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# AQUATIC FIELD SURVEYS AT VOLUNTEER ARMY AMMUITION PLANT CHATTANOOGA, TENNESSEE

# FINAL REPORT

J.H.SULLIVAN, JR., H.D.PUTNAM, M.A.KEIRN,

D.R.SWIFT AND B.C. PRUITT, JR.

**JUNE, 1977** 

SUPPORTED BY

U.S. ARMY RESEARCH AND DEVELOPMENT COMMAND WASHINGTON, D.C. 20314

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reference bays unaffected by munitions plant effluent. Both chemical and biological analyses were conducted. Water and sediment samples were analyzed for major ions, nutrients, and munitions residues. Biological components selected were periphyton, phytoplankton, and benthic macroinvertebrates.

Data gathered during the summer show that nitrobodies, nitrogen compounds, and dissolved solids discharged in the VAAP waste maintain a chemical environment in upper Waconda Bay significantly different from the other study areas., Distinct gradients were observed for conductivity, hardness, sulfate, and chlorides beginning where wastes entered the receiving waters to downbay areas.

TV Munitions residues during the survey ranged to 345 ppb at the point of waste entry to near zero, one-half to three-quarters of a mile downbay. The data reported as the median and concentration ranges for 2,4-DNT, 2,6-DNT, and TNT showed no consistent pattern for any of the three compounds related to distance from the outfall. Dinitrotoluene compounds, however, made up the largest portion of the total residue concentration.

—Biologic response by periphyton and macroinvertebrates was observed in similar areas of Waconda Bay where chemical characteristics were altered. Diatoms, a major part of the periphyton assemblage, were more sensitive to munitions wastes than were macroinvertebrates. In both surveys, population densities of diatoms on artificial substrates were reduced at the bayhead and elevated where NO<sub>3</sub>-N levels promoted biostimulation.

Because the observed biologic responses were to a mixed waste input, it was not possible to determine precisely the concentration of the individual materials causing the responses. However, at the bayhead where toxicity was noted in both the periphyton and macrobenthic communities, median munitions concentrations ( $\alpha$ -TNT + 2,4-DNT + 2,6-DNT) in June and August were 123 and 56 ppb, respectively, with individual samples as high as 345 ppb. Little reduction was noted in the concentrations of the specific munitions measured from the outfall to distances downbay of approximately three-eighths of a mile. Since the biologic response significantly shifted from toxic to biostimulatory, it is unlikely that the toxicity was due specifically to any of these three compounds which persisted in the environment. Nevertheless, it was observed that when munitions concentrations dropped below 20 ppb, no further biologic responses were evident. At munitions concentrations between 40 and 80 ppb, slight biostimulatory effects were noted.

Based on these results, it is concluded that environmental impact of TNT waste effluent would be minimal if the combined concentration of lpha-TNT, 2,4-DNT, and 2,6-DNT did not exceed 20 ppb in the receiving waters.

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#### AQUATIC FIELD SURVEY

AT

VOLUNTEER ARMY AMMUNITION PLANT, CHATTANOOGA, TENNESSEE

#### FINAL REPORT

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JUNE, 1977

#### SUPPORTED BY

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  - J. GARETH PEARSON, PROJECT OFFICER

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#### **EXECUTIVE SUMMARY**

In June and August, 1975, water quality surveys were conducted to determine the impact of liquid wastes from Volunteer Army Ammunition Plant at Chattanooga, Tennessee. This TNT manufacturing plant discharges wastes into Waconda Bay which is a part of the Lake Chickamauga Reservoir on the Tennessee River. The purpose of the study was to develop information that would aid in the rational development of effluent standards for the munitiors industry.

In each of the two surveys, sampling was carried out at 20 locations in Waconda Bay and in two similar reference bays unaffected by munitions-plant effluent. Both chemical and biological analyses were conducted. Water and sediment samples were analyzed for major ions, nutrients, and munitions residues. Biological components selected were periphyton, phytoplankton, and benthic macroinvertebrates.

The relative sizes of Waconda Bay, Reference Bay A, and Huss Lowe Slough are 235, 52, and 71 acres, respectively. Mean depth is about 10 feet for all three bays. Stormwater runoff from residential areas and VAAP enter Waconda Bay. The watershed of Reference Bay A is largely residential and that of Huss Lowe Slough forested.

Flushing of the three bays is caused by: 1) runoff, 2) stage fluctuations, and, in the case of Waconda Bay, 3) effluent from VAAP. Drainage areas for Waconda Bay and Huss Lowe Slough are each approximately nine times the area of the bay. For Reference Bay A the drainage basin is about 21 times the bay area. Present estimates for flushing of Waconda Bay are about 20 percent per month.

One week prior to the June survey the plant was closed by a labor dispute. Production was not resumed until after the August survey. During May, however, plant production was 1.3 million pounds of TNT. Effluent from this manufacture caused significant water quality changes in Waconda Bay. Data gathered during the summer show that nitrobodies, nitrogen compounds, and dissolved solids discharged in the VAAP waste maintain a chemical environment in upper Waconda Bay significantly different from the other study areas. Distinct gradients were observed for conductivity, hardness, sulfate, and chlorides beginning at Station A (effluent input) to downbay areas. Effluent from the plant appeared to move principally along the west shore of Waconda Bay. The data reflect this trend which is consistent with the morphometry of the upper bay, viz. deep water to the west and shallow water to the east.

Periodic discharges of TNT and nitrobodies historically have exceeded the 0.3 mg/l NPDES permit limits. Analyses conducted for monitoring purposes in 1974 showed TNT levels ranging above 1 mg/l in

the upper bay. Munitions residues during the 1975 June and August survey ranged to 345 ppb at Station A to near zero, one-half to three-quarters of a mile downbay. The data reported, as the median value and concentration ranges for 2,4-DNT, 2,6-DNT, and TNT, showed no consistent pattern for any of the three compounds related to distance from the outfall. Dinitrotoluene compounds, however, made up the largest portion of the total residue concentration. Highest levels of compounds were always observed on the west side of Waconda Bay.

Significant nitrogen enrichment occurred in Waconda Bay. Gradients of nitrate, nitrite, organic nitrogen, and ammonia were observed in the upper bay with maximum concentrations up to 5 mg N/l. Elevated levels of these substances existed from the waste entry to a point one-half mile downbay.

Selected heavy metal analyses were made which included iron, lead, cadmium, hexavalent chromium, copper, nickel, and zinc. Results indicated that the concentration was insufficient to cause biotoxicity.

Biologic response by periphyton and macroinvertebrates was observed in similar areas of Waconda Bay where chemical characteristics were altered. At Station A, the bayhead, toxic effects were noted in both communities. Diatoms, a major part of the periphyton assemblage, were more sensitive to munitions wastes than were macroinvertebrates. In both the June and August data, population densities of diatoms on artificial substrates were reduced at Station A and elevated at Station B-1 and B-2. Analysis of variance showed significant differences between station means in three of the four samplings.

Diatom community structure was altered at the bayhead relative to the other study areas. One hundred species of diatoms representing 23 genera were recorded from the artificial substrates placed in Lake Chickamauga in June. Achnanthes-Fragilaria-Synedra-Gomphonema was the most common diatom association. At Station A this pattern shifted to Achnanthes-Nitzschia-Synedra. In terms of total numbers of species, Station A had the least (14) compared to an average of 36 for Waconda Bay, 37 for Huss Lowe Slough, and 28 for Reference Bay A. Shifts in diatom species associations, viz. the reduction of species number and the increase of those pollutant tolerant, correlate with total munitions residues and NO<sub>3</sub>-N at Station A.

Data for chlorophyll a and biomass for periphyton collected on artificial substrates generally agreed with the cell count data. In both June and August, Station A had low chlorophyll a and biomass as compared to all other stations. Correspondingly, Stations 8-1 and B-2 generally showed the highest levels of chlorophyll a and biomass. Analysis of variance showed that in most cases for the June and August 2-week and 4-week incubation periods the differences in mean values between stations for these parameters were significant. Further statistical analysis using Tukey's least significant difference technique showed that in most cases Station A was significantly different from Stations B-1 and B-2. These data suggest that inhibition of microbial growth occurs in the area of the waste outfall. Biomass of both autotrophs and heterotrophs is reduced

although the former is suppressed to a greater degree. The increase in cell counts, biomass, and chlorophyll  $\underline{a}$  at Stations B-1 and B-2 suggest biostimulation.

Data from Hester-Dendy artificial substrates and bay sediments showed that chironomids and oligochaetes were the predominate macroinvertebrates. Chironomids preferentially colonized the artificial substrates and accounted for nearly 80 percent of the total population. Oligochaetes attained higher populations in the sediments. As observed in the periphyton, the density and number of macroinvertebrate taxa were lowest at Station A ranging from 100 to 400 and 6 to 13, respectively, on Hester-Dendy units. Effects of the munitions waste were also evident at Station B via a biostimulation response of photosynthetic autotrophs. Algae colonizing the H-D plates served as food for chironomids and populations increased two orders of magnitude over Station A.

Invertebrates inhabiting natural substrates were reduced in number of organisms and species in the area of the outfall suggesting inhibition by VAAP waste. Results of both surveys indicate a residual toxicity in bay sediments which continued during the summer period following plant shutdown in May. Examination of the data suggests that recovery in natural substrates lags behind that on the Hester-Dendy plates and that the increase in chironomids associated with primary production at Station B was not reflected in the sediments. Sediment chemistry shows munitions residues at Stations A and B. Other materials from VAAP likely deposit in the sediments contributing either by themselves to growth inhibition or behaving synergistically with other compounds.

An analysis of the phytoplankton showed a total of 71 species as representative of the study area. Algal association shifted during the investigation. During June diatoms were the predominant groups with common to dominant species of Melosira ambigua, M. distans, and Fragilaria crotonensis. In August, blue-green algae dominated the plankton. Important species were Schizothrix calicola and Anacystis incerta.

Phytoplankton populations (cells/ml) were generally higher during the second survey. Mean cell densities ranged from approximately 1,300 to 2,600 in June as compared to 2,100 to 5,400 in August. Application of an analysis of variance and Tukey's test show that a significant difference at the l percent level existed during August. Since the cell number was higher in these upper bay stations, biostimulation from available nitrogen is suggested. In June cell counts were slightly reduced at Station A and elevated at Stations B-1, B-2, and C-1. However, these differences were not statistically significant. Possibly the combined effects of toxicity and biostimulation were more nearly balanced. By August, the toxicity factors had diminished in the absence of new loading, and phytoplankton populations were stimulated by the residual high level of nitrogen compounds. This trend was also reflected in the other biological compartments examined.

The response is more subtle in the phytoplankton, and, in terms of the number of plankton species recorded per station, no significant trends were observed that could be attributable to munitions waste. Mean values for numbers of species per station ranged from 44 to 53 during the June survey and from 52 to 55 species per station for the August survey. The highest value recorded through the study was 71 species at Station B-1 (August 11, 1975); the lowest, 35 species at Station U-2 (June 9, 1975).

Because the observed biologic responses were to a mixed waste input, it is not possible to determine precisely the concentrations of the individual materials causing the responses. However, at the bayhead where toxicity was noted in both the periphyton and macrobenthic communities, median munitions concentrations ( $\alpha$ -TNT + 2,4-DNT + 2,6-DNT) in June and August were 123 and 56 ppb, respectively, with individual samples as high as 345 ppb. Little reduction was noted in the concentrations of the specific munitions measured from the bayhead to transect B. Since the biologic response significantly shifted from toxic to biostimulatory, it is unlikely that the toxicity was due specifically to any of these three compounds which persisted in the environment. Nevertheless, it was observed that when munitions concentrations dropped below 20 ppb, no further biologic responses were evident. At munitions concentrations between 40 and 80 ppb, as at transect C, slight biostimulatory effects were noted.

Based on these results, it is concluded that environmental impact of TNT waste effluent would be minimal if the combined concentration of  $\alpha$ -TNT, 2,4-DNT, and 2,6-DNT did not exceed 20 ppb in the receiving waters.

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#### INTRODUCTION

The U.S. Army Medical Research and Development Command has supported field and laboratory research for the development of environmental standards for munitions related residues. A significant portion of this effort has been directed toward field assessments at the various munitions facilities within the United States. These studies have been for the purpose of evaluating impact on the various biotic compartments in freshwater systems in order to establish effluent levels in receiving waters consistent with the maintenance of environmental quality.

To meet these objectives, Water and Air Research, Inc. conducted field investigations at Volunteer Army Ammunition Plant during the summer of 1975.

Volunteer Army Ammunition Plant is a TNT manufacturing facility located northwest of Chattanooga, Tennessee (Figures 1 and 2). This facility is a 7300 acre contractor-operated government munitions manufacturing plant that produces trinitrotoluene. A portion of the reservation is leased to Farmers Chemical Associates, Inc., manufacturers of nitric acid, ammonia and nitrogenous fertilizers. Wastewater from VAAP drains northward into a series of treatment lagoons and is discharged into the head of Waconda Bay after undergoing PH adjustment with lime.

The purpose of this investigation was to define the environmental impact of effluent from Volunteer Army Ammunition Plant on Lake Chickamauga during the summer period. The assessment was conducted using selected physical, biological and chemical techniques for the development of guidelines to prevent water quality degradation.

The scope of work included both biologic and chemical sampling during the months of June and August. Water and sediments were characterized to include major ions, nutrient concentrations, and munitions specific residue. Biological compartments selected for examination were periphyton, phytoplankton, and benthic macroinvertebrates.

Since Waconda Bay is impacted by munitions effluents, an intensive survey was conducted in this bay. Two additional bays were selected as reference areas; these were Huss Lowe Slough and the un-named bay immediately east of Waconda Bay (Reference Bay A).

In the selection of the reference bays several factors were considered including the bay's area and shape, depth, physical orientation (north-south axis, etc.) and drainage basin. The land use of the basin, its location relative to munitions discharge, as well as its flushing and sediment characteristics were also considered.

The principal differences in the three bays can be described as follows:

Waconda Bay -- receives munitions wastes plus runoff from residential and industrial areas.

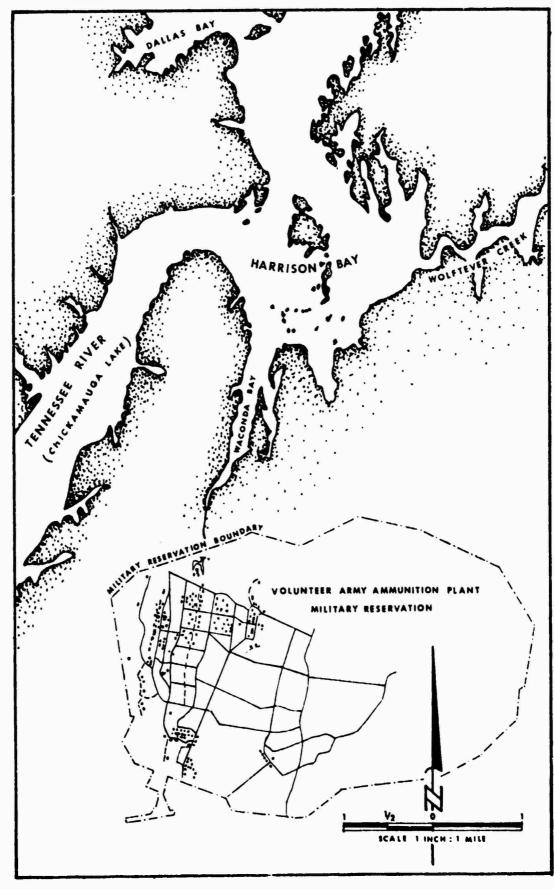


FIGURE 1. VICINITY MAP OF VAAP STUDY AREA.

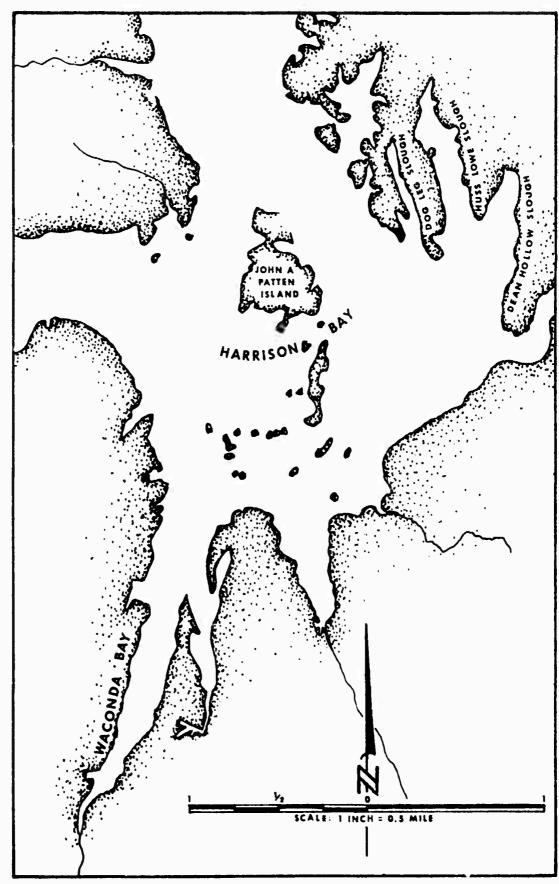


FIGURE 2. LOCALITY MAP OF VAAP STUDY AREA.

Reference Bay A -- receives only runoff from residential areas

Huss Lowe Slough -- receives only runoff from forested plus small residential area

Twenty sampling stations were selected within the three bays (Figures 3 and 4). A brief description of each site is given in Table 1.

The combined lune and August surveys, considering as they did effluent discharge, distribution, mixing characteristics, changes and/or alterations in the chemical, biological or sediment characteristics of the three bays provide a data base for the recommendation of effluent guidelines for munitions wastes into receiving bodies.

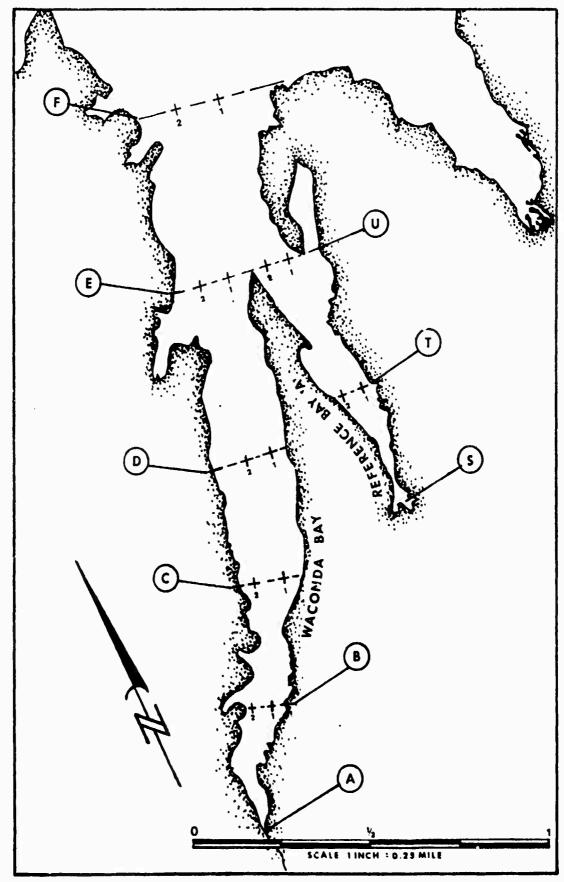


FIGURE 3. SAMPLING STATIONS IN WACONDA BAY AND ADJACENT REFERENCE BAY 'A.'

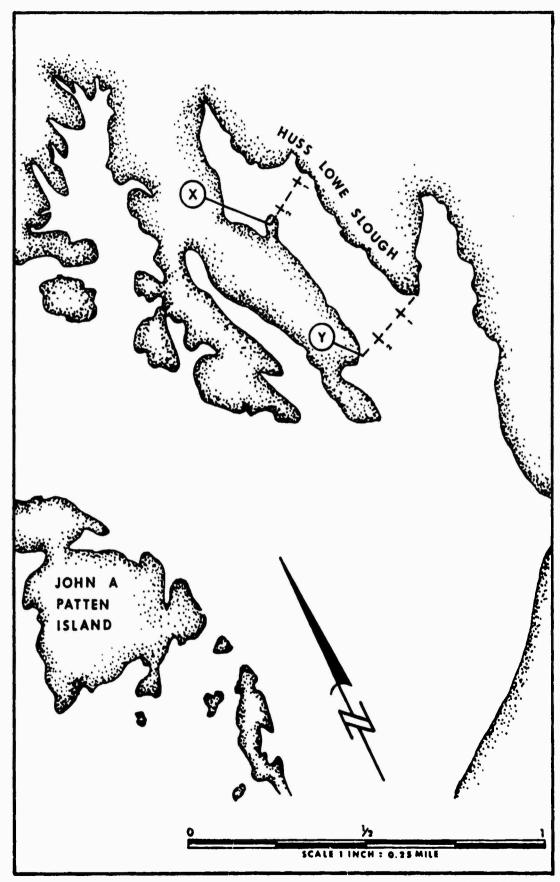


FIGURE 4. SAMPLING STATIONS IN HUSS LOWE SLOUGH (REFERENCE BAY 'B').

TABLE 1
SAMPLING SITE DESCRIPTION

Station Number	Approximate Depth (Feet)	Bottom Type
Α	10	Black mud, leaf litter
B-1	2	Clay and sand
B-2	11	Clay and small pebbles
C-1		Soft clay and silt, Some leaf litter
C-2	5 6 7	Soft clay and silt
D-1	7	Pebbles, leaf litter, twigs, some clay
D-2	12	Soft clay and silt
E-1	19	Rock, pebbles, coarse sand, some clay
E-2	11	Clay, pebbles, and sand
F-1	26	Leaf litter
F-2	18	Sand and shell, little clay
S	5	Pebbles, coarse sand, and clay
T-1	6	Leaf litter, clay
T-2	5 6 3 7	Leaf litter, gravel, clay
U-1	7	Gravel, sand, leaf litter, pine needles, rocks
U-2	12	Gravel, sand, some clay
X-1	6	Clay, pebbles, sand, some twigs and leaf litter
X-2	8	Clay and pebbles
Ÿ-1	8 7	Very hard clay
Y-2	15	Clay and pebbles

#### WATER QUALITY

#### Introduction

A considerable data base is being developed through contract research by the Army Medical Research and Development Command to characterize the environmental effects of TNT discharges. Major impact effects from a water quality standpoint are increased color, "pink water," resulting from photochemical action on nitrobodies; large increases in dissolved solids, mostly chlorides and sulfates; nitrogen enrichment; as well as munitions release. Failure of waste neutralization processes can result in significant pH changes. Effects on sediments include nitrogen, organic carbon, and salt enrichment, as well as accumulation of munitions compounds and their degradation products.

The Wapora study (1975) characterized water quality and biologic conditions in Waconda Bay and concluded that discharges from VAAP containing high levels of color, sulfate, solids, nitrates, nitrites, and organic material caused fish avoidance of upper Waconda Bay and destructuring of zooplankton and invertebrate communities.

#### Methods

During two 5-day survey periods, June 9 - 13 and August 11 - 15, 1975, samples and field measurements were taken at 20 stations located on nine transects of the three study bays, at the head of Waconda Bay, and Reference Bay A. Locations and descriptions of the stations are contained in the Introduction. The samples were preserved in accordance with EPA (1974) methods and shipped to WAR, Inc., in Gainesville, Florida, for processing.

Field Parameters. Dissolved oxygen (D.O.). temperature, pH, and specific conductance were monitored at various times throughout the day. Dissolved oxygen was measured with a YSI Model 51B D.O. meter. Specific conductance and temperature were measured with a YSI Model 33 Salinity-Conductivity-Temperature meter. Measurement of pH was made with a Photovolt Model 126A portable battery-operated pH meter.

<u>Laboratory Analyses</u>. Samples were collected as described and shipped refrigerated to WAR, Inc. The water quality parameters monitored included the following:

#### Major Ions

# Oxygen Demand

Total Alkalinity Chloride Total Hardness Sulfate Total Dissolved Solids (TDS) Chemical Oxygen Demand (COD) Total Organic Carbon (TOC)

#### Suspended Materials

# Trace Metals

Suspended Solids (SS) Total Solids

Cadmium (Cd)
Copper (Cu)
Chromium, Hexavalent (Cr<sup>+6</sup>)
Iron (Fe)
Lead (Pb)
Mercury (Hg)
Nickel (Ni)
Zinc (Zn)

# Plant Nutrients

# Munitions Compounds

Ammonia Nitrogen (NH<sub>3</sub>-N) Total Kjeldahl Nitrogen (TKN) Nitrite Nitrogen (NO<sub>2</sub>-N) Nitrate Nitrogen (NO<sub>3</sub>-N) Total Phosphorus (Total P) 2,4-Dinitrotoluene (2,4-DNT) 2,6-Dinitrotoluene (2,6-DNT) α-Trinitrotoluene (TNT)

The sediments were characterized by analyzing the following parameters:

#### Nutrients

# Trace Metals

Chemical Oxygen Demand (COD)
Total Kjeldahl Nitrogen (TKN)
Nitrate Nitrogen (NO3-N)
Nitrite Nitrogen (NO2-N)
Total Phosphorus (Total P)
Total Solids
Total Volatile Solids

Cadmium (Cd)
Copper (Cu)
Iron (Fe)
Lead (Pb)
Manganese (Mn)
Mercury (Hg)
Nickel (Ni)
Zinc (Zn)

# Munitions Compounds

2,4-Dinitrotoluene (2,4-DNT) 2,6-Dinitrotoluene (2,6-DNT) α-Trinitrotoluene (TNT)

The methods employed for collecting, preserving, and analyzing the routine water quality parameters followed accepted <u>Standard Methods</u> (APHA, 1971) or EPA (1974) procedures. <u>Chemistry Laboratory Manual Bottom Sediments</u> (EPA, 1969) was the source of the routine methods utilized for collection, preservation, and analysis of the sediment samples. Where existing methods, particularly for trace metals and munitions were insufficient to provide the desired levels of detection, alternate analytical procedures were employed after accuracy and precision had been verified. Details on analytical procedures are presented in Appendix A-1.

# Waste Loading and VAAP Permitted Discharge

Table 2 tabulates the permitted discharges of wastes from VAAP. A labor strike, shutting down operations one week before the June survey, lasted through the summer. Therefore, no effluent was being discharged during either the June or August surveys. During the month of May, however, plant production was 1.3 million pounds of TNT. Significant water quality changes resulted in upper Waconda Bay. Comparison of the permitted discharges with plant records for the period October, 1974 (when the permit came into effect) to December, 1974 showed a number of parameters exceeding the set limits. During 1972, an EPA study summarized in STORET examined the characteristics of VAAP wastes. The parameters exceeding NPDES specifications during the two periods are listed below:

NPDES Moni	toring (1974)	EPA Study (19)	72)
Ammonia BOD COD Nitrite Nitrate	Chromium Copper Manganese	BOD Ammonia Suspended Solids Phosphate Iron	Chromium Copper Lead Manganese

Table 3 shows ranges of receiving water parameters monitored during early spring (1974) in conjunction with NPDES requirements.

#### Water Exchange Characteristics

The relative size of Waconda Bay, Reference Bay A, and Huss Lowe Slough (235, 52, and 71 acres, respectively) can be seen in Figures 3 and 4. Mean depth is about 10 feet for all three bays. Runoff from residential areas and VAAP enter Waconda Bay. The watershed of Reference Bay A is largely residential and that of Huss Lowe Slough forested.

Flushing of the bays will be caused by runoff, by stage fluctuations, and in the case of Waconda Bay by effluent imput. Drainage areas for Waconda Bay and Huss Lowe Slough are each about 9 times the area of the bay. For Reference A, the drainage area is about 21 times the bay area. Assuming typical runoff coefficients, flushing by rainfall will be roughly 10 percent of bay volume per month in Waconda Bay and Huss Lowe Slough and about 20 percent per month for Reference Bay A. At an effluent flow of 5 MGD, Waconda Bay will be about 20 percent flushed in one month. The relative effectiveness of flushing by stage variations is not known. However, Huss Lowe Slough because of morphometric differences, probably would be more affected by stage variations than the other two bays. Given the possible error in these estimates, it seems reasonable to conclude that there are no apparent major differences in potential flushing ability of the three bays. Wind velocity and direction would have only local temporary effects on mixing and flushing rather than creating a consistent pattern. This would be unimportant compared to rainfall and stage variation. In al. three the time for complete flushing probably is on the order of months.

# Characterization of Water Quality

The monitoring data presented in the previous section and the results of the 1975 survey by Wapora suggest that the impact from nitrobodies.

TABLE 2

NPDES PERMITTED DISCHARGE AND TYPICAL EFFLUENT DISCHARGE DATA AT VAAP

	NPDES Discharge Limitations	e Limitations	OPN	NPDES Monitoring Data Oct Dec. 1974	ring Data 1974		EPA Survey June, 1972	rvey 1972
	Oct. 1, 1974 - Ap (1bs/day) m	April 30, 1979 mg/l	Quantity lbs/day	ty Iy	Concentration mq/l	ation	Concentration mq/1	ration /l
Parameter	Daily Average	Daily Maximum	Minimum	Maximum	Minimum	Maximum	Minimum	Maximum
NH3-N BOD5 COD	(99) (-0	0.5 10 20	6.9 45 297	140 2890 713	0.21 1.6 7.1	3.5 12.0 25.7	0.39 5.0 Not R	0.92 11.0 Reported
Cr Cu Dissolved Solids	(0.33) (0.13) 750	0.05 0.02 1000	0 0 4670 3	4.32 3.0 36,000	0 0 130	0.10 0.06 1000	50.0 50.0 10.0	50.0 50.0 519
P P P P	(0.33)	0.05 0.05 0.05	3.2 0 1.4	3.7	0.09	0.1	0.9 0.10 0.13	4.5 0.10 0.25
Hg NO2-NO3-N Oil and Grease	(0.013) 0.002 99 (10)	0.005 10 15	0 76 Not Rep	0 1050 Reported	0 2.4 Not Rep	0 22.2 Reported	Not Re 6.2 Not Re	Reported 9.9 Reported
Phenol Phosphate	(0.0066)	0.001		Reported Reported	Not Reported Not Reported	Reported Reported		Reported Reported
Softleable Solids SO4 Suspended Solids	30	250	Not Kep 2020 159	Keported 15,000 1450	40 40 2.0	300 29.0	Not R	Reported Reported 88.0
INT + Nitrobodies Temperature Flow Rate	0.3	0.5 32.2°C	0 Not Repo	104 Reported MGD 6.4 MGD	0 3. Not Reported	3.2 orted	Not Re Not Re 23.6 M	Reported Reported MGD 36.9 MG

TABLE 3

CONCENTRATION RANGES DERIVED FROM NPDES MONITORING REPORTS
OF SELECTED MUNITIONS MANUFACTURE GENERATED COMPOUNDS IN
WACONDA BAY JANUARY - APRIL 1974

	STATION LOCATION				
PARAMETER (mg/l)	Dredge	Wilson Dock	Rod and Gun Club	Harbor Entrance	Main Channel
Ammonia-N COD Nitrate-N Sulfate TNT & Nitrobody Total Sus. Solids Chromium Copper Iron Lead	0.5-4.5 Ni1-52.42 6.1-9.8 80-360 Ni1-1.6 7-22 0.00-0.00 0.00-0.04 0.03-0.21 Ni1	5.4-8.6 113-326 Nil-1.0 4-17 0.00-0.00 0.00-0.03	0.3-2.0 Nil-29.95 1.9-5.5 26.0-196 Nil-0.2 2-19 0.00-0.00 0.00=0.03 0.06-0.47 Nil	Nil-1.5 Nil-37.7 1.8-3.5 13.1-240 Nil Nil-13.0 0.00-0.00 0.00-0.03 0.10-0.73 Nil	Nil-0.2 Nil-44.5 0.03-2.4 10.0-27.5 Nil-0.1 Nil-26.0 0.00-0.00 0.00-0.03 0.21-0.66 Nil

#### STATION LOCATIONS:

Dredge - Holding pond outlet prior to discharge into Waconda Bay

Wilson Dock - Approximately 500 meters down bay from the head of Waconda Bay (approximately at transect B)

Rod and Gun Club - Approximately 1200 meters down bay from the head at approximately transect D

Harbor Entrance - Off-shore at approximately location of transect F
Main Channel - Main Channel of Tennessee River

nitrogen compounds, carbon, and dissolved solids discharged in the VAAP waste maintain a chemical environment in upper Waconda Bay significantly different from the rest of Harrison Bay. The results of the June - August, 1975 survey further document these water quality changes and offer a unique opportunity to examine the recovery rate of a lacustrine ecosystem from such effects.

Water quality data for selected major inorganic ions, plant nutrients, trace metals, and munitions, as well as associated field measurements, are tabulated in appendices A-2 to A-5. Transect and station locations are shown in Figures 3 and 4. Mean values of selected parameters are graphically illustrated in the text. The impact of TNT wastes on Waconda Bay may be placed in perspective by comparison with two reference bays and background data (STORET) on Chickamauga Lake for summer (1972). Baseline data from 1960 - 1961 gathered by Tennessee Valley Authority also exist for the lower part of Chickamauga Lake. These latter background data were considered by Tennessee Valley Authority to still be relevant to 1974 conditions (Tennessee Valley Authority, 1974). These background materials are appended in Appendix A-6. In general, these data correspond closely to data taken in the reference bays during the June and August, 1975, study in terms of hardness, alkalinity, conductance, and iron. During the 1960 survey, chlorides dropped from 13 - 15 mg/l in 1960 to 4 - 7 mg/l in 1961. These latter levels agree with more recent data. Plant nutrient concentrations, nitrogen and phosphorus, are sufficient to support well developed phytoplankton populations but are not indicative of excessively eutrophic conditions. The background data also suggest that biotoxic metals exist in Lake Chickamauga, but at insignificant levels.

Field Measurements. During the June and August, 1975 surveys, field measurements of dissolved oxygen, temperature, pH, and conductivity were made at all sampling locations. These data are presented in Appendix A-2, Tables A-3 through A-10.

For the week prior to the June sampling rainfall was only 0.19 inches. However, on the second and third sampling days (June 10-11) 1.39 inches were recorded, primarily between 0700 and 1100 hours. Consequently, some runoff was entering the reservoir during the sampling period. Sunshine, expressed as percent of possible, for the five-day sampling period was 50, 4, 2, 65, and 95 percent.

In August, 3.94 inches of rain fell during the week prior to sampling. No rainfall occurred during the five-day sampling period. Sunshine, expressed as percent of possible, for the five-day sampling period was 53, 86, 75, 73, and 15 percent.

In June, surface dissolved oxygen varied from a low of 4.9 mg/l at Station S on day 4 to a high of 9.75 mg/l at Station T-l on day 1. Dissolved oxygen within one foot of the bottom varied from a low of 0.2 mg/l measured in a localized hole at Station Y-2 to a high of 8.5 mg/l at Station Y-1. Bottom D.O. was normally in the range of 3.5 - 8.0 mg/l. As might be expected, diurnal variations were evidenced by generally lower values in the early morning hours with a late afternoon maximum.

Surface dissolved oxygen values in August were generally higher than in June with concentrations as high as 10.6 mg/l. Bottom concentrations were similar to June observations. The higher surface levels probably are due to the increased amount of sunlight available in August which stimulated algal activity.

June water temperatures averaged  $24.4^{\circ}\text{C}$  with extremes of 20.9 and  $27.0^{\circ}\text{C}$ . In August, the average was  $27.6^{\circ}\text{C}$  with extremes during the photoperiod of  $24.0^{\circ}$  and  $32.0^{\circ}\text{C}$ .

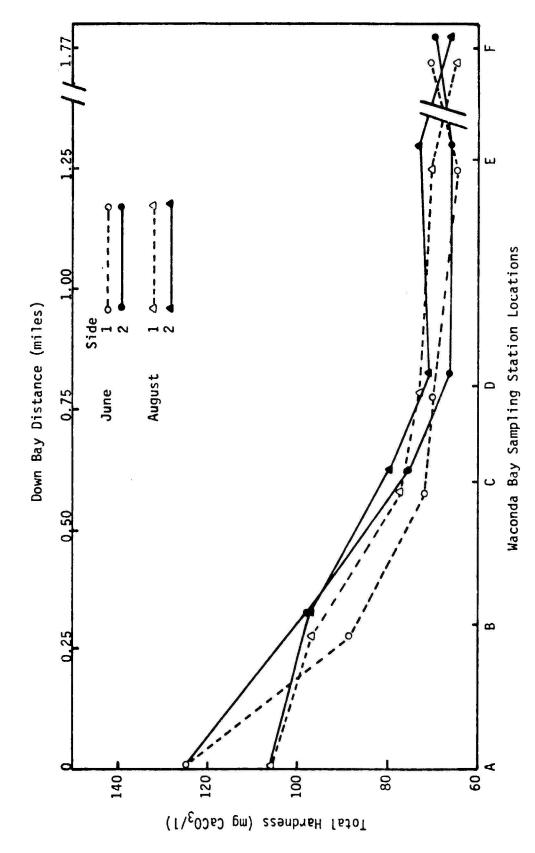
Conductivity averaged 226  $\,\mu mhos/cm$  in both June and August. A distinct gradient was noted from Station A (VAAP effluent input) to Stations F-1 and F-2 (approximately 500  $\,\mu mhos/cm$  down to about 180  $\,\mu mhos/cm$ ). During both surveys a surface to bottom conductivity gradient was observed at Station A. . Surface values ranged from 330 to 480. A variation of 465 to 1,350 was noted in bottom water. The highest conductivity was measured within a few inches of the bottom.

For both trips pH was in the range of 7.1 to 8.8. The higher values were measured in the afternoons during the peak algal activity. A slightly higher pH was observed during August probably because of the increased amount of sunlight available for photosynthetic activity.

<u>Major Ions</u>. The mean hardness, sulfate, and chloride levels in Waconda Bay are plotted in Figures 5, 6, and 7, respectively. Hardness shows a steep gradient, dropping from 125 mg/l as CaCO<sub>3</sub> at the head of the bay to about 65 mg/l offshore of transect D, a distance of 0.8 miles. An east (side 1) - west (side 2) gradient is also apparent with higher concentrations in the latter portion of the bay. This is a consistent trend throughout the data and reflects the morphometry of the upper bay -- deep water to the west, shallow water to the east. The steepness of the north - south gradient had decreased by the August survey viz maximum concentration 105 mg/l as CaCO<sub>3</sub>. This suggests that the discharges from the plant create major changes in hardness in the upper bay during operation. A slight gradient existed in Reference Bay A. Huss Lowe Slough showed very stable hardness concentrations and no downbay gradient. Hardness at Station S was slightly elevated (75 mg/l in June, 73 mg/l in August) as compared to background values of about 60 mg/l. These correspond to TVA (1974) data which ranged from 54 to 78 mg/l as CaCO<sub>2</sub>. Nearly four inches of rain occurred the week prior to the August sampling trip. This may have caused some suppression of concentrations in Waconda Bay. However, flushing characteristics and data on the reference bay hardness do not support a hypothesis of major dilution effects.

Alkalinity was unaffected by discharges from VAAP and showed no down bay changes. Concentrations were about the same in all three bays and corresponded with historic data on Lake Chickamauga. Alkalinity was slightly higher during August.

Sulfate ions are also a major discharge from VAAP as shown in Figure 6. The shape of the gradient is identical to hardness, dropping to ambient reservoir levels between transects C and D. Maximum values at A were 108 mg/l in June, 55 mg/l in August. The maximum sulfate concentration found at the bay head by Wapora, Inc. (1974) was 72 mg/l. NPDES monitoring



MEAN TOTAL HARDNESS CONCENTRATION GRADIENTS IN WACONDA BAY. FIGURE 5.

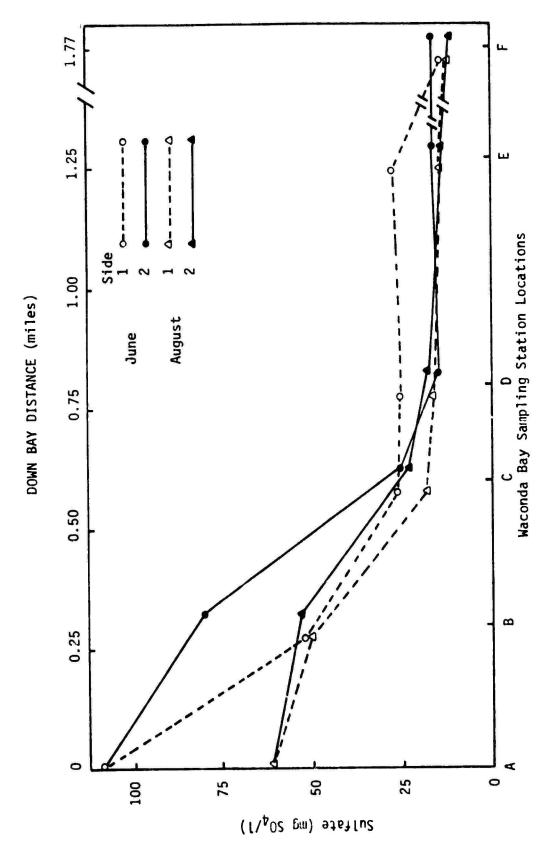


FIGURE 6. MEAN SULFATE CONCENTRATION GRADIENTS IN WACONDA BAY.

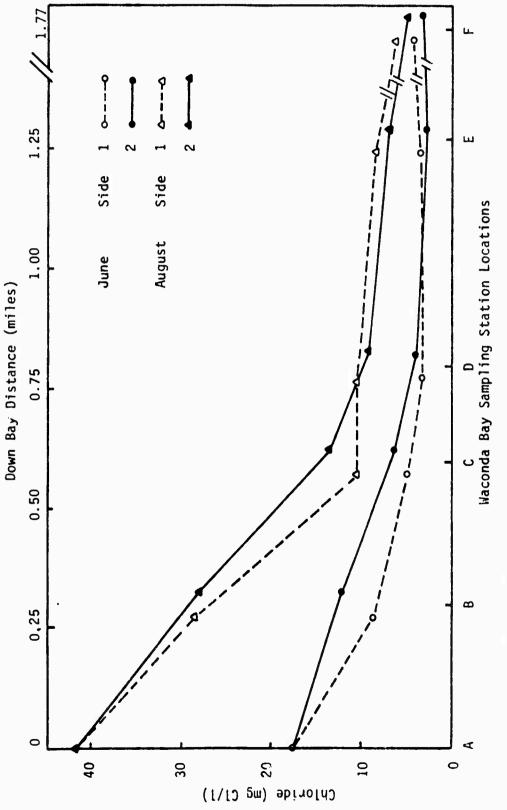


FIGURE 7. MEAN CHLORIDE CONCENTRATION GRADIENTS IN WACONDA BAY.

data indicate a range of  $80 - 360 \text{ mg } SO_4/1$  for the discharge into the bay. Station S again showed a slightly higher  $SO_4$  concentration (27.3 mg/1) in June than the offshore stations in the reference bays. The background ranges for this ion agree with TVA data from 1960 - 1961.

The background chloride concentration corresponds to the historical Tennessee Valley Authority data. However, the concentration gradients in Waconda Bay differed significantly from those for sulfate and hardness in that the maximum was found during August rather than in June. Maximum June concentrations were only slightly higher than ambient (17.4 mg Cl/l) while the August concentration at A was over 40 mg Cl/l. None of the other major ions measured show a retrograde pattern between June and August levels.

Total dissolved solids ranged from 275 mg/l at A to approximately 100 mg/l at transect F in June. TDS data agree generally with the historical background for dissolved materials in Chickamauga Lake.

The overall ranges for several parameters for the reference bays are tabulated below:

	Concentration Range			
	Reference Bay A		Huss Lowe Slough	
	June	August	June	August
Total Hardness (mg CaCO <sub>3</sub> /1) Alkalinity (mg CaCO <sub>3</sub> /1) Sulfate (mg SO <sub>4</sub> /1) Chloride (mg CI/1 Total Dissolved Solids (mg/1)	59-79 44-54 12-28 2-5 76-141	64-81 53-70 12-14 2-4 70-227	59-65 46-54 10-11 6-8 64-105	60-67 55-60 10-11 5-6 61-243

Except for samples taken on August 15, 1975, suspended materials were fairly low <2 - 15 mg/l. A slight gradient appeared to exist in Reference Bay A and in Waconda Bay. In Reference Bay A a general decrease in values offshore from S through U occurred probably due to runoff and dilution effects. In Waconda Bay a pattern existed with transects B and C showing consistently higher values than Station A or the offshore transects. The increase may be due to the higher plankton populations found at these two stations. Levels of suspended solids were relatively the same for both surveys. The gradient (for both the June and August surveys) is shown in Figure 8.

Carbon. The VAAP facility is permitted to discharge 20 mg/l chemical oxygen demand (COD) and up to 66 pounds per day biochemical oxygen demand (BOD) into Waconda Bay. Table 2 shows that up to 25 mg/l COD is likely to be discharged during plant operation. This carbonaceous material would include munitions residues, solvents, process impurities, and wasted starting materials. Apparently, these compounds degrade rapidly as only a slight COD gradient was apparent in Waconda Bay in June (Figure 9). During August the COD has the same pattern as suspended solids, i.e. low at A with highest concentrations at transect B. This suggests the suspended solids pulse in mid-bay is organic, possibly from increased algal populations.

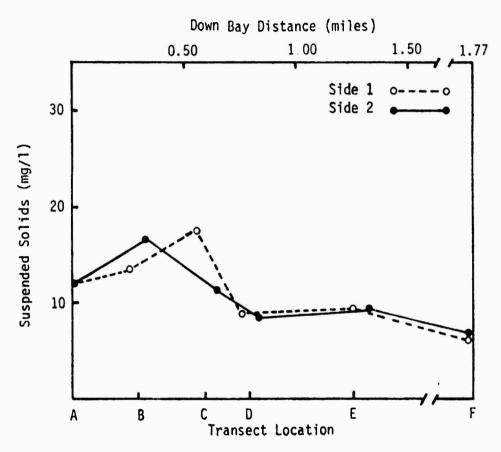
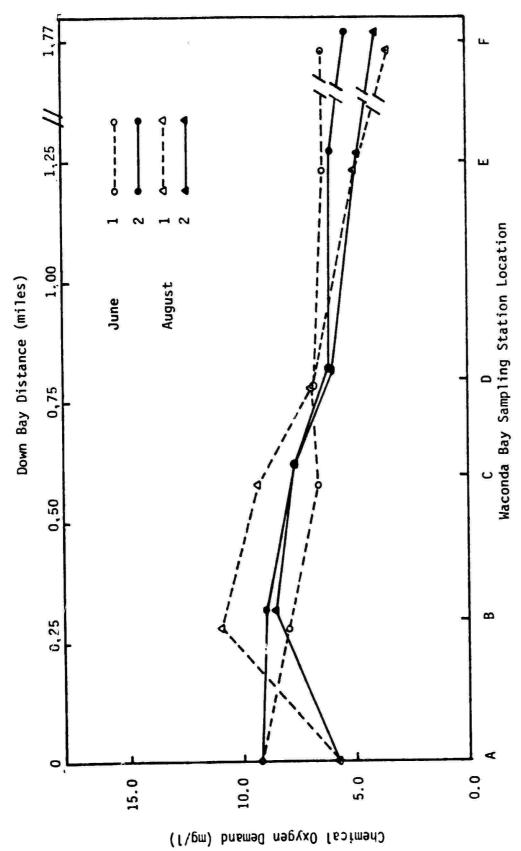


FIGURE 8. MEAN SUSPENDED SOLIDS CONCENTRATIONS IN WACONDA BAY.



