>Ludwigia peruviana</u>	2	2		2	2	2	3	3	2		2	2	3
<u>Nuphar luteum</u>													
<u>Panicum hemitomom</u>													
<u>Panicum repens</u>													
<u>Pontederia lanceolata</u>							2						
<u>Sagittaria lancifolia</u>								2		2			2
<u>Typha latifolia</u>													
No other vegetation present	3	3	2	3	3	3			3	3	3	3	
Bare bottom													
Beach													
Substratum	3	4	4	3	3	4	3	3	5	4	3	4	3
Date of habitat change													

(Continued)

(Continued)

(Sheet 3 of 4)

Table B3 (Continued)

Habitat Condition	Trap Station							
	1400	1410	1420	1430	1440	1450	1460	1470
<u>Eichhornia crassipes</u>					2	1	1	1
<u>Fuirena scirpoides</u>								
<u>Ludwigia peruviana</u>	3				2	2	2	2
<u>Nuphar luteum</u>								
<u>Panicum hemitomon</u>								
<u>Panicum repens</u>								
<u>Pontederia lanceolata</u>	2				3			
<u>Sagittaria lancifolia</u>		2	2	2		3		
<u>Typha latifolia</u>			3	3				
No other vegetation present		3				3		
Bay bottom	1	1	1	1	1	1		3
Beach								
Substratum	3	3	3	3	3	3	2	3
Date of habitat change								

Table B4

West Pool

Habitat Condition	Trap Station											
	0	10	20	30	40*	50	60	70	80	90	100	110 120
<u>Eichhornia crassipes</u>						1			1	1	1	1
<u>Fuirena scirpoides</u>	2	2	2	2	2	2		2				
<u>Ludwigia peruviana</u>										2	2	
<u>Nuphar luteum</u>												
<u>Panicum hemitomon</u>												
<u>Panicum repens</u>				3	3	1	1	1				
<u>Pontederia lanceolata</u>									2			
<u>Typha latifolia</u>												2
No other vegetation present	3	3	3		3	3	3	3	3	3	3	3
Bare bottom												
Beach	1	1		1	1							
Substratum	1	1	1	1	1	1	1	1	2	2	2	3
Date of habitat change												

(Continued)

*Station 50 was located under a boat dock and this trap was not set.

(Sheet 1 of 4)

Table B4 (Continued)

Habitat Condition	Trap Station											
	130	140	150	160	170	180	190	200	210	220	230	
<u>Eichhornia crassipes</u>	1	1	1	1	2	2			3			
<u>Fuirena scirpoides</u>												
<u>Ludwigia peruviana</u>				2			3	2		2	2	
<u>Nuphar luteum</u>												
<u>Panicum hemitomom</u>												
<u>Panicum repens</u>												
<u>Pontederia lanceolata</u>							2		2			
<u>Typha latifolia</u>	2			3	1	1						
No other vegetation present	3	2	2		3			3		3	3	
Bare bottom						3	1	1	1	1	1	
Beach												
Substratum	4	3	2	2	3	2	1	2	2	2	2	
Date of habitat change												

(Continued)

(Sheet 2 of 4)

Table B4 (Continued)

Habitat Condition	Trap Station										
	240	250	260	270	280	290	300	310	320	330	
<u>Eichhornia crassipes</u>		3	1	1	1	1	1	1	1	1	1
<u>Fuirena scirpoides</u>											
<u>Ludwigia peruviana</u>	2	1	2		2	2	3	2	2	2	2
<u>Nuphar luteum</u>											
<u>Panicum hemitomom</u>											
<u>Panicum repens</u>											
<u>Pontederia lanceolata</u>											
<u>Typha latifolia</u>											
No other vegetation present	3			2	3	3		3			
Bare bottom	1	2	3				2		3	3	
Beach											
Substratum											
Date of habitat change	2	2	2	3	2	2	3	2	2	2	2

(Continued)

(Sheet 3 of 4)

Table B4 (Concluded)

Habitat Condition	Trap Station		
	340	350	360 370
<u>Eichhornia crassipes</u>	2	2	2
<u>Fuirena scirpoides</u>			
<u>Ludwigia peruviana</u>	3		
<u>Nuphar luteum</u>			
<u>Panicum hemitomon</u>			
<u>Panicum repens</u>			
<u>Pontederia lanceolata</u>			
<u>Typha latifolia</u>			
No other vegetation present		3	2 3
Bare bottom	1	1	1 1
Beach			
Substratum	2	1	1 1
Date of habitat change			

Table B5

Gatlin Canal

Habitat Condition	Trap Station											
	0	10	20	30	40	50	60	70	80	90	100	110
<u>Eichhornia crassipes</u>		1	1	1						2	3	
<u>Fuirena scirpoides</u>												
<u>Nuphar luteum</u>		2	2	2	1	2	2	3				
<u>Panicum hemitomom</u>												
<u>Panicum repens</u>			3	3	3	2	1	1	1	1	2	2
<u>Pontederia lanceolata</u>										3	1	1
<u>Typha latifolia</u>							3	2	2			
No other vegetation present	2				3	3			3			3
Bare bottom	1											
Beach												
Substratum												
Date of habitat change	2	2	4	4	3	2	2	2	3	3	2	2

(Continued)

(Sheet 1 of 2)

Table B5 (Concluded)

Habitat Condition	Trap Station								
	120	130	140	150	160	170	180	190	
<u>Eichhornia crassipes</u>	3				3	1	1	3	
<u>Fuirena scirpoides</u>									
<u>Nuphar luteum</u>		3							
<u>Panicum hemitomom</u>									
<u>Panicum repens</u>	2	2	2	2	1			1	
<u>Pontederia lanceolata</u>	1	1	1	1					
<u>Typha latifolia</u>					2	2	2	2	
No other vegetation present			3	3		3	3		
Bare bottom									
Beach									
Substratum	2	3	3	2	2	3	3	2	
Date of habitat change									

DATE
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28