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MEAN CHEMICAL OXYGEN DEMAND IN WACONDA BAY.

FIGURE 9.

Background COD for transects E and F and the reference bays was 2.3 to 8.7 mg/l in June and <0.5 to 8.2 in August. Station S at the head of Reference Bay A contained slightly elevated COD levels, 6.1 - 10 mg/l in June and 4.2 - 8.5 mg/l in August. At the low carbon levels encountered, both the COD and TOC test lack resolving ability.

The TOC and COD results of Wapora, Inc. (1975) reveal inconsistencies. These data show high COD without a corresponding high TOC value at the bay head and high TOC without similar COD at their transect E located offshore. The upper ranges of COD reported in Table 4 (NPDES monitoring) are high for lake waters (30 - 45 mg/l).

Nitrogen and Phosphorus. Total Kjeldahl nitrogen (TKN) and ammonia levels were similar in Waconda Bay during the two surveys. Ammonia nitrogen discharge is permitted to 0.1 mg N/l. The discharge monitored under operation in 1974 ranged from 0.21 to 3.5 mg N/l. This resulted in concentrations above 1 mg N/l in the bay (Table 3). During the June survey (WAR, Inc.) ammonia averaged 0.32 mg N/l at Station A, but decreased to <0.1 mg N/l by transect C. Background ammonia concentrations in Reference Bay A and Huss Lowe Slough were generally less than 0.1 mg N/l. Ammonia values observed in these bays correspond quite closely to similar low levels reported in STORET for other areas of Lake Chickamauga. During the August sampling period, similar levels of NH3-N were observed except for very high values encountered on August 15, the last sampling date. Decreasing lake stage may have caused water containing large amounts of NH3-N from the shallow outfall ditch to enter the upper end of Waconda Bay.

	Ammonia N	mg N/l
Station	Range August 11 - 14	August 15
A	0.07 - 0.33	4.99
B-1	0.04 - 0.25	1.22
B-2	0.06 - 0.26	1.02

Increased levels were also observed on August 15 for total Kjeldahl nitrogen, munitions, and total dissolved solids. Most of the increase in Kjeldahl nitrogen was as ammonia.

The concentration profile of reduced nitrogen species is illustrated for Waconda Bay as TKN in Figure 10. No significant differences occur between June - August and concentrations are at ambient for Harrison Bay at transect D. In contrast, Wapora, Inc. (1975) found TKN considerably elevated only at transect B while the plant was operating. The background values (0.2 to 0.6 mg/l) suggest mesotrophic lake conditions for the Harrison Bay segment of the lake. No gradient was observed in the reference bays. The data suggest that reduced nitrogen is principally in the form of biomass as NH<sub>3</sub>/TKN ratios ranged 0.1 - 0.3.

NPDES monitoring data suggest high NO3-N values in Waconda Bay (Table 3) although discharge concentrations were below the permitting requirements of 10 mg NO3-N/1. Historically, NO3-N varies from a negligible amount to -0.5 mg N/1 (Tennessee Valley Authority, 1974, STORET). Reference Bay NO3-N values fall within this range. Nitrate-nitrogen is higher in Reference Bay A (which receives some urban runoff) than in Huss Lowe Slough which receives only drainage

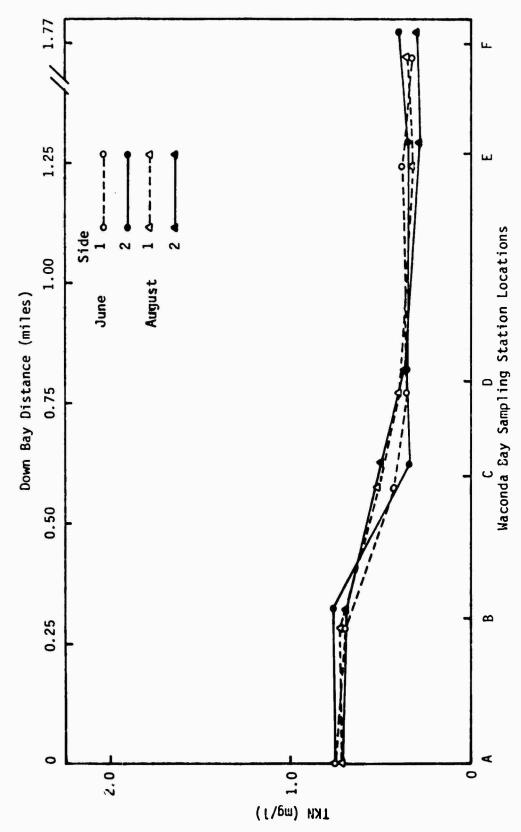


FIGURE 10. MEAN TOTAL KJELDAHL NITROGEN CONCENTRATIONS IN WACONDA BAY.

from a wooded watershed. Station S with an average NO $_3$ -N value of 1.14 mg N/1 in June an 0.66 mg N/1 in August was nearly 2 times the levels at transects T and U.

Significant gradients and enrichment of the upper end of Waconda Bay by nitrite and nitrate-nitrogen were observed. Figure 11 shows the gradient for  $NO_3-N$ . Nitrite distributed in the same pattern. Elevated  $NO_3$  and  $NO_2$  nitrogen concentrations existed down bay to transect C.

Phosphorus showed no apparent differential among the bays or gradations offshore. No differences overall were seen between the June and August surveys. The range of total PO<sub>4</sub>-P for both was 0.012-0.096~mg/l which agrees with the historical data (STORET).

Trace Metals. Volunteer Army Ammunition Plant is limited to low level discharges of chromium, 50  $\mu$ g/l; copper, 20  $\mu$ g/l; lead, 50  $\mu$ g/l; mercury, 5  $\mu$ g/l; manganese, 50  $\mu$ g/l; and iron, 0.3 mg/l. VAAP discharges during 1972 exceeded these limits (STORET). The data extracted from the NPDES reports showed some manganese concentrations in excess of the permitted limits. The receiving water data showed no environmentally significant concentrations of the metals studied (Table 3).

Selected heavy metal analyses sufficient to characterize the burden in Waconda Bay and the reference bays are tabulated in Appendix A-4. These showed no heavy metals in potentially biotoxic concentrations. Ranges of metal concentrations are recorded below:

Metals	Concentration $(\mu g/1)$
Iron Lead	82 - 310 <5 - 21
Cadmium	>5
Hexavalent Chromium Copper	>5 <5 <b>-</b> 7
Nickel Zinc	< <b>5 -</b> 20 <2 <b>-</b> 85

Most analyses showed undetectable heavy metals. Mercury was undetectable at all stations in June and was not sampled in August. Wapora, Inc. (1974) reported similar metal concentrations in Waconda Bay confirming that VAAP metal discharges do not significantly affect the bay ecosystem.

Munitions Residues. Periodic discharges of TNT and nitrobodies exceeded the 0.3~mg/l permit limit in fall and winter, 1974 (Table 2). However, during the study by Wapora, Inc., in September, 1974, no residues of greater than 100~mg/l were observed. NPDES monitoring of Waconda Bay (Table 3) showed TNT concentrations ranging above 1 mg/l in the upper bay during spring 1974. These were virtually undetectable beyond 0.5~miles down bay.

Munitions residues were detected during both June and August, 1975 surveys at concentrations below 400 ppb. These ranged to 345 ppb at A in June to near zero between transects C and D. Most upper Waconda Bay concentrations were below 200 ppb. Occasional residues ranging from traces to 15 ppb

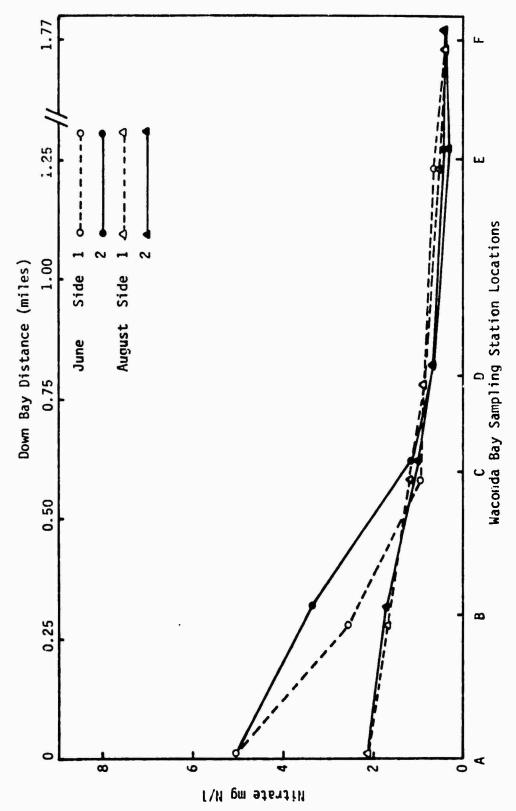


FIGURE 11. MEAN NITRATE CONCENTRATION GRADIENTS IN WACGNDA BAY.

were detected as far down bay as transect F, 1.77 miles from the point of discharge. No munitions residues were detected in either reference bays. Figures 12 and 13 depict gradients of munitions residues in Waconda Bay for the June and August trips, respectively. These data are reported as the median and concentration ranges for 2,4-DNT+2,6-DNT+ $\alpha$ -TNT. There were no consistent patterns in any of the three compounds related to distance from the outfall. Dinitrotoluene compounds, however, made up the largest proportion of the total residues concentrations. Median munitions concentrations were maximum at B-2 for both studies.

Cluster Analysis. Six parameters were chosen to represent the major facets of water quality. They were total Kjeldahl nitrogen, nitrite-nitrate, chloride, sulfate, total hardness, and the sum of nitrated toluenes. They were clustered as described in "Computational Methods" using the average distance formula, in order to determine water quality relationships and impacted areas. Figures 14 and 15 present the phenograms for June and August, respectively. The phenograms show that the upper 0.3 to 0.5 mile of Waconda Bay (Transect B and Station A) contains water quality characteristics which cause it to be unrelated to the remainder of Waconda Bay and Harrison Bay. Conditions were very similar among the other transects and the reference bays; they group very closely. All of the parameters examined are different for transects A and B. Station S contains some values slightly higher for the parameters considered but not sufficient to significantly alter the cluster pattern. Among the non-impact transects no consistent relationships were observable.

# Characterization of Sediments

Except in the upper reaches of Waconda Bay the sediments consisted of a reddish clay matrix mixed with gravel and sand. Overall, the clay is covered with a mixture of detritus in the form of leaves and sticks. The lack of deep organic sediment in Harrison Bay is probably a result of the relatively short time that Lake Chickamauga has been in existence (construction began in 1936). Sediment chemical characteristics, nutrients, percent solids, and metal concentrations are tabulated in Appendix A-47.

Sediment Nutrients. Table 4 compares the June and August means for the first 0.6 miles of Waconda Bay versus the offshore sediments and the reference bays' sediments. The sediments taken in June and August, 1975 were found to be richer in phosphorus than reported by Wapora, Inc. (1975). The Waconda Bay contains more organic material, nitrogen, and phosphorus in the down bay area or the reference bays. Reference Bay A was found to have an average NO3-N content nearly as high as upper Waconda Bay, primarily as a result of two elevated values from the August survey at Station U-2 and T-2. These samples may not be representative of the bay in terms of NO3-N. If these are ignored, the average drops to about half that for upper Waconda Bay. Essentially the same situation exists for Huss Lowe Slough.

Trace Metals. The trace materials from urban runoff normally are deposited into aquatic sediments. Certain of these are metal ions which may express toxic effects through bioaccumulation. Little is known about the fate and distribution of such materials although several in-depth studies have been conducted. Iskandar and Keeney (1974), and Holmes,

TABLE 4
SUMMARIZED MEAN VALUES FOR SEDIMENTS

Parameter	Upper Waconda Bay*	Lower Waconda Bay**	Reference Bay "A"	Huss Lowe Slough
Volatile Solids (%)	7.9	4.4	4.6	4.0
Chemical Oxygen Demand (gm/kg dry wt.)	8.9	5 <b>.</b> 5	4.1	3.2
Total Kjeldahl Nitrogen (gm N/kg dry wt.)	1.6	1.0	0.7	0.4
Nitrate Nitrogen (mg N/kg dry wt.)	62	33	54(29)***	49(27)***
Total Phosphorus (gm P/kg dry wt.)	1.07	0.49	0.34	0.17

<sup>\*</sup>Upper Waconda Bay - A, Transects B & C

<sup>\*\*</sup>Lower Waconda Bay - Transects D, E, F

<sup>\*\*\*</sup>Excludes two values that appear to be unreasonably high.

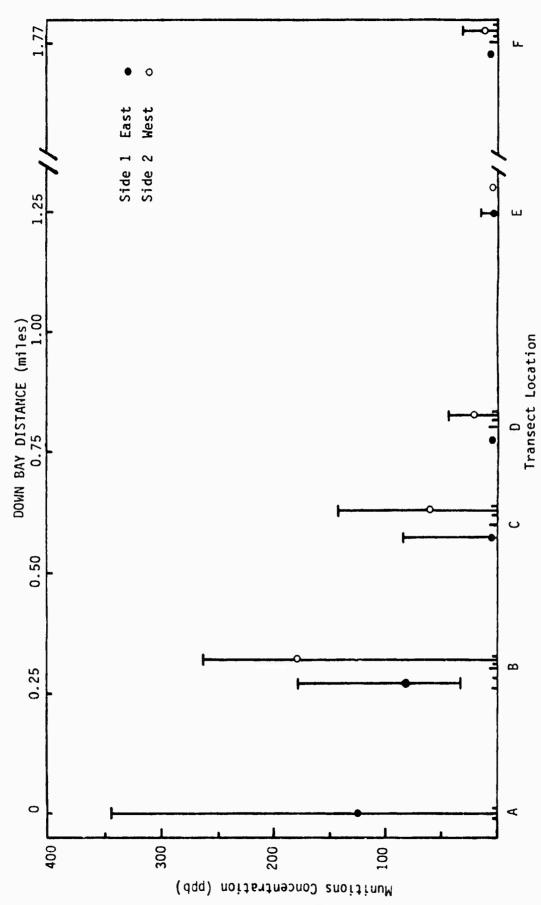


FIGURE 12. MEDIAN VALUES AND CONCENTRATION RANGES FOR MUNITIONS RESIDUES IN WACONDA BAY, JUNE 1975.

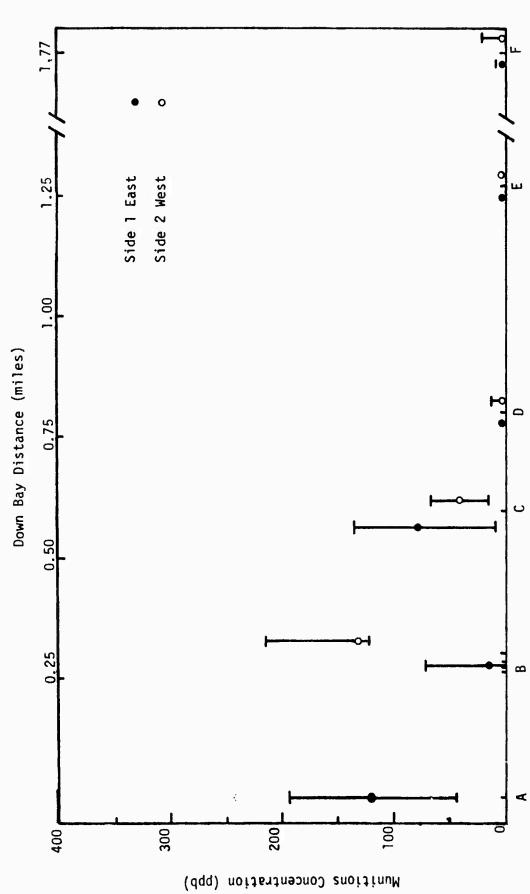
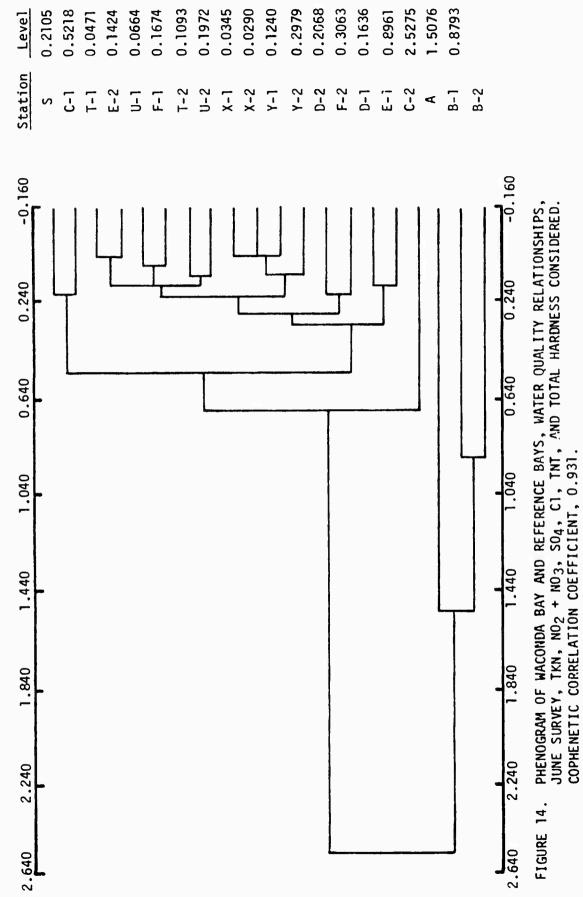


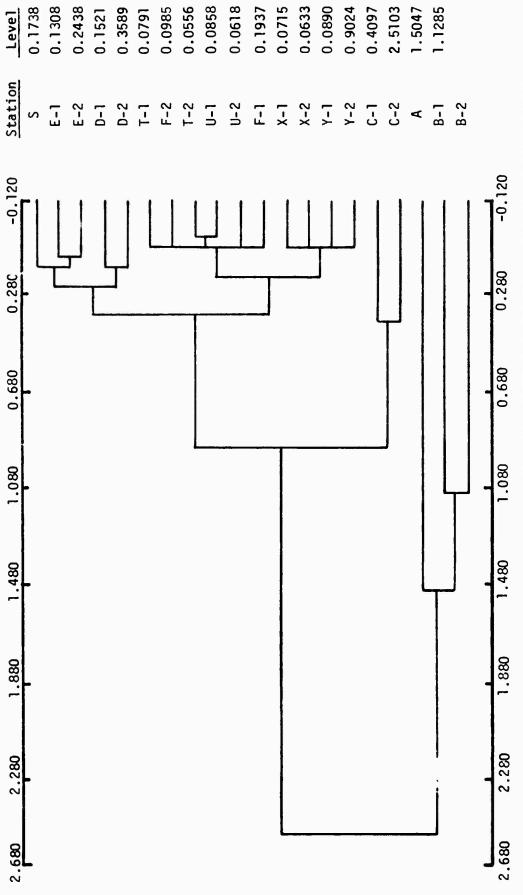
FIGURE 13. MEDIAN VALUES AND CONCENTRATION RANGES FOR MUNITIONS RESIDUES IN WACONDA BAY, AUGUST, 1975

Transect Location



41

FIGURE 14.



PHENOGRAM OF WACONDA BAY AND REFERENCE BAYS, WATER QUALITY RELATIONSHIPS, AUGUST SURVEY, TKN, NO2 + NO3, SO4, C1, TNT, AND TOTAL HARDNESS CONSIDERED. COPHENETIC CORRELATION COEFFICIENT, 0.947. FIGURE 15.

et al. (1974) found that redistribution of metals between sediments and overlying waters in lakes and estuaries occurs seasonally, paralleling variations in pH, D.O., temperature, etc. Precipitation of insoluble species and scavenging of metals by sorption to suspended particulates are two main pathways for sediment enrichment. Appendix A-7 tabulates the concentrations found in selected Harrison Bay sediments in June and August, 1975. As expected, iron and manganese ranged an order of magnitude higher than the other metals. Concentration ranges for the various species are tabulated below:

Element	Concentration Range mg/Kg Dry Weight
Fe	1,400 - 140,000
Mn	30 - 10,000
Cd Cr+6	<1 - 3.5
Cr <sup>+6</sup>	23 - 740
Cu	<1 - 130
Hg	<0.1 - 1.90
Ni	2 - 110
Pb	5 - 860
Zn	4 - 690

Iskandar and Keeney (1974) were able to show distribution of copper, lead, and zinc in Wisconsin lake sediments related to cultural activities on the watersheds. Ranges for copper were 0-400~mg/kg; lead 5-160; and zinc 10-200, roughly similar to those found in the Harrison Ba, sediments. The upper bound of the ranges for Fe, Mn,  $\text{Cr}^{+6}$ , Cu, Fb, and Zn, were found at A. Table 2 shows that occasional discharges of Mn, Fe, Cr. and Cu at levels above permit specifications may occur. Overall, the data agree with levels found by Wapora, Inc. (1975), except for higher values at A observed during the present study.

The environmental significance of metal concentrations measured by destructive analysis is unclear. Elutriation (Keeley and Engler, 1974) and ammonia acetate extraction measure the amounts leachable from the sediment interstitial water. The relative amounts of the various fractions available or which can act on benthic organisms are undefined. Little evidence exists (WAR, 1975, Keeney and Iskandar, 1974) for biotic effects at the levels reported in this study.

Munitions. Table 5 presents the distribution of munitions residues in selected Bay sediment samples. Single sediment samples were extracted and analyzed for the presence of munitions residues. The distribution of TNT parallels the concentration gradient of munitions in the water column. Wapora, Inc. (1975) were unable to detect residues in sediment samples. (However, their detection limit was 10 mg/kg). Selected samples from offshore transects D to F and the reference bays showed no detectable concentrations.

The TNT concentrations reported may considerable underestimate the actual munitions related biotoxic residues present. Microbial breakdown occurs stepwise, giving rise to dinitrated amines and hydroxyl amines such as 4-Amino-2, 6-dinitrotoluene. Resistance of aromatic rings to attack is measured by increased numbers of nitro-radicals (McCormick, 1974) and toxicity of intermediate breakdown products may be as great or greater than

TABLE 5
MUNITIONS RESIDUES IN HARRISON BAY SEDIMENTS
LAKE CHICKAMAUGA, 1975

Station	TNT (mg/kg June	Dry Weight) August
Waconda Bay		
Α	3.1	2.4
B-1	0.3	0.2
B-2	0.5	<0.1
C-1	0.32	
C-2	<0.1	0.2
D-2	<0.1	<0.1
E-2	<0.1	<0.1
F-2	<0.1	<0.1
Reference Bays		
\$	<0.1	<0.1
X <b>-</b> 1	<0.1	

the parent compound. Weitzel, et al. (1975) were able to characterize a major TNT breakdown in sediment at IAAP as a monohydroxylamino-dinitrotoluene which exhibited comparable biotoxic properties with TNT and further found concentrations of this daughter in sediments equal to or greater than corresponding TNT concentrations.

#### PERIPHYTON

### Introduction

The periphyton or "Aufwuchs" community is an assemblage of attached microorganisms (primarily algae) growing on the free surfaces of submerged substrates forming a slimy green or brown coating. The attached periphyton community consists of an assemblage of both autotrophic (i.e. diatoms and filamentous algae) and heterotrophic (bacteria, protozoa, rotifers, etc.) organisms.

Odum (1956) reports that under clear, clean water conditions, the algal component of the periphyton community can develop a large standing crop (e.g. 100 gm dry wt/m², Silver Springs, Florida) and, therefore, represents an important food source for a wide variety of aquatic organisms. As levels of organic pollution increase, the algae are replaced by filamentous bacteria and other non-chlorophyll "consumer-type" organisms (such as <u>Sphaerotilus</u>) resulting in significant increases in biomass-chlorophyll ratios and the establishment of a heterotrophic periphyton community (Weber and McFarland, 1969).

Major factors which limit or control periphyton growth are light, turbidity levels, temperature, current, and nutrient and substrate availability. Current provides a constant nutrient supply and carries away dead organic matter. Adequate light penetration is essential for growth of periphytic algae. Turbid waters strongly affect periphyton development by limiting the light necessary for their photosynthetic development.

The periphyton community lends itself well to biological investigations of water pollution. These organisms remain at fixed locations and are sensitive to changing environmental conditions. Their populations and biomass are relatively easy to quantify using standard laboratory procedures and are adaptable to a variety of statistical analyses.

A detailed study of the Lake Chickamauga periphyton community was conducted during the summer of 1975 in the vicinity of the Volunteer Army Ammunition Plant, Chattanooga, Tennessee. The purpose of this investigation was to investigate and report the effects of TNT residues and other munitions related waste compounds upon lacustrine periphyton flora.

Munitions wastes from the VAAP TNT manufacturing process enter the head of Waconda Bay producing impacts on both water quality and biological communities. Principal components of the discharge consist of TNT residues, organic nitrogen, nitrate and nitrite, chloride, hardness, ammonia, and sulfates. Although the plant had ceased discharge one week before the June sampling, elevated concentrations of nitrate, sulfate, chloride, and hardness existed from transect A to transect c. Average nitrate concentrations at transect A were 5 mg N/1 during the June survey. "Pink-water" conditions characteristic of nitrobody discharges were observed at the head of Waconda Bay. TNT and DNT were recorded from transects A through F. Reference bays contained no

TNT or DNT residues. Reference Bay A receives urban runoff at Station S, while Huss Lowe Slough drains forested land. The reference bays exhibited little difference in water quality except that ambient nitrate levels are generally higher on the west side of Harrison Bay (i.e. in Reference Bay A and the offshore transect F).

Previous studies of the effects of munitions related compounds upon natural periphyton communities have varied widely. For example, Wapora, Inc. (1975) surveying New River, Virginia and the Obion River in Tennessee was unable to correlate periphyton community structure with pollutant loading. Field studies by Water and Air Research, Inc. (1975) at the Longhorn, Texas and Louisiana Army Ammunition Plants indicated no significant effects on stream periphyton due to munitions waste effluents. However, results of Battelle Columbus Laboratories (1975) on three munitions facilities (Badger Army Ammunition Plant, Baraboo, Wisconsin; Joliet Army Ammunition Plant, Joliet, Illinois; and Lake City Ammunition Plant, near Kansas City, Missouri) reported observable effects of munitions waste discharges on the periphyton community. Weitzel, et al. (1975) observed shifts in periphyton species diversity corresponding to variations in nutrient levels and TNT concentrations in the streams adjacent to the Iowa Army Ammunition Plant. Periphyton studies conducted on the Holston River in the area of the Holston Army Ammunition Plant have indicated various degrees of environmental stress due to the discharge of ammunition wastes. Wapora (1975) indicated diatom populations percentages to be considerably lower on the north bank as opposed to the far shore. Periphyton distribution patterns also differed in respect to the two banks of the river.

Recent studies on the Holston River (Water and Air Research, Inc. 1976)\* noted effects of munitions waste on the periphyton in the vicinity of the HAAP waste outfalls. Significant increases in heterotrophic biomass and reductions in chlorophyll bearing species were noted. Reduction in diatom cell densities, species diversity, and shifts in diatom associations indicated toxic impacts from RDX and associated residues. Effects on periphyton were observed in waters containing RDX concentrations in a range of 20-100 µg/l.

Wapora, Inc.'s (1975) study of Lake Chickamauga, Chattanooga, Tennessee, reported differences in periphyton community structure in the area of Waconda Bay in the vicinity of Volunteer Army Ammunition Plant waste discharge.

#### Methods

Both natural and artificial substrate periphyton communities were studied during June-July and August-September, 1975. Due to a lack of comparable

<sup>\*</sup>Draft in preparation for Army Medical Research and Development Command.

natural substrates from which to sample, periphyton collections from artificial substrates probably give a more reliable station to station comparison. Standard microscope glass slides were placed in periphyton samplers one inch below the water surface at 20 selected VAAP station locations (Figures 3 and 4) according to the methods outlined in Standard Methods for the Examination of Water and Wastewater 13th Edition (APHA, 1971) and Biological Field and Laboratory Methods for Measuring the Quality of Surface Water and Effluents (Weber, 1973). Table 1 gives the depth and bottom type for each biological station. Due to heavy boat traffic and vandalism, a number of periphyton samples were lost or pulled out of the water at Stations E-1, E-2, F-1, and Y-1 throughout the study period. As a result, periphyton data from these stations were not available during certain sampling periods. As a backup to lost or vandalized stations, two additional periphyton sampling sites were utilized; station "no wake" (a navigation buoy) located midway between transects D and E in Waconda Bay; and Station "F-buoy" located on a channel-marker buoy in the vicinity of Station F-2.

At the end of each incubation period, slides were removed from each sampler for the following laboratory analyses:

(1) diatom community structure (artificial and natural substrates);

(2) filamentous organisms; and

(3) determination of organic biomass (ash-free dry wt.) and chlorophyll a for the determination of an Autotrophic Index and net primary production.

Diatom Community Structure. In terms of cost effectiveness and information content, the diatom (Bacillariophyceae) component of the periphyton community represents the most important group of algae studied in general water quality monitoring surveys. Cairns, et al. (1972) report that other groups of algae (e.g. Cyanophyceae and Chlorophyceae) are sensitive to pollution stresses but the difficulty and high cost of identifying them to the species level (Archibald and Bold, 1970) precludes their use in most field bioassay studies.

Diatom cell density estimates (cells/mm²) were determined as follows. Replicate samples were analyzed for community structure by first scraping the glass slides into 50 ml of distilled water. The glass slides have an area of 3,871 mm². The algal suspension was thoroughly mixed after which a 10 ml aliquot was pipetted into a tall 1 liter beaker and oxidized by 10 ml of hydrogen peroxide and 60 mg of potassium dichromate. The solution was cooled and centrifuged for 15 minutes, decanted, and expanded to a volume of 25 ml (2.5 times the original 10 ml sample volume). The 2.5 fold dilution permitted a distribution of 10-15 organisms per microscope field when magnified 1,250 diameters.

Permanent diatom mounts were prepared by pipetting 0.4 ml of the "cleaned" material onto a  $18 \times 18 \text{ nm}$  coverslip ( $324 \text{ mm}^2$ ) and allowing the

sample to dry at 65°C (150°F) on a laboratory hot plate. The dried coverslip was placed on a standard microscope slide containing one drop of HYRAX mounting medium (refractive index, 1.71) and the slide was gently heated to drive off the toluene solvent. When couled, the permanent slide was labeled with station number, date, location, incubation period, and dilution factor. Under an oil immersion lens (Zeiss microscope, 1250X) diatoms were identified and enumerated to the species level where possible utilizing the following standard taxonomic references: Hustedt, 1930, 1962; Cleve-Euler, 1952; Schmidt, et al., 1874-1959; Huber-Pestalozzi, and F. Hustedt, 1942; and Patrick and Reimer, 1966.

All densities were estimated by performing 30 field counts while randomly scanning each slide from left to right (15 field counts) and from top to bottom (15 field counts). At 1,250 magnification each microscope field represented an area of 0.038 mm², with a total area examined of 1.14 mm² (i.e.  $0.038 \text{ nm}^2 \times 30 \text{ field counts}$ ). Therefore, cell densities were estimated using the following formula:

The rationale for performing the "short count" 30 field method is supported by the following data. Figure B-1 (Appendix B) shows the effect of increasing sample size (i.e. area examined, mm²) on diatom species diversity on successive 0.038 mm² microscopic field counts. As illustrated in Figure B-1, diatom community structure, as estimated by the Shannon-Weaver Index, was largely established after counting 30 fields. Diatom counts from Stations F-1 and X-1 in Lake Chickamauga indicated that after counting 150 fields the diversity index was approximately the same as a 30 field count (Figure B-1). At Station F-1, after the examination of 150 fields, the index increased 0.065 units; a 3.4 percent increase.

Table B-1 (Appendix B) presents a summary of microscopic field count data from Station T-1, indicating the total numbers of "new" species recorded when counting more than 30 fields. The most important aspect of Table B-1 is the actual densities of these "new" species and their percentage contributions to the total population. These data indicate that after increasing sample size to 150 fields, the number of "new" species only represent 1.0 percent of the total population. As a result these species may be considered rare or "unimportant" in terms of energy flow through the periphyton community.

At Stations X-l and F-l, the percentage contribution of "new" species to the total population was slightly higher -- 2 to 4 percent. In conclusion, the performance of 30 field "short counts" seemed to describe adequately the primarily dominant or "important" diatom species present at VAAP; i.e. those species which probably account for the largest percentage of energy flow through the periphyton community.

In an effort to compare diatom populations from station to station, cell density estimates were used to calculate community indices such as the Shannon-Weaver Species Diversity Index  $(\bar{H})$ , (Shannon and Weaver, 1949; Margalef, 1968) to the base e. In addition, stations were compared by measuring the degrees of similarity between species associations at different stations utilizing the Pinkham-Pearson (1974) Index of Biotic Similarity (see Computational Methods for detailed explanations of these indices).

Collections of periphytic algae were also made from natural substrate materials. Periphyton was scraped from the surface areas of rocks into 50 ml of water at six station locations.

Filamentous Organisms. Attempts were made to estimate quantitatively the filamentous algae component of the periphyton community. Artificial substrates (glass slides) were incubated for four weeks during June 11 to July 10, 1975. At the end of the incubation period, the periphyton slides were removed and preserved in 5 percent Formalin in a light excluding sample box. In the laboratory, the slides were scraped and preserved in labeled sample bottles containing 50 ml of a 5 percent Formalin solution. Filamentous algae were identified and enumerated by the Utermohl (1958) sedimentation technique utilizing a 50 ml plankton counting chamber and a Zeiss inverted D microscope. Species identifications were carried to species level where possible, utilizing the following standard references: Drouet (1968); Prescott (1962); and Desikachary (1950).

Organic Biomass and Chlorophyll a. Growth of periphyton on artificial substrates was measured as organic biomass (ash-free dry weight) and chlorophyll a (corrected for phaeopigments) after 4-wick incubation. Standard procedures were used for assessing biomass and chlorophyll a levels as outlined below. Primary production estimates were made using chlorophyll a as indicated. This estimated the standing crop of both the autotrophic and heterotrophic components of the attached community and characterized summer production levels in the lacustrine system.

Periphyton communities in unpolluted waters tend to be dominated by algae, especially diatoms. The organic weight, or biomass to chlorophyll ratio, in such communities approaches that of an algal culture, i.e. 50-100. Organic pollution, particularly, causes an increase in the ratio due to increase in the heterotrophic component (bacteria such as <u>Sphaerotilus natans</u>, fungi, and protozoa) while toxic effects may decrease total mass of either heterotrophic component, autotrophic component, or both.

The levels of periphyton biomass, therefore, serve as an overall index of the level of biological activity in the producer and decomposer compartments as influenced by environmental conditions. Weber and McFarland (1969) have examined artificial substrate data from a number of environments, both polluted and unpolluted, and arrived at an "Autotrophic Ratio" of 100 or less as being indicative of clean-water conditions.

For organic biomass determinations, each slide was rehydrated for 15 minutes, accumulated material scraped from the slide into a graduated cylinder, then resuspended in a total volume of 50 ml of distilled water. An aliquot of the suspension was filtered on a tared, fired-glass filter (Gelman, GFA), the ash-free dry weight determined (APHA, 1971), and converted to grams of organic matter per square meter as ash-free dry weight.

Net production based on biomass accumulation was calculated by converting organic biomass to equivalent carbon\* then dividing by the incubation period.

Slides collected for chlorophyll  $\underline{a}$  were placed in 50 ml of a 90 percent acetone v/v, and 10 percent of a saturated MgCO3 solution in the field and immediately stored in the dark in dry ice. Prior to analysis, chlorophyll was extracted for 24 hours in the dark at 4°C. To facilitate extraction, slides were scraped and the acetone suspension ground 30 seconds at 500 rpm in a Potter-type tissue homogenizer.

Following extraction, chlorophyll <u>a</u>, corrected for phaeophytin, was determined fluorometrically after the methods of Yentsch and Menzel (1963), Holm-Hansen, et al. (1965), Lorenzen (1967), and Moss (1968), using a Turner Design Model 10 fluorometer. Fluorometric determination of chlorophyll depends on red fluoroscence emitted by the chlorophyll <u>a</u> molecule when excited by ultrabiolet light and is 100 times more sensitive than spectrophotometric analysis. The method is limited to chlorophyll <u>a</u> only; chlorophyll <u>b</u> and <u>c</u> cannot be determined.

The chlorophyll a reference solution was a purified spinach chlorophyll standard (Product No. C5753, Sigma Chemicals, St. Louis, MO) calibrated by spectrophotometric chlorophyll analysis.

Acidification of chlorophyll  $\underline{a}$  converts it quantitatively to phaeophytin. Reading the fluorescence before and after adding one drop of 1N HCl to the sample cuvette allows calculation of an acid factor related to the interference. Periphytic chlorophyll  $\underline{a}$  was calculated as follows:

Chlorophyll 
$$\underline{a}$$
 (mg/m<sup>2</sup>) =  $\frac{(F)(r)(Ca)(ml \ extract)}{(r-1)[substrate area (mm)] 10-3]}$ 

where: Ca = fluorometer reading before - fluorometer reading after acidification)

r = standard before acidification standard after acidification

<sup>\*</sup>Gram organic matter = (2) (grams carbon), Odum, 1971.

$$F = \begin{bmatrix} Ca \\ \hline Fluorometer reading \end{bmatrix} \begin{bmatrix} \frac{\text{dilution ratio fluorometer}}{\text{dilution ratio spectrophotometer}} \end{bmatrix}$$

Net primary productivity based on chlorophyll  $\underline{a}$  accumulation on slides was computed for the 2- and 4-week incubation in terms of grams carbon/square meter based on a chlorophyll to plant carbon ratio of 60 (Strickland and Parson, 1960).

$$\frac{[Chlorophyl] \underline{a} (gm/m^2)] [60]}{days incubated} = Net Primary Production (gm C/m2 day-1)$$

Autotrophic Index. The autotrophic index (Weber, 1973a) indicates the relative composition of the developing periphyton community. this ratio is expressed as:

and has been used to indicate organic pollution and effluent toxicities. The numerical value of this index increases with an increase in non-algal or heterotrophic biomass and decreases with increasing algal biomass. Systems receiving inputs of organic materials will be likely to show elevated heterotrophic biomass and thus a higher index due to proliferation of attached bacteria and protozoa. Nutrient enriched or autotrophic dominated systems on the other hand will approach autotrophic indices of 109-500 (Weber, 1973a) reflecting the ratio of chlorophyll to plant carbon. An autotrophic index greater than 100 indicates organic pollution, less than 100 "clean water" conditions (Weber, 1973b).

# Presentation of Data

A total of 18 filamentous green and blue-green algae species and 100 species of diatoms representing 23 genera were recorded from the VAAP artificial substrate sampling stations. Tables B-2 through B-7, Appendix B, provide an alphabetical listing of these species including cell densities and distribution patterns among stations.

In addition to artificial substrates, a selected number of natural substrates were also analyzed for diatom community structure. Table B-8 presents a list of the diatom species recorded from Lake Chickamauga natural substrates. To quantify further the periphyton community, biomass and chlorophyll <u>a</u> were monitored for the determination of an autotrophic index and net primary production (Tables B-19 through B-31).

Artificial Substrate Colonization Studies. One of the first effects of pollution on periphyton community structure is a change in the reproduction rates of diatom populations (Patrick, 1967). As a result, certain species are unable to reproduce and may become extinct, while tolerant species become more common because of less competition for nutrients and space associated with a reduction in predator pressure.

To ascertain the effects of munitions waste dicharges on the Waconda Bay periphyton communities, diatom artificial substrate cell densities (ceels/mm²) were plotted for the 2- and 4-week surveys conducted during June through July and August through September, 1975. Tables B-13 and B-14 (Appendix B) present mean diatom cell densities for VAAP artificial substrates incubated for the June through July, and August through September 2- and 4-week incubation periods. Means and ranges for diatom cell densities are shown in Figures B-2 through B-5 (Appendix B) for these periods.

Diatom cell densities in Waconda Bay during the June 11-25, 1975, survey averaged 1.06 x  $10^4$  cells/mm², while stations located in Reference Bay A and Huss Lowe Slough averaged 1.56 x  $10^4$  and 1.10 x  $10^4$  cells/mm², respectively. Overall, diatom cell density was 1.19 x  $10^4$  cells/mm². Analysis of variance was utilized to test the hypothesis that all the results belong to populations with a common mean. The calculated F value is significant at the 1 percent level indicating statistically significant differences in the results (F[15, 39] = 14.8, F0.01 [15,39] = 2.54). Ranking of the results shows the lowest cell density at Station A (708 cells/mm²) and two of the three highest densities at Stations B-1 and B-2 (27,902 and 18,531 cells/mm²).

Similar trends were noted during June- July 4-week incubation period (Figure B-3). Analysis of variance gives an F value which is significant at the 1 percent level ( $F_{[10,12]} = 5.05$ ,  $F_{0.01}$  [10,12] = 4.30). Again Station A had the lowest cell density (6,328 cells/mm²). In both the 2-week and 4-week June data, Tukey's w-procedure (Steel and Torrie, 1960) shows the densities at Stations B-1 and B-2 to be significantly different from the density observed at Station A at the 5 percent level.

Somewhat similar trends were seen in the August data. These results suggest toxicity at Station A and biostimulation at Stations B-1 and B-2.

Cairns, Scheier, and Hess (1963) have suggested that a 50 percent reduction in the growth rates of diatoms compares favorably with a 50 percent survival (or TLm) for fish and snails. Cairns (1972) states "this idea is based on the assumption that a toxicant concentration producing a 50 percent reduction in division rate for a microbial population under otherwise optimal conditions would be approximately equivalent to lethal effects on a static population of fish or invertebrates." This assumption was adopted by Patrick, et al. (1968) when comparing the sensitivities of diatoms, snails, and fish to the effects of industrial waste materials. Utilizing these criteria, the impact of munitions waste upon Waconda Bay periphyton can be assessed. Chemical data collected from Station A indicated high concentrations of total munitions waste residues throughout the study period. High levels of total munitions waste residues have produced a localized toxic or inhibitory effect upon diatom populations in the vicinity of Station A during the June - July 2-week and 4-week surveys.

The presence of relatively large diatom populations at the downbay stations suggests that concentrations of total munition wastes were below chronic effect levels. Biostimulatory trends noted at Stations B-1 and B-2 may reflect the impact of  $NO_3$ -N being discharged into Waconda Bay. This nutrient enrichment effect was also apparent at other trophic levels (see Macroinvertebrates).

Similar trends were noted during the August 2-week incubation period. Diatom populations at Station A were again low: 906 cells/mm², 95 percent less than overall mean of Reference Bay A and Huss Lowe Slough (Figure B-4). No significant biostimulation was seen at Stations B-1 or B-2. However, a trend of increasing cell densities was noted at the downbay stations "no wake,"\* E-1, and E-2.

During the August-September 4-week incubation period, populations at Station A were lower than at other stations, but not as much as in June. For example, diatom cell densities at Scation A in June were reduced 84-95 percent below the mean of reference bay stations, whereas in the August 4-week data this reduction was 70-71 percent.

Artificial Substrate Diatom Community Structure. One hundred species of diatoms representing 23 genera were recorded from the Lake Chickamauga artificial substrate sampling stations. Tables B-2 through B-5 (Appendix B) provide a taxonomic list of all species identified, including cell densities (cells/mm²) and distribution patterns among stations during the summer surveys. Tables B-9 through B-12 (Appendix B) present percent relative abundances for the common to dominant species present at each station location during the survey periods. With the exception of two stations (A and F-2), Achnanthes minutissima comprised 75 - 94 percent of the diatom populations during the initial June 2-week incubation period.

Early studies by Geitler (1927) found that A. minutissima and certain species of Cocconeis are usually the first colonizers of glass slide communities. This is probably an indication that these species are not selective about their substrate requirements. The literature indicates A. minutissima to be "one of the most ubiquitous diatoms known" (Hustedt, 1949). Budde (1930a) and Fjerdinstadt (1950) reported this species from eutrophic (i.e. characteristic of waters containing high nutrient concentrations) river systems; Swift (1972) found it as a dominant species in both artificial and natural substrate communities; Scheele (1952) reports A. minutissima to be tolerant of a broad spectrum of environmental conditions. Other common species reported from artificial substrates included:

Fragilaria capucina
Synedra rumpens
Gomphonema parvulum
Achnanthes nollii
Cymbella affinis
Synedra delicatissima
Synedra ulna
Melosira ambigua
Navicula cryptocephala
Nitzschia denticula
Anomoeoneis vitrea

<sup>\*</sup>Station "no wake" refers to an alternate sampling station located midway between transects D and E in Waconda Bay near the navigational sign entitled "no wake."

June Survey. With the exception of Stations A and F-2, diatom populations during the June 2-week incubation period included the following species associations (Table B-9): Achnanthes minutissima (88 percent); Fragilaria capucina (2.1 percent); Synedra ulna (0.9 percent); and Navicula cryptocephala (0.7 percent). A shift in diatom dominance was observed at Station A where the relative abundance of A. minutissima was reduced to 59 percent and the normal Fragilaria-Synedra-Gomphonema flora was replaced by Nitzschia palea (15 percent), Nitzschia cf. capitellata (8 percent) and Synedra rumpens (4 percent). A number of investigators have reported large populations of N. palea as an indicator of organic or toxic pollution (Butcher, 1947; Schroeder, 1939; and Patrick, 1967). Cholnoky (1968) reports N. palea to be an obligate nitrogen heterotroph; tolerant of a wide range of environmental conditions.

In terms of total numbers of species, Station A had the least, with only 14 taxa present, compared to an average of 36 species for Waconda Bay artificial substrates; 28 species for Reference Bay A; and 37 species for Huss Lowe Slough (Table B-2).

Shifts in diatom species associations -- reduction of total species number -- and the increase of those pollution tolerant, correspond with total munitions residues and  $NO_3-N$  being discharged into Waconda Bay at Station A during the June 2-week incubation period.

A shift in diatom dominance was also noted at Station F-2 located at the far end of Waconda Bay. Diatom relative abundance at Station F-2 was dissimilar to the shallow bay stations located in Waconda Bay, Reference Bay A, and Huss Lowe Slough. A. minutissima populations comprised only 57 percent of the diatom flora followed by Navicula cryptocephala (6 percent), Synedra ulna (5 percent), Cymbella prostata (3.7 percent), Fragilaria vaucheriae (3.0 percent) and Gomphonema parvulum (2.9 percent). The distribution of individuals among species present was high and, therefore, the species diversity index (2.07) at Station F-2 was above other shallow bay stations in the study area. The physical and chemical data indicate that Station F-2 was not affected by urban runoff or munitions wastes.

The relative abundance of diatoms (Table B-10) for the June 4-week incubation period was similar to the June 2-week data and included the following species associations: A. minutissima (88 percent); Navicula cryptocephala (1.1 percent); Achnanthes nollii (1.0 percent); Cymbella affinis (0.8 percent); Melosira ambigua (0.7 percent); Gomphonema parvulum (5 percent); Synedra rumpens (0.4 percent); and Nitzschia denticula (0.4 percent).

Station A had a total of 24 species. This represented the lowest number reported and can be compared to an average of 37 species in Waconda Bay; 31 species in Reference Bay A; and 34 species in Huss Lowe Slough (Table B-3).

These data indicated that during the June 2- and 4-week incubation period severe toxicity is manifested at Station A from munitions effluent with the following effects: (1) the elimination of sensitive species; (2) the normal colonizing species (e.g. A. minutissima) are not killed but have had their reproductive rates sharply reduced; (3) tolerant species (i.e. Nitzschia palea) became more common due to reduced competition for nutrients, space, and reduction of predator pressures.

August Survey. Trends established for the June - July period are similar for the August 2-week incubation period. With the exceptions of Station A and "F-Buoy," A. minutissima was again the dominant taxon -- ranging from 70 - 94 percent in relative abundance (Table B-11). Other common species were Fragilaria capucina, Cymbella affinis, Melosira ambigua, Synedra delicatissima, and Achnanthes nollii.

A major shift in diatom dominance was observed at Station A where the relative abundance of A. minutissima was reduced to 48 percent. Nitzschia palea and Nitzschia kutzingiana increased to 11 percent followed by Navicula cf. heufleri v. leptocephala and Melosira granulata at 7 percent. Station A exhibited the lowest total number of species present (14) compared to Waconda Bay, Reference Bay A, and Huss Lowe Slough stations (Table B-4).

Again, shifts in diatom dominance; reduction of total numbers of species; and increases of pollution-tolerant species (N. palea) correspond with munition-waste residues at Station A. The relative abundance of A. minutissima was reduced at Stations D-2, E-1, and F-1. However, population increases of Achnanthes nollii, Synedra delicatissima, Fragilaria capucina, and Synedra ulna were noted (Table B-4). Stations D-2 and E-1 also had higher numbers of species present with corresponding higher diversity indices.

Populations during the August - September 4-week incubation period exhibited similar trends in diatom dominance and relative abundance as noted during the June - July and August surveys (Table B-12). However, at Station A a recovery trend was noted as total numbers of species and individuals showed an increase over populations collected during the previous surveys. The relative abundance of common species present at Station A during the August - September 4-week incubation period were: A. minutissima (83 percent); Cyclotella stelligera (4 percent); Fragilaria capucina (2.0 percent); Nitzschia kutzingiana (1.4 percent), and Achnanthes nollii (1.2 percent). Populations of the pollution "indicator" species, Nitzschia palea, were reduced to 0.6 percent.

The munitions plant halted TNT production near the end of May, 1975. Periphyton recovery from VAAP munitions wastes was not observed until September. These data indicate that toxic conditions persisted at Station A for a period of three months after the cessation of munition waste discharge.

Comparison of Diatom Assemblages. In an effort to compare the various diatom assemblages at each station, the Shannon-Weaver Species Diversity Index (Odum, 1971) and the Pinkham-Pearson (1974) Index of Biotic Similarity were employed. For a review of the indices, the reader is referred to the Computational Methods section. Tables B-16 and B-17 (Appendix B) present a summary of the mean Shannon-Weaver species diversity indices for artificial substrate diatoms during the June - July and August - September 2- and 4-week incubation periods.

Species giversity indices were low throughout the study period with overall mean values ranging from 0.77 to 0.87 for the June - July and August - September sampling periods. Except for Station A, species diversity was generally very low at most bay transects due to the overwhelming dominance of Achnanthes minutissima - comprising 80 - 90 percent of the population. Large populations of A. minutissima produced a low evenness among the species present and therefore a lower Shannon-Weaver index.

Data from this study suggest that the Shannon-Weaver index does not reflect subtle changes in community structure (Tables B-2 - B-5). The diatom population at Station A reflected low numbers of taxa, but each taxa had a relatively even distribution of individuals. This phenomenon is probably a result of the low levels of toxicity reducing the reproduction rates of otherwise dominant forms (i.e. A. minucissima). Therefore, Station A appeared to have a higher diversity although it possessed the lowest number of taxa and total organisms.

These results compare well with those of Patrick (1968) concerning the toxic effects of pH on the structure of diatom communities and those reported by Water and Air Research, Inc. (1977) for munitions waste impact on periphyton populations in the Holston River, Tennessee.

Figures 16 through 19 present biotic similarity expressed in a phenographic display. Each phenogram illustrates the VAAP artificial substrate diatom data with stations clustered on the basis of species occurrence and abundance. In these analyses, it was considered unimportant if a species was mutually absent from two stations and, therefore, 0/0 matches were given a value of zero (mutual absence, unimportant).

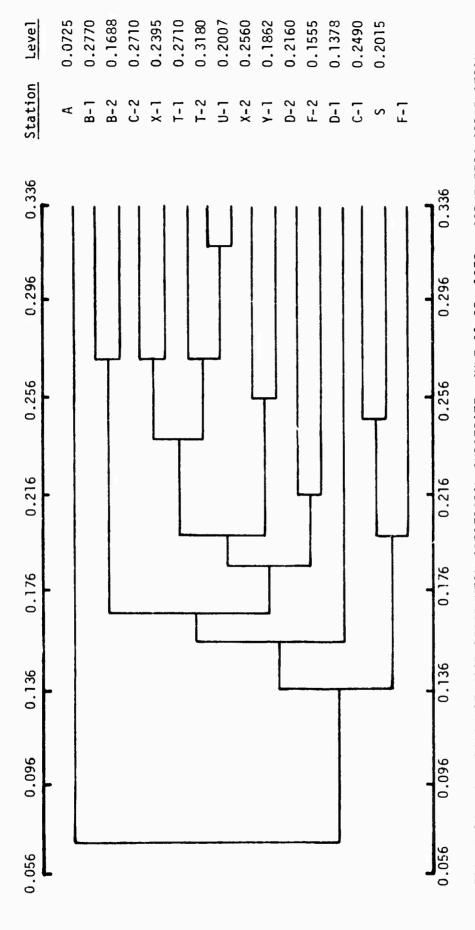
The phenogram (Figure 16) developed from periphyton populations collected during the June 2-week incubation period illustrates the uniqueness of Station A in relation to all others. Diatom populations at Station A exhibited shifts in relative abundance, with significant reductions in cell densities, total numbers of species present, and organic biomass. Total munitions waste residues at Station A during this time ranged to 172 ppb and probably represented the principal stress to the system.

The reference bay stations, T-1, T-2, and U-1, clustered at relatively high levels of similarity whereas stations C-2, X-1, X-2, Y-1, D-2, and F-2 grouped at somewhat lower levels. In the latter group, the X and Y stations were located in Huss Lowe Slough and, therefore, outside the influence of munitions waste.

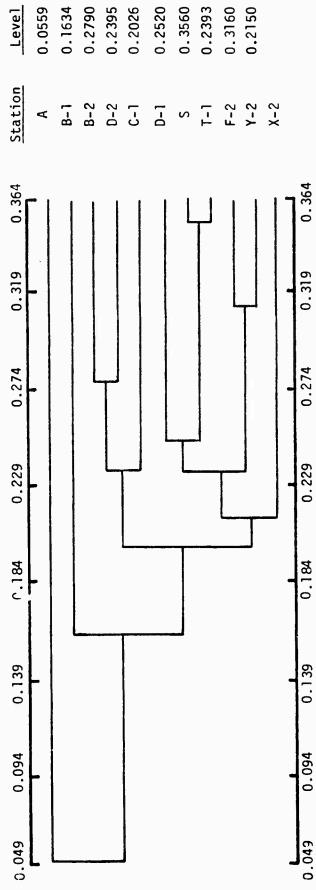
Stations B-1 and B-2 (Figure 16) illustrate a low correlation to the reference stations. These low similarity coefficients were influenced primarily by significant increases in diatom cell densities at Stations B-1 and B-2. As a result, Figure 16 may indicate a trend of biostimulation corresponding to increased levels of NO<sub>3</sub>-N at Stations B-1 and B-2 during the June - July survey.

Figure 17 developed from the June - July 4-week incubation period helps to confirm those trends observed during the June 2-week survey. Again, Station A exhibited a unique diatom assemblage, producing the lowest similarity coefficient. Diatom populations at Station B-1 also indicated a low similarity suggesting biostimulation to  $NO_3$ -N.

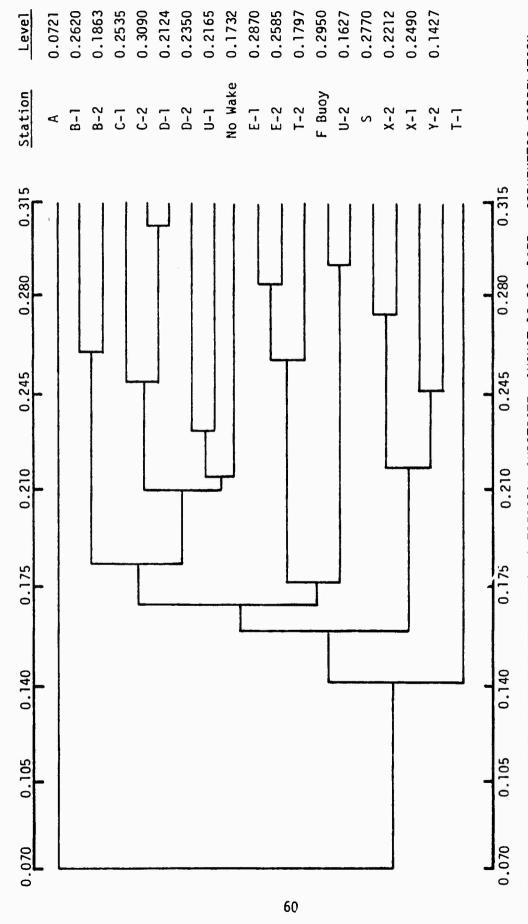
During the August 2-week survey, Station A continued to be markedly dissimilar from all other stations (Figure 18). However, when the incubation period was increased to 4 weeks, the dissimilarity almost completely disappeared (Figure 19). These data suggest that the toxicity at Station A is reduced from June to August.



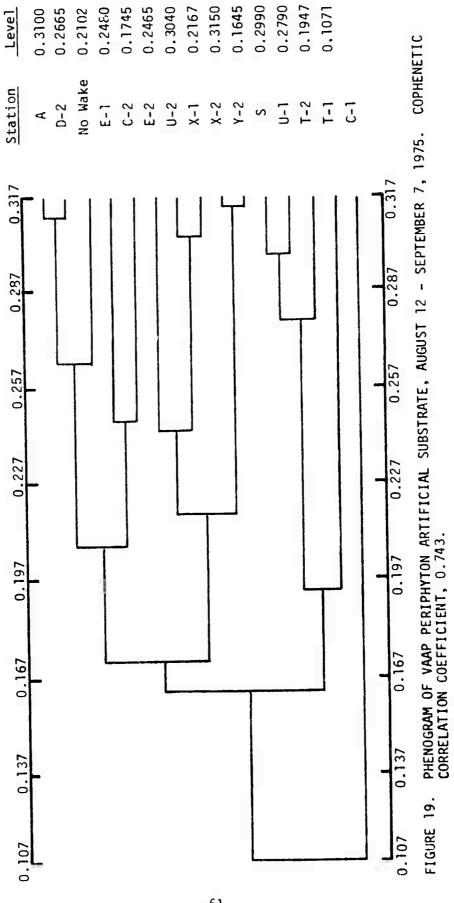
PHENOGRAM OF VAAP PERIPHYTON ARTIFICIAL SUBSTRATE, JUNE 11-25, 1975. COPHENETIC CORRELATION COEFFICIENT, 0.812. FIGURE 16.



PHENOGRAM OF VAAP PERIPHYTON ARTIFICIAL SUBSTRATE, JUNE 11 - JULY 10, 1975. COPHENETIC CORRELATION COEFFICIENT, 0.897. FIGURE 17.



PHENOGRAM OF VAAP PERIPHYTON ARTIFICIAL SUBSTRATE, AUGUST 12-26, 1975. COPHENETIC CORRELATION COEFFICIENT, 0.806. FIGU. 18.



61

FIGURE 19.

Natural Substrate Diatom Community Structure. Table B-8 (Appendix B) presents a detailed taxonomic list of diatom species recorded from VAAP natural periphyton substrates (i.e. rock scrapings). Diatom species (65) representing 18 genera were observed on natural substrates collected from Waconda Bay (Stations A, B-1, and F-2), Reference Bay A (Stations S and T-2) and Huss Lowe Slough (Station X-2).

The dominant species were Achnanthes minutissima -- ranging in relative abundance from 49 - 89 percent. Other common species included: Nitzschia sinuata v. tabellaria, Nitzschia denticula, Navicula cincta, and Cymbella microcephala. The natural substrate community exhibited a more even distribution of individuals among species than the glass slide flora As a result, species diversity indices were higher (Table B-16).

The number of species per station were in general agreement to those reported from artificial substrates during June (Table B-18). A species area curve (Figure B-6, Appendix B) was developed to illustrate the effect of increasing sample size (area microscopically examined in mm²) on total numbers of species. The data graphically illustrated represent the number of species observed per 0.1 mm² interval to a total area of 2.0 mm² (i.e. 64.5 microscope fields under oil immersion, 1000 X). Total species numbers were highest at Stations F-2 and T-2 where 27 - 30 species/mm² were observed (Figure B-6). Lowest values were observed at Station A (10.5 species/mm²), Station B-1 (13 species/mm²), and Station S (13.5 species/mm²).

Diatoms from natural substrates showed a number of similarities to the diatom flora collected from the glass slide communities. Both populations were dominated by  $\underline{A}$ . minutissima; total numbers of species present/station were comparable to the artificial substrates. Quantitatively, however, diatom cell densities/mm² were lower for natural substrates. In addition, species diversity indices were generally higher for the natural substrate flora (Table B-18).

Filamentous Organisms. Attempts to quantify periphytic filamentous organisms on a per unit area basis by the use of a number of standard counting techniques (i.e. Sedgewick-Rafter and inverted microscopic counting chambers) were generally unsuccessful. Most filamentous species clumped with the ends of filaments firmly embedded within detrital particles or entangled with filaments of Oedogonium sp. or Stigoclonium sp. As a result, a qualitative examination was made based on presence-absence criteria at eight selected VAAP station locations (i.e. Stations A, B-1, B-2, C-1, F-2, S, T-1, and X-2) during the June - July and August - September 4-week incubation periods. Tables B-6 and B-7 (Appendix B) lists the most common species present at each of the stations during each survey.

With the exception of Station A, the most common organisms were green algae (Chlorophyceae) comprised of <u>Oedogonium spp.</u>, <u>Mougeotia spp.</u>, <u>Choleachaetae spp.</u>, and <u>Ulothrix spp.</u> Blue-green (Cyanophyceae) algae were represented by <u>Schizothrix calcicola</u>.

Table B-6 shows only one species, <u>Oedogonium</u> spp., at Station A, in comparison to an average 5.3 species per sample from stations located in Waconda Bay and the reference bays, and may indicate toxic effects.

By the end of the August - September 4-week incubation period, some recovery was noted at Station A. Total species increased and a number of heterotrophic organisms (i.e. stalked protozoans) was observed as common components of the periphytic community at Stations A and B-1. These were <u>Vorticella</u> sp., <u>Rhabdostyla</u> sp., and Opisthostyla sp.

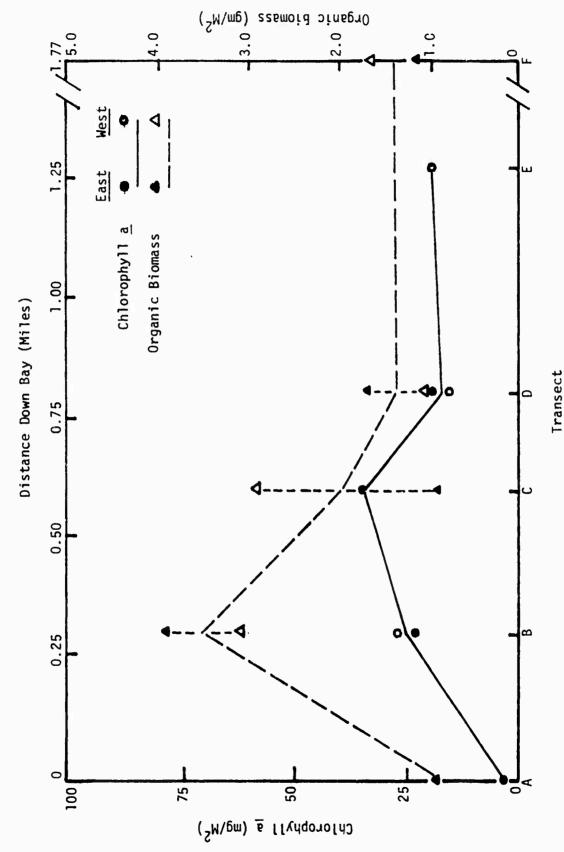
Organic Biomass and Chlorophyll. Loss of periphytometers due to vandalism hampered analysis of periphytic biomass in Waconda Bay and the reference bays. Morphometric conditions in Waconda Bay produced unequal mixing of the VAAP waste discharge. Consequently, higher waste ion concentrations were reported on the west side of the impact bay transects. Incomplete data for both sides of the bays required that the periphyton community be analyzed in terms of transects. This is especially true for the August - September trip. Analysis of variance was carried out on all chlorophyll and biomass data. In order to be consistent with the statistical analysis carried out on the diatom cell count data, only individual stations were considered. Replicates consisted of values obtained for separate individual slides located at a station.

Mean chlorophyll and biomass values as well as corresponding autotrophic indices are presented in Tables B-19 to B-22 and B-31 (Appendix B), and Figures 20 through 23. Tables B-22 through B-29 provide the raw chlorophyll  $\underline{a}$  and organic biomass (AFDW) data for each transect replicate.

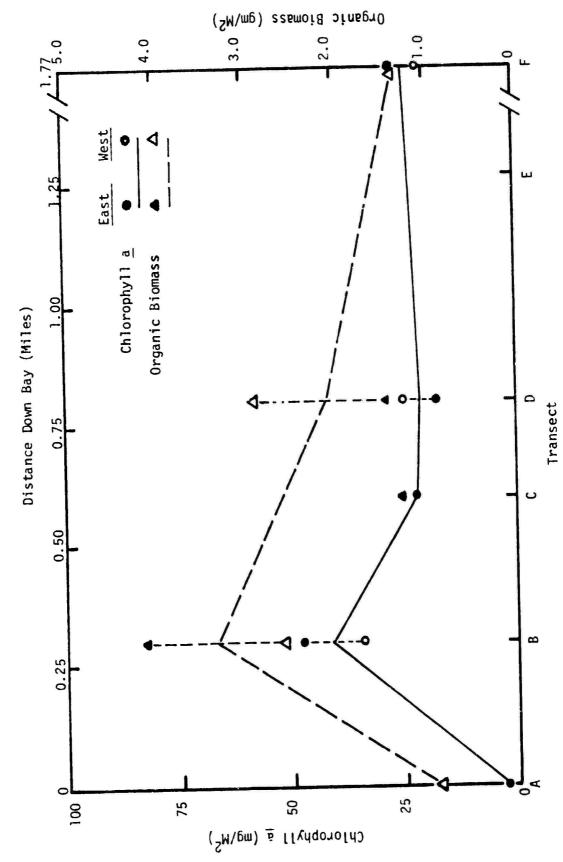
Table B-19 and Figure 20 illustrate the distribution of chlorophyll a and biomass incubated for two weeks in June. Analysis of variance showed that significant differences existed in the biomass values, but not the chlorophyll (Biomass,  $F_{[18,72]} = 10.5$ ,  $F_{0.01} [18,72] = 2.23$ ; Chlorophyll,  $F_{[11,9]} = 1.35$ ,  $F_{0.05} [11,9] = 3.10$ . As with the cell count data, Station A showed uniformly low values while the B stations were uniformly high. Tukey's test (Steel and Torrie, 1960) showed that significant differences existed between Stations A and both B-1 and B-2 at the 1 percent confidence level.

The June 4-week incubation data (Table B-20 and Figure 21) showed significant differences for chlorophyll  $\underline{a}$ , and biomass (Chlorophyll,  $F_{[10,27]}$ = 10.7,  $F_{0.01}$  [10,27] = 3.06; Biomass,  $F_{[10,21]}$  = 4.08,  $F_{0.01}$  [10,21] = 3.31) In both cases Station A ranked lowest or next to lowest. Stations B-1 and B-2 were either the two highest or two of the three highest values. These results are consistent with the diatom cell count results. Tukey's test applied to the chlorophyll  $\underline{a}$  data showed Station A to be different from both Stations B-1 and B-2 at the 1 percent level. With the biomass data Station A was different from B-1 at the 5 percent level, but not significantly different from Station B-2.

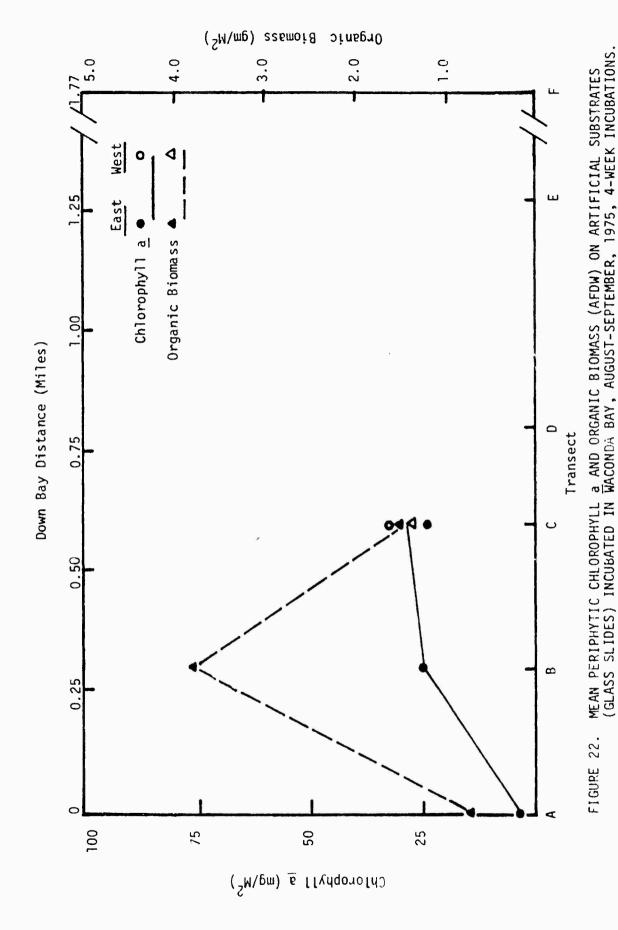
The August data for 2-week incubation (Table B-21) are less conclusive for chlorophyll <u>a</u> and biomass in that these values were not determined for Station A. Station B-1 had the third highest chlorophyll <u>a</u> value and the second highest biomass value. However, cell count data show Station B-1 to be in the lower third when stations were ranked low to high.



MEAN PERIPHYTIC CHLORGPHYLL a AND ORGANIC BIOMASS AS ASH-FREE DRY WEIGHT ON ARTIFICIAL SUBSTRATES (GLASS SLIDES) INCUBATED IN WACONDA BAY, JUNE-JULY, 1975, 2-WEEK INCUBATIONS. FIGURE 20.



MEAN PERIPHYTIC CHLOROPHYLL a AND ORGANIC BIOMASS (AFDW) ON ARTIFICIAL SUBSTRATES (GLASS SLIDES) INCUBATED IN WACONDA BAY, JUNE-JULY, 1975, 4-WEEK INCUBATIONS. FIGURE 21.



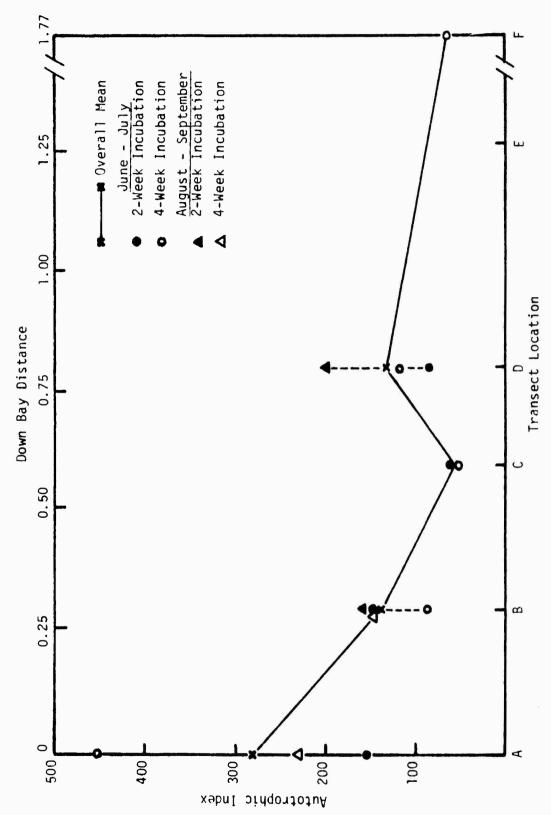


FIGURE 23 . MEAN TRANSECT AUTOTROPHIC INDICES FOR WACONDA BAY, JUNE - SEPTEMBER, 1975.

The August data for 4-week incubation (Table B-22 and Figure 22) are similar to results obtained in June. Station A had the lowest values for cell count, biomass, and chlorophyll <u>a</u>. Station B-1 had either the highest or next to the highest values. Analysis of variance showed significant differences in both the chlorophyll <u>a</u> and biomass data (Chlorophyll,  $F_{[10,19]} = 13.4$ ,  $F_{[0,01]} = 3.43$ ; Biomass,  $F_{[8,11]} = 16.3$ ,  $F_{[0,01]} = 4.74$ ). Tukey's test showed that in both cases Station A was different from Station B-1.

Autotrophic index means for the 1975 summer survey are shown in Figure 23. There is a general downbay trend from stressed conditions at Station A to no stress at transect F. The increase in autotrophic index at transect D may be the result of the marina at the location. Domestic waste from houseboats as well as gasoline engine related wastes could enter the bay at that point. Except at Station A, the mean autotrophic index ratio at 4 weeks generally ranged from 62 to 150, typical of algal-dominated material. Transect X in Huss Lowe Slough was an anomaly, however, in that the autotrophic index in June was 1000. The mean autotrophic index at 2 weeks versus 4 weeks for all stations showed the 4-week data to be lower. These results indicate that the heterotrophic component of the periphyton develops more rapidly than the autotrophic in this environment rather than indicating the presence of pollution stress. Station A indices did not decrease between 2- and 4-week incubation, showing that stress persists.

# Productivity Data

Generally, net production based on chlorophyll and biomass accumulation showed that the periphyton component on glass slides was relatively low as compared to studies previously conducted in waters impacted by munitions wastes (Weitzel, 1975, WAR, Inc., 1976). The exception to this was transect B where relatively high productivity was measured. Production rates of organic biomass tended to be lower for the 4-week incubations compared to the two week while chlorophyll a based productivity rates remained more constant. This is further evidence that heterotrophs colonized the slides more rapidly, building stable population sizes by the end of 2 weeks, while autotrophs continued to increase in mass for the entire 4-week incubation period.

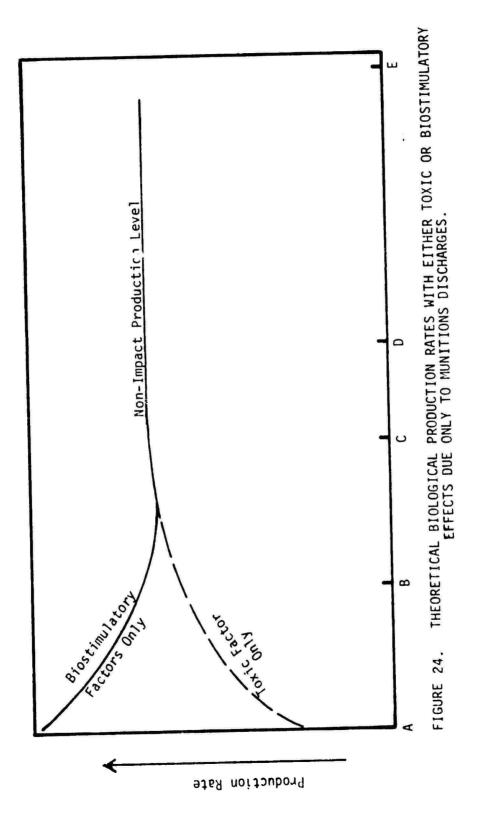
The chlorophyll <u>a</u> and biomass data for artificial substrates suggest several conclusions regarding autotrophic and heterotrophic microbial communities in Chickamauga Lake. Primarily inhibition of microbial growth occurs at the impact point, Station A. Biomass of both autotrophs and heterotrophs is reduced, however, autotrophic metabolism seems to be suppressed to the greater degree. Biomass of both components is increased at transect B, 0.3 miles downbay from A and some stimulation is evident at C, 0.6 miles downbay from A. TNT and DNT residues, while found to be highly variable in concentration, decreased from a median value of 123 ppb at A to about 50 ppb beyond transect C in June and 121 to 60 ppb for the same reach in August. Nitrite plus nitrate-nitrogen decreased over the same distance from >5 mg N/1 to about 1.0 mg N/1 in June and 2.3 to 1.0 mg N/1 in August. Total reduced nitrogen, sulfate, and hardness showed similar

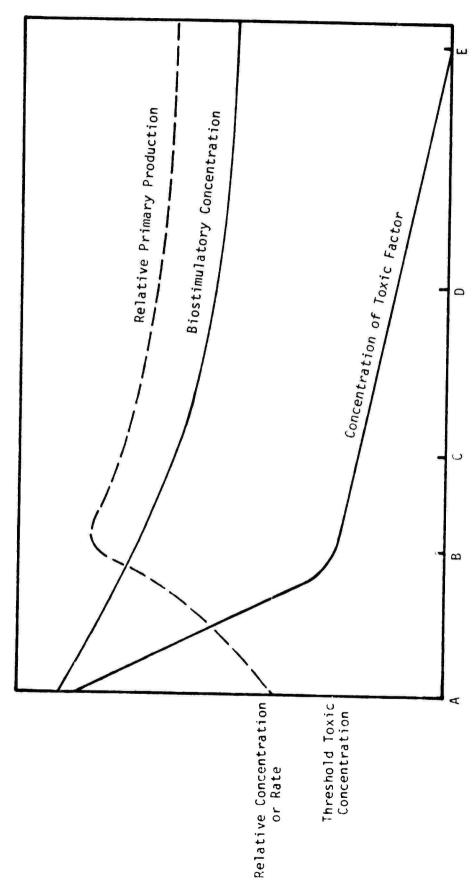
patterns of concentration. The concentrations of the nitrogen forms, anions, and cations measured, however, are unlikely to be totally responsible for the biological effects. The nitrate-nitrite and Kjeldahl nitrogen concentrations should cause a definite biostimulatory effect. Pink-water conditions at Station A may reduce light, causing a reduction of algal biomass at that station. This does not, however, explain the effect on the heterotrophs. Elevated sulfate concentrations do not likely explain the inhibition as it existed in the August survey when sulfate concentration had dropped at A to the same levels as was found at transect B in both June and August in conjunction with biostimulation. The enrichment at transects B and C is an indication that Harrison Bay may be nitrogen limited overall, however, the enrichment shown may be due to a combination of factors rather than simple nitrogen limitation since slightly elevated nitrogen concentrations at the offshore stations on the west side of Harrison Bay do not increase chlorophyll or biomass over eastern stations.

It is likely that two major munitions related factors interact to produce the patterns of periphytic growth in Waconda Bay. These are illustrated conceptually in Figures 24 and 25. In Figure 24, the pattern of growth rate or production as affected by biostimulatory factors only is illustrated as a decrease downbay to non-impact production levels. Such factors might include nitrogen compounds (TKN, NH3, and NO3) from the effluent. Growth rate would parallel nutrient concentrations which would decrease with dilution downbay. Concentrations of toxic factors would also be diluted downbay. The effect on periphyton growth would be the opposite of the biostimulatory effect such that as the toxic factor was diluted the growth rate would increase. Toxic or suppressive effects could result from munitions residues, physical effects of "pink water," or some synergistic effect related to SO4, NO3. TKN, and munitions. Figure 25 illustrates the probable effect of both biostimulation and toxicity on primary production. Below a critical level biostimulation or suppression of overall productivity becomes negligible and levels of production are not different from offshore by conditions. The interaction of these two effects is shown as a dotted line in Figure 25, where production level balances between suppressive and biostimulatory effects. Initially, suppression of production occurs in the presence of strong biostimulatory concentrations of nutrients. As the concentration of toxic materials is reduced downbay. biostimulation overrides suppression. Primary production then peaks and decreases with the lower concentrations offshore.

Examination of the decreases in munitions concentration indicates that above 100 to 150 ppb TNT toxic effects may occur, below about 50 ppb there is no effect on overall periphytic community production. Concentrations of munitions at A and B-2 appear to range to similar levels although community response is markedly different. This suggests that other factors than just the strict TNT, DNT concentrations may be active, for example a synergism between nitrate-nitrite level and munitions, or that the more consistent presence of high levels munitions residues at A than in transect B causes a more integrative effect than intermittent munition levels above 100 ppb at B.

Shading due to "pink water" and turbidity is also a likely suppressive factor at A which has been reduced at B, allowing biostimulation due to nutrients to be expressed.





THEORETICAL BIOLOGICAL PRODUCTION RESULTING FROM BOTH BIOSTIMULATORY AND TOXIC EFFECTS FROM MUNITIONS DISCHARGES. FIGURE 25.

There are also some downbay differences in biological production evident in the reference bays; some of the effects in Waconda Bay may be due to runoff factors such as turbidity. These biological effects in the reference bays do not exhibit the consistency or magnitude of the effects in Waconda Bay.

### Introduction

Plankton are free floating organisms suspended within the water column of aquatic systems. They span all trophic levels and consist primarily of three components: 1) the phytoplankton -- free floating microscopic plants (algae); 2) the zooplankton -- free living microscopic animals (protista, rotatoria, mollusca and crustacea); and 3) the meroplankton -- free floating eggs and/or larvae of certain invertebrates and fishes. Chlorophyll-bearing plants such as algae usually constitute the greatest portion of plankton biomass. Phytoplankton use the energy of sunlight to metabolize inorganic nutrients and convert them to complex organic materials. Zooplankton and other herbivores graze upon the phytoplankton and, in turn, are preyed upon by other organisms -- thus transferring stored energy to larger, more complex organisms. In this manner, nutrients become available to large organisms such as macroinvertebrates and fish (Weber, 1973).

The phytoplankton community response to pollution stress is similar to that of other microbial populations. In waters severely affected by organic pollution, heterotrophs may be extremely abundant; sometimes with a biomass exceeding that of algae. As a result of heterotrophic metabolism, high concentrations of inorganic nutrients become available and massive algal blooms may develop. Plankton blooms can cause extreme fluctuations of dissolved oxygen in water, may cause taste and odor problems -- and if present in large numbers, are aesthetically objectionable. In some cases, plankton may be of limited value as indicator organisms as they move with the water currents; thus, the origin of the plankton may be obscure and the duration of exposure to pollutants may be unknown (Weber, 1973.)

The quantity of phytoplankton occurring at a particular station depends upon many factors including sampling depth, time of day, season of year, nutrient content of water, and the presence of toxic materials.

### Methods

Phytoplankton samples were collected June 9-13 and August 11-15, 1975, at 20 stations in Lake Chickamauga, (Figures 3 and 4). Surface samples were taken and preserved with five percent neutralized formalin and placed in a cool, light-excluding box for shipment back to the laboratory. Plankton enumeration was made according to the Utermohl (1958) sedimentation technique. Phytoplankton counts were performed on the following samples: 6/9, 6/10, 6/13, 8/11, 8/13, and 8/15, 1975. Species identifications were carried to the lowest taxonomic level, utilizing a Zeiss inverted "D" microscope at 400 diameters, while species identifications were carried out utilizing an oil immersion lens (1250 magnification). Four strip counts were made of each settled sample for an estimate of numbers of organisms per liter. Phytoplankton species occurring commonly as groups, clumps, or filaments were considered as one unit (e.g. Micractinium, Gomphosphaeria, and Chroococcus). The diatom Melosira, however, was a noted exception as filaments were usually shorter than 6 cells and were therefore counted as individual units. Major taxonomic references used were Drouet and Daily (1956); Drouet (1968);

Prescott (1954, 1962); Witford and Schumacher (1969); and Patrick and Reimer (1966).

Shannon-Weaver Species Diversity Index  $(\bar{H})$  (Shannon and Weaver, 1949; Margalef, 1968) and the Pearson-Pinkham (1974) Index of Biotic Similarity were used to make station to station comparisons. These indices are explained in detail in the Computational Methods section of this report.

# Presentation of Data

A total of 71 phytoplankton species representing at least 70 genera were present from Lake Chickamauga during the two 5-day sampling periods. Tables C-1 through C-4 present taxonomic lists of species occurrence and distribution patterns between stations for four typical sampling trips. Figures 26 through 31 illustrate the percent relative abundance of the various phytoplankton groups at each station during the two surveys. The remainder of the abundance data is illustrated in Appendix C. Means, standard deviations, and coefficients of variation for phytoplankton cell densities (cells/ml); total numbers of species per station; as well as Shannon-Weaver species diversity estimates\* are also tabulated in Appendix C.

Diatoms (Bacillariophyceae) were the predominant organisms at Stations A through D (Waconda Bay) during the June survey -- accounting for 38 - 68 percent of the phytoplankton population. The dominant to common diatom species during the period were approximately: Melosira ambigua (15 percent), Melosira distans (14 percent) and Fragilaria crotonensis (5 percent).

Other important groups were: the Chlorophyceae (green algae) represented by Ankistrodesmus falcatus, Scenedesmus bijuga, S. quadricauda, Tetraedron minum, and Chlamydomonas sp.; the Chrysophyceae, including Dinobryon divergens, and D. bavaricum; the Cryptophyceae, with Cryptomonas sp. and Rhodomonas sp.; and the Dinophyceae represented by Peridinium pusillum.

Diatom dominance decreased from Station A to D (Waconda Bay) with equal numbers of diatoms and Chlorophyceae present at transects E and F. The relative abundances of diatom populations were generally lower at the offshore transects E-F and higher in Waconda Bay and Huss Lowe Slough. The relative abundance of Chlorophyceae populations increased at transects E-F and X-Y (Figures 26 and 27).

With the exception of the June 9 sampling period (Figure 27), similar population trends of diatom-chlorophycean dominance were observed in Reference Bay A and Huss Lowe Slough.

During the August 5-day survey, the phytoplankton was dominated by the Cyanophyceae (blue-green algae), (Figures 30-31) comprising 36-43 percent of population followed by the diatoms and Chlorophyceae representing about 21-20 percent.

<sup>\*</sup>Based on two replicates.

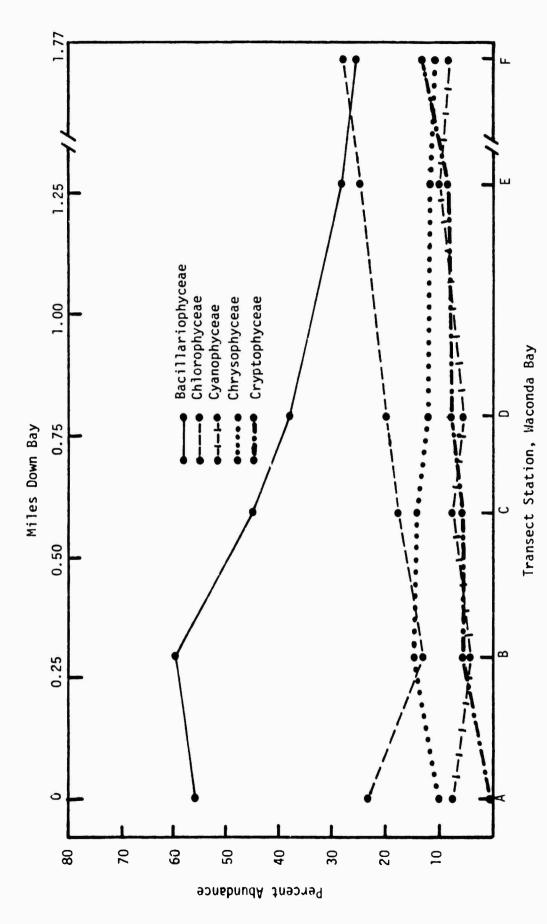


FIGURE 26. DISTRIBUTION OF MAJOR GROUPS OF PHYTOPLANKTON IN WACONDA BAY, JUNE 9, 1975

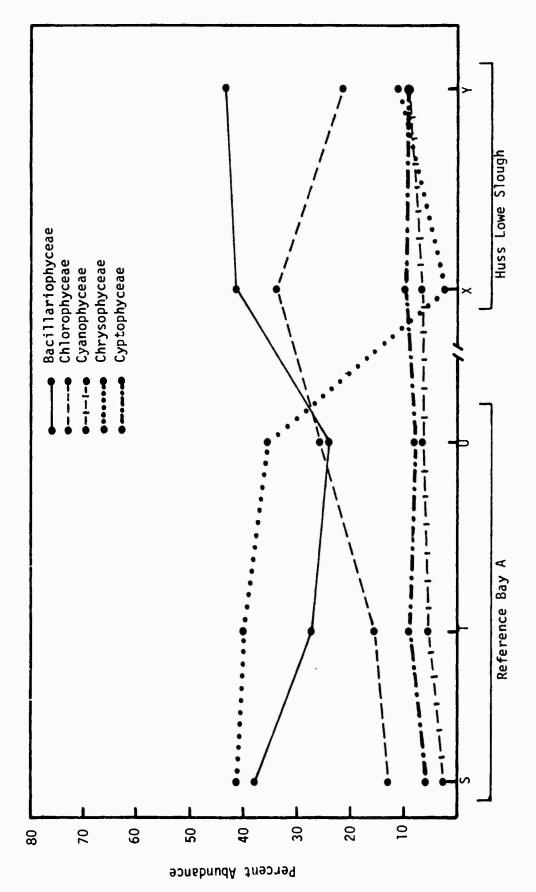


FIGURE 27. DISTRIBUTION OF MAJOR GROUPS OF PHYTOPLANKTON IN THE REFERENCE BAYS, JUNE 9, 1975.

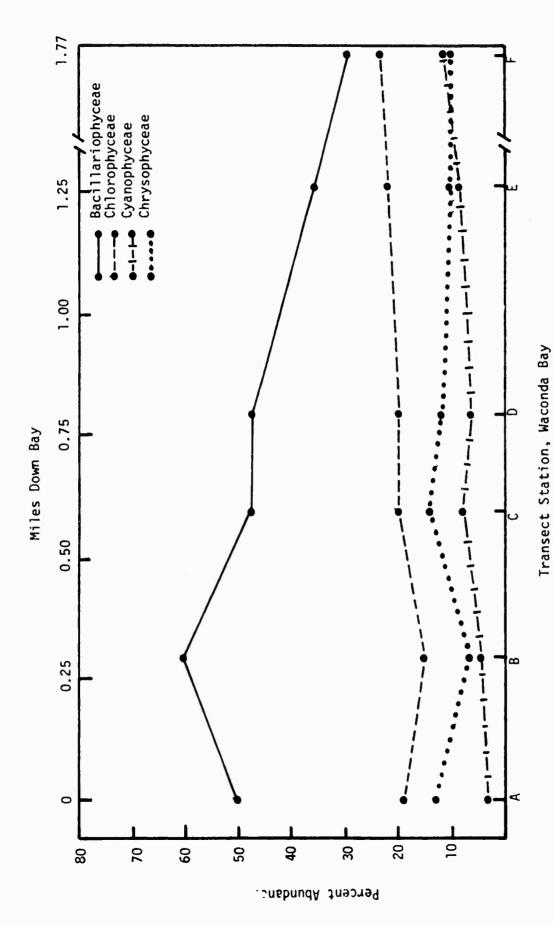


FIGURE 28. DISTRIBUTION OF MAJOR GROUPS OF PHYTOPLANKTON IN WACONDA BAY, JUNE 10, 1975.

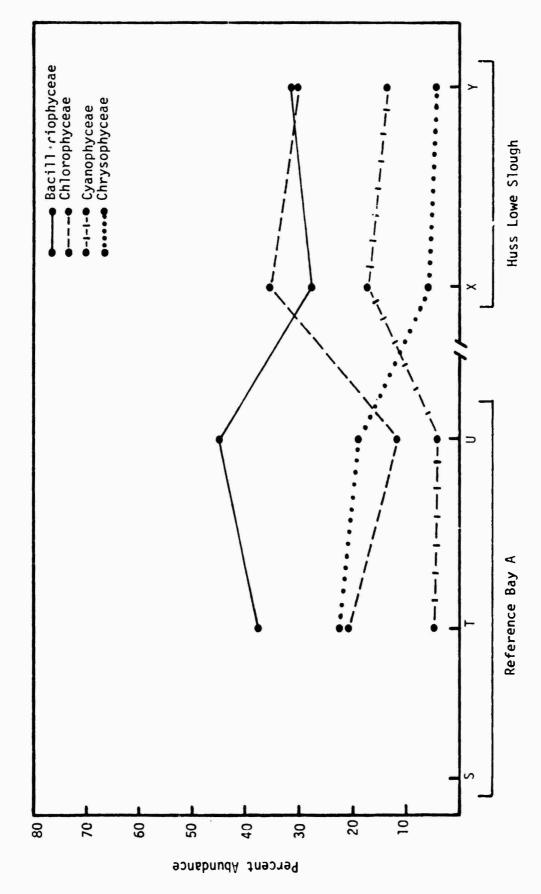


FIGURE 29. DISTRIBUTION OF MAJOR GROUPS OF PHYTOPLANKTON IN THE REFERENCE BAYS, JUNE 10, 1975.

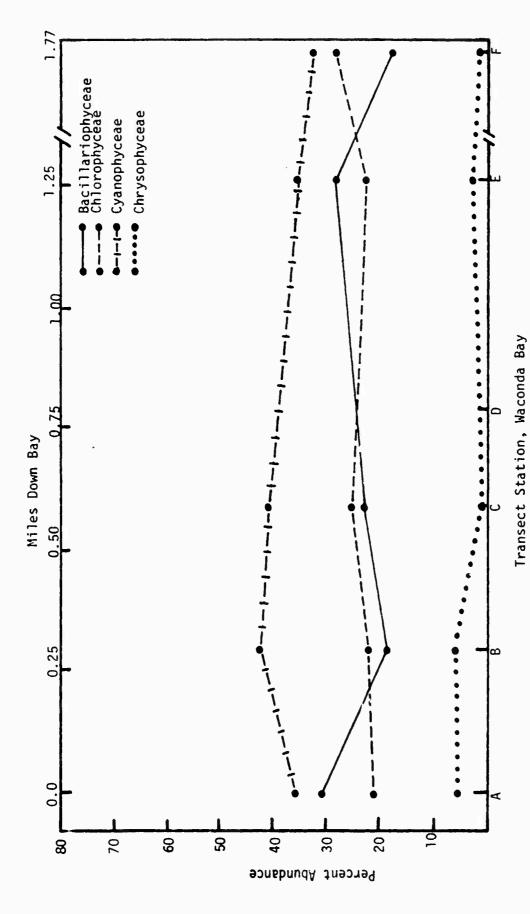


FIGURE 30. DISTRIBUTION OF MAJOR GROUPS OF PHYTOPLANKTON IN WACONDA BAY, AUGUST 11, 1975.

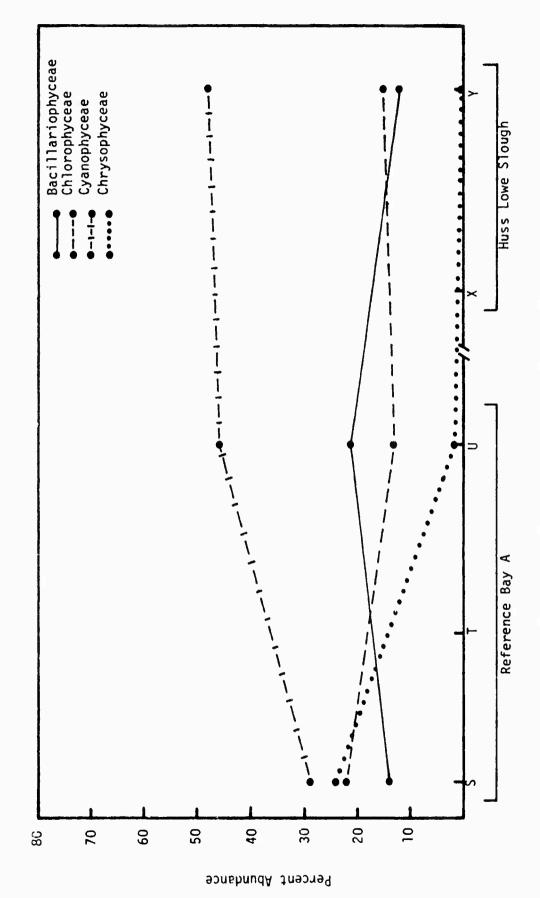


FIGURE 31. DISTRIBUTION OF MAJOR GROUPS OF PHYTOPLANKTON IN THE REFERENCE BAYS, AUGUST 11, 1975.

Bloom conditions (i.e. >500 cells/ml) were reported for the blue-green species, Spirulina laxissima and Gomphosphaeria lacustris at Stations A, B-1, C, and Y-1. Other important blue-green species were: Oscillatoria geminata (Schizothrix calcicola)\* and Aphanocapsa delicatissima (Anacystis incerta). Prominent non-Cyanophytes were: the diatoms Melosira distans, M. ambigua, Chamydomonas sp. (Chlorophyceae), Rhodomonas sp. and Cryptomonas sp. (Cryptophyceae).

Phytoplankton populations (cells/ml) were generally higher during the second survey. Mean cell densities ranged from approximately 1300 to 2600 in June as compared to 2100 to 5400 in August. Analysis of variance on the June data showed no significant differences among station mean values at the 5 percent level ( $F_{[19,38]} = 1.48$ ,  $F_{[0.05]} = 1.87$ ). However, the three highest cell densities were at Stations B-1, B-2 and C-1. Station A had a density slightly below the overall mean. The elevated populations at Stations B-1, B-2 and C-1 suggest possible biostimulation in the upper end of Waconda Bay. In August greater differences existed between stations. Analysis of variance indicated that significant differences did exist at the 1 percent level ( $F_{[10,27]} = 4.62$ ,  $F_{[0,01]} = 3.06$ ). The three highest densities were found at Stations B-1, C-1, and A, respectively. If any plankton toxicity ever existed at Station A, it was not evident in August, some three months after plant shut-down. However, biostimulation in the upper end of Waconda Bay was indicated. This trend was also reflected in the other biological compartments examined.

The response is more subtle in the phytoplankton, and, in terms of the number of plankton species recorded per station, no significant trends were observed that could be attributable to munitions waste. Mean values for numbers of species/station ranged from 44 to 53 during the June survey and from 52 to 55 species/station for the August survey. The highest value recorded through the study was 71 species at Station B-1 (August 11, 1975); the lowest, 35 species at Station U-2 (June 9, 1975).

Phytoplankton populations were compared on a station-to-station basis by employing the Pearson-Pinkham Biotic Similarity Index (1974). Each phenogram illustrates the VAAP phytoplankton with stations clustered on the basis of species occurrence and abundance. In these analyses, it was considered unimportant if a particular species was mutually absent at two stations (i.e. mutual absence, unimportant).

With the exception of phytoplankton data collected on June 9, no significant trends were observed as a result of biotic similarity cluster analysis during June and August. The phenograms plotted in Appendix C depict impact stations readily clustering with control stations in both Reference Bay A and Huss Lowe Slough. On June 9, Stations A, B-1, and B-2 exhibited relatively low levels of biotic similarity to all other stations, and indicated a

<sup>\*</sup>Drouet (1968) classification.

possible impact. Using this technique of data analysis, the trend was not observed on the following day's analyses (June 10) -- nor during other sampling periods. Based on these results, it appears that the phytoplankton is a marginal biological compartment to discriminate waste impact in Waconda Bay from VAAP during plant operation. There is limited evidence, however, that limnoplankton may be useful to delineate areas of biostimulation associated with TNT decomposition in aquatic systems.

### **MACROINVERTEBRATES**

## Introduction

Aquatic macroinvertebrates are a diverse group of small aquatic animals too large to pass through a U. S. Standard No. 30 mesh screen. They are comprised of snails, clams, arthropods, annelids (segmented worms and leeches), planarians, and coelenterates. Of these, oligochaetes and chironomid (midge fly) larvae account for the majority of the organisms in this study.

Aquatic macroinvertebrates are a major biological component of aquatic systems and form an important part of the food chain. They feed on detritus and microscopic plants and animals. They are in turn eaten by small fish which support the larger recreationally and economically important species. They are of special importance in stream environments because of their role in recycling large amounts of organic detritus introduced from uplands.

Macroinvertebrate species composition (density, number of taxa, and diversity) is primarily dependent on three factors -- water quantity, water quality, and substrate composition.

Water quantity limits species within a site. For example, some organisms prefer large, deep lakes while others are found in smaller, shallower lakes.

Water quality is a significant factor in determining the assemblage of macroinvertebrates. Principal parameters include oxygen, temperature, hardness, and dissolved solids. The most important of these is oxygen. While many species require oxygen-saturated water in order to thrive, others can tolerate reduced oxygen tensions. Aquatic macroinvertebrates are also affected by temperature extremes. The Aquatic Life Advisory Committee (1956) indicates that benthic communities in temperature zones are adapted to seasonal fluctuations of temperature between 0 and 32°C (32 - 90°F).

Substrate is the most important determinant in species composition (Hynes, 1960). There is a direct relationship between amounts of available surface area and species abundance and diversity. That is to say, there are more hiding and foraging places in a rock or pebble bottom than in a sand or mud bottom. The amount of organic matter, particularly from plants, is also important. Aquatic plants increase the abundance and diversity of benthic organisms viz. there is more surface area, periphytic food organisms, food from the plants themselves, and detritus on which to feed. Beck (1954) states, "...after careful examination of many streams, diversity of fauna was primarily the result of one factor -- the diversity of habitat."

Aquatic macroinvertebrates were chosen as a parameter for this study because they are sensitive to environmental changes and thus are important indicators of water quality. Natural or man-induced fluctuations in the physical-chemical characteristics of a lentic system are reflected by shifts in benthic community structure. They are useful as an integrated monitor of the environment. They tend to remain at fixed locations and they have a relatively short life span -- usually a year or less -- therefore, reflecting both the present and recent past environmental conditions.

### Methods

Aquatic macroinvertebrates were collected from natural and artificial substrates during June - September, 1975. Sampling information is tabulated in Appendix F. Natural substrates were sampled with a petite Ponar dredge. Hester-Dendy artificial substrates were suspended approximately 1.5 - 3.0 feet below the surface. Five replicates of the natural substrate and three Hester-Dendy units were collected to minimize natural variability. The number of replicates were determined utilizing the information presented in Appendix E.

In the field, dredge samples of the natural substrate were washed in a bucket sieve (U.S. Standard No. 30 mesh) and bottled. Rose Bengal dye was then added to facilitate laboratory sorting. Samples were preserved in 10 percent formalin. Natural substrate samples were rewashed in the laboratory and picked in a white enamel pan partially filled with water. After sorting, organisms were placed in vials containing 95 percent ethanol. Chironomid larvae were mounted in polyvinyl-lactophenol for microscopic identifications. Identifications were made to the lowest practical taxonomic level. Verification of chironomid identifications were made by Mr. W.C. Beck of Florida A & M University. Key taxonomic references used in this study were Edmondson (1959) and Pennak (1953).

The community structure indices computed for aquatic macroinvertebrates are the Shannon-Weaver Species Diversity Index and Pearson-Pinkham Index of Similarity.

# Presentation of Data

Chironomid (midge fly) larvae and oligochaetes (sludgeworms) were the dominant groups in the Chickamauga Reservoir bays.

Midge larvae were more than 50 percent of the population during the survey. They colonized artificial substrates abundantly and accounted for nearly 80 percent of the total number of organisms (Tables 6 through 9).

Of the 54 taxa enumerated, 44 were associated with the surfaces of Hester-Dendy plates. This suggests that chironomids preferentially colonize artificial substrates of this type to graze on periphyton. Lake sediments generally were composed of clay, silt, and detritus, which are also conducive to chironomid colonization. Lower population and taxa (37) in the latter environment probably reflects periphyton enrichment on artificial substrates.

In contrast to the chironomid distribution, oligochaetes reached higher populations in sediments. This is to be expected since these organisms burrow into the substrate and recycle organic materials. This kind of habitat is not available on artificial substrates as used in this study.

Both of these groups exhibit similarities in their response to environmental conditions. The chironomids, which dominate, are adaptable to changes in pH, oxygen concentrations, and turbidity. They prefer moderate-to-high nutrient concentrations and will tolerate some organic enrichment.

TABLE 6
VAAP MACROBENTHOS ARTIFICIAL SUBSTRATE, JUNE, 1975, POPULATION SIZE EXPRESSED PER M2 BASED ON POOLED REPLICATES

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CMANICCHIMUMCHAR ANTANA CMANICCHIMUMCHAR MTWMINY CMICOTOMAR & W	,;	.;	ųį			•					
OICHOTO-MCIOPS LEUCTOSCELIS CH-HTTTPMCIPES SH CH-HTTTPMCIPES SH CH-HTTTPMCIPES SH	15	,				*****					
DICHOTENGINES MODESTUS DICHOTENGINES NEGRODESTUS DICHOTENDINES SP 4	•	:	.1								
SECICLADIUS ON BLTSTCTSHDIRES SO HAMNISCHIS OF		:	:			•					
Macmorpicala se Mancolabius se Mancolabius se items:1	:		:	•	• • •••••	•	•	: :::::			
Banechimchous & Strs. Panechacchim + cm Panechacchim + cm	1	: :	1	• • • • • • • • • • • • • • • • • • • •	• • • • • • •	•	•	:			
BARATPHDERS CIMMPCTENS POLTHODELUM HALTOBALP MARTE HILLS SO		7	,,		•		: :::::				
PS0(THDCL0D)US 4P PSP(THDCH)ANDCHIS 5P BH6(TANTIARUS 5P		:	,,,	: :::::		: :::::		• • • • • • • • • • • • • • • • • • • •			
SOLITIA OU STRONGLING SP STRONGLING SP TRANSUS OU		:	:	• ••••	• ••••	• ••••	• ••••	•			
 	103	100	***			• • • • • • • • • • • • • • • • • • • •	• ••••				
RENECHIOCHOUG 46 CHIOCHOUG SH & CHISCHOUGH G B		,		• ••••			• ••••	•			
CHEMOMORIDA UNIDPHTIFIANT CHEMOMORIDA UNIDPHTIFIANT	: '	•	. "		•••••	• ••••	• • • • • • • • • • • • • • • • • • • •	•••••			
CHACOPOUS AP CEROSCONGOMEDAE UNIDENSISSISSIS CULTCIDAE UNITENSISSISSISSI		.;	•	•		•		•			
••••••	<u>:</u>	······	••••••		•	••••••	•				
America Chellettress	:	:	• •	:		•	•				
HYTR6 4P			-	*****		•		••••			
	<b>!.</b>		· · · · · · · · · · · · · · · · · · · ·		•		: ••••••••••••••••••••••••••••••••••••				
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v1v1040t-q •m						••••	•••••				
CLASS PYLSCYPTOS		: ;;			• • • • •		• ••••	•••••			
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**************************************				•	•	•					
				•							
TUMPRELARIS AM 8	:		•		• ••••	•	•	•			
TOTAL MANAGER OF COCCUSES	** 10	.,	4044	•••••							
terming in the + a.a.d.	• ••	1*	• ••	•••••	•••••	• • • • • •	•••••	•••••			

TABLE 8

TAXONOMIC LIST OF VAAP MACROINVERTEBRATES:
ARTIFICIAL SUBSTRATE, AUGUST 1975
POPULATION SIZE EXPRESSED PER M<sup>2</sup> BASED ON POOLED REPLICATES

PHYLUM ANNELIDA	Α	B1	C1	D2	El	Fl	S	TI
· OL 1 GOCHAETA	•	-	-	-	-	-	221	-
PHYLUM ARTHROPODA - CLASS INSEC	ta l							
ORDER EPHENEROPTERA								
BAETUS SP	-	_	-	-	-	-	_	
ČAĒNĪS ŠP UXYETHIRA SP		1281	-	-	-	-	.=	=
STE NONEMA SP	-	-	23	-	11	127	-	-
ORDER ODONATA								
NEHALENN (A SP	11	34	11	-	-	-	-	-
ORDER TRICUPTERA								
CYRNELLUS MARGINALIS	56	114	1186	710	519	190	218	591
URDER COLEOPTERA								
DINEUTUS SP	-	-	-	•	•	-	-	•
ORDER DIPTERA - FAMILY CHIRUNOMIDA	E							
PULABESMYIA AMERICANA ABLABESMYIA ANNULATA ABLABESMYIA MALLOCHI	11	-	=	-	-	-	-	-
ABLAGESMY IA PARAJANTA	34	22	687	44	222	80	136	329
CHIRONUMUS ATTENUATUS CUELUTANYPUS CONCINNUS	11	11	-	-	•		26	=
CRICUTOPUS SP A CRYPTOCHIRONOMUS FULVUS	-	-	45	_	•	-	13	45
CRYPTOCHIRONOMUS SP A DICROTENDIPES LEUCUSCELIS	150	. 126	234	90	11	13	- 26	186
DICRUTENDIPES NEOMODESTUS DICRUTENDIPES NERVOSUS		94	437	281	i 27	27	13	34
DICRUTENDIPES SP A GLYPTUTENDIPES SP		6421	580		162	136	275	- 92
GOELDICHIRONOMUS HOLOPRASINUS	11	-	•	-	-	-	-	-
PARACHIRUNUNUS CARINATUS PARACHIRUNUNUS MUNUCHHUMUS PARACHIRUNUMUS PECTINATELLAE		23	-	11	•	13		
PARALAUTERHORNIELLA NIGRUHALTERALÉ POLYPEDILUM FALLAX	11	•	•	-	•		40	
PSECTRUCLADIUS ŠP	ii	-		-	-	13	-	5.6 1.1
PSEUDICHIRONOMUS SP RHEGTANYTARSUS SP ITANYTARSUS SP	=	=	90	=	•	40 13 233	26	1.1
THIENEMANNIELLA SP	-	-	-	-		_	.	-
TRIBELUS SP A Chironomid Sp A	-		45	68	91	13	13	22
CHIRONOMIDAE UNIDENTIFIABLE	-	-	11	-	•	13	-	1.1
ONDER DIPTERA - UTHER								
CERATOPOGONIDAE UNIDENTIFIABLE	-	•				<u> </u>	13	•

TABLE 8 (CONTINUED)

PHYLUM MOLLUSCA	A	B1	C1	DŽ	El	Fl	5	Tl
CLASS GASTHOPODA Physa sp	46	9	•	•	•	•	•	•
CLASS PELECYPUTA SPHAERIUM SP	•	•	•	•	•	•0	13	-
PHYLUM NEMATUDA	11	366	•	•	•	•	•	•
PHYLUM TURBELLARIA								
TURBELLARIA SP B	-	-	•	•	•	. 13	-	•
TOTAL NUMBER OF GREANISMS NUMBER OF TAXA	406 13	1	3361 13	1204	1315 9	1051	1086	1600

TABLE 8 (CONTINUED)

PHYLUM ANNELIDA		VI	V1
OLI GOCHAETA	υ1 -	X1 104	Y1
PHYLUM ARTHROPODA - CLASS INSECTA			
DRDER EPHEMERDPTERA			
BAETUS SP	925	-	197
CAENIS SP OXYETHIRA SP	174	-	-
STENONEMA SP	11	-	-
ORDER COONATA			
NEHALENNIA SP	11	-	-
ORDER THICOPTERA			
CYRNELLUS MARGINALIS	45	425	102
ORDER COLEOPTERA			•
DINEUTUS SP	-	11	-
ORDER DIPTERA - FAMILY CHIRONOMIDAE			
ABLABESHYIA AMERICANA ABLABESHYIA ANNULATA ABLABESHYIA MALLOCHI	11	11 11	-
ABLABESMYIA PARAJANTA Chirənghus Attenuatus	116	269	22
COELOTANYPUS CONCINNUS	-		t. –
CRICOTOPUS SP A CRYPTOCHIRONOMUS FULVUS CRYPTOCHIRONOMUS SP A	ι <u>ī</u>	11	11
DICROTENDIPES LEUCOSCELIS	55 -	23	56
DICHOTENDIPES NERVOSUS	128	• 46 •	h 81
DICROTENDIPES SP A GLYPTOTENDIPES SP GOELDICHIRONOMUS HOLOPRASINUS	484	1185	389
PARACHIRUNOMUS CARINATUS PARACHIRUNCMUS MONJCHROMUS PARACHIRONOMUS PECTINATELLAE	22	-	114
PARALAUTERBOHNIELLA NIGROHALTERALE POLYPEDILUM FALLAX PSECTROCLADIUS SP	- 11	-	=
PSEUDOCH LACNONUS SP RHEOTANY TAR SUS SP	• 23 • 11	567 11	-
TANYTANSUS SP		11	
THIENEMANNIELLA SP TRIBELOS SP A CHIRONUMID SP A	-	213	-
CHIRONCHIDAE UNIDENTIFIABLE	• • 11	• • -	-
ORDER OIPTERA - UTHER	•	•	•
CERATOPOGONIDAE UNIDENTIFIABLE	-	-	• •

TABLE 8 (CONTINUED)

PHYLUM MOLLUSCA	บา	ΧΊ	Υl
CLASS GASTROPODA	•		
PHYSA SP	-	-	-
CLASS PELECYPUDA	•	•	
SPHAERIUM SP	•	• - ·	-
PHYLUM NEMATODA	• •	•	
NEMAT ODA	• • •	-	-
PHYLUM TURBELLARIA	* *	•	,
TURBELLARIA SP B	• • •	_	-
TOTAL NUMBER OF ORGANISMS NUMBER OF TAXA	2049	2898 14	1039 11
	•	•	

TABLE 9

TAXONOMIC LIST OF VAAP MACROINVERTEBRATES: NATURAL SUBSTRATE,
AUGUST SURVEY, 1975.
POPULATION SIZE EXPRESSED PER M2 BASED ON POOLED REPLICATES

POPULATION SIZE EXPRESSEL	) ILK I	I- DAS	LD ON	PUULEL	KEPL	TUATE;	)	
PHYLUM ANNELIDA					1			
	A	B1	C1	DI	D2	El	Fl	S
OL I GOCHAETA	-	394	94	531	110	1410	1117	2271
PHYLUM ARTHROPODA - CLASS INSEC	TA							
ORDER EPHEMEROPTERA								r,
CAENIS SP HEXAGENIA SP	-	-			42		16	33.
ORDER ODDNATA								
UNIDENTIFIABLE ODONATA	8	-	-	-	-	-	-	-
ORDER PLECOPTERA								•
UNIDENTIFIABLE PLECOPTERA	-	-	-	-	-	-	-	-
ORDER TRICOPTERA	·							
AGRAYLEA SP Cyhnellus Marginalis	:	=		-	=	•	:	8
ONDER DIPTERA - FAMILY CHIRONOMIDA	E			4				
ABLABESMYIA ANNULATA ABLABESMYIA PARAJANTA Chironomus attenuatus	=	- -	-	- -	8	=	- 6	8 -
CLADUTANYTARSUS SP CUELUTANYPUS CONCINNUS COELUTANYPUS SCAPULARIS	=	17	- 17	- 8	16	=	25	16
CUELOTANYPUS TRICOLOR CRYPTUCHIRONOMUS BLARINA CRYPTUCHIRUNOMUS FULVUS	:	17	-	34	=		17	25 25
HARNISCHIA SP PAGASTIELLA SP (TENT.) PARALAUTEHUURNIELLA NIGRUHALTERALE	:	. 25	17	=	=	=	-	Ξ
PARATEND PES SP POLYPEDILUM HALTERALE PRUCLADIUS SP	:	, E	51	24	=	- 6	16	16
PSEUDOCHIRUNOMUS SP RHEUTANYTARSUS SP SERGENTIA SP (TENT.)	:		43	50	=	=	=	`- -
TAMYPUS CARINATUS TAMYPUS NEUPUNCTIPENNIS TAMYTARSUS SP		8 -	=	 16	=	=		-
CHIRONOMID SP A CHIRONOMID SP B CHIRONOMIO SP B CHIRONOMIOAE UNIDENTIFIABLE	:	-	=	=	=	:	=	=
ORDER DIPTERA - OTHER								
CHAOBORUS SP CERATOPOGONIDAE UNIDENTIFIABLE	115	:	24	59	205	=	161	85

TABLE 9 (CONTINUED)

PHYLUM MOLLUSCA	Α	В1	C1	DI	D2	ΕΊ	F1	S
CLASS GASTROPODA								
PHYSA SP  CLASS PELECYPUDA	•		•	-	•	•	•	•
SPHAERIUM SP	-	-	50	8	•	93	5	74
PHYLUM NEMATODA .								
NEMATODA	-	33	-	16	8	-	<b>-</b>	. 8
PHYLUM TURBELLARIA								
TURBELLARIA SP B	-	-	-	-	-	-	-	-
TOTAL NUMBER OF DRGANISMS Number of Taxa	1 35 3		312	746 9	389 6	1527 5	1376	25 <b>0</b> 5

TABLE 9 (CONTINUED)

PHYLUM ANNELIDA	Tl	U2	Хl	Υl
ULI GOCHAE TA	16	246	1100	1986
PHYLUM ARTHROPODA - CLASS INSECTA				
ORDER EPHEMEROPTERA				
CAENIS SP HEXAGENIA SP	16	42	42 41	532
ORDER ODGNATA			,	
UNIDENTIFIABLE ODONATA	•	-	-	-
ORDER PLECOPTERA				
UNICENTIFIABLE PLECOPTERA	-	-	6	-
CRDER TRICOPTERA				
AGRAYLEA SP Cyhnellus Marginalis	, =	10	=	:
ORDER DIPTERA - FAMILY CHIRONOMIDAE				
ABLABESMYIA ANNULATA ABLABESMYIA PARAJANTA CHIRUNUMUS ATTENUATUS	34	21 10 10	=	=
CLADUTANYTARSUS'SP COELOTANYPU3 CONCINNUS COELOTANYPUS SCAPULARIS	- 8 16	10	-	110
CUELOTANYPUS TRICULOR CRYPTOCHIRUNUNUS BLARINA CRYPTOCHIRONOMUS FULVUS	25 . 8	20	34	25 50
MARNISCHIA SP Pagastiella SP (Tent.) Paralauterdurniella nigr <del>oma</del> lter <b>ale</b>	Ξ	:	=	:
PARATENDIPES SP FOLYPEDILUN HALTERALE PRUCLADIUS SP	34	10	:	8 6 -
PSEUDOCHIRONOMUS SP RHEOTANYTARSUS SP SERGENTIA SP (TENT.)	:	=	51 17	273
TANYPUS CAHINATUS TANYPUS NEUPUNCTIPENNIS TANYTAHSUS SP	-	32	- - 8	-
CHIRONOMIO SP A CHIRONOMIO SP E CHIRONOMIDAE UNIDENTIFIABLE	:	10		-
ORDER OIPTERA OTHER				
CHAQUORUS SP CERATOPOGONIDAE UNIDENTIPIABLE	42	:	:	

TABLE 9 (CONTINUED)

PHYLUM MOLLUSCA	TI	U2	Хl	Υl
CLASS GASTROPODA		_		
PHYSA SP  CLASS PELECYPODA	•	•	•	17
SPHAER IUN SP	•	10	43	25
PHYLUN NEMATODA				
NEMATODA	-	•	33	<b>-</b>
PHYLUM TURBELLARIA				
TURDELLARIA SP 8	25	21	<b>-</b>	-
TOTAL NUMBER OF URGANISMS Number of Taxa	10	452 13	1617	3066 14

Most of them are scavengers -- filtering algae, bacteria, and suspended detritus from the water.

Oligochaetes are commonly found in standing or slow-moving waters over sediments enriched in organic matter. The presence of aquatic plants increases numbers of taxa and organisms. Plants provide shelter from current and predators and produce detritus as food. Most species are able to tolerate or even thrive in low concentrations of dissolved oxygen. Many can survive anaerobic conditions for extended periods of time.

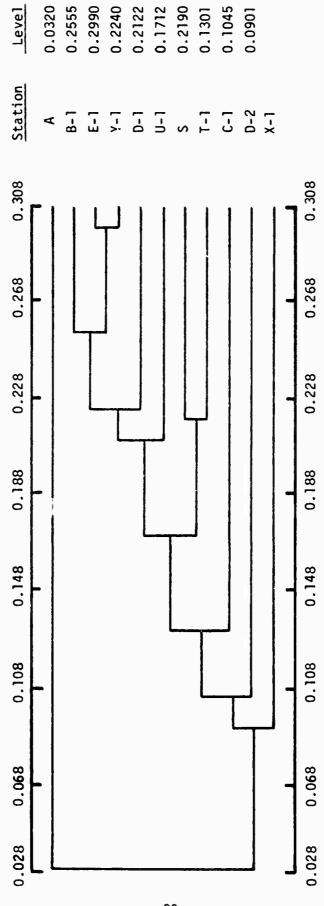
Artificial Substrates. To document macroinvertebrate response to TNT residues from VAAP, a series of Hester-Dendy (H-D) samplers were placed in Waconda Bay from Stations A through F.

The density and number of taxa were lowest at Station A and ranged from 100 - 400 and 6 - 13, respectively. Chironomids were the most common organisms with <u>Procladius</u> dominant in June. In August, <u>Procladius</u> was replaced by <u>Dicrotendipes leucoscelis</u>. Cyrenelus, a tricopteran, also increased markedly in numbers. <u>Procladius</u> and <u>Cyrnellus</u> are both pollution-tolerant organisms (Weber, 1973). Species diversity, commonly used as an indication of community structure, was employed during this study. Results indicate (Table 10) that the relatively elevated (1.67 and 2.00) values at Station A are misleading due to low densities and numbers of taxa. This phenomenon has been discussed previously in the section on periphyton.

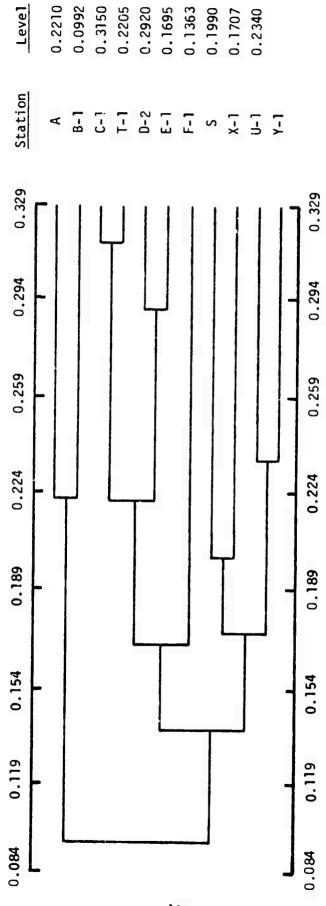
The Pearson-Pinkham Index of Biotic Similarity was also utilized in assessing biological conditions. In this analysis, Station A was unlike all others. In addition, no defined clustering patterns were evident among impact stations or reference areas in the initial survey. In August, Stations A and B-I were similar and together displayed characteristics different from all other stations (Figures 32 and 33).

Effects of VAAP wastes are evident at Station B and are manifested mainly via biostimulation of photosynthetic autotrophs. Growth of periphyton was greatly increased at Station B and in June supported approximately 60 times as many organisms as at Station A and 3 times as many taxa. A similar trend was reflected by macroinvertebrates where the population increased two orders of magnitude and the number of taxa increased from 6 to 13 at Station A. Principal chironomids were Glyptotendipes, Dicrotendipes spp., and Ablabesmyia mallochi. Oligochaetes were also present in relatively low numbers. Species diversity ranged from 0.83 to 1 (Table 10).

Down bay at transect C, the maximum number of organisms colonized the H-D plates and amounted to  $36 \times 10^3$  per m². Glyptotendipes accounted for 88 percent of the population. Water quality data do not reveal an association with VAAP wastes or cultural influences which would account for these high numbers. Physical characteristics may be influencing invertebrates in this area, but the data base is insufficient for further conclusions. The invertebrate community at this station may be showing a response to munitions wastes but the level of response is sufficiently subtle that separation of effects from VAAP wastes and other cultural activities occurring in this area (upland erosion and the presence of domestic ducks) cannot be made without further investigation. At transects D and F the macroinvertebrate community



PHENOGRAM OF WACONDA BAY AND REFERENCE BAY MACROINVERTEBRATE COMMUNITY RELATIONSHIPS, JUNE, 1975. BASED ON PEARSON-PINKHAM SIMILARITY INDEX, MUTUAL ABSENCE UNIMPORTANT, ARTIFICIAL SUBSTRATE, COPHENETIC CORRELATION COEFFICIENT, 0.882. 32. FIGURE



PHENOGRAM OF WACONDA BAY AND REFERENCE BAY MACROINYERTEBRATE COMMUNITY RELATIONSHIPS, AUGUST, 1975. BASED ON PEARSON-PINKHAM SIMILARITY INDEX, MUTUAL ABSENCE UNIMPORTANT, ARTIFICIAL SUBSTRATE, COPHENETIC CORRELATION COEFFICIENT, 0.735. FIGURE 33.

TABLE 10

SHANNON-WEAVER SPECIES DIVERSITY INDICES VAAP MACROINVERTEBRATES, 1975\*

	Artificial	Substrates	Natural Substrates		
Station	June	August	June	August	
Α	1.671	2.080	1.173	0.446	
B-1 B-2	0.843	0.997	2.390 0.322	0.947	
C-1	0 <b>.62</b> 8	1.760	2.059	1.926	
D-1 D-2	1.494 1.135	1.171	2.099 0.928	1.137 1.226	
E-1 E-2	2.053	1.771	1.395 1.543	0.327	
F-1 F-2		2.359	1.763 1.890	0.741	
S	1.334	2.056	1.031	0.626	
T-1 T-2	1.933	1.838	1.376 1.348	2.179	
U-1 U-2	1.456	1.722	1.749 2.289	1.753	
X-1 X-2	1.905 	1.729 	2.242 2.577	1.064	
Y-1	1.292	1.881	2.160	1.187	

<sup>\*</sup>Based on pooled data.

exhibits characteristics of the open lake and therefore these Waconda Bay stations represent areas where VAAP wastes have no discernible impact.

It should be emphasized that the above described effects on the macroinvertebrate community in Waconda Bay are directly correlated with loading from VAAP and therefore can be expected to be manifested further down bay as production levels increase by a reduction in numbers of species and individuals. At the present time impact is evident only at Stations A and B.

Natural Substrates. Invertebrates at Station A respond in a similar manner to those colonizing artificial substrates. That is to say population and taxa were reduced to low levels suggesting inhibition to TNT wastes. The number of organisms and species observed during both surveys may reflect residual toxicity in bay sediments which continued following plant shutdown in May (Tables 8 and 9).

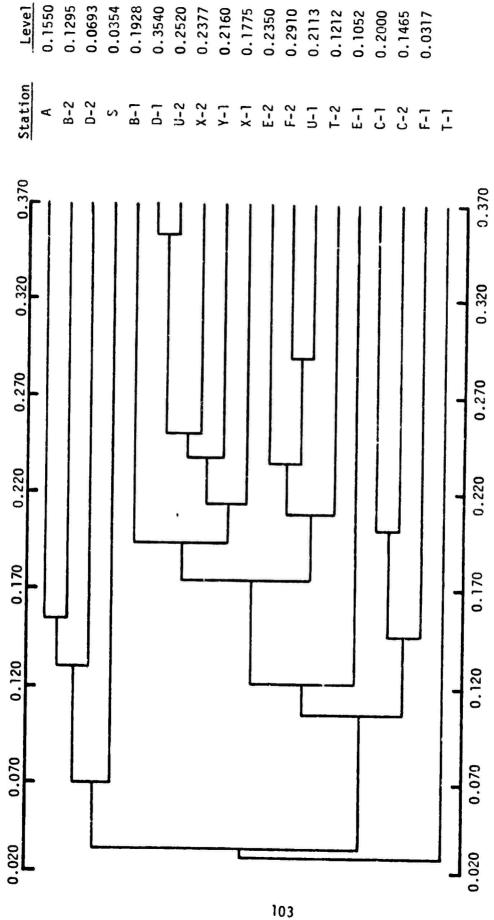
The data show that <u>Chaoborus</u> was the most abundant organism at Station A. This contrasts with oligochaetes and chironomids which dominated other station locations, but were observed at relatively low numbers at Station A. It should be noted that the chironomid, <u>Tanypus neopunctipennis</u>, was observed only at Station A and may reflect tolerance to VAAP wastes.

Cluster analysis as shown in Figures 34 and 35 for the two surveys characterizes the head of Waconda Bay as dissimilar to upper bay or reference areas. The phenograms suggest, considering the levels of VAAP wastes in sediments at Station A and transect B, that munitions compounds may influence the population level and taxa number.

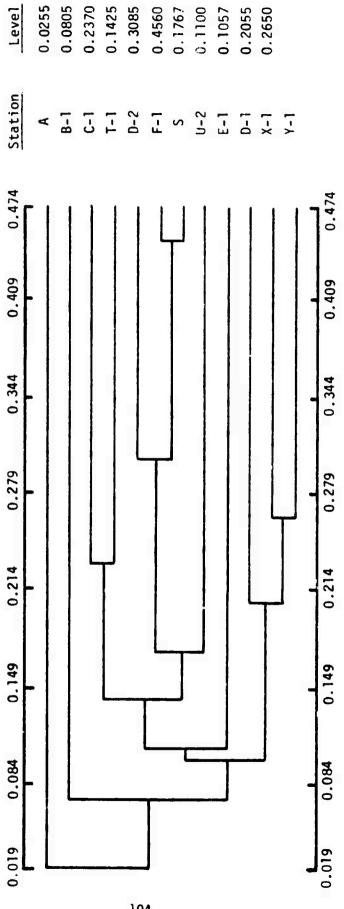
It is significant to note that densities in sediments varied little between June and August at Stations A, B-1, and C-1. Evidently sediment munitions residues affected benthic macroinvertebrates even after VAAP ceased its discharge. It is also worthy of note that macroinvertebrate densities were depressed further downbay in the natural substrates than in the artificial substrates. A comparison of munitions levels in water and sediments show higher concentrations in the latter which could influence this trend.

Stations D-1, E-1, and F-1 average 2 - 3 times the densities found at B-1 and C-1. Also, numbers of taxa fluctuate but their averages are approximately equivalent from Stations B-1 to F-1. The present techniques were therefore unable to detect effects from munitions on benthic macroinvertebrates downbay from Station C-1.

Examination of the data suggests that recovery in the natural substrates lags behind that on the Hester-Dendy samples. This can be seen by comparing populations and taxa downbay from Stations A through D. There is a general trend (mean value) of increasing numbers of macroinvertebrates and selected taxa at these stations with a peak occurring at Station D (Figures 36 and 37). The results of the study at VAAP indicate that munitions wastes impact the invertebrate community at the head of Waconda Bay. This is represented by an evident toxic response in both natural and artificial substrates. The increase in macroinvertebrates on H-D plates accompanying an increase in

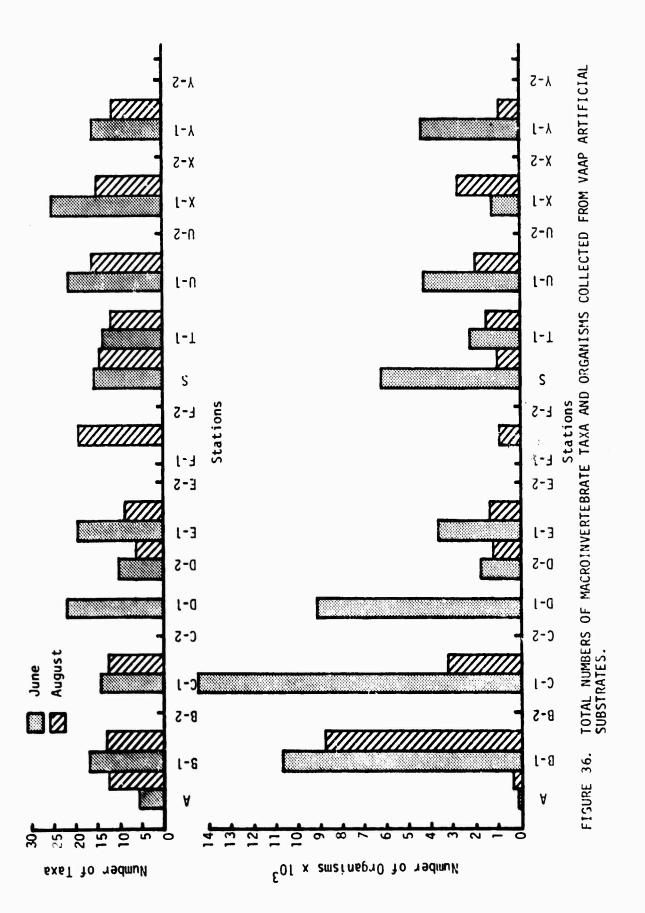


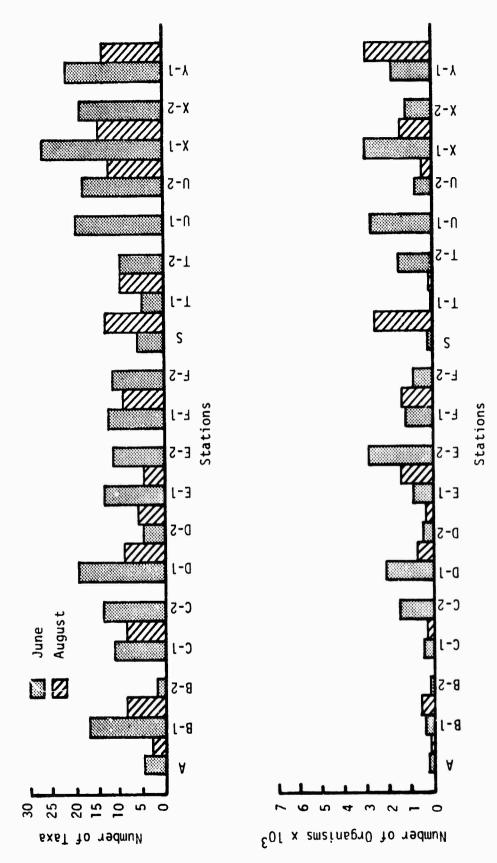
PHENOGRAM OF WACONDA BAY AND REFERENCE BAY MACROINVERTEBRATE COMMUNITY RELATIONSHIPS, JUNE, 1975. BASED ON PEARSON-PINKHAM SIMILARITY INDEX, MUTUAL ABSENCE UNIMPORTANT, NATURAL SUBSTRATE. COPHENETIC CORRELATION COEFFICIENT, 0.877. 34. FI GURE



PHENOGRAM OF WACONDA BAY AND REFERENCE BAY MACROINVERTEBRATE COMMUNITY RELATIONSHIPS, AUGUST, 1975. BASED ON PEARSON-PINKHAM SIMILARITY INDEX, MUTUAL ABSENCE UNIMPORTANT, NATURAL SUBSTRATE. COPHENETIC CORRELATION COEFFICIENT, 0.853. FIGURE 35.

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TOTAL NUMBERS OF MACROINVERTEBRATE TAXA AND ORGANISMS COLLECTED FROM VAAP NATURAL SUBSTRATES. FIGURE 37.

primary production is not reflected in the sediments. The variation observed in these two systems probably is caused by differences in habitat and response time to environmental stress, therefore, H-D devices would appear the most appropriate technique for assessing impact of munitions residues over relatively discrete time periods. Careful examination of natural substrates, however, will be the most useful in historically characterizing the system to determine the limits of munitions impact downbay.

#### CONCLUSIONS

The summer 1975 sampling at VAAP offered a unique opportunity to observe the response of the Waconda Bay ecosystem to the cessation of TNT production waste input. The chemical data showed that the VAAP discharge affected water quality in the upper portion of Waconda Bay by increasing the concentration of TNT, 2,4-DNT, 2,6-DNT, total dissolved solids, hardness, chlorides, ammonia, organic nitrogen, nitrates, and nitrites. Chemically, this effect was noticeable downbay to at least transect C some 0.6 miles from the bayhead. The chemical effects in the upper bay persisted through the August sampling, some 9 weeks after the plant was closed due to a labor dispute. However, the concentrations of munitions and associated materials in August were reduced as compared to June. Samples taken in two reference bays verified that the elevated concentrations observed in Waconda Bay were due to the VAAP discharge.

Biologic response by the periphyton and macrobenthic communities was observed in the same area of Waconda Bay where chemical characteristics were altered. At Station A, the bayhead, toxic effects were noted in both communities. However, at transect B, about 0.3 miles downbay, biostimulation for both periphyton and those macrobenthos colonizing artificial substrates was observed. At transect C, biologic effects were less clear-cut, although there was some suggestion of biostimulation. This pattern of toxicity followed by biostimulation suggests that one or more components in the waste stream are at toxic concentrations in bayhead water. As this material moves downbay, it is either diluted or degrades. When this occurs, nutrients become the controlling factor resulting in biostimulation.

Munitions wastes from VAAP have deposited in bay sediments during the years of plant operation. Analysis revealed a concentration gradient evident from the bayhead to transect C ranging as high as 3.1 mg/kg dry weight. Macroinvertebrate data indicate that the immediate area where VAAP effluent enters Waconda Bay represents a zone of severe inhibition in population size and number of taxa. TNT levels are reduced by an order of magnitude at transect B which is 0.3 miles from Station A.

June data project a picture of inhibition downbay as far as transect C. This trend although not as evident remained during the second survey. When community structure relationships between artificial and natural substrates were compared, there was strong evidence of residual toxicity in bay sediments especially at Station A.

Diatoms seem to be extremely sensitive to munitions wastes. Between the June and August samplings, significantly more recovery occurred in the macroinvertebrate community colonizing artificial substrates than in the periphyton community. It is apparent that diatoms are the most sensitive of the two components to munitions waste and would be the most effective early warning indicator. On the other hand, the above facet of the benthic community would be the most sensitive indicator of recovery. Insufficient time passed between the two studies to observe recovery among the natural sediment community which had been significantly stressed by waste discharges.

Phytoplankton identification and enumeration showed that Waconda Bay and the two reference bays contained sufficient algal populations to change dissolved oxygen values by several units diurnally. However, at this time phytoplankton appears to be of limited use to discriminate waste impact during plant operation. There is some suggestion based on the results of this study that the limnoplankton may be useful to delineate areas of biostimulation associated with TNT decomposition.

Because the observed biologic responses were to a mixed waste milieu, it is not possible to determine precise cause and effect relationships. However, at the bayhead where toxicity was noted in both the periphyton and macrobenthic communities, the median munitions concentrations in June and August were 123 and 56 ppb with individual samples as high as 345 ppb. Little reduction was noted in concentrations of the specific munitions from the bayhead to transect B. Since the biologic response of those organisms in the water column shifted from toxic to biostimulatory, it seems unlikely that the toxicity was due specifically to these compounds. The biotoxic response at A is suggestive of a separate factor, perhaps one of the breakdown products of the munitions specific materials. Concentration gradients of these compounds, not quantitated in this survey may be responsible for the differential response at Stations A and B. Nevertheless, it was observed that when munitions concentration dropped below 20 ppb, no further biologic responses were evident. At munitions concentration between 40 and 80 ppb, as at transect C, slight biostimulatory effects were noted.

Based on these results, it is concluded that the environmental impact of TNT waste effluent at VAAP would be minimal if the combined concentration of munitions residues did not exceed 20 ppb in the receiving waters or 100 ppb in sediments.

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APPENDIX A
WATER QUALITY DATA

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APPENDIX A-1
ANALYTICAL METHODOLOGY

Table A-1 lists the specific <u>Standard Methods</u> or EPA procedures used to characterize the background water <u>quality</u>. Metals, except for hardness (calcium and magnesium), were run by flame atomic absorption spectrophotometry on acidified water samples. The expected low concentrations indicated that extraction using ammonia 1-pyrrolidinedithiocarbamate (APDC) and methyl isobutyl ketone (MIBK) (after Nix and Goodwin, 1970) would be necessary to gain the required sensitivity. Mercury was analyzed by cold vapor atomic absorption spectrophotometry.

The nutrient parameters (nitrogen, phosphorus, and carbon) measured were those usually used to characterize trophic state. These results were used to assess what, if any, factors might limit plant growth, whether biostimulatory effects might be expected; and, secondly, to determine what oxygen demands might result. Selection of these parameters was also keyed to those specific to munitions manufacture waste impacts.

In the case of the metal analyses run on the sediments, the digestate contained sufficiently high concentrations of metals to be run directly by flame atomic absorption spectrophotometry making MIBK-APDC extraction unnecessary. The sediment mercuries were analyzed employing cold vapor atomic absorption spectrophotometry after the digestion of a portion of sediment with aqua regia as described in EPA (1974).

Analytical procedures (Table A-2) utilized on the sediment samples came mostly from the Chemistry Laboratory Manual Bottom Sediments (EPA, 1969), with the exceptions of mercury (as described above) and total phosphorus. The total phosphorus procedure employed sulfuric acid-potassium persulfate digestion in an autoclave as specified in "Sludge-Sediment Analysis" (EPA, Region IV, 1973). The numbers of samples collected and analyzed are tabulated in Appendix F by station and analysis.

#### <u>Munitions Analysis</u>

The munitions samples were collected in amber glass reagent bottles that were pre-rinsed in acetone. The samples were refrigerated until analysis, which consisted of extraction concentration and gas-liquid chromatography.

Extraction of Water Samples. A sample of 250 ml was measured into a clean 500 ml separatory funnel equipped with a Teflon stopcock. Seventy-five ml of ethyl acetate (pesticide grade) was added, the flask stoppered, and shaken for 2 to 3 minutes. The layers were allowed to separate and the lower (water) layer drained into a second 500 ml separatory funnel and again extracted with 50 ml ethyl acetate. The water layer was discarded. The extracts were combined and filtered through a plug of cotton previously wetted with ethyl acetate. The separatory funnels were rinsed with an additional 10 ml of ethyl acetate and filtered through the cotton plug. The ethyl acetate was evaporated to a volume of 2.5 ml under reduced pressure with the flask temperature not exceeding 40°C.

Extraction of Sediment Samples. In order to dry the wet sediments, 80 gm of sodium sulfate was added to 20 gm of wet sediment. This was then packed into a chromatographic column and extracted for one hour with ethyl acetate. The extraction was followed by evaporation of the ethyl acetate extract to

a volume of 2.0 ml under the same conditions as described earlier. The one hour of extraction with ethyl acetate was proven to be sufficient by the fact that quantitative recovery of sediment samples spiked with 2,4-DNT, 2,6-DNT, and  $\alpha$ -TNT was obtained.

Three levels of spiking were used as follows:

#### SEDIMENT SAMPLE WPA 40

Component	Quantity Added mg/kg (Wet Wt.)	Quantity Recovered mg/kg	Percent Recovery
2,4-DNT	0.5 1.0	0.48 1.03	96 103
	10.0	9.0	90
2,6-DNT	0.5	0.46	92
	1.0	0.97	97 94
	10.0	9.4	94
2,4,6-DNT	0.5	0.52	104
-, . ,	1.0	0.93	93
	10.0	9.3	93

Chromatography of Extracts. Samples were chromatographed on a 5 ft. x 1/8 in. glass column packed with 3 percent Dexsil 300 on 80/100 mesh Gas Chrom Q. A Varian Model 1840 Gas Chromatograph with electron capture (EC) and Thermionic (Alkali Flame Ionization Detector) (AFID) detectors was chosen. The readout was obtained by using a Varian Model 285 Electronic Integrator which was recorded permanently by a Beckman 1 mv, 10 inch scale recorder. Peak areas were automatically printed by integrator. Electron capture was chosen as the prime detector with AFID as back-up and confirmation detector.

An alternate column used for confirmatory information was a 4 ft. x 1/8 in. glass column packed with 8 percent UCW 98 on 80/100 mesh Gas Chrom. Q. Instrument conditions for both columns and detectors were:

Column temperature: 185°C, isothermally

Injector temperature: 220°C Detector temperature: 215°C

Carrier Gas: Nitrogen @ 40 ml/min.

Electrometer setting:  $10^{-10}$  afs at 1 x attentuation into

integrator with appropriate attenua-

tion setting for recorder.

Five microliter portions of standards and samples were injected. The peak heights, peak areas, and retention times were recorded for comparison.

<u>Preparation of Standards</u>. Purified standards of 2,4-Dinitrotoluene; 2,6-Dinitrotoluene; 1,3,5-Trinitrobenzene; and 2,4,6-Trinitrotoluene (TNT) were supplied by the Army Medical Research and Development Command.

Discussion of Procedure. Under test conditions, 2,4-DNT, 2,6-DNT, and 2,4,6-TNT were adequately resolved by both Dexsil 300 and UCW 98 columns. However, 1,3,5-TNB and 2,4,6-TNT were not differentiated by the Dexsil column and only partially by the UCW 98 column; consequently 1,3,5-TNB, if present, was combined with and reported as 2,4,6-TNT.

Five µl injections of sample extracts and standards were first injected onto the Dexsil 300 column using the EC Detector. Peaks corresponding to standards were noted and the areas compared. Samples and standards were next injected onto UCW-98 column and like comparisons were made. Likewise, samples and standards were injected onto the Dexsil 300 column using the Thermionic or Alkali Flame Ionization Detector. Again, peaks corresponding to the standards were noted and the areas were compared. Sample peaks which did not elute at the same times as the standards on both sets of columns and detectors were rejected, and only those that were peaks confirmed on both sets were quantitated.

The AFID was used primarily for confirmation of the presence or absence of various compounds in the samples. However, results were calculated and compared with results from the EC detector. In most cases, quantitative results were comparable with both detectors. Where agreement was not within limits of  $^{\ddagger}$  10 percent, additional injections were made until agreement could be obtained within these limits, or it was determined that substrate interference effected response from one or the other of the detectors. This was normally determined by spiking the sample with the appropriate standard and noting the recovery.

Recovery Studies. Initial recovery studies were made by the addition of standards to tap water and then carrying through the entire extraction, concentration, and gas chromatographic procedures, as previously outlined.

Three levels of spiking were used, as follows:

Component	Quantity Added	Quantity Recovered	Percent Recovery
2,4-DNT	1.00 mg/l	0.95 mg/l	95
	5.00 mg/l	4.80 mg/l	96
	10.00 mg/l	10.05 mg/l	101
2,6-DNT	1.00 mg/l	0.93 mg/l	93
	5.00 mg/l	4.95 mg/l	99
	10.00 mg/l	10.20 mg/l	102
2,4,6-TNT	1.00 mg/l	0.96 mg/l	96
	5.00 mg/l	5.15 mg/l	163
	10.00 mg/l	9.80 mg/l	98

Selected samples containing low levels or no munitions residues were spiked with corresponding low levels of TNT to assess recovery under these conditions. Results of these experiments were:

#### TABLE A-1

## SUMMARY OF ROUTINE LABORATORY ANALYTICAL PROCEDURES FOR WATER SAMPLES

Parameter	Procedure
Total Alkalinity	Standard Methods, 201: Potentiometric Titration, p. 370.
Chloride	Standard Methods, 112B: Mercuric Nitrate Method, p. 97
Total Hardness	Standard Methods, 112B: EDTA Titrimetric Method, p. 179.
Sulfate	Standard Methods, 156C: Turbidimetric Method, BaCL <sub>2</sub> , p. 334.
Solids - Total Solids	Standard Methods, 148A: Gravimetric Method
Suspended Solids	Method, p. 288.  Standard Methods, 148C: Gravimetric Method
Total Dissolved Solids	Method, p. 291. <u>Standard Methods</u> , 148B: Gravimetric Method  Method, p. 290
Ammonia Nitrogen	EPA, STORET #00610: Distillation and Nesslerization, p. 159.
Total Kjeldahl Nitrogen	EPA, STORET #00625: Acid Digestion, Distillation, Nesslerization, p. 175.
Nitrite Nitrogen	EPA, STORET #00630: Automated Analyses, Diazotization, Sulfanilic Acid-Napthylamine Hydrochlorida Method, p. 207.
Nitrate Nitrogen	EPA, STORET #00630: Automated Analyses, Cadmium Reduction Method, p. 207.
Total Phosphorus	Standard Methods, 223C.III: Persulfate Digestion Method, p. 526. EPA, STORET #00671: Automated Colorimetric Ascorbic Acid Single Reagent Method, p. 256.
Total Organic Carbon	EPA, STORET #00680, Infrared CO <sub>2</sub> Detection, Carbon Analyzer, p. 236.
Chemical Oxygen Demand	EPA, STORET #00335, Low Level 0.025 <u>N</u> K <sub>2</sub> Cr <sub>2</sub> O <sub>7</sub> , p. 21.

EPA 1974, "Manual of Methods for Chemical Analysis of Water and Wastes."

Standard Methods for the Examination of Water and Wastewater, 13th Ed., 1971, APHA, AWWA, WPCF.

#### TABLE A- 2

### SUMMARY OF ROUTINE LABORATORY ANALYTICAL PROCEDURES FOR SEDIMENT SAMPLES

Parameter	Procedure
Chemical Oxygen Demand	Bottom Sediments - Great Lakes: High Level 0.250N K <sub>2</sub> Cr <sub>2</sub> O <sub>7</sub> , p. 5.
Total Kjeldahl Nitrogen	Bottom Sediments - Great Lakes: Acid Digestion, Distillation, and Titration with 0.02N H <sub>2</sub> SO <sub>4</sub> , p. 38.
Nitrate Nitrogen	Bottom Sediments - Great Lakes: Acid Digestion, p. 32. Standard Methods, 213B: Cadmium Reduction Method, p. 458.
Nitrite Nitrogen	Bottom Sediments - Great Lakes: Acid Digestion, p. 32. EPA, STORET #00630: Automated Analyses, Diazotization, Sulfanilic Acid- Napthylamine Hydrochloride Method, p.207.
Total Phosphorus	EPA, Region IV, "Sludge-Sediment Analyses," 1973: Sulfuric Acid-Persulfate Digestion using an Autoclave. EPA, STORET #00671: Automated Colori- metric Ascorbic Acid Single Reagent Method, p. 256.
Total Solids	Bottom Sediments - Great Lakes: Gravimetric Method, p. 85.
Total Volatile Solids	Bottom Sediments - Great Lakes: Gravimetric Method, p. 85.
Mercury	EPA, 1974: Aqua Regia Digestion, Potassium Permanganate Oxidation, and Cold Vapor Technique Atomic Absorption Spectrophotometer, p. 134.
Trace Metals (Cd, Cu, Cr, Fe, Pb, Mn, Ni, Zn)	Bottom Sediments - Great Lakes: Nitric Acid - Hydrogen Peroxide Digestion, p. 18. Atomic Absorption Spectrophotometry.

<u>Chemistry Laboratory Manual Bottom Sediments</u>, EPA 1969, compiled by Great Lakes Region Committee on Analytical Methods.

EPA 1974, "Manual of Methods for Chemical Analysis of Water and Wastes".

EPA, Region IV, Surveillance and Analysis Division, Chemical Services Branch, "Sludge-Sediment Analyses," June 7, 1973, mimeograph courtesy of James Finger, EPA, Region IV.

Standard Methods for the Examination of Water and Wastewater, 13th Ed., 1971, APHA, AWWA, WPCF.

TABLE A-2 (CONTINUED)

Sample No.	Resid Component &	dual Concentration	Compor Added & Cor	nent ncentration	Percent Recovery
B-41	2,4-DNT	19 μg/l	2,4,6-TNT	4 μg/1 8 μg/1 10 μg/1	94 92 98
B-59		<del></del>	2,4,6-TNT	l μg/l 2 μg/l 3 μg/l	89 93 96
B-39			2,4,6-TNT	2 μg/l 4 μg/l 6 μg/l	90 93 96

Recovery of TNT and its analogs by this procedure appears to be excellent, averaging more than 95 percent from tap water and 93 percent from spiked water samples.

#### APPENDIX A-2

#### FIELD MEASUREMENTS:

Dissolved Oxygen, Temperature, pH, and Specific Conductance Harrison Bay and Lake Chickamauga

> June 9-13, 1975 August 11-15, 1975

Waconda Bay Reference Bay A Huss Lowe Slough

TABLE A-3 DISSOLVED OXYGEN (ppm)

	Range	6.0 - 8.5	6.4 - 9.1	3.8 - 9.0	5.0 - 9.1	4.0 - 9.4	5.9 - 9.6	4.8 - 9.6	4.2 - 9.1	5.5 - 9.5	3.7 - 9.4	3.9 - 9.6
	Average	7.08	7.63	6.46	69.9	6.53	7.0	7.0	6.82	7.13	7.02	6.75
1/75	ν. Σ.	11	! !	1 1	! !	11	!!		! !	::		1 1
6/13/75	A. Ж.	6.0	6.4	3.8	6.2	6.4	5.9	6.7	7.3	7.3	3.7	7.3
2/75	P. M.	1 1	: :	; ;	1 1	1-1		1 1	: :	1 1		: :
6/12/75	A.M.	6.3	6.8	4.5	6.0	4.4	6.3	4.9	6.6	5.5	6.9	6.8
/75	φ. Ψ.	1 1	1 1	11	1 1	: :	1 1	; ;	: :		: :	::
6/11/75	Α.Μ.	7.1	7.2	6.8	6.7	6.5	6.7	6.1	7.0	6.6	6.8	6.7
175	P. M.	1 1	! !	1 1	8.9	8.6	9.4	8.2	8.1	8.4	8.7	7.8
6/10/75	A. <del>X</del> .	8.1	8.7	8.4	1 1	1 1	1 1	1 1	: :	1 1	1 1	: :
6/9/75	P. W.	8.5	9.1	9.0	9.1	9.4	9.6	9.6	9.1	9.5	9.4	9.6
6/9	A.M.	1 1	1 1	! !	1 1	1 1	! !	1 1	1 1	1 1	1 1	1 1
	Station	A N B	8-1 S	8-2 S	C-1 8	C-2 S	D-1 S	0-2 S	E-1 S	E-2 S	F-1 S	F-2 S

TABLE A-3(Continued)

	6/9	5//6/9	6/10/75	5//2	6/11/75	/75	6/12/75	/75	6/13/75	775		
Station	A.M.	P.M.	A.M.	P.M.	A.M.	Р.М.	А.М.	Р.М.	A.M.	Р.М.	Average	Range
S B	: :	9.3	7.2	1 1	1 1	6.2	3.5	1 1	6.3	1 1	5.92	3.5 - 9.3
T-1 S	! !	9.8	7.9	! !	1 1	6.1	5.6	11	6.8	! !	6.49	4.5 - 9.8
1-2 S	1 1	9.6	8.0	: :	1 1	6.2	5.9		5.6	! !	6.74	5.0 - 9.6
U-1 S	: :	9.5	8.3	!!	1 1	6.7	6.3		7.7	! !	6.94	5.0 - 9.5
U-2 S	: :	9.4	8.2	! !	1 1	6.9	6.6	1 1	7.5	! !	. 6.82	4.2 - 9.4
x-1 S	7.4	1 1	: :	9.0	8.3	! !	! !	8.8	8.8	! !	8.30	7.4 - 9.0
x-2 S	7.5	: :	: :	8.5	7.7	::	!!	8.8	8.3	!!	8.19	7.3 - 8.5
Y-1 S	8.2	: ;	: :	9.5	8.5	::	: :	8.6	8.8	1 1	8.61	8.2 - 95
Y-2 S	7.7	!!	: :	9.6	8.6	1 1	1 1	8.6	1.5	1 1	6.42	0.2 - 9.6

TABLE A-4 DISSOLVED OXYGEN (ppm)

	8/1]	8/11/75	8/12/75	775	8/13/75	/75	8/14/75	7.5	8/15/75	/75		
Station	A.M.	P. M.	A.M.	.¥.	A. A.	P. M.	A.M.	P.M.	A.M.	P.M.	Average	Range
A 88	8.4	::	: :	9.6	7.4	: ;	8.6 7.3	: :	7.6	: :	7.10	4.4-9.6
8-1 S	8.8	::	11	10.6	8.0	; ;	9.5	! !	8.8	! ;	9.14	8.0-10.6
B-2 S	8.2 5.5	; ;	1 1	10.2 5.8	3.2	::	9.2	!!	8.2	: :	6.24	1.4-10.2
C-1 S	7.3	; ;	; ;	9.7	9.2	: :	9.1	! !	7.9	; ;	7.36	5.0-9.7
C-2 8	7.5	; ;	: ;	9.5	8.9		9.3	: :	7.8	1 1	7.56	2.1-10.2
0-1 S	1 1	7.1	8.3		9.9	! ;	9.4	; ;	5.0	; ;	7.49	5.0-9.9
D-2 S	!!	7.6	8.5		9.7		9.2	: :	3.7	1 1	6.36	2.5-9.7
E-1 S	1 1	7.3	8.2	1 1	9.4	::	9.5	: :	2.1	1 1	5.82	1.6-9.4
E-2 S	1 1	7.5	8.2	! !	9.4		8.9	1 1	7.9	; ;	7.29	4.5-9.4
F-1 S	1 1	7.5	3.1	! !	9.2	! !	8.9	; ;	8.5	::	5.52	0.9-9.2
F-2 S	1 1	7.7	8.2 5.3	1 1	9.4		2.2	::	8.3	! !	6.36	2.2-9.4

TABLE A-4 (Continued)

	8/11	8/11/75	8/12/75	175	8/13/75	/75	8/14/75	775	8/15/75	/75		ć
Station	A.F.		. F. F.		. ш. ч	. M.	.E.	.E.	Υ. A.	E.	Average	Kange
S S	7.6	1 1	: :	7.9	6.8	1 1	6.7	1 1	6.4	11	6.27	4.4-7.9
T-1 S	6.4	! !	11	8.0 5.8	8.3		8.1 6.5	3 <b>1</b>	7.6	: :	6.87	4.8-8.3
T-2 S	7.0		1 1	8.2	8.5		8.0	1 1	7.5	1 1	7.78	7.0-8.2
u-1 S	5.2	1 I	1 1	8.3	8.8 5.6	11	8.7	1 1	7.9	1 1	7.04	5.2-8.8
U-2 S	5.2	1 1	! !	3.3	8.8	1 1	8.5	! !	8.0	! !	6.26	3.3-8.8
x-1 S	1 1	9.8	7.8		11	9.1 6.5		9.2	9.1	! !	8.40	6.1-9.5
X-2 S	: :	9.1	7.2	!!	: :	9.2	1 1	9.0	8.0	1 1	7.34	4.3-9.2
Y-1 S	: :	9.0	8.8		: :	9.0	1 1	9.3 10.9	9.5	1 1	8.87	7.0-10.9
Y-2 S		9.3	8.7	::	: :	9.2 0.9	: 1	9.4	9.3	1 1	6.58	0.2-9.4

TABLE A-5

# TEMPERATURE (°C)

	5/9	6/9/75	6/10/75	/75	6/11/75	/75	6/12/75	/75	6/13/75	/75		
Station	A.M.	P.M.	A.M.	P.M.	A.M.	P.M.	A.M.	Р.М.	A.M.	P.M.	Average	Range
A B	1 1	26.6 23.5	25.0	::	24.0	: :	24.1	; ;	24.6	1 1	24.0	23.0 - 26.6
B-1 S	1 1	27.0 25.5	24.5		11	1 1	24.0		24.6	; ;	25.0	24.0 - 27.0
8-2 \$	1 1	26.5 25.0	24.5	1 1		! !	24.0	; ;	24.6	; ;	24.2	22.9 - 26.5
C-1 S	1 1	26.5 25.0	11	25.0		1 1	24.0		24.1	! !	24.4	23.0 - 26.5
C-2 S	1 1	27.0	1 1	25.0 23.5	23.5	11	24.2	1 1	24.1	I I	24.3	23.1 - 27.0
0-1 S	: :	27.0 25.1	1 1	25.0	23.5	1 1	24.0	: :	24.5	! !	24.4	23.5 - 27.0
D-2 S	1 1	2 <b>6.</b> 5 24.9	1 1	24.8	23.5	1 1	23.8		24.4	f 1	24.2	23.0 - 26.5
E-1 S	1 1	2 <b>6.0</b>	1 1	24.9	23.5	1 1	23.8	1 1	25.1 24.0		23.9	22.0 - 26.0
E-2 S	1 1	26.0 24.8		24.9	23.2	1 1	23.9	1 1	24.8	l †	24.2	23.0 - 26.0
F-1 S	1 1	25.5 21.9	1 1	24.9	23.5	1 1	24.0	! !	25.2		23.4	21.0 - 25.5
F-2 S	; ;	25.8 21.9		2 <b>4.8</b> 22.2	23.2	: :	24.0	; ;	25.1	- 1	23.6	21.9 - 25.8

TABLE A-5(Continued)

	Average Range		24.1 20.9 - 26.8	20.9 -	20.9 - 23.1 - 23.5 -	20.9 - 23.1 - 23.5 - 23.2 -	24.4 23.1 - 24.8 23.5 - 24.4 23.2 - 24.4 23.1 -	24.4 23.1 - 24.8 23.5 - 24.4 23.2 - 24.4 23.1 - 24.4 23.1 - 25.0 24.3 -	24.4 23.1 - 24.8 23.5 - 24.4 23.2 - 25.0 24.3 - 25.2 24.0 - 25.2 24.0 - 25.2	24.4 23.1 - 24.4 23.5 - 24.4 23.2 - 24.4 23.1 - 25.0 24.3 - 25.2 24.0 - 25.0 25.0 24.0 - 25.0 25.0 25.0 - 25.0 25.
	-							24.4 24.4 24.4 25.0	24.4 24.4 24.4 25.0	24.4 24.4 25.0 25.2 25.2
Σ.		2	-			<del></del>				
Ë		25.3	4,	7.4	4 · 6 7 · 8 · 6 7 · 6	4.6 6.3 6.0 7.0	, &ô &ô -8	7 8 9 FO -8 4 F	7 8 9 EO -8 4E V2	7 80 EO -8 4E V2 V6
	Ë	1 1	1 1	   	<del></del>			25.5 24.5	25.5 24.5 24.8	25.5 24.9 24.9
	A.3	21.8	23.5							
	Σ.	1 1			1 1	11 11	11 11 11	11 11 11 11		11 11 11 11 11
_	A. A	1 1	1 1		; ;			<u>∞</u> .∞.		
_ 2	Σ.	: :	: :		: :			   25.3	25.3 25.3 25.9	25.3 25.3 25.3 25.3 25.9
_	A.M.	24.8	24.5		24.8	3, 8 .2 3, 8 .2	8.4 4.8 8.3 8.8 6.0	4.4 4.6 4.8 1.2 8.6 1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0	44 4	44 4E 44 11 11 11 11 8E
	Ē.	26.8 25.0	26.5 24.8		6.5 5.5		26.5 25.5 26.0 24.9 24.9	26.5 25.5 26.0 24.9 24.9	26.5 25.5 26.0 26.0 24.9	າດເບິ່ ດ່4 ຄ 4 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1
_	F. 4		; ;		1 1			0.0	1 1 1 1 4 6 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0	1 1 1 1 4 4 6 7 4 6 7 9 9 9 9 9 9 9 9 9 9 9 9 9 9 9 9 9 9
		νæ	vω	_	~ B					
, , , ,	Station	S	7		1-2					

TABLE A-6

TEMPERATURE (°C)

30.3 30.5 29.6 29.0 24.0 - 28.9 0 24.5 - 32.0 26.0 - 31.0 29.1 28.1 28.8 Range 25.4 -26.0 25.9 26.0 25.9 25.2 26.0 26.1 Average 27.0 28.3 26.6 27.6 27.0 27.6 27.0 27.7 27.7 27.1 27.1 8/15/75 A.M. ! P.M. 1 1 | | 1 1 | | | | 1 1 | | 27.3 27.1 27.2 25.9 25.2 27.4 27.5 28.0 25.9 27.5 25.8 27.7 8/14/75 A.M. | P.M. 1 1 | | | | † I | | 1 1 1 1 1 1 1 1 1 1 29.0 28.8 28.8 27.0 28.8 27.0 28.0 26.6 28.0 26.1 29.9 27.9 28.9 28.9 œ.Ο. 27. 26. 8/13/75 A.M. | P.M. | | | | | | | | 1 1 | | | | 1 1 | | 1 1 1 1 28.0 27.9 25.0 28.2 28.0 29.6 29.1 28.9 28.6 28.0 27.5 25.2 28.1 26.2 32.0 25.1 31.0 31.0 30.3 30.5 8/12/75 A.M. | P.M. 1 1 1 3 | | 28.3 27.9 27.9 28.2 26.5 27.8 1 1 | | ! ! 27.7 26.0 28.0 26.0 28.5 26.0 29.0 26.0 28.1 26.0 28.5 8/11/75 A.M. | P.M. 1 1 1 1 1 1 26.4 26.0 26.0 25.0 26.2 25.4 Station SB SB ഗമ SB S  $\sim \infty$ Sug Sa လ ဆ **B-2** F-2 8-1 C-5 0-2 E-2 F-1 ٦ 0-1 1 ⋖

TABLE A-6(Continued)

	8/1	8/11/75	8/1	8/12/75	8/1.	8/13/75	8/1/	8/14/75	8/1	8/15/75		
Station	A.M.	P.M.	A.M.	Р.М.	A.M.	Р.М.	A.M.	P.M.	А.М.	Р.М.	Average	Range
S	25.0 24.1	1 1	: :	30.1 27.0	28.0 27.0	; ;	28.3	: :	28.1		27.4	24.1 - 30.1
T-1 8	26.0 25.0	!!	1 1	29.1 27.5	28.0	! !	27.3	: :	27.9	! !	27.0	25.0 - 29.1
T-2 S	25.9		: :	29.9	28.0	! !	28.8	; ;	28.0	1 1	28.1	25.9 - 29.9
U-1 S	26.0		: :	29.1 26.9	28.1	; ; ; ;	28.9		28.0	1 1	27.4	26.0 - 29.1
U-2 S	26.1 26.0			29.0	28.0		28.9		27.9		. 27.2	26.0 - 29.0
x-1 S	1 1	30.5	26.9	! !		30.5	1 1	31.0	30.0		28.9	26.5 - 31.0
X-2 S	1 1	29.1	26.9	!!!		30.2 28.0	1 1	30.8	28.8	1 1	28.3	26.5 - 30.8
y-1 S	1 1	29.8	27.0		1 1	29.7	1 1	31.0	30.0		28.6	26.8 - 31.0
Y-2 S	; ;	28.0 26.3	27.1 26.8	: :	1 1	30.5 26.1		30.1 26.2	30.0	! !	27.9	26.1 - 30.5

TABLE A-7

핌

8.4 8.8 8.3 5.5 8.4 7.9 8.4 7.9 7.8 8.1 0 2 3 ر 4 م 4 4 m . 8.8 Range  $\infty \infty \infty$ ထထထ 1 1 1 1 1 1 1 1 1 1 1 7.5 7.4 7.4 7.5 7.6 7.6 7.7 4.9. 7.4 7.4 7.5 က်က်က 9.10 ~ Median 7.8 7.7 7.8 7.5 7.5 7.6 7.7 7.8 7.6 7.4 7.6 8.2 8.1 8.2 7.7 7.7 7.6 A.M. | P.M. 1 1 1 111 111 1 1 1 1 1 1 1 1 1 6/13/75 7.4 7.8 7.7 7.8 7.5 7.6 7.4 7.6 8.2 8.1 8.2 <u>σ</u> 1 1 1 1 1 1 8.4 8.4 111 1 1 1 6/12/75 A.M. 7.5 7.5 7.6 7.6 7.8 7.8 7.6 7.7 7.6 7.7 7.8 7.8 7.1 7.4 7.4 7.5 Δ d 7.4 7.5 7.5 7.6 1 1 1 6/11/75 A.M. 7.5 7.5 7.5 7.6 7.5 7.5 7.5 7.5 7.5 7.4 8.2 8.2 <u>Σ</u> 8.0 8.0 7.7 7.9 8.4 4.6 6/10/75 ထထ A.M. 7.6 7.8 7.7 1 1 1 111 7.8 8.2 8.0 8.4 7.9 8.4 7.8 w 6 4 8.4 A.M. | P.M 4 0 5 6/9/75 ထ်ထဲထံ ထ်ထဲထ 7.9 111 111 1 1 1 111 111 Station A B-1 B-2 -5-G D-2 E-1 E-2 F-1 F-2 S T-1 T-2 U-1 U-2 X-1 X-2 Y-1 Y-2

TABLE A-8

Hd

	8/1	-	8/12/75	775	8/13/75	//5	8/14//5	//5	8/15/75	ر / /		
Station	А.М.	P. M.	Α.Μ.	۳. <del>۳</del>	A.A.	۳. چ	A.A.	Ξ.	A.M.	F. F.	median	Range
A	7.7	:	!	7.9	•	;		1		1		5 - 7.
8-1	7.7	;	;	8.3	7.6		8.2	!	8.0	;	8.0	7.6 - 8.3
B-2	7.6	-	-	8.1	•	1		:	•	!		6 - 8.
,	1			¢								0
ا -	/ · /	-	:	4.0		!	ο . Ο .	!		!	٠	000
C-2 D-1	7.6	7.6	8.2	4.8	χ 2. α. 4. τυ	1 1	- m 	1 1	8 8 		- m - m	7.6 - 8.5
2-0		7 7	α	-	α	;	α	!		;	7.7	ا ب
F-1		7.7	7.0		7.7	;	7.6	:	9.7	<del>`</del>	7.6	7.5 - 7.7
E-2	1	7.7	7.9	ļ	8.3	!	7.9	:	•	-	7.9	.7 -
	ŀ	7 5	7	;		;	7.5	!	•	:		.5 - 7.
F-2	1	7.7	7.7	1	7.9	:	7.8	:	7.5	:	7.7	7.5 - 7.9
S	7.5	;	ł	7.8	•	!	7.4	!	•	:		.3 - 7.
[-]	7.5	!	ł	7.9		:	8.0	!	7.8	1	7.9	.5 - 8.
T-2	7.6	!	1	7.9	8.0	;	7.9	;	7.8	1	7.9	7.6 - 8.0
U-1	7.4	;	1	8.0		;	0.8	1	0.8	:	8.0	.4 - 8.
U-2	7.4	!	:	7.8	7.8	;	7.8	ļ		:		.4 - 7
X1	ł	8.3	8.0	<u> </u>	;	8.5	ļ	8.7	8.6	;	8.5	8.0 - 8.7
x-2	!	8.4	7.7	;	!		1	8.7		1		.7 - 8
٧-١	-	8,1	8.4	-			!	8.7	8.8	;		.1 - 8.
Y-2	}	8.0	8.3	-	-	7.7	1	7.6	8.8	:	8.0	7.6 - 8.8
<u></u>			A	-	Ź							

TABLE A-9
SPECIFIC CONDUCTANCE (umhos/cm)

1/9	6/9/75	6/10/75	/75	5/11/75	/75	6/12/75	6/13/75		
A.M.	P. X	А.М.	Р.М.	A.M.	Σ.	A.M. P.M.	A.M. P.M.	Average	Range
; ;	330 1350	370 1150	::	1100				774	330 - 1350
1 1	305 260	290	; ;	11 1				287	260 - 305
1 1	320 275	255 700	: !	; ;		N	N	388	255 - 700
! !	210	! !	215	: :		NCTIO	NCTIO	221	170 - 265
: :	210 700	! !	260	230		MALFU	MALFU	283	160 - 700
1 1	175 175	! !	220	185	,	MENT	MENT	185	170 - 220
: :	175 225	; ;	170	170		EQUIP	EÓNIB	180	170 - 225
1 1	160	! !	188	172				170	150 - 188
1 1	158 155		160	170				191	150 - 172
1 1	163	1 1	160	168 270				181	155 - 270
11	157 175	1 1	155	170		-	-	164	148 - 177

TABLE A-9 (Continued)

	Range	185 - 290	155 - 190	158 - 167	156 - 165	150 - 165	148 - 170	148 - 175	148 - 170	148 - 190
	Average	223	169	162	162	160	162	163	159	168
6/13/75	A.M. P.M.			иоттэ	NUJJA	м Тиз	MqIUD	3		
9/15//9	A.M. P.M.			CTION	ALFUN	M TN3	MAIUQ	3		
6/11/75	P.M.									
.1/9	A.M.	1 1	!!		1 1	1 1	165	162	168	170 187
/75	P.M.	: :	: :	::	: :	: :	148	148 149	148 149	148 152
52/01/9	A.M.	198 240	161 155	158	156	160 150	1 1	1 1	1 1	1 1
51/6/9	P.M.	185 190	170	167 165	165 165	165 165	: :	: :	: :	: :
6/9	A.M.	1 1	1 1	1 1	1 1	!!	170	175	160	163 190
	Station	S B B	T-1 8	1-2 S	U-1 S	U-2 S	X-1 S	X-2 S	Y-1 S	Y-2 S

TABLE A-10

SPECIFIC CONDUCTANCE (umhos/cm)

	Range	340-700	276-380	265-640	200-296	190-490	188-253	171-340	170-280	168-185	168-260	165-183
	Average	498.0	323.2	412.5	243.0	244.2	205.4	238.5	208.6	176.1	192.9	173.4
8/15/75	Р.М.	1 1	! !	! !	!!	: !	1 1	!!	1 1		1 1	
8/1	A.M.	340 600	310	300	260	198 225	188 190	171 188	171 235	180	179 197	180
8/14/75	P.M.	: :	t 1 1 I	!!	! !	1 1	!!	: :	! !	; ;	1 1	
8/1/	A.M.	430 650	380	360 460	265 296	270 190	253 20 <b>4</b>	211 282	199	179 168	191 260	170
8/13/75	P.M.	t	! !	: :	! !	1 1	! ! ! \$	1 1	1 1	! !	1 1	1 1
8/1	A.M.	480 465	276	325 485	251 245	205	230	290	191 255	185 170	175 189	178
8/12/75	P.M.	340	280	265 480	200	200	! !	! !	t !		1 1	! !
8/1	А.М.	1 1	1 ;	! !	! !		190	183	170	170 171	168 230	168
8/11/75	P.M.	: :	! !	1 1	1 1	1 1	190 198	240 340	170 175	175 183	170 170	175 165
8/1	А.М.	375 600	370	330 640	203 220	210 490		1 1	1 1	! !	1	! °
	Station	A B	8-1-8	8-2 S	C-1 S	C-2 8	D-1 S	0-2 S B	E-1 8	E-2 S	F-1 8 8	F-2 S

TABLE A-10 (Continued)

	8/1]	8/11/75	8/12/75	1/75	8/13/75	/75	8/14/75	/75	8/15/75	/75		
Station	A.M.	P.M.	A.M.	P. M.	A.M.	P.M.	A.M.	Р.М.	A.M.	Р.М.	Average	Range
S &	168 192		1 1	181 205	192 218	1 1	200 220		220 230	1 1	202.6	168-230
S B	170	1 1	! !	175 180	178 235		183	1 1	181	: :	191.3	170-235
S 8	168		: :	179	178		180	!!	178		176.6	168-180
S B	167 170		: :	175 172	175 173	1 1	177	! !	174		172.8	167-177
νæ	169 170	! !	: :	173 179	175 181	! !	178 175	: :	177	! !	. 176.7	169-190
S &	1 1	170 158	159	! !	; ;	173 165	1 1	175 168	172 175		167.6	158-175
S 8	! !	168 158	160	!!	: :	171 166	1 1	177 168	175 179	1 1	168.3	158-179
BS	! !	168 160	161 163	1 1	! !	178 165	: :	176 168	173	; ;	168.4	160-178
S &		165 165	162 165	1 1		175 170		173 169	175 170	: :	168.9	162-175

APPENDIX A-3

CHEMICAL WATER QUALITY

Waconda Bay Reference Bay "A" Huss Lowe Slough Chickamauga Lake

TABLE A-11
TOTAL HARDNESS
(mg CaCO<sub>3</sub>/1)
JUNE SURVEY

Station	6/9/75	6/10/75	6/11/75	6/12/75	6/13/75	Mean	Standard Deviation
A	87.0	103.0	147.0	146.0	143.0	125.0	78
8-1	79.0	80.0	85.0	89.0	110.0	88.6	13
8-2	82.0	95.0	75.0	114.0	123.0	97.8	20
<u>ا</u>	0.79	72.0	75.0	71.0	73.0	71.6	2.9
C-2	79.0	0.79	77.0	74.0	79.0	75.2	2
<u>-</u>	64.0	71.0	67.0	74.0	72.0	9.69	4
0-2	0.99	65.0	65.0	65.0	67.0	65.6	0.1
<u></u>	0.99	•	64.0	63.0	0.99	64.8	1.5
E-2	62.0	65.0	65.0	67.0	71.0	99	3.3
	63.0	65.0	63.0	70.0	63.0	64.8	3.0
F-2	63.0	62.0	65.0	0.99	62.0	63.6	1.8
		•					
S	69.0		81.0	71.0	76.0	5.	5.1
1	59.0	0.79	67.0	67.0	0.99	65.2	3.5
T-2	65.0	65.0	70.0	65.0	0.79	ô.	2.2
<u>-</u> -	65.0	65.0	67.0	62.0	67.0	5.	2.0
n-2	67.0	0.39	64.0	65.0	63.0	99	9.1
							1
- X	61.0	61.0	65.0	61.0	59.0	61.4	2.5
x-2	67.0	63. C	59.0		90.0	$\sim$ 1	3.2
_ ( } :	60.0	63.0	60.0		0.19	19	2
Y-2	63.0	0.19	0.19		02.0		٠

TABLE A-12
TOTAL HARDNESS
(mg CaCO<sub>2</sub>/1)
AUGUST SURVEY

Standard Deviation	4.0 12.0 8.2 9.2 6.9 3.6 1.5	7.0 1.5 0.6 1.2 1.2 1.2
Mean	105 97.7 97.7 77.0 79.3 73.0 71.0 70.0 72.7 64.7	73.0 67.3 67.3 66.7 65.3 65.0 64.0
8/15/75		
8/14/75	103 99.0 90.0 87.0 82.0 81.0 70.0 71.0 65.0	81.0 69.0 67.0 66.0 66.0 65.0
8/13/75		
8/12/75	103 85.0 97.0 69.0 70.0 69.0 73.0 66.0 66.0	68.0 66.0 67.0 66.0 66.0 65.0
8/11/75	110 109 106 75.0 86.0 69.0 73.0 66.0 64.0	70.0 67.0 66.0 66.0 64.0 65.0 65.0 67.0
Station	A B-1 C-2 D-2 E-1 F-2	S T-1 U-1 V-2 X-2 Y-2

TABLE A-13 SULFATE (mg SO<sub>4</sub>/1)

Mean	62.2 50.3 53.0 18.7 23.5 16.9 17.7 14.3	14.1 12.8 12.0 12.5 10.0 10.3
August 8/12/75	55.4 33.0 49.6 17.2 16.8 16.8 17.0	13.8 12.8 13.5 13.5
8/11/75	69.0 67.5 56.4 30.4 16.9 11.5	14.4 - 11.6 9.5 9.5 9.9
Mean	108 52.6 80.0 26.6 25.5 25.3 14.7 16.3 16.4	27.3 17.4 17.2 15.4 13.4 10.9 10.7
6/12/75	115 48.6 83.2 27.4 23.6 16.7 15.0 13.3	27.0 16.2 15.0 12.8 10.5 10.6
June 6/10/75	100 55.8 76.8 23.4 23.6 12.7 39.7 17.2 15.3	27.6 18.6 18.0 15.7 14.0 10.7 10.7
Station	A B-1 C-2 1-2 F-2	S L-1 L-2 L-3 L-3 2

TABLE A-14

5 <b>c</b> 0			
Standard Deviation	0.8.8.8.9.9.8.6.8.8. 8.5.6.6.9.8.8.8.8.8.8.8.8.8.8.8.8.8.8.8.8.8		3.3
Mean	46.5 48.8 49.8 49.0 48.8 48.2 49.0	49.8 49.1 47.9 50.2	48.5 50.1 49.4 49.6
6/13/75	39 44 45 44 44 45	48 46 45 47	43 47 46 45
6/12/75	47 53 52 52 52 53 53 53.5	50 53 52 51.5	51.5 51 53 54
6/11/75	40.5 51.5 48.5 52.5 51.5 51.5 49.5	54 53 51.5 48 51.5	50 52.5 51.5 50.5
6/10/75	51 52.5 49.5 50 51 51 49.5 50	51 48.5 47.5 51 50.5	50 50 50.5 50.5
6/9/75	55 50 52 48 47 48 46 46	46 45 44 50	48 50 46 48
Station	A B B	S T-1 T-2 U-1	X-1 X-2 Y-1

TABLE A-15

TOTAL ALKALINITY (mg CaCO3/1) AUGUST SURVEY

Standard Deviation	4.50.0.0.0.0.0.0.0.0.0.0.0.0.0.0.0.0.0.0	3.0	23.22.2.5.3.5
Mean	50.4 54.8 57.2 55.8 55.0 53.8 53.8	63.0 56.6 56.4 55.0 55.2	57.2 55.2 58.6 55.2
8/15/75	55 57 60 57 57 55 58 58	70 60 59 55 56	59 60 62 57
8/14/75	59 58 59 57 57 57 56	67 57 60 57 58	59 55 55
8/13/75	48 557 553 563 563	55 55 55 55 55 55 55 55 55 55 55 55 55	54 51 53
8/12/75	4 C C C C C C C C C C C C C C C C C C C	2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2	56 59 56
8/11/75	47 52 57 57 54 55 55	64 53 54 54	58 55 57 55
Station	A B-1 C-2 D-2 F-1 F-2	S 1-1 1-2 0-1	x-1 x-2 Y-1 Y-2

TABLE A-16 CHLORIDE (mg Cl/1)

Standard Deviation	5.6 1.7 1.6 0.5 0.7 0.7	1.2 0.9 0.7 0.8	0.7 0.7 0.8
Mean	17. 8.5 8.5 4.8 3.3 3.3 3.3 4.3	3.8.8 3.6.9 6.6	2.3.2.9 2.8.2.9 8.2.8
6/13/75	15.0 14.5 3.5 3.5 3.0 2.5 5.5 5.5	2.0 2.0 2.5 5.0	6.6.4 0.0.0
6/12/75	21.5 7.5 11.5 6.0 3.0 4.5 3.0 4.5	2.3.5 2.0 2.5 2.0	2.5 2.5 5.5
6/11/75	25.0 8.5 6.5 3.0 3.5 3.5	0.4.8.6.0 0.5.6.0	8.0 8.5 9.5
6/10/75	6.000 4 0	0.44 E.S	0.0.6.6.
9/1/5	20 20 20 20 20 20 20 20 20 20 20 20 20 2		2.0 2.5 2.5 5.5
Station	488-0-0-0-0-1-1-1-1-1-1-1-1-1-1-1-1-1-1-1	S 1-1 1-2 1-1	×××-2 2 2

TABLE A-17 CHLORIDE (mg C1/1)

Standard Deviation	8.01 8.28 8.00 9.0 9.0 8.0 0.3	0.7.00	1.0 0.3 0.5 0.5
Mean	41.7 28.5 28.0 11.3 13.5 10.3 9.3 6.8	5.8 6.7 6.7	5.7 5.8 6.0 5.5
8/15/75	1 1 1 1 1 1 1 1 1 1 1	1111	1 1 1 1
8/14/75	51.0 36.0 24.5 14.5 12.5 9.5 9.0 5.0	0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0	6.5 5.5 5.0
8/13/75		1 1 1 1 1	1 1 1 1
8/12/75	36.0 16.0 29.5 9.5 9.5 7.7 7.5	.5 6.0 7.5 5.5	4.5 6.0 5.0
8/11/75	38.0 33.5 30.0 17.0 9.5 9.5 6.5	7.0 8.0 7.0 6.5	6.0 6.0 5.5 6.0
Station	A B-1 C-2 0-1 E-1 F-1	27-1-1 1-1-2 1-1-2	x-1 x-2 Y-1

TABLE A-18
TOTAL DISSOLVED SOLIDS
(mg/l)

284 208 242 96
-
•
***************************************
•
•

TABLE A-19 TOTAL DISSOLVED SOLIDS (mg/l)

-
2
_
_
2
_
_
_
_
_
205
192
_
_
_
32
<u> </u>

TABLE A-20 SUSPENDED SOLIDS (mg/1)

			<del></del>
Standard Deviation	0.04.0.04.0.0.0.0 0.0000000000000000000	0.8 0.0.0.0 0.0.0.0	5.2 6.5 11 7.2
Mean	01 01 13 8 8 6.9 7 8.7	4 8 0 6 6 6	0 6 6
6/13/75	13 22 13 8	14 7 11 12 11	11 2 7 1 7
6/12/75		1 1 1 1 1	1 1 1 1
6/11/75	8 10 8 7 8 7	21 10 6	27 2 8 4
6/10/75		1111	1 1 1 1
6/9/75	41 6 7 8 8 8 8	8 12 8 10 11	12 24 21 5
Station	P-2-12-12-12-12-12-12-12-12-12-12-12-12-1	S 1-1 U-1 U-2	x-1 x-2 Y-1 Y-2

TABLE A-21 SUSPENDED SOLIDS (mg/l)

Standard Deviation	12 13 20 11 14 10	0 0 8 4 4	10 2 2 15
Mean	14 22 22 14 10 10 8	121 11 4 8	12 4 5 12
8/15/75	26 24 35 24 26 18 20	32 22 18 2	22 2 6 30
8/14/75	1 1 1 1 1 1 1 1 1 1 1		1 1 1 1
8/13/75	4 4 0 8 8 8 8 8 8 8 8 8 8 8 8 8 8 8 8 8	71 22 22 25	ოოოო
8/12/75		1 1 1 1	1 1 1 1
8/11/75	3 15 13 16 3 4 5 9	15 7 13 8	11 6 6
Station	A B-1 C-2 D-1 F-1 F-2	S T-1 1-2 U-1	x-1 x-2 Y-1 Y-2

TABLE A-22 TOTAL SOLIDS (mg/1)

ρμ		
Standard Deviation	42 20 32 17 6.1 9.1 8.4 7.2 7.2	9.3 8.5 7.2 7.2 6.6 10 11
Mean	285 192 211 211 122 138 117 111 103 96	145 111 105 108 94 95
6/13/75	297 214 248 110 135 103 97 98 91	134 83 110 100 102 75 82 84 84
6/12/75		
6/11/75	320 186 183 142 145 125 104 101 111	151 108 106 101 107 108 95
6/10/75		
6/9/75	138 176 203 114 134 117 103 100 128	149 142 123 113 115 98 103 94
Station	A 8 8 - 2 C C - 2 C C - 2 C C C C C C C C C C	N

TABLE A-23 TOTAL SOLIDS (mg/l)

·	T	
Standard Deviation	65 32 32 44 79 41 68 56 95	36 66 30 30 45 64 56
Mean	198 243 215 215 159 202 97 160 164 127 175	170 161 147 152 140 135 134 122
8/15/76	272 288 288 251 208 262 127 233 130 158 239	191 193 192 226 173 173 196 196 176
8/14/75		
8/13/75	151 252 204 150 232 113 154 164 181 124	202 167 173 134 134 134 125
8/12/75		
8/11/75	170 190 191 120 112 49 84 84 96 68	119 122 76 97 114 76 65
Station	R B B - 1 C - 2 D - 2 F - 2 F - 2	X X X C C C C C C C C C C C C C C C C C

TABLE A-24 CHEMICAL OXYGEN DEMAND (mg/l)

Standard Deviation	60-668-8-27	æoo.	0 4 9 0 o
Sta Devi			-0
Mean	99.2 9.0 6.0 6.0 6.9 6.9	6.8 6.7 6.1	6.8 6.9 6.6
Σ Σ			
6/13/75	0.6 2.2 2.3 8.9 5.7 5.9	6.6 7.7 6.9 6.3	7.9 7.6 6.7
/9	20/20/20/20/20/20		
6/12/75	~	<b>ന</b> ന ന <b>ന</b> ന	V V 85
6/12		<u>ઌઌઌ૱</u> ઌ	6.0.4.8
/75			
6/11/75	8.7.7.88.8.4.8.8.8.8.8.8.8.8.8.8.8.8.8.8	10.2 4.5 6.9 5.7 4.5	5.5 6.5 7.7
75			
6/10	6.00 6.00 6.00 6.00 6.00 6.00 6.00 6.00	6.1 7.7 7.7 7.7	5.3 6.9 4.4
75		- 60 00 7	7 2 8
6/9/75	10. 8. 7. 8. 8. 7. 6.	8. 7. 7. 8.	9.8.7.
Station	FF-20-20-2-1	S T-1 U-1 U-2	x-1 x-2 Y-1 Y-2
Sté	<b>နောတ်တု</b> တ်တုတ်ကိုင်း	25.57.7	****

TABLE A-25 CHEMICAL OXYGEN DEMAND (mg/1)

Standard Deviation	24.24.9. 8.8.2.0.0.4.8. 8.9.0.0.0.4.9.	1.7 2.7 2.9 2.0	2.3 2.3 2.5 5.5
Meal	5.8 8.6 8.6 7.7 7.0 6.1 6.1 4.9	7.1 4.4 4.2 4.5	4 4 4 4 4 . 4 . 4 . 4 . 4 . 4 . 4 . 4 .
8/15/75	7.8 9.7 9.1 11.9 12.5 7.5 8.7 8.4 10.3	7.2 1.6 4.1 5.3	2.5 5.6 0.9
8/14/75	5.0 12.8 10.0 10.0 8.1 5.6 5.9 5.9	7.5 6.6 3.4	2.5 2.8 3.7
8/13/75	3.1 7.8 7.8 7.2 3.4 9.9 0.9	6.9 6.3 7.3	6.2 6.9 6.9
8/12/75	3.5 9.0 3.2 3.2 6.7 6.7 2.6	4.2 0.5 0.3 1.6	3.8 4.2 5.4
8/11/75	9.5 10.2 11.8 11.8 7.2 5.6 3.0 0.7	88.83.2 20.000	6.9 6.2 7.5
Station	A B-1 C-2 C-2 D-1 E-1 F-1	5 1-1 1-2 U-1 U-2	x - 1 x - 2 x - 2

TABLE A-26
TOTAL ORGANIC CARBON
(mg C/1)

Standard Deviation	2	2.3	0.8 0.9 1.2
Mean	7.00.00.00.00.00.00.00.00.00.00.00.00.00	7.2 5.8 5.9 5.9	5.9 6.1 5.5
6/13/75	0.0.04.04.44.04 0.0.0.0.0.0.0.0.0.0.0.0.		5.0 5.0 4.5
6/12/75	დ. დ.დ. 4.დ. ბი. ი.	ა. 4 ფიაი დ.	5.5 5.5 6
6/11/75	7. 5. 6. 5. 5. 5. 5. 5. 5. 5. 5. 5. 5. 5. 5. 5.	6.5 5.5 5.5 5.5	6.5 6 4
6/10/75	6 6 7 7 8 8 8	6 6.5 7.5 6.5	5.5 7 6
6/9/75	11 8.5 8.5 7.5 6.5 7.5 10	9.5 8.5 7 11 7.5	7
Station	P-1 C-2 P-2 F-2 F-2	S T-1 L-2 L-2	X-1 X-2 Y-1

TABLE A-27
TOTAL ORGANIC CARBON (mg C/1)

Station	8/11/75	8/12/75	8/13/75	8/14/75	8/15/75	Mean	Standard Deviation
A 18	4.5 8.5	1 1	6.0		5.0 5.5	4.8 6.1	1.0
8-2 C-1	6.0	1 1	0.0		5.50	5.9 .0.	0.6 2.5
0-1 0-2	5.5°	1 1 1	8.0 8.0		5.0 5.0	5.8 5.6	- 0. 9.6.9.
E-1 F-2	4.0 6.5 0.0	1 1 1 1		5.5 5.5 17.5 7.5		6.5 6.0	0.6 0.7 1.1
S T-1 T-2 U-2	5.5 7.5 5.0		7.0 8.0 5.0 7.0 6.0	7.0 5.5 6.0 12.0 6.0	7.0 5.0 5.0 5.5	6.6 6.0 5.9 5.6	0.1.2 3.1.2 0.5
x-1 x-2 Y-1 Y-2	6.0 7.5 7.0 6.0	1 1 1 1	7.0 6.0 5.0 5.0	4.5 5.0 4.0 6.5	4.0 5.5 5.0	5.4 6.0 5.3 5.6	1.1

TABLE A-28

TOTAL KJELDAHL NITROGEN (mg N/1)

Station	6/9/75	6/10/75	6/11/75	6/12/75	6/13/75	Mean	Standard Deviation
A		5	0.84				
8-1		2	0.61				
8-2		9	0.65				
<u>ا</u> ۔		4	0.47				•
C-2		က	0.27				
D-1		က	0.37				
D-2		4	0.34				
		က	0.40				
<b>E-</b> 2		2	0.37				
		0.33	0.27	0.42	0.30	0.33	90.0
F-2	0.32		0.50				
s	0.32	0.38	0.41	0.52	0.53	0.43	0.09
-		$\sim$	2	က	0.55		
1-2		$\sim$	_	$\sim$	0.50		
-i		$\sim$	2	2	0.47		
n-2		~	$\sim$	က	ı		
x-1		4	$\sim$		0.31	က	0.07
x-2		က	က	0.39	0.34	က	0.05
\ <del>-</del> 1	0.30	0.38	0.26	0.38	0.31	0.33	0.05
Y-2		4	~	0.43	0.38	က	0.09

TABLE A-29 TOTAL KJELDAHL NITROGEN (mg N/1)

							,
Station	8/11/75	8/12/75	8/13/75	8/14/75	8/15/75	Mean	Standard Deviation
A		7.	0.72	0.75	6.45		2.55
8-1	0.65	0.68	0.76	0.78	1.58	0.89	0.39
8-2	•	3	0.71	0.78	1.43		0.33
C)	•	ω.	0.41	0.68	0.74		0.19
C-2	•	4.	0.39	0.62	0.64		0.12
D-1	•	7	0.37	0.62	0.41		0.13
D-2	•	.7	0.43	•	0.38		0.10
E-1	•	2	0.29		0.34		0.04
E-2	•	٦.	0.27		0.32		0.07
F-1	•	Ÿ	0.27	0.52	0.46		0.13
F-2	0.23	0.22	0.28		0.40		0.08
S	0.42	0.23	0.40	0.32	0.38	0.35	0.08
	2	_	0.32	0.45	0.41		
T-2	٧.	_	0.37	0.52	0.44		
<u>-</u> -	4.	?	0.37	0.45	0.47		
n-2	۳.		0.30	0.43	0.48		
X-1	0.45	0.21	0.30	0.43	0.47	0.37	0.1
x-2	4.	<u>ب</u>	0.34		0.42		0.13
- ·	.5	N	0.25		0.65		0.19
۲-2	9.	ú	0.33		0.67		0.21
j							

TABLE A-30
AMMONIA
(mg N/1)

							Standard
Station	6/9/75	6/10/75	6/11/75	6/12/75	6/13/75	Mean	Deviation
¥	0.08	0	0.36	0.48	09.0	0.32	0.24
8-1	0.03	က	90.0	0.13	0.37	0.19	0.16
B-2	0.03	0	0.12	0.25	0.37	0.17	0.14
د-1	0.01	0	90.0	0.04	0.12	0.05	0.04
C-5	0.01	0	0.05	90.0	0.02	0.03	0.02
0-1	0.01	0.03	0.04	0.04	0.01	0.03	0.02
D-2	0.01	$\mathbf{c}$	0.02	90.0	0.01	0.03	0.02
E-1	0.01	0	0.04	0.04	0.01	0.02	0.02
E-2	0.01	0	0.05	٠.04	0.01	0.02	0.02
<u>_</u>	<0.01	0.01	0.02	0.03	0.01	0.01	0.01
F-2	0.01		0.05	0.04	0.01	0.02	0.02
S	0.01	0	90.0	0.11	•	90.0	•
	•	0	90.0	0.07	•	0.03	
1-2	•	0	0.03	0.07	•	0.03	
<u>-</u> -	•	0.04	0.04	90.0	0.37	0.11	0.14
n-2	0.01	0	0.04	0.03	•	0.03	
	90.0	0.01	0.02	0.01	0.01	0.02	0.02
	•	0.01	0.01	<0.01	0.01	0.02	0.01
<b>-</b>	•	0.01	0.05	0.02	0.05	0.02	0.01
	•	0.02	10.0	0.02	0.02	0.02	0.004

TABLE A-31
AMMONIA
(mg N/1)

Standard Deviation	2.16 0.03 0.03 0.05 0.05 0.05 0.05	0.03 0.05 0.03 0.06	0.03 0.04 0.05 0.05
Mean	0.00 0.034 0.00 0.008 0.008 0.009 0.009	0.04 0.03 0.03 0.05	0.05 0.05 0.04 0.04
8/15/75	4.99 1.22 1.02 6.23 0.09 0.06 0.01 0.08	0.08 0.05 0.05 0.05	0.04 0.04 0.02 0.04
8/14/75	0.13 0.06 0.05 0.05 0.07 0.09 0.08	0.00 0.00 0.03 0.03	0.07 0.05 0.02 0.03
8/13/75	0.07 0.04 0.05 0.05 0.02 0.03 0.08	0.04 0.04 0.02 0.02 0.02	<pre>&lt;0.02 &lt;0.02 &lt;0.02 &lt;0.02 &lt;0.02</pre>
8/12/75	0.10 0.20 0.06 0.05 0.13 0.13 0.16 0.16	<pre>&lt;0.02 0.05 0.05 &lt;0.02 &lt;0.02</pre>	0.06 0.04 <0.02 <0.02
8/11/75	0.33 0.25 0.25 0.06 0.07 0.04 0.03	0.07 0.16 0.14 0.06 0.15	0.06 0.12 0.12 0.15
Station	A 10-2 10-2 10-2 1-2 1-2	S T-1 U-2 U-2	x-1 x-2 Y-1 Y-2

TABLE A-32 NITRITE (mg N/1)

Standard Deviation	0.048	0.007	0.007	0.001	0.001	0.004	0.002	0.002	0.001	0.001	0.002
Mean	0.251 0.117 0.116	0.038	0.036	0.023	0.021	0.024	0.020	0.021	600.0	0.008	0.013
6/13/75	0.270	0.037	0.040	0.025	0.020	0.028	0.022	0.023	0.009	0.008	0.015
6/12/75	0.310	0.038	0.040	0.023	0.021 0.020	0.027	0.022	0.022	0.009	0.008	0.014
6/11/75	0.271	0.039	0.029	0.022	0.021	0.025	0.020 0.019	0.019	0.008	0.008	0.009
6/10/75	0.217	0.047	0.043	0.025	0.020	02	00	0.020	0.010	0.009	0.013
9/1/2	0.189	0.027	0.028	0.023	0.022	0.018	0.018	0.019	0.00	0.009	0.013
Station	A B-1	2-1 2-2	3 - C	п — 2 п — 2	F-1 F-2	S	1-1	U-1	- X	x-2 Y-1	Y-2

TABLE A-33 NITRITE (mg N/1)

Station	8/11/75	8/12/75	8/13/75	8/14/75	8/15/75	Mean	Standard Deviation
A B-1 C-2 C-2 D-1 E-2 F-1	0.183 0.179 0.149 0.034 0.067 0.027 0.014 0.012 0.011	0.158 0.072 0.072 0.026 0.020 0.023 0.023 0.012 0.009	0.216 0.053 0.102 0.045 0.027 0.030 0.018 0.018 0.010	0.178 0.141 0.092 0.051 0.048 0.043 0.022 0.029 0.009	0.170 0.113 0.100 0.057 0.025 0.019 0.009 0.014 0.012	0.181 0.112 0.115 0.043 0.028 0.028 0.017 0.010	0.021 0.051 0.024 0.020 0.009 0.005 0.001
S T-1 U-1 U-2	0.010 0.012 0.012 0.013	0.010 0.010 0.010 0.010	0.010 0.000 0.009 0.009	0.008 0.008 0.008 0.008	0.014 0.009 0.009 0.009	0.00 0.00 0.00 0.00 0.00	0.002 0.001 0.002 0.002
x-1 x-2 Y-1	0.005 0.005 0.005 0.005	0.005 0.005 0.005 0.005	0.004 0.004 0.004 0.005	0.004 0.003 0.004 0.004	0.005 0.004 0.005 0.005	0.005 0.004 0.005 0.005	0.00 0.00 0.00 0.00 0.00

TABLE A-34 NITRATE (mg N/1)

Standard Deviation	1.52 0.71 1.28 0.21 0.30 0.12 0.18 0.76 0.05	0.34 0.09 0.11 0.04 0.06 0.06
Mean	5.04 3.34 0.92 0.93 0.63 0.40 0.40	1.15 0.54 0.50 0.53 0.23 0.24 0.26
6/13/75	7.17 3.82 5.33 1.04 0.95 0.82 0.68 0.43	0.58 0.68 0.65 0.19 0.22 0.33
6/12/75	2.2.8 0.0.0.0 0.98 0.99 0.55 0.55 0.56 0.56 0.56 0.56 0.56 0.56	1.29 0.58 0.58 0.47 0.26 0.25 0.25
6/11/75	4.96 2.23 2.48 0.95 0.56 0.50 0.51 0.42	1.34 0.59 0.48 0.23 0.23
6/10/75	3.41 2.40 2.50 1.13 0.85 0.69 0.63 0.44	1.03 0.56 0.46 0.47 0.23 0.25 0.26
9/1/2	3.84 2.18 2.45 0.59 0.55 0.55 0.27 0.35	0.61 0.39 0.37 0.39 0.39 0.15 0.15
Station	FF-2-12-12-12-13-13-13-13-13-13-13-13-13-13-13-13-13-	S 1-1 1-2 1-2 1-2 4-2

TABLE A-35 NITRATE (mg N/1)

TABLE A-36 TOTAL PHOSPHORUS (mg P/1)

Standard Deviation	0.005 0.012 0.005 0.007 0.004 0.003 0.007 0.005	0.012 0.008 0.007 0.007 0.005	0.011 0.008 0.005 0.005
Mean	0.018 0.025 0.024 0.030 0.027 0.022 0.023 0.023	0.029 0.022 0.020 0.020 0.020	0.026 0.024 0.022 0.022
6/13/75	0.016 0.019 0.022 0.022 0.022 0.018 0.021 0.021	0.027 0.920 0.017 0.018 0.019	0.019 0.016 0.017 0.020
6/12/75	0.015 0.043 0.042 0.034 0.034 0.028 0.027 0.029 0.031	0.043 0.033 0.027 0.030 0.028	0.031 0.026 0.028 0.028
6/11/75	0.012 0.032 0.027 0.033 0.029 0.024 0.025 0.023	0.037 0.022 0.025 0.021 0.022	0.042 0.026 0.028
6/10/75	0.023 0.017 0.025 0.025 0.024 0.021 0.021 0.020 0.018	0.025 0.021 0.020 0.019 0.020	0.019 0.016 0.019 0.021
6/9/75	0.022 0.017 0.020 0.020 0.020 0.020 0.023 0.017 0.017	0.011 0.002 0.009 0.011	0.018 0.036 0.019 0.017
Station	A BB-1 C-2 1-2 F-2	S T-2 U-1 U-2	x-1 x-2 Y-2

TABLE A-37 TOTAL PHOSPHORUS (mg P/1)

033 033 033 033 029 020 020 000 000 000	0.74 0.053 0.051 0.065 0.024 0.037 0.037 0.037 0.037 0.037 0.038 0.038 0.028 0.029 0.029 0.025 0.024 0.029	0.046 0.037 0.051 0.065 0.037 0.033 0.033 0.033 0.028 0.029 0.029 0.029	38 0.074 0.053 0.053 0.053 0.051 0.053 0.051 0.065 0.037 0.051 0.065 0.033 0.033 0.033 0.033 0.028 0.029 0.029 0.025 0.025 0.024 0.024 0.025 0.024 0.025 0.024 0.025 0.024 0.024 0.025 0.024 0.024 0.025 0.024 0.025 0.024 0.024 0.025 0.025 0.024 0.025 0.0	38 0.074 0.053 0.053 0.053 0.051 0.053 0.051 0.065 0.037 0.051 0.065 0.033 0.020 0.033 0.029 0.029 0.025 0.025 0.024 0.025	0.027         0.038         0.074         0.053         0.025           0.031         0.032         0.046         0.037         0.038           0.042         0.030         0.051         0.065         0.048           0.031         0.032         0.024         0.037         0.054           0.025         0.033         0.020         0.031         0.031           0.030         0.034         0.034         0.032           0.019         0.034         0.028         0.029         0.027           0.030         0.024         0.029         0.027         0.022           0.021         0.024         0.025         0.024         0.024           0.019         0.029         0.024         0.024         0.024           0.019         0.029         0.024         0.024         0.024
033 034 029 027 027 064 06	020 031 028 028 016 016 025 0.024 0.029 0.096	020 0.033 0.031 0.031 0.031 0.028 0.029 0.027 0.027 0.025	33 0.020 0.033 0.033 0.033 0.033 0.033 0.034 0.028 0.029 0.027 0.025 0.024 0.025	33 0.020 0.033 0.033 0.033 0.033 0.034 0.023 0.028 0.029 0.027 0.025 0.024 0.025 0.024 0.026	025 0.033 0.020 0.033 0.03 030 0.034 0.031 0.034 0.03 019 0.034 0.028 0.029 0.03 030 0.075 0.016 0.027 0.02 021 0.024 0.025 0.024 0.03
024 0.	025 0.024 0. 029 0.096 0.	025 0.024 0.0	24 0.025 0.024 0.	24 0.025 0.024 0.	021 0.024 0.025 0.024 0. 019 0.029 0.029 0.096 0.
		029 0.096 0.	29 0.029 0.096 0.		
0.025 0.041 0.023 0.030 0.193 0.027	041 0.025 0.026 0.026 0.025 0.023 0.025 0.	041 0.025 0.026 0.026 0.025 0.023 0.025 0.	34 0.041 0.025 0. 27 0.026 0.023 0.	34 0.041 0.025 0. 27 0.026 0.023 0.	030 0.034 0.041 0.025 0.029 0.027 0.026 0.023 0.028 0.028
0.023 0.193 0.115 0.039	026 025 022 0. 022 0.	026 025 022 0. 022 0.	27 0.026 0. 32 0.025 0. 27 0.022 0. 27 0.022 0.	27 0.026 0. 32 0.025 0. 27 0.022 0. 27 0.022 0.	029 0.027 0.026 0.028 0.028 0.023 0.025 0.025 0.025 0.022 0.023 0.022 0.032 0.027 0.022 0.02
0.039	022 0.	0022	27 0.022 0.	27 0.022 0.	0.027 0.022 0 0.027 0.022 0
	041 026 025 022	029 025 022 022	29 0.029 34 0.041 27 0.025 27 0.022 27 0.022	23 27 27 27 27 20.025 27 27 0.022	0.034 0.041 0.027 0.026 0.032 0.025 0.027 0.022 0.027 0.022
	0.041 0.026 0.025 0.022	0.029 0.026 0.025 0.022 0.022	23 27 27 27 27 27 27 20 0.0	23 32 52 60 6. 52 52 60 60 60 60 60 60 60 60 60 60 60 60 60	0.034 0.027 0.032 0.027 0.027 0.027
00000	00000	0.030 0.029 0.028 0.021 0.032	0.030 0.029 0.028 0.021 0.032		

#### APPENDIX A-4

TRACE METALS IN HARRISON BAY, LAKE CHICKAMAUGA

JUNE 9-13, 1975 AUGUST 11-15, 1975

Waconda Bay Reference Bay A Huss Lowe Slough

TABLE A-38

# DAY TO DAY VARIATION IN IRON AND LEAD AT SELECTED STATIONS IN WACONDA BAY JUNE 2-6, 1975

### Iron (mg Fe/1)

Date	А	Station C-1	C-2
June 2 June 3 June 4 June 5 June 6	0.31	0.17	0.30
	0.27	0.21	0.23
	0.19	0.17	0.29
	0.25	0.25	0.31
	0.27	0.22	0.28

## Lead (mg Pb/1)

Date	А	Station C-1	C-2
June 2 June 3 June 4 June 5 June 6	0.006	0.014	0.008
	0.018	0.014	0.008
	0.015	0.017	0.012
	0.012	0.006	0.012
	0.010	0.008	<0.002

TABLE A-39

SELECTED METALS IN HARRISON BAY, CHICKAMAUGA LAKES, JUNE 4, 1975

		16	Meta	1 Concenti (µg/1) Fe	ration		
Station	Cd	Cr <sup>+6</sup>	Cu	Fe	Pb	Ni	Zn
A C-1 F-1 S X-1	<5 <5 <5 <5 <5	<5 <5 <5 <5 <5	7 <5 <5 <5 <5	190 170 82 160 120	15 17 17 21 17	6 5 5 <5 <5	<2 2 <2 <2 <2 <2

TABLE A-40

SELECTED METALS IN HARRISON BAY, CHICKAMAUGA LAKE, AUGUST SURVEY, 1975

al	Metal Date		A	B-1	B-2	C-1	C-2	D-1	D-2	F-1	F-2	S	γ-1	Y-2	
	August 1		<15 21/15	<15 <15	<15	<15	<15	<5		<15		<15 <15	<15 <15	>/91>	<15/<15/<15
	_ ,- ;-	<u></u> Δ.	<15 18 15/<15	8   c  >   18   15/<15		<15	<15			<15		\$\frac{1}{5} \frac{1}{5} \frac{1}{5}\$	<15 <15 <15/15		
	August 1	- 20	<5 <5/6	17	, 5	S	16	, 5		<5		9 2 2 9	\$\$	<5/<5/<5	-/<5
		24 rd	6 9 <5/<5	<2 6/5		7	14			20		- <del>(</del> , <del>(</del>	25 8 12/12		
Zn	August 1	22	26 18/30	21	46	28	21	57		48		19			,
		5 4 5	22 38/40	34/26		35	22			30		19	48 35/48		
Fe	August 12	2		259	221			126					113		
Cr+6	August 12	2		<del>-</del>	, 5			<u>،</u> 5			, ,		<5		
n O	August 12	2		ç	<.5			\$					, 5		

# APPENDIX A-5

MUNITIONS RESIDUES IN CHICKAMAUGA LAKE JUNE AND AUGUST, 1975

> Waconda Bay Reference Bay A Huss Lowe Slough

TABLE A-41

VOLUNTEER ARMY AMMUNITIONS PLANT MUNITIONS RESIDUES
JUNE, 1975

STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT. ppb	A 6-9-75 B-161 40 81 2	6-10-75 B-58 O O	6-11-75 B-205 130 144 71	6-12-75 B-190 172 0 51	6-13-75 B-124 O O
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	B-1 6-9-75 B-27 19 19	6-10-75 B-57 56 0 <2	6-11-75 B-45 26 4	6-12-75 B-183 52 16 17	6-13-75 B-132 87 40 51
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb	B-2 6-9-75 B-25 0	6-10-75 B-54 O	6-11-75 B-46 80 87	6-12-75 B-191 137 0	6-13-75 B-128 114 45
TNT, ppb	<2	0	10	63	102
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,0 DNT, ppb TNT, ppb	C-1 6-9-75 B-157 O O	6-10-75 B-166 2 0	6-11-75 B-38 < 2 0	0-12-75 B-198 < 2 08 12	0-13-75 B-44 © 2 0
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	C-2 6-9-75 B-160 < 2 0	6-10-75 B-49 O O	6-11-75 B-42 31 64 2	6-12-75 B-203 51 96 0	0-13-75 B-43 52 0 2

TABLE A-41 (Continued)

STATION	D-1				
DATE	6-9-75	6-10-75	6-11-75	6-12-75	6-13-75
SAMPLE NO.	B-34	B-93	B-37	B-188	B-92
2,4 DNT, ppb	0	< 2	< 2	0	< 2
2,6 DNT, ppb	0	0	0	0	ō
TNT. ppb	0	0	<2	0	0
STATION	D-2				
DATE	6-9-75	6-10-75	6-11-75	6-12-75	6-13-75
SAMPLE NO.	B-30	B-90	B-41	B-184	B-85
2,4 DNT, ppb	16	40	19	32	17
2,6 DNT, ppb	0	<2	0	0	0
TNT, ppb	0	4	0	6	3
STATION DATE	E-1 6-9-75	6-10-75	6-11-75	6-12-75	6-13-75
SAMPLE NO.	B-32	B-94	B-211	B-195	B-31
2,4 DNT, ppb	< 2	0	0	0	<2
2,6 DNT, ppb	<2	0	0	0	0
INT, ppb	2	0	< 2	15	2
STATION	E-2			<del></del>	<del></del>
DATE	6-9-75	6-10-75	6-11-75	6-12-75	6-13-75
SAMPLE NO.	B-168	<b>B-</b> 96	B-214	B-187	B-197
2,4 DNT, ppb	0	0	0	0	0
2,6 DNT, ppb	0	0	0	0	0
INT, ppb	0	0	0	0	0
STATION	F-1				
DATE	6-9-75	6-10-75	6-11-75	6-12-75	6-13-75
SAMPLE NO.	B-35	B-95	B-206	B-202	B-193
2,4 DNT, ppb	0	0	0	0	Sample broker
2,6 DNT, ppb	0	0	0	0	in shipment
INT, ppb	0	0	0	0	•

TABLE A-41 (Continued)

STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	F-2 6-9-75 B-33 12 16 0	6-10-75 B-53 O O 8	6-11-75 B-208 10 0 6	6-12-75 B-196 O O 4	6-13-75 B-201 13 21
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	S-, 6-9-75 B-26 0 0 <-2	6-10-75 B-56 O O	6-11-75 B-39 O O	6-12-75 B-181 O O	6-13-75 B-129 0 0 <2
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	T-1 6-9-75 B-29 O O	6-10-75 B-51 O O	6-11-75 B-47 O O	6-12-75 B-182 O O	6-13-75 B-130 0 0
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,0 DNT, ppb TNT, ppb	T-2 6-9-75 B-159 0 0	6-10-75 B-52 O O	6-11-75 B-40 0 0	6-12-75 B-189 O O	6-13-75 B-126 O O
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	U-1 6-9-75 B-36 O O	6-10-75 B-55 O O	6-11-75 B-209 O O	6-12-75 B-185 O O	0-13-75 B-131 O O

TABLE A-41 (Continued)

STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	U-2 6-9-75 B-162 0 0	6-10-75 B-50 2 0	6-11-75 B-48 O O	6-12-75 B-186 O O	6-13-75 B-127 O <2
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,0 DNT, ppb TNT, ppb	X-1 6-9-75 B-165 0 0 <2	6-10-75 B-86 O O	6-11-75 B-210 O O	6-12-75 B-192 O O	6-13-75 B-209 O O
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	X-2 6-9-75 B-165 O O	6-10-75 B-86 0 0 <2	6-11-75 B-210 0 0	6-12-75 B-192 O O	6-13-75 B-209 O O
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	Y-1 6-9-75 B-164 <2 0	6-10-75 B-91 0 0	6-11-75 B-216 <2 0	6-12-75 B-199 O O	6-13-75 B-213 O O
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	Y-2 6-9-75 B-158 0 0	6-10-75 B-89 O O	6-11-75 3-215 0 <2	6-12-75 B-204 O O	6-13-75 B-194 O O

TABLE A-42

VOLUNTEER ARMY AMMUNITIONS PLANT MUNITIONS RESIDUES AUGUST, 1975

STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	A 8-11-75 BA-186 12 20 14	8-12-75 BA-153 71 100 24	8-13-75 BA-96 17 28 38	8-14-75 BA-47 58 31 70	8-15-75 BA-16 Sample broken in shipment
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,0 DNT, ppb TNT ppb	B-1 8-11-75 BA-182 30 10	8-12-75 BA-152 <2 < 2 < 2 < 2	8-13-75 Ba-89 <2 <2 <2	8-14-75 BA-41 < 2 < 2 < 2	8-15-75 BA-10 45 11 12
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	B-2 8-11-75 BA-187 51 65 8	8-12-75 BA-151 86 116 9	8-13-75 BA-88 45 60 15	8-14-75 BA-44 27 72 28	8-15-75 BA-11 62 88 8
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	C-1 8-11-75 BA-183 8 <2 0	8-12-75 BA-146 0 70 7	8-13-75 BA-129 44 71 0	8-14-75 BA-46 38 97	8-15-75 BA-18 42 < 2 0
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,0 DNT, ppb TNT. opb	C-2 8-11-75 BA-189 27 30 0	8-12-75 BA-145 < 2 28 < 2	8-13-75 BA-95 11 0	8-14-75 BA-37 < 2 38 < 2	8-15-75 BA-19 28 35 0

TABLE A-42 (Continued)

STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	D-1 8-11-75 BA-217 O O	8-12-75 BA-94 0 0	8-13-75 BA-195 O O	8-14-75 BA-38 0 0	8-15-75 BA-22 Sample broken in shipment
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	D-2 8-11-75 BA-220 15 0	8-12-75 BA-90 < 2 < 2 0	8-13-75 BA-198 < 2 < 2 0	8-14-75 BA-39 0 0 < 2	8-15-75 BA-14 0 0
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,0 DNT, ppb TNT, ppb	E-1 8-11-75 BA-224 < 2 < 2 0	8-12-75 BA-87 0 0	8-13-75 BA-196 ≺2 0 < 2	8-14-75 BA-48 0 0	8-15-75 BA-8 0 0
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	E-2 8-11-75 BA-225 0 0	8-12-75 BA-91 < 2 0	8-13-75 BA-197 O O	8-14-75 BA-45 0 0	8-15-75 BA-1 0 0
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	F-1 8-11-75 BA-226 0 10 <2	8-12-75 BA-147 0 0 < 2	8-13-75 BA-202 0 0	8-14-75 BA-43 O O	8-15-75 BA-2 <2 0

TABLE A-42 (Continued)

STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,0 DNT, ppb TNT. ppb	F-2 8-11-75 BA-221 <2 15 <2	8-12-75 BA-148 < 2 20 2	8-13-75 BA-203 Sample broken in shipment	8-14-75 BA-40 0 0	8-15-75 BA-5 < 2 0 < 2
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNR, ppb	S 8-11-75 BA-181 0 0	8-12-75 BA-157 O O	8-13-75 BA-201 0 0	8-14-75 BA-42 0 0	8-15-75 BA-20 0 0
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	T-1 8-11-75 BA-191 O	8-12-75 BA-149 0 0	8-13-75 BA-184 0 0	8-14-75 BA-68 0 0	8-15-75 BA-3 0 0
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,0 DNT, ppb TNT, ppb	T-2 8-11-75 BA-192 0 0	8-12-75 BA-155 <2 0 0	8-13-75 BA-223 0 0	8-14-75 BA-61 0 0	8-15-75 BA-9 0 0
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT. ppb	U-1 8-11-75 BA-190 <2 0	8-12-75 BA-150 <2 0	8-13-75 BA-188 0 0	8-14-75 BA-67 0 0	8-15-75 BA-4 0 0

TABLE A-42 (Continued)

STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	U-2 8-11-75 BA-185 0 0	8-12-75 BA-154 O O	8-13-75 BA-130 < 2 0	8-14-75 BA-62 O O	8-15-75 BA-6 0 0
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,0 DNT, ppb TNT, ppb	X-1 8-11-75 BA-222 0 0 <2	8-12-75 BA-86 0 0	8-13-75 BA-199 0 0	8-14-75 BA-64 O O	8-15-75 BA-24 0 0
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT ppb	X-2 8-11-75 BA-219 0 0	8-12-75 BA-92 0 0	8-13-75 BA-204 0 0	8-14-75 BA-63 0 0 <2	8-15-75 BA-23 0 0
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,6 DNT, ppb TNT, ppb	Y-1 8-11-75 BA-227 0 0	8-12-75 BA-93 Sample broken in shipment	8-13-75 BA-193 0 0	8-14-75 BA-66 O O	8-15-75 BA-21 0 0
STATION DATE SAMPLE NO. 2,4 DNT, ppb 2,0 DNT, ppb TNT, ppb	Y-2 8-11-75 BA-218 0 0	8-12-75 BA-85 O O	8-13-75 BA-200 0 0	8-14-75 BA-65 0 0	8-15-75 BA-15 0 0

#### APPENDIX A-6

# HISTORICAL WATER QUALITY CHICKAMAUGA LAKE

TABLE A-43

SUMMARY OF WATER OUALITY DATA
CHICKAMAUGA LAKE, EPA
JUNE - OCTOBER, 1973\*

Parameter	Mean	Maximum	Minimum
Water Temperature (°C)	21.6	26.5	18.9
Transparency (Secchi in inches)	38.0	42.0	33.0
Field Conductivity (µmhos/cm)	162	176	150
Dissolved Oxygen (mg/1)	6.9	8.0	5.8
pH (SU)	7.5	7.6	7.3
Total Alkalinity (mg CaCO <sub>3</sub> /1)	54.6	62.0	46.0
Ammonia (mg N/l)	0.08	0.11	0.05
Total Kjeldahl N (mg N/l)	0.34	0.7	0.2
Nitrite plus Nitrate (mg N/1)	0.37	0.43	0.32
Total Phosphorus (mg P/l)	0.03	0.05	0.21
Chlorophyll <u>a</u> (µg/l)	4.07	5.8	2.3
Water Depth (feet)	55.0	65.0	50.0

\*Mile 477.3, Mid-channel South of Black Can Buoy. At Jct. of River Channels at tip of Harrison Bluff, 240° from center of John A. Patten Island, 100 yards East of bluff with caves at water level.

TABLE A-44

SUMMARY OF WATER QUALITY DATA CHICKAMAUGA LAKE AT DAM WALL JUNE - OCTOBER, 1973, EPA \*

Parameter	Mean	Maximum	Minimum
Water Temperature (°C)	21.6	26.6	19.0
Transparency (Secchi in inches)	38.3	40.0	36.0
Field Conductivity (µmhos/cm)	157	120	120
Dissolved Oxygen (mg/l)	6.95	8.4	5.8
pH (SU)	7.46	7.6	7.3
Total Alkalinity (mg CaCo <sub>3</sub> /l)	59.4	61.0	57.0
Ammonia (mg N/l)	0.07	0.09	0.05
Total Kjeldahl N (mg N/l)	0.28	0.60	0.20
Nitrite plus Nitrate (mg N/l)	0.37	0.40	0.34
Total Phosphorus (mg P/1)	0.03	0.04	0.02
Chlorophyll <u>a</u> (µg/l)	3.0	3.7	2.3
Water Depth (feet)	50.0	54.0	47.0

<sup>\*130°</sup> from flags at lock of Chickamauga Dam, adjacent to middle of 3 "danger" buoys, east of sluice gates about 150 yards.

TABLE A-45 CHICKAMAUGA LAKE WATER QUALITY DATA TVA, (1974) TENNESSEE RIVER MILE 472.3

Date	Depth, Feet	Total Coliforms, MPN/100 ml	Temperature, °C	D.O., mg/1	80D, mg/l	Color, PCU	Turbidity, JTU	Alkalinity, mg/l	Hard. CaCO3, mg/l	рн, в	Specific Resistance, µmhos	Cl, mg/l
1960												
7/12	Surf. 10 20 30 40 50 55	Ξ	28.0 27.2 26.2 24.4 24.2 24.3	7.70 7.40 6.49 5.15 4.40 4.00	1.90	ഹ	6.0	48.4	75.2	8.2 7.9 7.6 7.5 7.5	4,800	13.4
8/2	Surf. 10 20	5.2	88.69.79	7.96 7.78 5.40	1.37	10	5.2	53.0	74.4	8.2	5,300	12.7
	50 50 54		25.3 24.9 24.9	3.57 3.57 3.50	1.39	10	9.4	53.9	78.3	44.6	2,000	12.9
8/23	Surf. 5 10 20	2.6	29.1 27.2 25.8 25.7	8.11 8.06 5.63	1.74	10	3.8	51.2	75.0	8.3 7.9 7.7	5,400	9.93
	30 50 55		n m m er	4.83 4.73 4.64	0.77	10	21	50.5	74.6	7.6	5,400	10.2
9/22	Surf. 5 10	011		6.63 6.63 6.31	1.33	01	6.3	47.2	75.0	7.6	4,800	18.9
	20 30 50 55		24.1 24.2 24.1 24.1	6.33 6.33 6.14 6.17	1.38	10	נו	47.0	72.2	7.6	4,900	18.8

TABLE A-45 (Continued)

1 Si02 S04 Suspended mg/1 mg/1 mg/1 mg/1 mg/1 mg/1 mg/1 s.56 18.2 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4								Nitrite+			Š	Solids	
Surf.       23.4       8.60       7.40       0.80       0.19       0.00       0.80       17.6         Surf.       22.6       5.09       6.90       0.80       0.10       0.00       3.56       18.2         40       22.8       6.10       7.10       0.85       0.28       0.18       3.88       18.2         Surf.       22.1       5.66       6.60       0.85       0.03       0.00       3.70       17.3         40       23.0       4.84       6.10       0.85       0.07       0.42       3.90       18.2         Surf.       22.3       5.54       8.90       0.90       0.24       0.08       5.92       15.5         30       22.0       4.91       9.40       1.00       0.33       0.22       6.64       15.5	Date 1960		Ca mg/1		Na mg/1	K mg/l	Fe, Total	Nitrogen mg/l	Si02 mg/l	S04 mg/1	Suspended mg/l	Dissolved mg/l	Total mg/l
Surf.       22.6       5.09       6.90       0.80       0.10       0.00       3.56       18.2         40       22.8       6.10       7.10       0.85       0.28       0.18       3.88       18.2         Surf.       22.1       5.66       6.60       0.85       0.03       0.00       3.70       17.3         40       23.0       4.84       6.10       0.85       0.07       0.42       3.90       18.2         Surf.       22.3       5.54       8.90       0.90       0.24       0.08       5.92       15.5         30       22.0       4.91       9.40       1.00       0.33       0.22       6.64       15.5	21/7	Surf.	23.4	8.60	7.40		0.19	0.00	08.0	17.6	21	83	104
4022.86.107.100.850.280.183.8818.2Surf.22.15.666.600.850.030.003.7017.34023.04.846.100.850.070.423.9018.2Surf.22.35.548.900.900.240.085.9215.53022.04.919.401.000.330.226.6415.5	8/5	Surf.	22.6	5.09	6.90		0.10	00.00	3.56	18.2	4	114	118
Surf. 22.1 5.66 6.60 0.85 0.03 0.00 3.70 17.3 40 23.0 4.84 6.10 0.85 0.07 0.42 3.90 18.2 Surf. 22.3 5.54 8.90 0.90 0.24 0.08 5.92 15.5 30 22.0 4.91 9.40 1.00 0.33 0.22 6.64 15.5		40	22.8	6.10	7.10	0.85	0.28	0.18	3.88	18.2	28	96	124
40 23.0 4.84 6.10 0.85 0.07 0.42 3.90 18.2 Surf. 22.3 5.54 8.90 0.90 0.24 0.08 5.92 15.5 30 22.0 4.91 9.40 1.00 0.33 0.22 6.64 15.5	8/23	Surf.	22.1	5.66	6.60	0.85	0.03	00.00	3.70	17.3	4	35	96
Surf. 22.3 5.54 8.90 0.90 0.24 0.08 5.92 15.5 30 22.0 4.91 9.40 1.00 0.33 0.22 6.64 15.5		40	23.0		6.10	0.85	0.07	0.42	3.90	18.2	9	112	118
22.0 4.91 9.40 1.00 0.33 0.22 6.64 15.5	9/22	Surf.	22.3		8.90	0.90	0.24	0.08	5.92	15.5	12	128	140
		30	22.0	4.91	9.40	1.00	0.33	0.22	6.64	15.5	38	109	147

TABLE A-45 (Continued)

101	101	-							Hand		Spacific	
Depth Coli Feet MPN/	Coli MPN/	Coliforms, MPN/100 ml	Temperature ° C	D.O., mg/l	86U, mg/l	Color, PCU	Turbidity, JTU	Alkalinity, mg/l	CaCO3, mg/l	¥.	Resistance, umhos	Cl mg/l
Surf.		160	m'r	1 .	1.84	10	6.3	50.3	74.4	7.9	4,500	21.6
10 20 30 40			222.22 222.44 223.44	7.24 6.67 6.66 6.72	0.99	10	9.8	49.2	73.5	7.7 7.8 7.7 7.7	4,700	19.3
50 Surf.		280	, ww.		1.64	15	6.7	46.4	70.2	7.6	4,800	19.0
20 30 42			12.8 12.8 12.8	8.71 9.31 8.65 8.60	1.50	15	9.8	47.0	70.2	7.5 7.5 7.5	4,700	20.1
Surf. 10 20		6,200	& & & & & &	10.28 10.30 10.20	2.94	15	Ξ	48.0	70.0	7.6	2,000	18.6
30 40 50				000	1.46	15	14	48.5	70.0	7.6	2,000	18.1
	1	000.07		11.35	1.34	15	9.4	46.0	64.0	7.3	2,600	13.6
200 400 800 800			00000	11.35 11.35 11.54	1.36	15	Ξ	45.0	65.9	7 1 1 1 2	2,600	13.6
Surf.		220			1.73	20	14	55.4	76.2	7.5	4,800	15.2
20 30 50 50			၀ ထ ထ <b>ထ</b> ထ ဂ က က က က	11.98 12.12 11.99 11.02	1.38	15	14	56.0	76.2	7.5	4,800	15.4
	1											

TABLE A-45 (Continued)

!							Nitrite+			S	Solids	
Date	Depth Feet	Ca mg/1	Mg mg/l	Na mg/l	K mg/l	Fe, Total mg/l	Nitrogen mg/l	Si0 <sub>2</sub> mg/1	504 mg/1	Suspended mg/l	Dissolved mg/l	Total mg/l
.1960 10/18	Surf.	22.1	5.47	9.50	1.00	0.16	0.48	6.12	15.5	21	29	88
	8	22.1	5.22	9.50	1.00	0.28	0.08	7.20	15.5	2	96	101
11/22	Surf.	18.9	6.54	10.0	0.95	0.32	0.10	6.12	18.2	27	70	97
	30	19.0	6.48	11.0	1.00	0.44	0.15	6.80	15.5	16	36	108
12/13	Surf.	21.1	4.91	9.80	1.00	0.39	0.10	7.04	19.1	4	16	95
	20	21.7	4.53	9.80	1.00	0.42	0.10	7.54	18.2	37	100	137
1961 1/19	Surf.	18.8	4.72	7.50	0.85	0.29	0.22	7.00	18.2	19	75	94
	30	19.2	4.09	6.90	0.85	0.27	0.32	6.64	16.4	14	75	68
12/2	Surf.	22.6	5.41	7.90	0.90	0.19	0.22	6.48	20.9	16	107	123
	30	22.7	5.35	7.50	0.85	0.25	0.28	6.28	20.9	33	93	126

TABLE A-45 (Continued)

Date	Depth te Feet	Total ith Coliforms, et MPM/100 ml		Temperature ° C	D.0.,	80D, mg/l	color, Pcu	Turbidity, JIU	Alkalinity, mg/l	Hard. CaCO3, mg/l	pH,	Specific Resistance, umhos	Cl mg/l
1961		Surf. 1,300 10		11.3	9.68 9.62	1.12	10	30	45.8	58.7	8.1	009,9	5.84
	20 30 40 50	0000			9.54 9.54 9.61 9.59	1.58	15	29	46.5	59.2	8.0 8.0 7.9	009*9	5.84
4/18	Su 1	11.f. 36		00000	10.17 10.21 10.07	1.44	15	21	40.8	58.1	8.0 8.0 7.9	7,200	7.42
189	0.4 4 0.0 0.0 0.0	2000			10.00 10.00 10.00	1.36	15	23	41.4	58.3	2.7 2.8 8.7 8.8	7,100	7.56
5/16	16 Surf 5 10 20	6.0		21.0 20.8 20.5 19.6	9.08 9.10 9.03 8.80	1.14	10	5.0	44.9	55.6	7.9	7,790	4.30
	W 4 W	0.0:0		13.6 17.8 17.4	8.31 8.57 8.27	1.53	15	17	46.0	54.1	7.7	7,300	4.50
6/14	V	rf. 2.6		26.2 26.2 25.7	8.40 8.42 8.33	1.40	15	5.0	48.9	63.6	8 8 8 6 8 8 9 9	7,200	6.99
<del></del>	0.0 4 °C	2000		23.7 22.6 22.1 21.9	7.82 6.17 5.26 4.97	0.86	15	Ξ	48.8	63.6	8.0 7.9 7.8	006*9	7.09
Max Val	Maximum Values Minimum	70,00	c	29.1	12.14	2.94	20	30	56.0	78.3	8.8	7,700	21.6
٧a	Values	2.6	9	6.0	3.50	0.77	5	3.8	40.8	54.1	7.3	4,500	4.30

TABLE A-45 (Continued)

	:	,	:	1	;		Nitrite+ Nitrate	0.50	9		Solids	
Date	<b>Depth</b> Feet	ca mg/l	Mg mg/1	Na mg/l	K mg/l	Fe, Total mg/l	Nitrogen mg/l	5102 mg/l	304 mg/1	Suspended mg/l	Dissolved mg/l	Total mg/l
1961 3/21	Surf.	17.6	2.86	3.50	1.05	1.88	0.28	7.90	30.4	11	103	114
	30	17.1	3.32	3.00	0.95	1.58	0.22	8.40	30.4	11	109	120
4/18	Surf.	14.9	4.31	4.30	0.85	0.64	0.18	5.92	13.6	9	80	86
	40	15.2	4.43	4.50	0.80	0.77	0.18	6.02	13.6	4	72	9/
91/9	Surf.	16.9	4.03	4.64	0.95	0.08	00.00	5.20	11.8	56	99	36
	40	17.8	2.96	4.20	0.80	0.12	0.10	6.44	27.9	9	85	16
6/14	Surf.	19.4	5.85	4.55	0.95	0.01	0.10	4.74	12.8	59	06	119
	30	18.2	5.41	4.35	0.90	0.15	0.15	5.40	12.8	16	11	87
Maximum Values	₽.,	23.4	8.60	11.0	1.00	1.88	0.48	8.40	30.4	38	128	147
Minimum Values	E.	14.9	2.86	3.00	3.00 0.80	0.01	00.00	0.80	11.8	4	99	76

OBSERVED TRACE METAL CONCENTRATIONS (TOTAL) IN VICINITY OF WATER INTAKE SEQUOYAH NUCLEAR PLANT TABLE A-46

Tennessee River Mile 484.1

All results expressed as micrograms per liter (ug/l)

Sample	Depth ft.	Fe	£	7,7	Zn	Cr*	ž	LA L	Ag	P	Нд	Ba	As	Р	Se	Be
5/3/71	3	310		01^	82	<sup>^</sup> 50	\$50									
8/2/71	က	300		<10	70	<50	<50									
11/8/11	ო	370	0.06	<10	30	<50	< <del>5</del> 0									
2/1/72	٣	510		<10	20	12	<sup>2</sup> 50									
5/2/72	ю	280	50.0	10	20	<10	<50 <									
8/1/72	ю	310	60.0	01>	50	^5 ^5	<50	<del></del>								
11/8/72	က	290	80.0	~10	50	^5 -	<50									
2/28/73	39	690 710	60.0 70.0 <10	0,0	150	\$ \$ \$	<50 <50	700	°10 °10	°10 °10	0.2	°100 °100	<5 <5	~~		<10 <10 <10
5/21/73	39	450 620	60.0	20	808	နို လိ	<50 <50	1000	<10 <10	<10 <0.2 <10 <0.2	<0.2 :0.2	\$100 \$100	<5 <5	~~	~ √	مام مام

\*Precision of analysis was improved during the sampling period

TABLE A-47

SEDIMENT CHEMICAL CHARACTERISTICS HARRISON BAY, LAKE CHICKAMAUGA AT VAAP, JUNE 1975

NO <sub>3</sub> -N (mg/Kg Dry Wt	41 54 92	- 46 42 47	26 10 21 11	37 190 100 31	43 110 50 120
NO <sub>2</sub> -N (mg/Kg Dry Wt)	1.5 4.1	2.3 2.3 2.3		2 ~ ~ ~ ~	2.0 1.4 1.3
Total P (gm Kg Dry Wt)	2.40 0.91 1.80	0.57 0.45 0.47 1.10	0.44 0.18 0.59 0.26	0.26 0.34 0.46 0.41	0.13 0.12 0.22 0.24
COD (gm/Kg Dry Wt)	8.3 12.7 8.6	9.9 14.6 5.7 11.2	9.6 2.5 11.1	4.5 6.2 1.4 3.3	6.4 3.6 2.3
TKN (gm/Kg Dry Wt)	1.4	1.2 0.4 1.8	0.7	0.3 0.8 0.6	0.3 0.7 0.4 0.5
Solids % 11   Volatile	6.00 8.00 8.00 8.00	4.2 6.9 1.4	6.7 1.1 5.0 3.9	3.9 4.6 7.7	1.3 3.2 5.7 5.3
Soli Total	41.9 59.6 50.7	62.7 54.4 74.0 49.1	63.4 77.4 51.9 59.9	68.1 62.4 65.0 47.3 53.8	78.9 67.7 63.3 63.6
Station	A B-1 B-2	C-1 0-1 0-2	F-12-1	S T-1 U-1 U-2	x-1 x-2 Y-1 Y-2

TABLE A-48

SEDIMENT CHEMICAL CHARACTERISTICS HARRISON BAY, LAKE CHICKAMAUGA AT VAAP, AUGUST 1975

	Soli	, ; Q	TKN (am/Ka	COD (am/Ka 1	Total P	NO2-N	NO <sub>3</sub> -N
Station	Total Vola	Volatile	Dry Wt)	Dry Wt)	Dry Wt)		
A 1_g	23.5	16.5	•	15.0	0.51	•	66 45
8-2		5.5	0.4	1.2		0.83	120
C-1 C-2	54.0	5.7		4.4	0.41	2.0	65 26
0-1	62.0 41.5		0.5 2.1	3.5 82.	0.34	1.2	16 180
-2-	64.0 75.0 33.0	3.23 4.23	0.5 8.2 8.2	1.8	6.00 6.4	1.2 2.5 1.7	112
2-4		•	0.0	0.4	•	•	/
S T-1				2.5			39
1-2 U-1	59.0 63.0 62.0	ა. ი. ა. ი. ი.	0.7	2.0 2.0 3.0	0.42 0.34 0.41	3.9 0.51	8 23 57
c		•	•	•	٦, ٦	0.78	26 6
X-1 X-1 X-2	66.0 72.0	. w.c.	0.7	0.6	0.18	0.66	16 22

TABLE A-49

SEDIMENT METAL CONCENTRATIONS HARRISON BAY, LAKE CHICKAMAUGA, JUNE 1975

				Metal	Concentr	ation			
Station	P	Cr <sup>+6</sup>	Cu (mg/Kg	Hg Dr 3	Ni Pb	Pb	Zn	Fe (gm/Kg	Mn Dry Wt)
A B-1 B-2	3.5 2.1 3.1	140 58 72	22 14 32	0.67 0.54 0.38	36 16 20	200 310 370	360 130 270	60 29 43	0.50 0.71 0.46
C-1 C-2 D-1	2.2 1.6 2.4	23 61 79	25 5 21	1.90	18 34 7 25	75 240 57 310	70 140 43 340	25 30 38	0.7 <b>6</b> 1.60 0.52 0.70
E-2 F-1	9.1 9.6 9.6	- 49 41 80	20 1 15	0.17	24 3 25 4	110 9 120 16	280 44 420 28	40 23 39.5 21	0.73 1.60 1.40 0.66
S T-1 T-2 U-1 U-2	1.9 1.9 2.5 2.5	99	e5E€^		13 10 15 30 22	79 71 130 140	89 123 140 280 16	22 47 47 35 35	2.60 1.20 10.00 4.90 4.30
x-1 x-2 y-1 Y-2	1.5	42	2 8 9	0.10	4 2 17 19	5 12 25 13	4 13 42 50	14 22 40 37	0.20 0.03 0.31 0.20

TABLE A-50

SEDIMENT METAL CONCENTRATIONS HARRISON BAY, LAKE CHICKAMAUGA, AUGUST 1, 1975

				Meta	Metal Concentration	ation			
Station	<b>P</b> 3	Cr	n) n)	(mg/Kg Dry Wt)	Ë	Pb	Zn	Fe (gm/Kg	Mn Dry Wt)
A 8-1	~⊽	740 53	130 15	0.77	110 14	860 360	690 140	140 33	6.1 0.9
8-2	i	ı	•	<0.10	ı	ı	•	1	•
C-3	,	31	_	1	10	120	120	28	0.59
0-1	~	98	,`	0.17	4	79	73	32	0.74
E-1	⊽	34	13	•	9[	78	150	34	0.68
-1	~	55	33	ı	19	170	540	88	2.00
S	\[\frac{1}{\sqrt{1}}\]	19	6	0.17	9	82	9/	34	2.00
	⊽	21	12	0.1	<sub>∞</sub>	96	140	28	1.30
[-n	/ <del>-</del>	140	<u>8</u>	1.3	15	140	110	48	2.80
x-1	⊽	30	<b>,</b>	0.10	က	19	15	17	1.90
۲-۱	٢>	38	6	1	∞	8	36	37	2.20

#### APPENDIX B

VAAP Periphyton Collections from Lake Chickamauga, Tennessee

June - July and August - September 1975

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Mo. of Fields Counted	* *	Total No. of Species	"New" Species Recorded After 30 Fields	Total No. of Individuals	Density of "New" Species Recorded After 30 Fields	Density of "New" Species Recorded After 30 Fields Total No. of Individuals	100
	0.689	19	ပ	662	0	0	
45	0.667	23	4	825	4	0.48%	
09	0.615	52	()	1004	თ	0.90%	
75	0.637	27	80	1165	13	1.12%	
06	0.644	31	12	1345	17	1.26%	
120	0.597	34	15	1680	21	1.25%	
150	0.626	34	15	1983	21	1.06%	
						٠	

\*Shannon-Weaver Species Diversity Index.

# TABLE B-2 VAAP PERIPHYTON ARTIFICIAL SUBSTRATE, JUNE 11-25, 1975

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TABLE B-3
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TABLE B-4
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TABLE B-5
VAAP PERIPHYTON ARTIFICIAL SUBSTRATE, AUGUST 12 - SEPTEMBER 7, 1975

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TABLE B-5 (CONTINUED).

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TABLE B-6

PRESENCE-ABSENCE DATA FOR FILAMENTOUS ORGANISMS COLLECTED FROM VAAP ARTIFICIAL SUBSTRATES, JUNE-JULY 4-WEEK INCUBATION PERIOD, 1975, LAKE CHICKAMAUGA, TENNESSEE

Species	Α	B-1	B-2	C-1	F-2	S	T-1	X-2
Cyanophyceae  Microcoleus lynbyaccous Oscillatoria lutea Oscillatoria submembranceae Schizothrix arenaria Schizothrix calcicola Spirulina subsala		X X	X	x	X	X	X	X X
Chlorophyceae  Chaetopeltis sp. Choleochaete spp. Cylindrocapsa geminella Mougeotia spp. Oedogonium sp. l Ulothrix spp.	X	X X X	X	х	X X	X	X X X X	X
Ciliophora (stalked protozoans) Opisthostyla sp. Vorticella sp.		Х		Х			X	X
Total No. Autotrophic Species Total No. Heterotrophic Species Total No. of Filamentous Species	1 0 1	5 0 5	4 1 5	3 1 4	5 0 5	3 0 3	7 2 9	5 1 6

TABLE B-7
PRESENCE-ABSENCE DATA FOR FILAMENTOUS ORGANISMS

COLLECTED FROM SELECTED VAAP ARTIFICIAL SUBSTRATES AUGUST-SEPTEMBER 4-WEEK INCUBATION PERIOD 1975, LAKE CHICKAMAUGA, TENNESSEE

Species	Α	B-1	C-1	F-2	S	T-1	X-2
Cyanophyceae							
Anabaena sp. Microcoleus lynbyaceous Oscillatoria lutea Oscillatoria submembranceae Schizothrix arenaria Schizothrix calcicola Spirulina subsala	X	X X	x x	X X	X	x x	X X
Aphanochaete polychaete Bulbochaete sp. Choleochaete spp. Cylindrocapsa geminella Mougeotia spp. Oedogonium sp. l Oedogonium lautummiarum Stigeoclonium attenuatum Stigeoclonium tenue Ulothrix spp.	X X X	x x x x	X X	x x x x	X X X	X X	X
Ciliophora (stalked protozoans)  Opisthostyla sp. Rhabdostyla sp. Thuricolopsis sp. Thuricola sp. Vorticella sp.	X X X	X X	х		X X	X	X
Total No. of Autotrophic Species Total No. of Heterotrophic	6	7	5	7	4	5	4
Species Total No. of Filamentous Species	3 9	9 10	1 6	7	2 6	1 6	5

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TABLE B-8	DIATOMS
	SUBSTRATE DIATOMS (RAW O
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TABLI VAAP NATURAL SUBSTRATE DIA	TABLE B-8 DIATOMS (	(RAW COUNTS), JUNE,	UNTS)	JUN.	E, 1	1975
Bacillariophyta (Diatoms)	1-2	X-2	∢	<b>B-1</b>	S	F-2
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CYMBELLA DELICATULA Cymbella Ventricosa Cymbella prostata	เพต	191	• • •	1 10 1	111	~ © M
CYPTELLA MICACCFPNALA Cyptella lafvis Cypgella tumida	<b>⊢</b> 1 =	<b>‡</b> ''	œI I	NIP	• • •	-01
CYMSELLA SP I DIPLUNEIS SMITHII DIPLONEIS SP A	<b>-11</b>		111	111	1-1	11-
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FRAGILARIA PINNATA FRAGILARIA VAUCHERIAE GEPFHUNEMA ACUMINATUM V MONTAMUM	N <b>9</b> →	N I =		111	110	កាលព
GOWPHONE MA GRACILE V LANCEGLATA GCHHGME MA MONTANUN V SUBCLAVATA GCMPHONEMA PAPVULUM		11-	111	<b>0</b> 1 →	<b>=</b> N =	l ⇔Pi
GCEFFCREVA SP 1 GCEFFCREVA SP 2 RELUSIKA DISTANS	100	-11	111	nn i	<b>NN</b> I	• • •
MELOSIRA GPAPULATA N Melcsira vapians Navicula aupiculata		0-1	1	111	<b>411</b>	611
NAVICULA CPVPTOCFPHALA V INTERMEDIA Navicula Cpvftncfphala v veneta Navicula lancenlata	N 0 =	Ini	-11	101	101	N N I
NAVICULA MENISCILUS V UPSALIS NAVICULA MINIMA NAVICULA MUTICA	1-1	111	1 J -	111	111	-11
NAVICULA PUPULA NAVICULA PUPULA V RECTANGULARIS NAVICULA SALINARIUM V INTERMEDIA.	N I N		111	111	+ 1 1	~ N ~
NAVIGULA VIRTOULA NAVIGULA SP 1 NAVIGULA SP 2	-11		117	110	11-	1-1
NAVIGULA SP 5 NAVIGULA SP 6 NITZSCHIA AMPHIBIA	110	11-	117	1-1	117	-14
NITZSCHIA DFNTTCULA NITZSCHIA KUTZINGIANA NITZSCHIA PALFA	21-	<b>5</b> 11	111	111	1 == 1	0 1 1
NITZSCMIA SP. 1 MITZSCMIA SINUATA V TABELLARIA MITZSCMIA TRYBI IONELLA V VICTORIAE	-41	121	11-	111	111	181

TABLE 8-8 (Continued)

Bacillariophyta (Diatoms)	T-2	T-2 X-2 A B-1 S	4	<b>8-1</b>	S	F-2
STEPHANDDISCUS ASTRAEA STEPPANDDISCUS SP 1 SURINELLA OVATA		-11	i	111	-11	mn I
SYNEDRA PULCHELLA Synedra delicatissima Synegra Humpens V Familaris	118	11-	-1-	110	1 - 1	
SYNEDRA ULNA TABELLARIA FERETRATA	<b>~</b> ~		• •	11	4 1	• 1
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TOTAL NUMBER OF ORGANISMS		445. 569	51.	51. 202. 287. 1536.	287	1536
MUNBER OF TAXA	36	22	16	90	0	<b>#</b>

TABLE B-9

RELATIVE ABUNDANCE OF DOMINANT TO COMMON VAAP ARTIFICIAL SUBSTRATE DIATOM SPECIES - JUNE 11-25, 1975

Station	Species	Density (Cells/mm <sup>2</sup> )	Percent Relative Abundance
А	Achnanthes minutissima Nitzschia palea Nitzschia cf. capitellata Synedra rumpens Gomphonema intricatum v. pumila Gomphonema parvulum Other species Total	420 105 60 30 22 15 56 708	59.3 14.8 8.5 4.2 3.1 2.1 7.9 99.9*
B-1	Achnanthes minutissima Fragilaria capucina Synedra rumpens Synedra ulna Other species Total	24,815 940 396 297 1,454 27,902	88.9 3.4 1.4 1.1 <u>5.2</u> 100
B-2	Achnanthes minutissima Fragilaria capucina Synedra rumpens Synedra ulna Other species Total	15,277 970 312 304 1,668 18,531	82.4 5.2 1.7 1.6 9.0 99.9*
C-1	Achnanthes minutissima Gomphonema intricatum v. pumila Fragilaria capucina Other species Total	5,687 93 54 282 6,116	93.0 1.5 0.8 4.6 99.9*
C-2	Achnanthes minutissima Fragilaria capucina Synedra rumpens Gomphonema intricatum v. pumila Cocconei Jacentula v. euglypta Other species Total	10,410 863 342 205 152 718	82.0 6.8 2.7 1.6 1.2 5.7
D-1	Achnanthes minutissima Fragilaria capucina Other species Total	13,218 75 247 13,540	97.6 0.5 1.8 99.9*

TABLE B-9 (CONTINUED)

Station	Species	Density <sub>2</sub> (Cells/mm <sup>2</sup> )	Percent Relative Abundance
D-2	Achnanthes minutissima Achnanthes nollii Synedra rumpens Other species Total	7,760 195 90 532 8,577	90.5 2.3 1.0 6.2 100
F-1	Achnanthes minutissima Gomphonema parvulum Synedra ulna Cymbella affinis Synedra rumpens Other species Total	1,083 79 32 30 25 90 1,339	80.9 5.9 2.4 2.2 1.8 6.7 99.9*
F-2	Achnanthes minutissima Navicula cryptocephala v. veneta Synedra ulna Cymbella prostata Fragilaria vaucheriae Gomphonema parvulum Cymbella ventricosa** Other species Total	3,287  358 274 212 175 166 105 1,160 5,737	57.3 6.2 4.8 3.7 3.0 2.9 1.8 20.2 99.9*
S	Achnauthes minutissima Gomphonema parvulum Fragilaria capucina Other species Total	10,995 314 118 271 11,698	94.0 2.7 1.0 2.3 100
T-1	Achnanthes minutissima Gomphonema parvulum Fragilaria capucina Gomphonema gracile v. lanceolata Other species	10,907 274 244 106 476	90.8 2.3 2.0 0.9 4.0
T-2	Total  Acnnanthes minutissima Nitzschia denticula Fragilaria capucina Other species Total	12,007 18,891 827 144 <u>802</u> 20,664	91.4 4.0 0.7 3.9 100

TABLE B-9 (CONTINUED)

Station	Species	Density <sub>2</sub> (Cells/mm <sup>2</sup> )	Percent Relative Abundance
U-1	Achnanthes minutissima Fragilaria capucina Other species Total	16,956 220 916 18,092	93.7 1.2 5.1 100
X-1	Achnanthes minutissima Fragilaria capucina Cymbella microcephala Anomoeneis vitrea Other species Total	16,746 668 332 236 1,385 19,367	86.5 3.4 1.7 1.2 <u>7.2</u> 100
X-2	Achnanthes minutissima Fragilaria capucina Synedra delicatissima Anomoeneis vitrea Cymbella microcephala Other species Total	5,788 173 167 151 70 430 6,779	85.4 2.6 2.5 2.2 1.0 <u>6.3</u>
Y-1	Achnanthes minutissima Cymbella affinis Cymbella microcephala Navicula cryptocephala v. veneta Nitzschia denticula Gyrosigma attenuatum Other species Total	5,170 228 221 174 136 106 793 6,828	75.7 3.3 3.2 2.5 2.0 1.6 11.6 99.9

<sup>\*99.9%</sup> relative abundance results from rounding off.

<sup>\*\*</sup>Cymbella ventricosa = Cymbella minuta.

TABLE B-10

RELATIVE ABUNDANCE OF DOMINANT TO COMMON VAAP ARTIFICIAL SUBSTRATE DIATOM SPECIES - JUNE 11 - JULY 10, 1975

Station	Species	Density (Cells/mm <sup>2</sup> )	Percent (%) Relative Abundance
А	Achnanthes minutissima Nitzschia palea Synedra rumpens Other species	5,614 273 97 344	88.7 4.3 1.5 5.4
	Total	6,328	99.9*
B-1	Achnanthes minutissima Melosira ambigua Nitzschia denticula Synedra rumpens Other species	73,102 1,651 1,055 733 <u>7,617</u>	86.7 2.0 1.2 0.9 9.1
	Total	84,158	99,9*
B-2	Achnanthes minutissima Fragilaria capucina Melosira ambigua Other species	84,345 2,202 1,284 4,199	91.6 2.4 1.4 4.6
	Total	92,030	100.0
C-1	Achnarthes minutissima Cocconeis placentula v. euglypta Gomphonema intricatum v. pumila Other species	54,039 618 585 2,737	93.2 1.1 1.0 4.7
	Total	57,979	100.0
D-1	Achnanthes minutissima Achnanthes nollii Cymbella affinis Synedra ulna v. danica Other species	60,772 1,551 700 650 4,400	89.3 2.3 1.0 0.9 6.5
	To tal	68,073	100.0
D-2	Achnanthes minutissima Achnanthes nollii Cymbella laevis Cymbella microcephala Cymbella prostata Other species Total	32,766 791 446 378 377 2,424 37,182	88.1 2.1 1.2 1.0 1.0 <u>6.5</u> 99.9*

Table B-10 (Continued)

Station	Species	Density (Cells/mm <sup>2</sup> )	Percent (%) Relative Abundance
F-2	Achnanthes minutissima Navicula cryptocephala v. veneta Achnanthes nollii Cymbella affinis Other species	42,404 1,617 1,616 550 4,857	83.1 3.2 3.2 1.0 9.5
	Total	51,044	100.0
S	Achnanthes minutissima Gomphonema parvulum Other species	45,709 1,032 2,529	92.8 2.1 5.1
	Total	49,270	100.0
T-1	Achnanthes minutissima Melosira ambigua Navicula cf. minima Navicula cryptocephala v. veneta Other species	28,499 585 413 343 2,320	88.6 1.8 1.3 1.1 
	Total	32,160	100.0
X-2	Achnanthes minutissima Cymbella microcephala Anomoeoneis vitrea Synedra delicatissima Other species	52,317 2,271 1,410 378 3,249	87.7 3.8 2.4 0.6 5.4
	Total	59,625	99.9*
Y-2	Achnanthes minutissima Cymbella affinis Achnanthes nollii Navicula cryptocephala v. veneta Navicula minima Other species	16,080 917 710 665 607 2,032	76.5 4.3 3.4 3.2 2.9 9.7
	Total	21,011	100.0

 $<sup>\</sup>star 99.9\%$  total relative abundance results from rounding off.

TABLE B-11

RELATIVE ABUNDANCE OF DOMINANT TO COMMON VAAP ARTIFICIAL SUBSTRATE DIATOM SPECIES - AUGUST 12-26, 1975

Station	Species	Density (Cells/mm <sup>2</sup> )	Percent (%) Relative Abundance
A	Achnanthes minutissima Nitzschia kutzingiana Nitzschia palea Navicula cf. heufleri v.	434 102 102	47.9 11.3 11.3
	leptocephala Melosira granulata Melosira ambigua	68 68 22	7.5 7.5 2.4
	Melosira distans Synedra delicatissima	22 22	2.4
	Other species	66	7.3
	Total	906	100.0
B-1	Achnanthes minutissima Navicula cf. heufleri v.	14,084	86.1
	leptocephala Gomphonema parvulum	228 228	1.4
	Melosira ambigua Other species	159 <u>1,661</u>	1.0 10.1
	Total	16,360	100.0
B-2	Achnanthes minutissima Cymbella affinis Fragilaria capucina Melosira ambigua Nitzschia kutzingiama Synedra delicatissima Other species	9,140 470 332 274 194 182 964	79.1 4.1 2.9 2.4 1.7 1.6 
	Total	11,556	100.1*
C-1	Achnanthes minutissima Fragilaria capucina Synedra ulna Other species	23,031 366 251 2,386	88.4 1.4 1.0 9.2
		26,034	100.0

TABLE B-11 (Continued)

Station	Species	Density (Cells/mm <sup>2</sup> )	Percent (%) Relative Abundance
C-2	Achnanthes minutissima Synedra delicatissima Fragilaria capucina Synedra rumpens Other species	13,438 397 198 198 <u>1,405</u>	85.9 2.5 1.3 1.3 9.0
	Total	15,636	100.0
D-1	Achnanthes minutissima Achnanthes nollii Melosira ambigua Other species	18,970 137 136 1,004	93.7 0.7 0.7 4.9
		20,247	100.0
D-2	Achnanthes minutissima Achnanthes nollii Fragilaria capucina Synedra delicatissima Synedra ulna Fragilaria crotonensis Other species	12,766 1,115 1,069 840 687 228 1,619	69.7 6.1 5.8 4.6 3.7 1.2 8.8
	Total	18,324	99.9*
E-1	Achnanthes minutissima Amphipleura pellucida Synedra delicatissima Achnanthes nollii Synedra ulna v. danica Fragilaria capucina Other species	18,997 825 687 550 503 458 4,181	72.5 3.1 2.6 2.1 1.9 1.7 16.0
	Total	26,201	99.9*
E-2	Achnanthes minutissima Fragilaria capucina Synedra delicatissima Achnanthes nollii Fragilaria construens Other species	27,441 1,514 1,330 504 413 2,678	81.0 4.5 3.9 1.5 1.2 7.9
	Total	33,880	100.0

TABLE B-11 (Continued)

Station	Species	Density (Cells/mm <sup>2</sup> )	Percent (%) Relative Abundance
"No Wake"	Achnanthes minutissima Synedra delicatissima Fragilaria capucina Synedra ulna v. danica Other species	33,858 1,009 641 504 1,191	91.0 2.7 1.7 1.3 3.2
	Total	37,203	99.9*
"F-Buoy"	Achnanthes minutissima Synedra delicatissima Synedra ulna v. danica Achnanthes nollii Fragilaria capucina Other species	6,766 1,719 825 389 206 510	65.0 16.5 7.9 3.7 2.0 4.9
	Total	10,415	100.0
S	Achnanthes minutissima Fragilaria capucina Synedra delicatissima Cymbella microcephala Other species	15,304 488 320 167 707	90.1 2.9 1.9 0.9 4.2
	Total	16,986	100.0
F-1	Achnanthes minutissima Melosira ambigua Other species	17,525 182 <u>784</u>	94.8 1.0 <u>4.2</u>
	Total	18,491	100.0
T-2	Achnanthes minutissima Fragilaria capucina Synedra delicatissima Melosira ambigua Synedra ulna v. danica Nitzschia kutzingiana Other species	15,346 1,697 687 618 251 228 1,887	74.1 8.2 3.3 3.0 1.2 1.1 9.1
		20,714	100.0

TABLE B-11 (Continued)

Station	Species	Density (Cells/mm <sup>2</sup> )	Percent (%) Relative Abundance
U-1	Achnanthes minutissima Synedra delicatissima Fragilaria capucina Synedra ulna Cymbella affinis Navicula cryptocephala v.	3,187 642 401 377 148	55.2 11.1 6.9 6.5 2.6
	venta Other species	137 	2.3 15.3
	Total	5,775	99. <b>9*</b>
U-2	Achnanthes minutissima Fragilaria capucina Synedra delicatissima Synedra ulna v. danica Other species	22,871 1,398 458 389 1,666	85.4 5.2 1.7 1.4 <u>6.2</u>
	Total	26,782	99.9*
X-1	Achnanthes minutissima Cymbella microcephala Fragilaria construens Fragilaria capucina Synedra delicatissima Cymbella affinis Anomoeoneis vitrea Other species	12,295 916 435 412 411 320 320 2,505	69.8 5.2 2.5 2.3 2.3 1.8 1.8
	Total	17,614	99.9*
X-2	Achnanthes minutissima Fragilaria capucina Fragilaria construens Other species	12,318 274 274 1,125	88.0 2.0 2.0 8.0
	Total	13,991	100.0
Y-2	Achnanthes minutissima Fragilaria capucina Synedra delicatissima Synedra ulna v. danica Cymbella microcephala Other species	18,374 664 412 251 228 1,122	87.3 3.2 1.9 1.2 1.1 5.3
	Total	21,051	100.0

<sup>\*</sup>Total relative abundances slightly below or above 100% are due to rounding off.

TABLE B-12

RELATIVE ABUNDANCE OF DOMINANT TO COMMON VAAP ARTIFICIAL SUBSTRATE DIATOM SPECIES - AUGUST 12 - SEPTEMBER 7, 1975

Station	Species	(Cells/mm <sup>2</sup> )	Percent (%) Relative Abundance
А	Achnanthes minutissima Cyclotella stelligera Fragilaria capucina Nitzchia kutzingiana Achnanthes nollii Other species	6,421 305 152 106 91 663	83.0 3.9 2.0 1.4 1.2 8.5
	Total	7,738	100.0
B-1	Achnanthes minutissima Cyclotella stelligera Other species	40,809 641 1,215	95.6 1.5 <u>2.8</u>
	Total	42,665	99.9*
C-1	Achnanthes minutissima	36,634 537	98.5 1.5
	Total	37.171	100.0
C-2	Achnanthes minutissima Fragilaria capucina Synedra delicatissima Cymbella laevis Other species	22,916 596 320 274 1,128	90.8 2.4 1.3 1.1 4.5
	Total	25,234	100.1*
D-2	Achnanthes minutissima Achnanthes nollii Cymbella affinis Other species	13,557 298 205 1,784	85.6 1.9 1.3 11.2
		15,844	100.0
NW	Achnanthes minutissima Fragilaria capucina Achnanthes nollii Nitzchia kutzingiana Other species	17,754 229 205 205 1,150	90.8 1.2 1.0 1.0 5.9

TABLE B-12 (continued)

Station	Species	(Cells/mm <sup>2</sup> )	Percent (%) Relative Abundance
E-2	Achnanthes minutissima Cymbella microcephala Fragilaria capucina Other species	21,838 687 228 1,328	90.7 2.9 0.9 5.5
	Total	24,081	100.0
S	Achnanthes minutissima Synedra delicatissima Fragilaria capucina Anomoeoneis vitra Cymbella microcephala Other species	21,242 871 664 573 366 <u>758</u>	86.8 3.6 2.7 2.3 1.5 3.1
	Total	24,474	100.0
T-1	Achnanthes minutissima Gomphonema intricatum v. pumila Achnanthes sp. A Other species	19,498 619 159 740	92.8 2.9 0.8 3.5
	Total	21,016	100.0
T-2	Achnanthes minutissima Synedra delicatissima Fragilaria capucina Anomoeoneis vitrea Navicula cryptocephala v. venta	17,250 550 527 435	85.7 2.7 2.6 2.1
	Other species	1,168	5.8
	Total	20,135	99.9*
U-1	Achnanthes minubissima Synedra delicatissima Fragilaria capucina Synedra ulna Other species	20,117 504 412 366 1,241	88.9 2.2 1.8 1.6 5.5
	Total	22,640	100.0

TABLE B-12 (Continued)

Station	Species	(Cells/mm <sup>2</sup> )	Percent (%) Relative Abundance
U-2	Achnanthes minutissima Cymbella microcephala Cymbella affinis Achnanthes nollii Fragilaria capucina Other species	37,460 549 504 458 366 1,757	91.1 1.3 1.2 1.1 0.9 4.3
	Total	41,094	99.9*
X-1	Achnanthes minutissima Anomoeneis vitrea Cymbella microcepha Fragilaria capucina Synedra delicatissima Cymbella affinis Other species	24,201 1,352 917 587 503 320 2,147	80.6 4.5 3.0 1.9 1.7 1.1
	Total	30,027	100.0
X-2	Achnanthes minutissima Cymbella microcephala Fragilaria construens Fragilaria pinnata Fragilaria construens v. venter	15,162 1,008 824 825 412	72.5 4.8 3.9 3.9
	Fragilaria capucina Achnanthes sp. A Other species	297 274 2,103	1.4 1.4 10.1
	Total	20,905	100.0
Y-2	Achnanthes minutissima Cymbella microcephala Fragilaria pinnata Melosira ambigua Nitzschia kutzingiana Cymbella affinis Fragilaria construens Synedra delicatissima Other species	22,389 1,284 619 481 458 458 412 320 2,870	76.4 4.4 2.1 1.6 1.6 1.4 1.1 9.8
	Total	29,291	100.0

<sup>\*</sup>Total relative abundances slightly above or below 100% are due to the effect of rounding off.

TABLE B-13

MEAN DIATOM CELL DENSITIES (CELLS/MM<sup>2</sup>) FOR VAAP ARTIFICIAL SUBSTRATES INCUBATED FOR 2- AND 4-WEEK INTERVALS JUNE - JULY, 1975, LAKE CHICKAMAUGA, TENNESSEE

Station <b>s</b>	2-Week* Incubation Interval June 11-25, 1975	4-Week** Incubation Period June 11 - July 10, 1975
A B-1 B-2 C-1 Waconda Bay C-2 D-1 D-2 F-1 Offshore F-2	708 27,902 18,531 6,116 12,690 13,540 8,577 1,339@ 5,737	6,328 84,158 92,030 57,979 L*** 68,073 37,182 51,044
S T-1 Reference Bay A T-2 U-1 U-2 X-! X-2 Y-1 Y-2	11,698 12,007 20,664 18,092 L***	49,270 32,160 L*** L*** L***  L*** 21,011

<sup>\*</sup>Means developed from 3 replicates

<sup>\*\*</sup>Means developed from 2 replicates @Station found washed ashore

L\*\*\*Station lost or andalized

TABLE B-14

MEAN DIATOM CELL DENSITIES (CELLS/MM<sup>2</sup>) FOR VAAP ARTIFICIAL SUBSTRATES INCUBATED FOR 2- AND 4-WEEK INTERVALS AUGUST - SEPTEMBER, 1975, LAKE CHICKAMAUGA, TENNESSEE

Stations	2-Week Incubation Interval * August 12-26, 1975	4-Week Incubation Interval* August 12-Sept. 7, 1975
A B-1 B-2 C-1 Waconda Bay C-2 D-1 D-2 No Wake E-1 E-2 F-1 Offshore F-2 F-Buoy	906 16,360 11,556 26,034 15,636 20,247 18,324 37,203 26,201 33,880 L** L** 10,415	7,738 42,665 L** 37,171 25,234 L** 15,844  L** 24,081 L** 19,543
S	16,986	24,474
T-1 Reference Bay A	18,491	21,016
T-2	20,714	20,135
U-1	5,775	22,640
U-2	26,782	41,094
X-1	17,614	30,027
X-2 Huss Lowe Slough	13,991	20,905
Y-1	L**	L**
Y-2	21,051	29,291

<sup>\*</sup>Means developed from 2 replicates L\*\*Lost or vandalized station

TABLE B-15 SHANNON-WEAVER SPECIES DIVERSITY INDICES (A) FOR VAAP ARTIFICIAL SUBSTRATE DIATOMS, JUNE - JULY 2- AND 4-WEEK INCUBATION PERIODS, LAKE CHICKAMAUGA, TENNESSEE

Stations	2-Week* Incubation Period June 11-25, 1975	4-Week** Incubation Period June 11-July 10, 1975
A B-I B-2 Waconda C-1 Bay C-2 D-1 D-2 F-1 Offshore F-2	1.49 0.64 0.64 0.44 0.91 0.17 0.59 0.93 2.07	0.62 0.83 0.52 0.45 L*** 0.66 0.71 0.97
S	0.34	0.45
T-1 Reference	0.54	0.70
T-2 Bay A	0.49	L***
U-1	0.42	L***
X-1 Huss	0.78	L***
X-2 Lowe	0.81	0.67
Y-1 Slough	1.28	L***
Y-2	L***	1.20

<sup>\*3</sup> replicates pooled. \*\*2 replicates pooled.

L\*\*\*Lost or vanualized station.

TABLE B-16 SHANNON-WEAVER SPECIES DIVERSITY INDICES (H) FOR VAAP ARTIFICIAL SUBSTRATE DIATOMS, AUGUST - SEPTEMBER 2- AND 4-WEEK INCUBATION PERIODS, LAKE CHICKAMAUGA, TENNESSEE

Stations	2-Week* Incubation Period August 12-26, 1975	4-Week* Incubation Period August 12-September 9, 1975
A B-1 B-2 Waconda C-1 Bay C-2 D-1* D-2 No Wake E-1 E-2 F-1 Offshore F-2 F Buoy	1.83 0.85 1.10 0.74 0.82 0.43 1.39 0.51 1.49 1.00 L** L**	0.91 0.29 L** 0.11 0.53 L** 0.82 0.55 L** 0.55 L**
S Reference	0.56	0.69
T-1 Bay A	0.34	0.42
T-2	1.27	0.78
U-1	1.89	0.63
U-2	0.78	0.55
X-1	1.54	1.03
X-2	0.69	1.34
Y-1 Lowe	L**	L**
Y-2 Slough	0.71	1.26

<sup>\*2</sup> replicates pooled. L\*\*Lost or vandalized station.

TABLE B-17

SHANNON-WEAVER SPECIES DIVERSITY INDICES, SHANNON EVENNESS VALUES, AND TOTAL NUMBER OF SPECIES FOR NATURAL SUBSTRATE DIATOMS, JUNE, 1975, LAKE CHICKAMAUGA, TENNESSEE

Station	Shannon-Weaver Index (H)	Shannon Evenness Values (J)	Total Numbers of Species
A	1.85	0.669	16
B-1 Waconda	0.13	0.337	20
F-2 Bay	0.99	0.280	34
S Reference	0.64	0.222	18
T-2 Bay A	1.84	0.512	36
X-2 Huss Lowe Slough	1.59	0.510	22

TABLE B-18

## COMPARISONS OF DIATOM CELL DENSITY ESTIMATES (CELLS/MM<sup>2</sup>) FOR NATURAL AND ARTIFICIAL SUBSTRATES COLLECTED DURING JUNE, 1975, LAKE CHICKAMAUGA

Station	Natural Substrates (Cells/mm <sup>2</sup> ) June, 1975	Artificial Substrates (Cells/mm <sup>2</sup> ) June 2-Week Incubation	Artificial Substrates (Cells/mm <sup>2</sup> ) June-July 4-Week Incubation
A	920	708	6,328
B-1 Waconda	14,540	27,902	84,158
F-2 Bay	11,510	5,737	1,044
S Reference	6,260	11,698	49,270
T-2 Bay A	17,620	20,664	32,168*
X-2 Huss Lowe Slough	10,185	6,779	59,625

<sup>\*</sup>Represents cells/mm $^2$  for Station T-1; T-2 was lost during the 4-week incubation period.

TABLE B-19

TRANSECT MEAN PERIPHYTON BIOMASS, CHLOROPHYLL a AND AUTO-TROPHIC INDICES, JUNE - JULY SURVEY, 2-WEEK INCUBATIONS\*

Station or Transect		Chlorophyll <u>a</u> (mg/M <sup>2</sup> )	Organic Biomass (gm/M <sup>2</sup> )	Autotrophic Index
A B Wad C Bay D E	conda y	4* 24 35 17 22	0.67 3.5 1.9 1.4 -	167 146 54 82 -
F Off	fshore	-	1.5	-
S Ref T Bay U A		12 14 -	2.1 2.9 1.5	175 207 -
	ss Lowe ough	- 13	2.0 1.3	100

<sup>\*</sup>Raw data in Tables B-23 and B-27.

TRANSECT MEAN PERIPHYTON BIOMASS, CHLOROPHYLL & AND AUTO-TROPHIC INDICES, JUNE - JULY SURVEY, 4-WEEK INCUBATIONS\*

TABLE B- 20

Statio Trans		Chlorophyll <u>a</u> (mg/M <sup>2</sup> )	Organic Biomass (gm/M <sup>2</sup> )	Autotrophic Index
Α		2.5	0.9	360
В	Waconda	38	3.6	95
C	Bay	23	1.2	52
D F	•	18	2.7	150
F	Offshore	21	1.3	62
S	Reference	12	1.1	92
T	Bay A	<b>2</b> 3	1.5	65
X Y	Huss Lowe Slough	2 12	2.0 1.2	1000 100

<sup>\*</sup>Raw data in Tables B-24 and B-28.

TABLE B-21

TRANSECT MEAN PERIPHYTON BIOMASS, CHLOROPHYLL a, AND AUTOTROPHIC INDICES, AUGUST - SEPTEMBER SURVEY 2-WEEK INCUBATIONS\*

Station Transec	1 67 1	Organic Biomass (gm/M <sup>2</sup> )	Autotrophic Index
	aconda 19	2.8	147
	ay 8	-	-
	18	2.5	138
	eference 25	3.1	124
	ay A 15	2.2	147
	uss Lowe 16 lough 18	1.6	100 111

<sup>\*</sup>Raw data in Tables B-25 and B-29.

TABLE B-22

TRANSECT MEAN PERIPHYTON BIOMASS, CHLOROPHYLL a AND AUTOTROPHIC INDICES, AUGUST - SEPTEMBER SURVEY, 4-WEEK INCUBATIONS\*

Station o Transect		Chlorophyll <u>a</u> (mg/M <sup>2</sup> )	Organic Biomass (gm/M <sup>2</sup> )	Autotrophic Index
A Wa B Ba C	conda y	3 25 29	0.7 3.8 1.4	233 152 48
_	ference y A	23 19	1.5	<b>65</b> 53
	ss Lowe ough	7 13	1.4 1.2	200 92

<sup>\*</sup>Raw data in Tables B-26 and B-30.

TABLE B-23

RAW CHLOROPHYLL a RESULTS JUNE 2 WEEK INCUBATIONS

	Replicate Results (mg/m²)			
Station	1	2	3	
А	3.7	5.2	-	
B-1 B-2	19 26	27 -	<u>-</u>	
C-1	35	35	-	
D-1 D-2	16 13	24 15	-	
No Wake	8	-	-	
E-2	6	15	44	
S	12			
T-1	14	-	•	
Y-1 Y-2	8 16	12 16	- -	

TABLE B-24

RAW CHLOROPHYLL a RESULTS - JUNE
4 WEEK INCUBATIONS

	Replicate Results (mg/m²)				
Station	]	2	3	4	5
Α	2.8	1.8	2.8	2.6	-
B-1 B-2	47 47	49 28	- 13	- 47	- 38
C-1	21	29	19	23	-
D-1	14	7	24	23	22
F-2	23	19	-	-	-
No Wake	34	-	-	-	_
s	14	10	11	-	-
T-1	22	36	18	28	12
X-2	4	2	2	1	-
Y-2	10	13	13	-	-

TABLE B-25

RAW CHLOROPHYLL a RESULTS - AUGUST 2 WEEK TNCUBATIONS

	Replicate Results (mg/m²)		
Station		2	3
B-1	19	-	-
C-1 C-2	7 9	-	-
D-1	17	17	21
S	25	21	29
T-2	15	-	-
U-1	6	10	-
X-2	14	19	15
Y-1 Y-2	23 13	18 16	20 -

TABLE B-26

RAW CHLOROPHYLL a RESULTS - AUGUST
4 WEEK INCUBATIONS

	Replicate Results (mg/m²)				
Station		2	3	4	5
Α	3	-	-	-	-
B-1	25	-	-	-	-
C-1 C-2	1 <i>7</i> 32	31 36	26 28	33	- -
S	30	24	21	19	-
T-1 T-2	20 20	17	- -	- -	-
U-1	18	17	-	-	-
X-1 X-2	7 7	6 11	16 6	7 2	- 1
Y-2	10	13	16	-	-

TABLE B-27

RAW ORGANIC BIOMASS DATA - JUNE 2 WEEK INCUBATIONS

	Replicate Results (gm/m <sup>2</sup> )				
Station	1	2	3	4	5
Α	0.75	0.96	0.21	0.48	0.96
B-1	6.2	3.0	5.0	2.1	3.1
B-2	3.0	3.3	4.0	2.6	2.4
C-1	0.69	1.2	0.64	0.85	1.3
C-2	3.4	3.3	2.8	2.8	2.3
D-1	1.7	1.7	2.1	1.7	1.3
D-2	0.85	1.5	0.75	1.4	1.1
F-1	1.3	1.0	1.1	1.2	-
F-2	2.1	2.2	1.4		1.5
No Wake	2.7	2.6	2.9	2.8	3.1
S	2.4	2.2	2.1	2.6	1.1
T-1	3.6	2.5	2.6	1.7	2.7
T-2	4.3	2.7	2.7	3.0	3.0
U-1	3.9	1.7	1.1	1.7	1.3
U-2	1.1	0.75	1.4	0.80	
X-1	2.6	2.1	3.3	3.4	2.6
X-2	1.3	0.53	1.7	1.4	0.91
Y-1	3.1	1.3	1.6	1.2	2.1
Y-2	0.85	0.59	0.43	0.48	

TABLE B-28

RAW ORGANIC BIOMASS DATA - JUNE
4 WEEK INCUBATIONS

	Replicate Results (gm/m <sup>2</sup> )		
Stations		2	3
A	1.3	0.80	0.59
B-1 B-2	2.6 3.6	3.3 1.7	6.7 -
C-1	1.1	1.5	1.1
D-1 D-2	1.1 5.3	1.8 2.7	1.2 4.3
F-2	0.53	1.0	2.3
S	0.80	1.3	1.3
T-1	1.2	0.69	2.7
X-2	2.1	1.3	2.6
Y-2	0.75	0.85	2.0

TABLE B-29

RAW ORGANIC BIOMASS DATA - AUGUST 2 WEEK INCUBATIONS

	Replicate Results (gm/m²)		
Station	1	2	3
B-1	3.1	2.9	2.4
D-1	1.3	4.6	1.6
S	3.1	-	-
T-1 T-2	1.7 2.7	2.1 2.7	1.8 2.5
U-1 U-2	2.8 3.0	2.8 1.6	2.3 3.0
X-2	0.48	2.5	1.8
Y-2	2.4	2.3	1.2

TABLE B-30

RAW ORGANIC BIOMASS RESULTS - AUGUST
4 WEEK INCUBATIONS

	Replicate Results (gm/m <sup>2</sup> )				
Station	1	2	3	4	5
A	0.70	0.82	0.58	-	-
B-1	3.8	-	-	-	-
C-1 C-2	1.5 1.3	1.6	1.0	- -	- -
s	1.5	1.0	1.6	1.5	1.9
T-1 T-2	0.80 1.6	0.81 -	0.79	- -	<u>-</u>
X-1	1.4	-	_	-	-
Y-2	1.0	1.4	-	-	-

TABLE B-31

AUTOTROPHIC INDEX DATA
CALCULATED FROM MEAN CHLOROPHYLL a
AND MEAN ORGANIC BIOMASS RESULTS

	June Trip		
Station	2 Week Incubation	4 Week Incubation	
Α	167	356	
B-1 B-2	170 120	88 77	
C-1	54	52	
D-1 D-2	85 79	76	
S	175	94	
T-1	187	65	
Y-1 Y-2	166 37	80 120	
X-2	-	1000	
	August Trip		
Station	2 Week Incubation	4 Week Incubation	
A	-	2333	
B-1 B-2	147 -	152	
C-1 C-2		61 41	
D-1 D-2	1 39 -	- -	
S	124	-	
T-1 T-2	- 175		

Table B-31 (Continued)

	Augus	t Trip
Station	2 Week Incubation	4 Week Incubation
U-1 U-2	-	- -
X-1 X-2	100	- -
Y-1 Y-2	100	- -

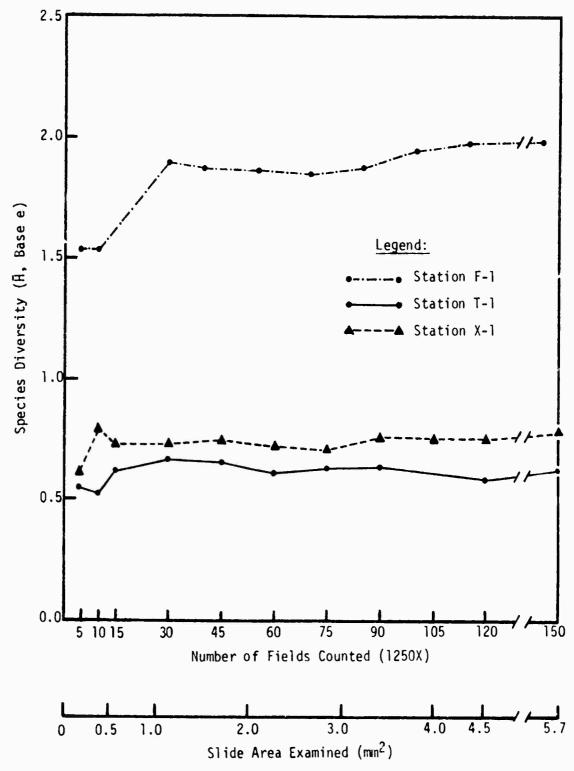
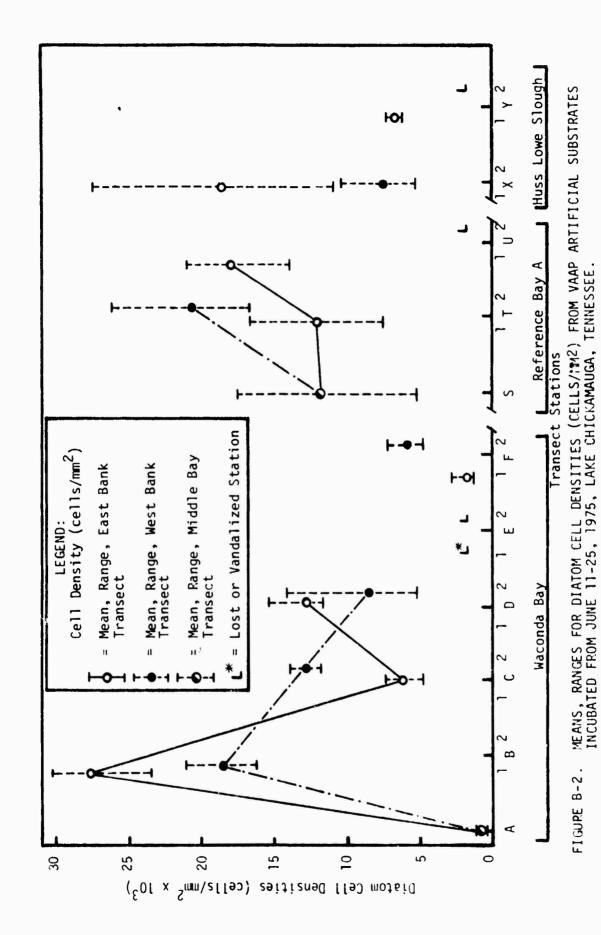
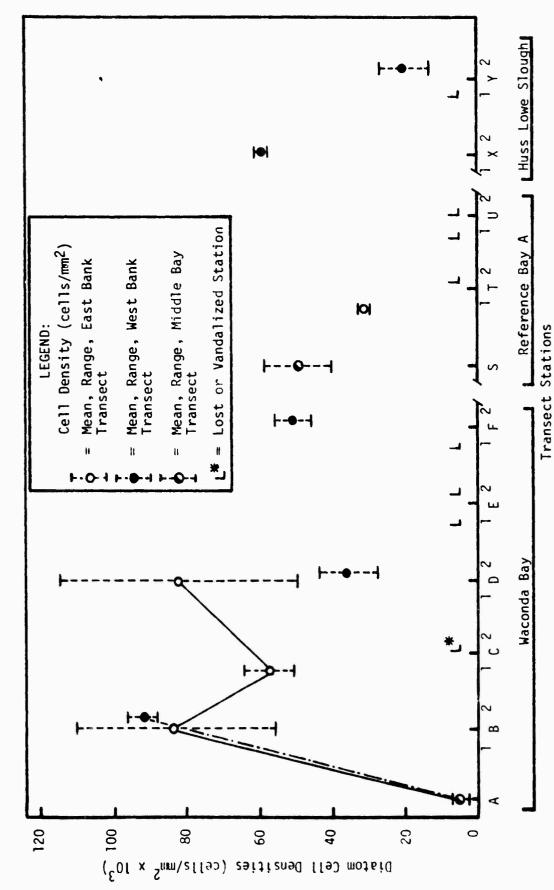
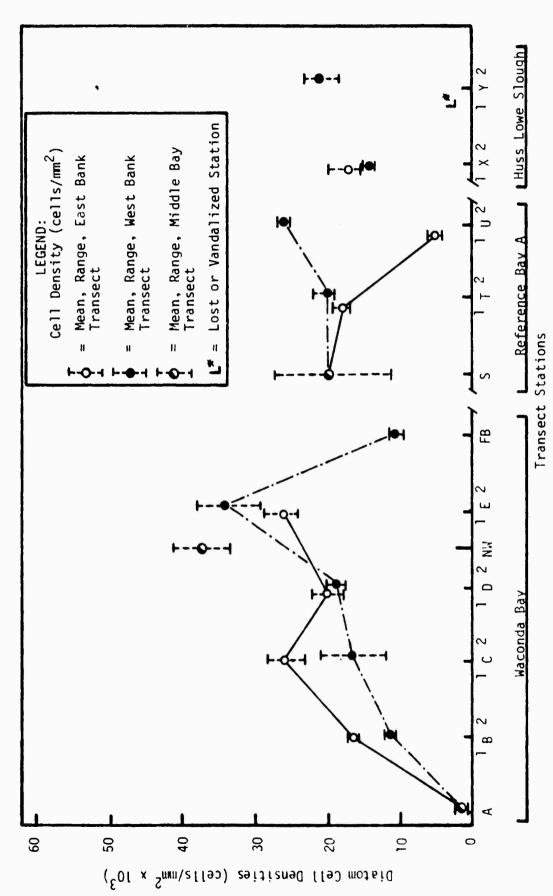


FIGURE B-1. THE EFFECTS OF INCREASING SAMPLE SIZE ON DIATOM SPECIES DIVERSITY IN SUCCESSIVE MICROSCOPIC FIELD COUNTS, VAAP PERIPHYTON DATA FROM STATIONS F-1, T-1, X-1, JUNE 11-25, 1975.



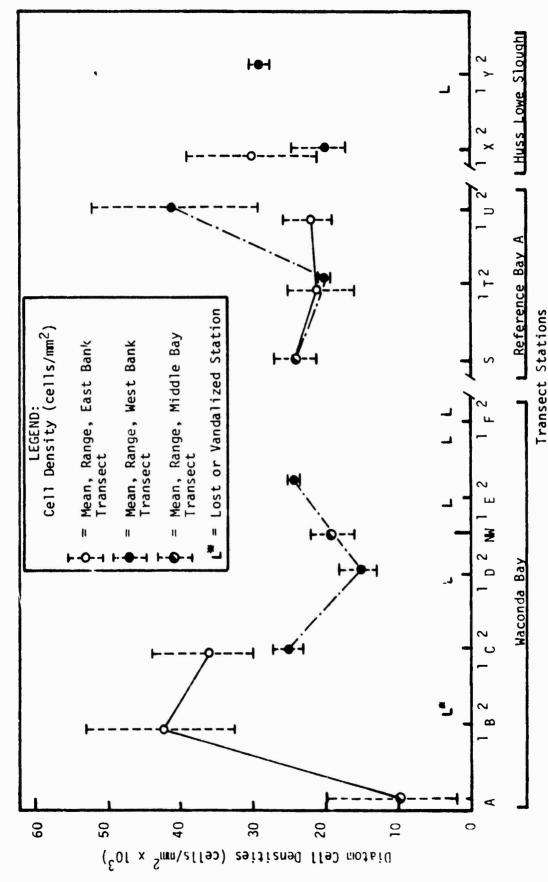


MEANS, RANGES FOR DIATOM CELL DENSITIES (CELLS/MM<sup>2</sup>) FROM VAAP ARTIFICIAL SUBSTRATES INCUBATED FROM JUNE 11-JULY 10, 1975, LAKE CHICKAMAUGA, TENNESSEE. FIGURE B-3.



MEANS, RANGES FOR DIATOM CELL DENSITIES (SELLS/MM $^2$ ) FROM VAAP ARTIFICIAL SUBSTRATES INCUBATED FROM AUGUST 12-26, 1975, LAKE CHICKAMAUGA, TENNESSEE FIGURE B-4.

\*Transect F, Stations F-1 and F-2, were lost during this sampling period; therefore Stations FB and NW were used as alternate sampling sites.



MEAN, RANGES FOR DIATOM CELL DENSITIES (CELLS/MM<sup>2</sup>) FROM VAAP ARTIFICIAL SUBSTRATES INCUBATED FROM AUGUST 12-SEPTEMBER 7, 1975, LAKE CHICKAMAUGA, TENNESSEE. FIGURE B-5.

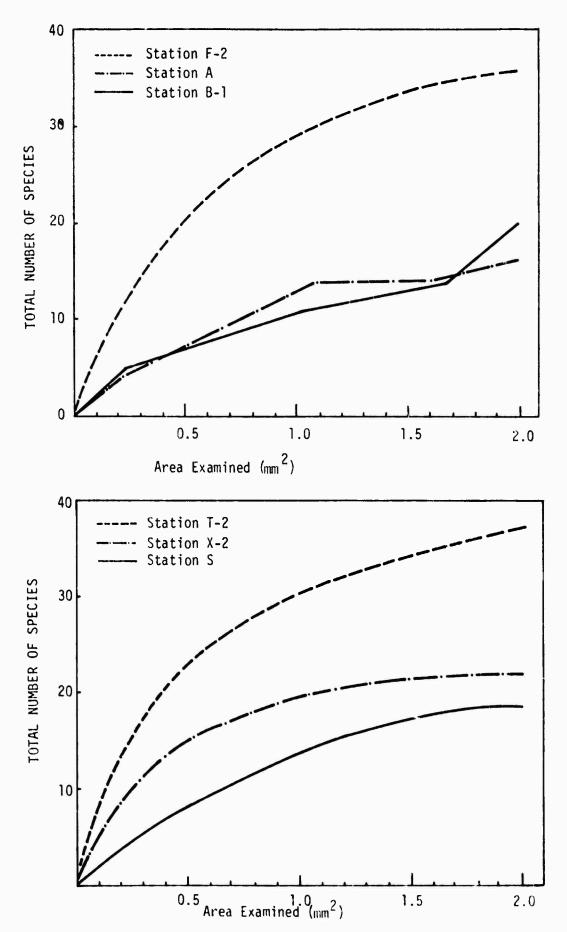


FIGURE B-6.DIATOM SPECIES AREA CURVES FOR VAAP NATURAL SUBSTRATES, LAKE CHICKAMAUGA, TENNESSEE, JUNE, 1975

APPENDIX C
PHYTOPLANKTON

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TABLE C-1
TAXONOMIC LIST OF VAAP FHYTOPLANKTON, JUNE 9, 1975
A ' B-1 B-2 C-1 C-2 D-1 D-2 E-1

	:		ı´´	•	)	•	<b>:</b>	- J
CHNANTHES SP A	1 1	0	•	:1	99	•	11	•
ACFRANTIES EXICUA Y METEROVALVE ACFRANTIES LANCEGLATA V DUBIA	' =	22	1	ıi	• •	1 1	22	22.
ACPNANTHES MINUTISSIMA . 6 ACHNANTHES MCLLII AMPHIPLEURA PELLUCIDA	876 126	33800	6754 388.	15315	17495.	15319.	3168	22830 182 90
AMFHORA CVALIS V FEDICULUS Amfhora Cvalis V Lievca Amfhora Perpussilla	121	111	iīı	111	111	121	121	811
AMFHORA SP 1 ANCEGENEIS VITREA CALONEIS BACILLUM	=''	1 88	111	101	121	1 25 1	121	1 60 1
COCCCNEIS PLACENTULA V EUGLYPTA CYCLGTELLA MENEGPINIANA CYCLGTELLA STELLIGERA	1 1 2	121	118	111	1 2 2	222	# 1 I	1 1 1
CYCLOTELLA SP 2 CYCLUTELLA SP 3 CYPBELLA AFFINIS	118	115	118	119	111	110	116	22.1
CYMBELLA DELICATULA Cypuella Sinuata Cymbella Ventficosa	111	111	811	110	1 2 1	4.0	111	115
CYMUELLA MICFOCEFMALA CYWBELLA LAEVIS CYPBELLA TUMICA	=''	1 44	4	791	221	©	0 T O	<b>;</b> ''
CVMBELLA SP 2 DIFLUNEIS OVALIS DIPLCNEIS SP A	111	111	• • •		111	122	111	111
FRACILARIA CAPUCINA Fublicaria Construens n Fracilaria Construens v Venter	103	9 1 1	200	# B B B B B B B B B B B B B B B B B B B		1691	397	1396
FRAGILAHIA CROTONENSIS Fragilaria Pinnata Fragilaria Vaucheriae	120	111	4 4	111	111	<b>\$</b> 11	""	114
GCBFHONERA ACUMINATUM V MONTANUM GCBPHONERA ANGUSTATUM V PRODUCTA GCBFHCNEMA AUGUS	1 1 2	22	111	1 20	1 10	' ' "	1 88	22 1
GCDFHCNEWA CCNSTRICTUM V. SUUCAPITATA GCDFHCNEWA GRACILE V LANCEULATA GGMPHUNEWA GRACILE V NAVICULDIDES	121	111	121	1 0 10	111	121	121	188
GCPFHUNEMA INTRICATUM V PUMILA GCPPHONEMA MCNTANUM V SUBCLAVATA GCMPHGNEMA PARVULLM	1 1 1	22	1 1 2	1 9 9	2 1 2	136	A 4	22 2
MELOSIRA AMDIGUA MELOSIWA DISTANS MELOSIWA GRANLLATA N	200	0 0 0 7 0 0 7 0 0	711	\$ TH	. 182	61.0 60.0 60.0	101	22
MELCSIRA GRANULATA V AMGUSTISSIMA ''' NAVICULA ARVENSIS NAVICULA CAPITATA	=''	1 1 2	111	115	111	111	111	
NAVICULA CHYPTOCEPHALA V INTERMEDIA Navicula chyptocephala v veneta Navicula decussis	1 0 1	111	111	1 1 1		1 6 1	- FE	162
NAVICULA MEUFLERI V LEPTOCEPHALA Navicula Lanceòlata Navicula Luzonensis	111	1 , (	22	111	111	<b>64</b>		111

TABLE C-1 (Continued)

NATIONA MINITA	B =	67	; 1	•	1	•	1	•	
AVICULA PUPULA AVICULA PUPULA V. ELLIPTICA AVICULA PYGMAFA	183	111	118	111	[8]	1 4 1	181	22.1	
VICULA RADIOSA <b>v Tene</b> lla Vicula Phyncecephala <b>v Gernani</b> Vicula Simplex	-1 1 1	111	111	111	111	111	111	111	
AVICULA SVEMPETRICA AVICULA SVEMETRICA AVICULA TENERA	="	111	• • •	115	•••	611	118	111	
VICULA THIERFRANMI VICULA THIDERTCLA VICULA SP 3	=''	111	111	iıı	811	121	111	N I I	
NAVICULA SP 6 NITZSCHIA ACICULARIS NITZSCHIA AMPFIRIA	111	111	1 🕻 1	111	111	1 0 1	111	1 🕻 1	
TZSCPIA DISSIPATA TZSCPIA DENTICULA TZSCPIA FUNTICOLA	=="	111	ıiı	181	111	19	1 <b>#</b> 1	1 10 1	
125CHIA KUTZINGIANA 125CHIA IGNCEATA 125CHIA LINEARIS	211	811	\$11	₩ I İ	iıī	228	611	22	
72SCHIA MICROCEPHALA 12SCHIA PALEA 72SCHIA RECTA	==1	111	111	1 50 1	1 1 1	111	101	1 0 1	
IZSCHIA SP 1 IZSCHIA SP 4 IZSCHIA SP 5	111	ill	N I I	1 1 1	811	811	110	811	
125CHIA SP 7 125CHIA SINUATA V TABELLARIA 125CHIA TRYBLICNELLA V VICTORIAF	181	111	111	111	111	1 4 8	111	1 6 1	
OPEPHONA MARTY! PINNULAHIA BRAUNI HHCICUSPHENIA CURVATA	111	1 1 1	1 1 1	111	110	1 1 1	111	•••	
STEPHARCOISCUS ASTRAEA Stepharodiscus SP 1 Sufirella cvata	23	121	1 % 1	111	1.6	113	121	181	
SYNEDHA ACUS Synecha Pulcmella Synecha uflicatissima	333	1000	1716	320	118	9	63.0	454	
SYNEDFA BUMPENS SYNEDRA ULNA SYNEDRA ULNA V DANICA	1 1 g	200	953	122	1 1 1	251	375	308	
ABELLARIA FENSTRATA		-	'	'			'	•	
0	8462.	8462,37138.10397	10397	16998.		20682	5741	.26737.	
NUMBER OF TAXA	34	24	22	23	21	37	33	31	

TABLE C-1 (Continued)

CRYPTOPMYCEAE	E-2	F-1	F-2	S	1-	T-2	U-1	N-2
CHYPTOMUNAS SP CRYPTOMONAS SP 3 RFCOCHCNAS SP	8 1 M	511	137	6 1 8	214	124	8-1 5-2	92
• • • • • • • • • • • • • • • • • • • •		*	•			•		•
CHLOROPHYCEAE						-		
CHLOPCCOCCALES								
ANNI STRUDE SMUS FALCATUS COELASTRUM MICROFORM CHUCICENIA TETHAPEDIA	<b>6</b> m m	K.w.+	6	110	2 · L	2 ° °	9''	<u></u>
DICTYDSPHAERIUM FLLCHELLUM GOLENKINIA PAUCISPINA KIFCHNERIELLA CCATORTA	1 1 89	• • •		• • •		•••	110	119
KIRCPMERIELLA LUMARIS KINCPMEMIELLA DJESA MICRACTINILM PUSILLUM	111,	31	2.5	110	210	6	12.	2 1 2
OCCYSTIS GLOFCCYSTIFORMIS PEGIASTHUM BIFACIATUM PLUIASTHUM DUFLEX	+ 17 +	• • •	101	110	1 MM	• • •		
PEDIASTRUM SIMPLEX PEDIASTRUM TETRAS Scenedesmus apuncans	• • •	•••	• • •	mit	• • •	• • •		<b>©</b> 1 1
SCENEDE SWUS ACUMINATUS SCENEDE SWUS AFCLATES SCENEDE SWUS BFFRAPPII	• • •	•••	1 1 M	• • •	• • •	• • •	• • •	•••
SCENEDESMUS PLUGA SCENEDESMUS DENTICULATUS SCENEDESMUS CUADRICAUDA	= 10	28 . 0	5101	H 10 10	2 ° ° ° ° ° ° ° ° ° ° ° ° ° ° ° ° ° ° °	:10	4 1 8	4 i Fi
SCENEDESMUS SERPATUS TETPAUESMUS WISCUNSINEMS! TETRAEDMON WINIMUM	118	116	163	118	110	110	1 1 15	5811
TREUBARIA SETIGERUM	m	•	•	n	0	•	•	٥
VCLVDCALFS								
CHLAMVOCHUMAS SP GENIUM FERMESOM PANCHINA CHARROWIENSIS	150	8n1	271	131	5 M 1	omn T	99 m	0 m m
PANDORINA MORUM SPHAEHELLIPSIS SP	• •	52	• •	10	3.5	• •	• •	• •
						•		•
CCSMARIEN PHASEOLUS CCSPARIUM SP 3 CCSPARIUM SP 4	• • •	1 1 11	• • •	• • •		• • •		• • •
EUASTRUM SP EUASTRUM SP 1 STAURASTRUM LEPTCCLADIUM	110	171	1 10 10	117	1177	•••	110	

TABLE C-1 (Continued)

XANTHOPHYCEAE	E-2	F-1	F-2	S	T-1	T-2	U-1	Ú-2
PH10CYTSU SEUDOTETR	10	-12	no	11	1.0	m i	11	
CHRYSOPHYCEAE		•		•			•	
CENTRITEACTLS DUPIUS CMFYSOCOCCUS PUSILLUM DINCLRYON BAVARICUM DINCBHYON DIVERGENS OCHROMGNAS GRANULCSA	112 81	1 1 S 4 P N	116 81	1222	110 81	110 1	124	383
DINCPHYCEAE	• • • •		•					
CERATIUM MIRUNDINFLLA GLENDOINIUM ACICLLIFERUM GLENDOINIUM ACICLLIFERUM PERIDINIUM ACICULIFFRUM PERIDINIUM INCONSFICUM PERIDINIUM PUSILLUM PERIDINIUM TATRICUM	11 7 011 41	911 IF 88	11 - 01 00 1	191 110 01	। ଏହି ମଧ୍ୟ ହା ମଧ୍ୟ	110 110 NI	IIO NII SI	IIO OIO NI
CY ANDPHYCF AE	•		•					
CHECCHCCALES COELDSPHAERIUM KELTZINGIANUM DACTYLUCCCCOPSIS RAPHIDIDIDES GCMPHOSPHAFRVA LACUSTRIS	114	1 92	119	MMO	202	916	110	1 . 2
MERCSMOPEDIA GLAUCA	<b>E</b>	•	ı	ı	1	•	en	ı
CSCILLATORIALES LYNGBYA MADTENSIANA DSCILLATGRIA GFWINATA RHAWHIDICPSIS SP	101	mıl	101	101	121	121	1 6 1	1 6 1
SPIRULINA LAXISSIMA TRICHUDESMIUM SP	7	16 C	21	₹'	81	KI I	<b>5</b> 1	;'
					_	_	_	

TABLE C-1 (Continued)

BACILLARICPHYTA (DIATOMS)	E-2	<u>:</u>	F-2	S	-	1-2	[-]	N-2
ACHNANTPES LANCEDLATA N ACHNANTHES LANCEDLATA V DUBIA ACHNANTHES SP 1	Inn	101	noi	111	111	101	111	111
ASTERIONELLA FCRMCS. Atteva zacharias Cucconeis placentula y Euglypta	110	011	115	111	010	111	118	110
CYCLOTELLA BODANICA CYCLOTELLA STELLIGERA CYCLOTELLA SP	111	101	111	111	111	111	111	111
CYPBELLA SP DIPLONEIS CCULATA FRAGILARIA CAPUCINA	mni	nın	011	128	20 L G	12	115	111
FHAGILARIA CHOTONENSIS Gyrcsigna Sp Melcsika ampicla	10.2	170	239	105	92	707	163	131
MELCSIRA DISTANS MELGSIRA GRANULATA N Navicula capitata	350	183	811	-60	303	278	286	108
NAVICULA CPYPTCCEPHALA V VENETA NAVICULA PHYNCHDCEPHALA NAVICULA VIRIGULA	011	1 ! !	111	911	111	111	n i i	211
NAVICULA SP 2 NAVICULA SP 3 NAVICULA SP 4	201	211	251	Ç11	mm (	\$ 1 1	nıı	011
NIIZSCHIA ACICULAPIS NIIZSCHIA PALEA RHIZUSCLENIA SP	N 0 1	nnw	110	0 12 1	101	251	211	011
SURIRELLA SP Synecha actimastrcides Syneora delicatissima	( N W)	1 2 2	100	113	1150	1501	110	110
SYNEDRA RUMPENS Synedra Ulna Tabellaria Fenstrata	944	1 2 2	mg i	· no	997	988	122	101
FUGIFNCEAE					•			
EUGLENA SP THACHELOMCNAS AUSTRALICA THACHELCMCNAS CRFEFA	110	Ø 1 Ø	nıı	210	oin	nın	10 t a	nio
TRACHELOMONAS VOLVECINA THACHELOMONAS SP 11 TRACHELOMONAS SP 111	111	111	Ø11	111		111	110	111
TRACMELCHONAS SP 1V	n	1	1	•	•	n	•	n
VEN [NEODO MO BEDERIN   14454		en en	9101	1928	7 7 7		2868	1575
NUMBER OF	:	9	•	45	*	38	39	3.5
•	1			•				

TABLE C-1 (Continued)

Y-2	11 ++++	· • • • •	*****		111	****	••••• •••••	****	159		101	п п
x-2 + Y-1	11	••••	110 7	• • • • • • • • • • • • • • • • • • •	• • • • • • • • • • • • • • • • • • •	111	++ + + • + + + 1	****	110	°	101	01 01
	m i	<b>, +++</b>	111 5	• • • • •		MI 9	M	••••	I ANUM + 1	***	• • • • •	• • • • •
HANTHOPHYCEAE	OPFICCYTIUM CAPITATUM PSEUDOTETRAEDRCN NEGLECTUM	CHRYSOPHYCEAE	CENTRITRACTUS DUBIUS CHEYSOCOCCUS PUSILLUM DINCLAYON RAVARICUM	OCHECHENAS GRANDLESA DINOPHYCEAE	CERATIUM HIPUNDINELLA GLENDDINIUM ACICULIFEHUM GLENDDINIUM QUADRIDENS	PERIDINIUM ACICULIFFRUM PERIDINIUM CUNNINGTONII PERIDINIUM INCINSFICUEM	PERIDINICA PUSILLUM PERIDINICA TATRICUM	CYANDPHYCEAE	COELOSPHAFRIUM KEUTZINGIAMUM DACTYLOCOCOPSIS RAPHIOIOIDES GCHPHOSPHAFRVA LACUSTRIS	MERGSMOFEDIA GLAUCA	OSCILLATORIALES LYNGHYA MARTENSIANA OSCILLATORIA GEMINATA RHAPMIDIOPSIS SP	SPIRULINA LAXISSIMA THICHODESMIUM SP

TABLE C-1 (Continued)

X-1 * X-2 * Y-1 * Y-2	* * * * * * * * * * * * * * * * * * *	12 + 12 + 12 + 12 + 12	• • •		55 + +	+ + + · · · · · · · · · · · · · · · · ·	+ + + + + + + + + + + + + + + + + + + +	**************************************	61	150 + 134 + 249 + 2399	•		••	N ++	+ + + + + + + + + + + + + + + + + + +	++ 0***	• •		· · · ·		22 + S1 + S2 + S2 + S2 + S2 + S2 + S2 +	· • • •	•	**	**	* 1	+++++++++++++++++++++++++++++++++++++++	•	1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1	•••	1 ++	. **	1450 + 1366 + 1306 + 1929	+
BACILLARIOPEVTA (DIATOMS)	ACFNANTFES LANCECLATA N ACHNANTHES LANCECLATA V UVBIA ACHNANTHES SP 1	STERICHELLA FORMOSA TPEYA ZACHARIAS	COCCONE IS PLACENTULA V EUGLYPTA	CYCLOTELLA BCCANICA	CYCLOTELLA SP	CYMBELLA SP	FEAGILARIA CAPUCINA	FRAGILARIA CECTURENSIS	MELCSIRA ANHIGUA	MELOSINA DISTANS MELOSINA GIANUIATA N	CAPITATA	NAVICULA CRYPTOCEPHALA V VENETA			NAVICULA SP 3		 MAISOSITEMIA SP	SURINELLA SP Symbola actinastroides	SYMLDRA DELICATISSIMA	SYNEDNA RUMPENS	SYMECKA ULNA Tareliagia Francidata		FUGLENDPHYCEAE				TRACHELOMONAS CREBEA		RACHELCHONAS		TRACHELOMONAS SP IV		TOTAL MUMBER OF ORGANISMS	

TABLE C-2

TAXONOMIC LIST OF VAAP PHYTOPLANKTON, JUNE 13, 1975

CAYPT OP HYCE AE		CHL CROPHYCE AE	CHLOPCCOCCALES	CINASTRUM MANTSSCHI ARISTRUDE SPUS FALCATUS	CAMPRICUM MICEGECEUM GUADRATA	TETRAPECIA PIUM PLICHELLUM PELLA LIMNETICA	DROESCHERI A PAUCISPINA ELLA CONTORTA	A LUMAPIS A OHESA QUADRISETA	PUSILLUM CCYSTIFORMIS	3	KAMPAISKY! OHILSLW SIMPLEX	AFUNCANS ACUMINATUS AFCUATUS	APPATUS HEUNARGII BIJUGA	DENTICULATUS OUADFICAUDA SERRATUS	SFTICERA VINIBUS FEGULAFE	TPIGONUM SPI SETIGERUM	TETRASPURALES X GELATINOSA 15 SCHPOEPTER!
	917	***************************************	r	110	111	111	1,11	111	101	110	111	nıı	100	125	ngı	111	11
B-1	<b>*</b> '5		o	n 9	111	mli	111		101	110	111	•	22.	n 80 -	101	111	11
B-2	105		,	114	115	חוו	110		111	110	110	011	110	141	nei	110	• •
13	10 10 10		ŗ	" 5	181	32.	123	111	101	111	111	mii	19	n <del>-</del> n	161	111	
C-2	63		•	410	mii	111	101		111	110	• • •	mm I	170	121	IDI	01 m	1 1
D-1	12.51		(	63	110	111	101	111	111	101	101	111	118	1 50 1	កស្ដ	110	1 6
0-2	212	<del>-</del>	,	63.	241		1 750	181	101	110	110	711	118	181	101	110	01
<u>.</u>	215	-	1	126	701	111	100	111	101	110	111	111	3, 1	181	101	111	<u></u>

TABLE C-2 (Continued)

E-1	. n	* * * *			8		11	•	0 m 0 1	•	111 18
3.5	· ·			' '			I M		n i e		110 11
1-0		•		4 4	•		m i	•	NOM 0	•	
C-2	I M	• • •		11	•	•	( P)	•	<b>6</b> 877 6	•	iii Mm
[-J	1 0		111	1 1	1		11		0 1 W 1	•	111 00
B-2	1 (		111	1 1			1 17		110 1		111 Pm
B-1		•	111	11	•		11	* * *	525	•	11P 00
A 6 1	. •		111	11	•		m I		911 0		111 20
VOLVOCALES HLAMYDCMONAS SP JUCORINA SP	GENIUM FERNOSOM Sphaerellopsis Sp	••••••••••••••••••••••••••••••••••••••	COSMARIUM SP 3 COSMARIUM SP 4 Elastrum SP	STAURASTRUM LEPTOCLADIUM Staurastrum Quadricuspidatum	ZYGNEMATEL OTIA SPP	XANTHOPHYCEAE	OPFICETTION CAPITATUM PSEUGOTETRAEDREN NEGLEGTUM		CHAYSOCOCCUS PUSILLUM Dinclayon Pavaricum Dincbayon Divergens Syruha Sp	DINOD HVCFAE	CERATION HIPUNDINFULA GLENDUNIUM ACICLLIFENUM PERIDINIUM CUNNINGTONII PERIDINIUM INCENSPICUM PERIDINIUM PUSTI LUM

TABLE C-2 (Continued)

( A   B-1   B-2   C-1   C-2	1 4	1 11	31 9 12 15	111	v	22 - 1 - 2 - 3 - 1 - 3 - 3 - 3 - 3 - 3 - 3 - 3 - 3	ı	***********	CDIATOMS	01 1 0 1 1 0	3 22 1 2 1 2 2 2 1 3 2 5 6	111	101	15 44 57 102 156	30 3 214 457 271 265 31 25 67 57 60	ANGUST 155 14A 25 - 6 76 - 15 - 15 - 15 - 15	67 87 22 31	31 0 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1	1
CYANDPMCEAE	CMEDCECEALES APHANDCAPSA DELICATISSIMA	KFUTZINGI IS ACICULA	CCMPHOSPHAERVA LACUSTRIS	MEFCSMOPEDIA GLAUCA Merismopedia tenuissima Microcystis aepuginosa	DSCILLATORIALE	AMABAEMA SPIPOIDES DSCILLATURIA GEMINATA SPIRULINA LAXISSIMA	TRICHODESMIUM SP	******	BACILLAPICPHYTA (C	ACHNANTHES LANCECLATA N ACHNANTHES LANCEOLATA V C ANCHOECNETS SP	ASTEMIUNFLLA FORMOSA ATTEVA ZACHARIAS CCCCCNEIS SP	CVCLOTELLA STELLIGERA (VCLOTELLA SP CVMUELLA AFFINIS	CYMBELLA VENTRICESA CYMBELLA TUMICA CYMBELLA SP	FRAGILARIA CAPUCINA FRAGILANIA CENTINENSIS GVECSIGNA SP	MELCSIRA AMPIGUA MELCSIKA DISTANS MELOSIKA GRANDLATA N	MELUSIFA CRANULATA V ANGL MELUSIFA VAHIANS NAVICULA CAPITATA	MAVICULA CPYPTOCEPHALA NAVICULA BHYNCHCCEPHALA NAVICULA SF 1	MAVICULA SP 3 MAVICULA SP 3 MAVICULA SP 4	NAVICULA SP 5

TABLE C-2 (Continued)

4 9 9 9 1 1 1 2 1 1 1 1 1 1 1 1 1 1 1 1 1	4	B-1	B-2	2	C-2	1-0 E	2-0	
NITSSCHIP SP 2 RHIZCSOLENIA SF	, m	1 1	1 1		1 1	n (		11
STEPHANDUISCUS ASTRAFA STEPHANDDISCUS ASTRAFA V. MINUTULA SUMIMELLA SP	111	111	110	111		181		111
SYNEDHA ACTINASTROIDES SYNEURA DELICATISSIMA SYNEDRA HUMPENS	<u>o</u> mm	Su I	205	101	71 -	on i		NE I
SYNEGRA ULNA SYNEGRA ULNA V DANICA TABELLARIA FENSTRATA	121	917	1010	IMM	211	121		121
T CUADRISEPTATA	•		1		22			1
EUGLENOPHYCEAE								
EUGLENA SP PHACUS SP 1 TRACHELCHINAS CREEEA	011	0.001	110	IID	111	010		111
TRACHELOMONAS COLENGA THACHELCHONAS VOLVOCINA TRACHELOMONAS SP 11	011	. 1 1	101	101	MII	111		111
TRACMELONONAS SP 1V	m	٥	12	ō	n	1		n
CNIDENTIFIED TAKA								<u> </u>
NIOE	1 1	11	ÐΙ	11	1 1	m i		11
TOTAL NUMBER OF URGANISMS	1277	159	2327	1040	1383	1789	1301	<del></del>
NUMBER OF TAXA	•	+3	53	4.4	20	4	43	Ð

	E-2	Ξ :	F-2	S	1-1	1-2	U-1	N-2	
CRYPTCHONAS SP CRYPTUMCNAS SP 3 RHCDGMCNAS SP	9 9	2 - 2	72.57	815	8 1 4	5,3	n n	50 T	
CHLCRCPHYCEAE							-		
CHLCACCOCCALES							-		
ACANTEDSFMAFFA ZACEARIASI ACTINASTRUM MANTZSCMII ANKI STRGUF, SMUS FALCATUS	: I KN	112		1 1 %	S 1 8		m 1 9	119	
COELASTRUM CAMMPICUM CUELASTRUM MICFOPORUM CALCIGENIA QUADRATA	mm ı	101	. m 2		ומו	101	101	102	
CALCIGENIA TETRAPFOIA DICTYOSPARPILM PULCMELLUM ECHINOSPARPELLA LIMNETICA	1	711	101	1 10 1	• • •		101	111	
FRANCEIA DROESCMEPI GULLNAINIA FAUCISFINA AIMCMNEFIFLLA CCNTCRTA	101	101	1 0 1	150	111	101	101	10.1	
KIRCPNEHIFLLA LUNARIS KINCHNEFIELLA CBESA LAGEMMEIMIA OLADRISETA	411	• • •	101	101	111	111	78	. 10	
LAGERHEIMIA SP Michactinium Rusillum Uocystis gloficestiformis	100	1 9 17	191	ומו	• 1 1	1 1 1	101	1 6 1	
DCCYSTIS SCLITARIA PEDIASTNUM RIRADIATUM PELIASTRUM DUPIFX	1 0 1	• • •	911	1 1 1				• • •	
PEDIASTRUM RANDAISKYI PECIASTRUM URTUSUM PECIASTRUM SIPPLEX	n.,	• • •		110	1:1	110		• • •	
SCENEDE SPUS ABUNDANS SCENEJESHUS ACLAINATUS SCENEJESHUS ARCUAILS	nın	• • •	211				i 1 1		
SCENEDESMUS ARMATUS SCENEDESMUS EFUNARDII SCENEDESMUS BIJUGA	n 1 ¢	1 1 82	1 1 20	n 1 0	118	912	918	1 100	
SCENEUESMUS DENTICLEATUS SCENEUESMUS QUADRICAUDA SCENEDESMUS SERRATUS	1 20	95.1	001	22.	1 25 1	101	181	16	
SCHRUFCEMIA SFIICEMA TETFAEDHUM PINIMUM TETFAEDHUM REGULAPE	121	101	<b>9</b> n1	nn i	nnı	121	181	កស្ព	
TETFAEDRON TRICONUM Tetraeumon sed Treubaria setigraum	011	mii	911	111	111	1 00	110	<b>0</b> 11	
4									
ELAKATDIMRIK GFLATINOSA Spraerucystis schroerteri	<u>4</u> 1	n ı	10	1 1	пı	• •	• •	<b>0</b> 1	
VOLVOCALES									
CMLAMYDOMDNIS SP EUCORIMA SP GDAIUM FORMOSUM	52 1 1	818	910	<b>3</b> 01	ก็ก็	# Im	901	810	
SPHAFHELLOPSIS SP	0	•	0	9.6		•	•	¢	

TABLE C-2 (Continued)

	DESMIDACEAE	E-2	F	F-2	S	1-1	T-2	L-U	N-2
AUPASTRUM QUADRICUSPIDATUM  ZYGNEMATELES  LGEOTIA SPP  CHRYSOBHYCEAE  CHRYSOBHYCEAE  CHRYSOBHYCEAE  CHRYSOBHYCEAE  CHRYSOBHYCEAE  DINOPHYCEAE  STOLING CONTINUE CHRYSOBHYCEAE  CHRYSOBHYCEAE  CHRYSOBHYCEAE  CHRYSOBHYCEAE  STOLING CONTINUE	888	111	111	111	• • • •	111	111	111	
TYGNEWATELES  UGEDTIA SPP  TICCYTUM CAPITATUM  TICCYTUM CAPITATUM  THOUGHT CAPITATUM  THO		• •	11	01	1 1	11	• •	11	
TANTHOPHYCEAE  TELECTIUM CAPITATUM  ELOGIETRACERROLL  CHPYSOPHYCEAE  TYSOCOCCUS PUSILLUM  TYSOCOCCUS PUSICUM  TYSOC	·	•		1	٠	1	•	67	······
TICCYTIUM CAPITATUM ELOGTETRAEGRON NEGLECTUM ELOGTETRAEGRON NEGLECTUM ELOGTETRAEGRON NEGLECTUM TYPERCENCY CHRYSOPHYCEAE  ALTIUM HIRUMEINELLA ENDOINIUM ACTOULIFERUM THOUSILLUM ACTOULIFERUM THOUSILLUM									
FLECTIUM CAPITATUM  ELOGOTETRAECRCN AFGLECTUM  CHPYSOBMYCEAE  TYSOCOCCUS PUSILLUM  NCLATCN BAVARICUM  NCLATCN BAVARICUM  NCLATCN BAVARICUM  NCLATCN BAVARICUM  NURA SP  DINOPHYCEAE  FATIUM HIRUNCINELLA  FATIUM HIRUNCINEL	OPHYCEAE		•						
CHPYSOPHYCEAE  CHPYSOPHYCEAE  TYSOCOCCUS PUSILLUM  CULKYCN GAVERICUM  COMMINCTONII  RIDINIUM PUSILLUM  RIDINIUM PUSILLUM  CYANGPHYCEAE  CYANGPHYCEAE  CHECCCCALES	DPFICCYTIUM CAPITATUM Pseudgtetraedron neglectum	71	10	• •	. 11	1 1	m ,	110	<del>and the same of t</del>
ANDRA SP  ENGUINUM MIRUNCINELLA  ENGUINUM MIRUNCINELLA  FIDINIUM PUSILLUM  RIDINIUM PUSILLUM  CVANDPHYCRAE  CMECCCCALES  CMECCCCALES	.+++++++++++++++++++++++++++++++++++++		•				•		
DINOPHYCEAE  SATIUM MIBUNCINELLA FLOINIUM ACICULIFFHUM FLOINIUM PUSILLUM RIOINIUM PUSILLUM CYANOPHYCEAE  CYANOPHYCEAE	CHAY SOCOCCUS PUSILLUM DINCLATCH BAVARICUM DINCUMATCH OIVERGENS SYNURA SP	min o	mod i	990 1	501 0	~ M & &	00n N	~ ~ F	നന്റെ ക
IUM HIBUNCINELLA INIUM ACICULIFFRUM INIUM CONNINCTONI INIUM PUSILLUM CYANDPHYCLAE CMECCCCALES	•								
INIUM PUSILLUM  CYANGPHYC.AE  CHECCCCALES	NO IN THE TO IN	111	111	111	110	110	115	111	
CY ANGPHYC'.AE		10	11	01	979	9.4	1 50	m o	••
CHRICCCCALES								•	
111 111 111	CMRCCCCALES APPANUCAPSA DELICATISSIMA CMRCGCCCUS DISPERSUS CGELDSPMAERIUM KEUTZIMGIANUM	111	Ø11	111	1 + 1		-		ØM 1

TABLE C-2 (Continued)

U-2	1 10 00	111		110	1			mii	110	Ini	111	91.	476 326 15	111	101	911	101	111	1:1-1
۱ <del>-</del> 1	116	111		104	ı	•		10 M	010	1501	111	87 I I	75 6 70 6 70 70 70 70 70 70 70 70 70 70 70 70 70 7	E' '	001	011	180	111	PII
T-2	22	IMI		188	1	:		mıl	1110	111	111	011	159	111	111	111	122	111	* * 1 1
-	23.1	111		22 1 1	1			111	111	101	111	111	204 214 35	811	101	111	IOM	111	mii
S	111	111		IMM	1			115	MII	1 27 m	IMI	011	175	, i	mol	22	222	111	111
F-2	100	omi	,	100	ı	:		111	521	101	111	911	115	SII	MM I	911	150	111	111
F-1	1 70	mmı		15	1			m l I	111	111	111	118	133	22 1	177	117	120	111	FFI
E-2	noi	mø i		1 9 %	ı			101	110	101	Î	28	278 195 159	111	101	mii	100	111	Ø11
	DACTYLOCUCCOPSIS ACICULARIS DACTYLOCCCOPSIS RAPHIDIOIDES GCEPHOSPHAERYA LACUSTRIS	MERCSMUPEDIA GLAUCA MERISMEPEDIA TENUISSIMA MICROCYSTIS AFPUGINOSA	DSCILLATORIALES	AMABAENA SPIPOIDES CSCILLATORIA GEMINATA SPIFULINA LAXISSIMA	THICHODESHIUM SP	*******	BACILLARIOPHYTA (DIATOMS)	ACHNANTHES LANCECLATA N ACHNANTHES LANCECLATA V DUBIA ANCHONEIS SP	ASTEMIONELLA FERMESA ATPEYA ZACHAPIAS CCCCCNEIS SP	CYCLOTELLA STELLIGERA CYCLUTE.LA SP CYMUELLA AFFINIS	CYPHELLA VENTRICOSA CYPHELLA TUMIDA CYPHELLA SP	FRAGILARIA CADUCINA FRAGILARIA COCTONENSIS GYFUSIGMA SP	MELOSIRA AMBICUA MELUSIRA DISTANS MELCSIRA GRANLLATA N	MELOSIRA GRANULATA V ANGUSTISSIMA MELCSIRA VARIANS NAVICULA CAPITATA	NAVICULA CPYPTREPHALA NAVICULA RHYNCHOCEPHALA NAVICULA SP 1	NAVICULA SP 2 HAVICULA SP 3 NAVICULA SP 4	NAVICULA SP 5 NITZSCHIA ACICULARIS NITZSCHIA PALFA	NITZSCHIA SP I NITZSCHIA SP 2 RHIZOSOLENIA SP	STEPHANODISCUS ASTRAEA STEPHANGDISCUS ASTRAEA V. MINUTULA SURIRELLA SP

TABLE C-2 (Continued)

	E-2	F-1	F-2	S	1-1	T-2	U-1	n-2	~
SYNEDRA ACTIMASTROIDES SYNEDMA DELICATISSIMA SYNEDRA RUMPFNS	57 I	111	111	211	ФM I	110	IMI	921	
STREDRA ULMA Stredra Ulma V Danica Tabellaria Fenstrata	100	101	121	IFR	101	<b>911</b>	152	121	
T CUADRISEPTATA	1	1	1	1	•	1	•	•	
•		•		•	:	****		•	
EUGLENDPHYCEAE			<del>,</del>	<del></del>					
EUCLENA SP PHACUS SP I THACHELCHONAS CREEFA	Ø11	111	<b>Q</b> 11	210	111	111	110	1 1 1	
TRACHELOMONAS OBLONGA TRACHELCHONAS VOLVECINA THACHELOMONAS SP 11	111	1 1 1	101	001	111	111	IFI	1 1 40	
TRACHELOMONAS SP IV	٠	1	1	25	•	1	1	•	
		•	•	•	* * *	•			
UNICENTIFIED SPECIES J	11		11	11	m i	1.1	m I	11	
• • • • • • • • • • • • • • • • • • • •				-			•	•	
TOTAL NUMBER OF ORGANISMS	1201	646	849 : 1071	1258	985	872	1360	1352	
NUMBER OF TAXA	53	38	24	0	38	37	20	:	

Y-2 :	31	111	•	• • •	'n	· ·	m	96	, <del>,</del>	<b>?</b> 1	•	m I	11	n f	. 52	 26		 Mm					11 1	 O I	150	1 2					 mm		· ·	
· · · · ·	0.	98		• • •	1	1 50		2 1	1	• • ·	••	• • ·	1		22 +	100	••	PI	m	10	· •	) m (		••	+ 62	me			90	•	15	• • •	1	1
X-2	ı,	17			1	68		na	1	12	•	<u>s</u> 1		N 1	1	520	,	11	1	11	F			1 1	73	10	1	•		m	10		2	1
X-1				• • •	1		n	•••	1	01	1	01	• • •	20	1	<b>*</b> 1	1	n •		••		1 1		••	E. ++	1 0	,	1 9	<b>2</b> 1	1	11	•••	1	1.
CPYPTOPHYCEAE	HYPTOMONAS SP	CRYPICHONAS SP 3 RHEDCHONAS SP	CHLCRCPHYCEAE	CHLCRCCCCALES	ACANTEGSPHAERA ZACHARIASI	ACTINASTRUM MANTSSCHII ANKISTROUFSHUS FALCATUS	DEL ASTRUM	COELASTRUM MICROPORUM Crucigenia Quacrata	CALCIGENIA TETRAPEDIA	DICTYCSFHAERILM FULCHELLUM Echingsphaerfla Limnetica	FHANCE IA DROE SCHER;	GOLENKINIA PAUCISPINA Kirchnemiella contopta	KIHCHNEH JELLA LUNARIS	KIRCHNEHITLLA CHESA LACERFEIMIA QUADRISETA	<	MICFACTINIUM FLSILLUM UDCYSTIS GLÜECCYSTIFORMIS	OCYSTIS SOL	PECIASIBUM BIRACIATUM PECIASIFUM DUPLEX	ASTRUM	PECIASTMUM CRILSCY FEDIASTMUM SIPPLEX	STATE OF STA	SCHEDENICS ACCURING SCHEDENICS ACCURING	CENEDESMOS	SCENEDE SAUS REPARTS	CENEUESMUS	SCENEUE SMUS CENTICULATUS	CENEUESMUS	CHROEDERIA S	TELTATOREN FIRST	W	TETAEDADM SPP TREUBARIA SETICEPUM	TETRASPCHALES	LAKATGTMP1X	SPHAEROCYSTIS SCHRCERTERI

TABLE C-2 (Continued)

VOLVOCALES	x-1	x-2	l-4 +-	* Y-2	
CHLAMYDOMONAS SP EUCOHINA SP GCNIUM FORMOSUM	กุกเ	M 1 1	<u>.</u>	211	
SPFAERELLOPSIS SP DESHIDACEAE	n		· • • • • • •	· • • • • • •	
CUSMARIUM SP 3 COSMARIUM SP 4 EUASIKUM SP	1 0 1	m o i	1 I M	110	
STAURASTRUM LEFTCCLADIUM Staurastrum Quadricuspidatum	81	N N N = 1	<u>8</u> 1	in I	
ZYGNEMATELES MOLGEOTIA SPP		38	m • • • • •	· · · · ·	
XANTHOPHYCEAE			· • • • •	· · • · •	
OPHIOCYTIUM CAPITATUM PSEUDOTETRAEORCM NEGLECTUM	11	11		11	
CHRYSOPHYCEAE		• • • • •			
U >>	100	\$ 4 W	F 00 1 + + + + + + + + + + + + + + + + +	M 21	
SYNURA SP DINCPHYCEAE	• • • •	<u>.</u>		· · · · · ·	
CERATION HIRUNDINELLA GLENDDINIUM ACICULIFFRUM PERIDINIUM CUNNINGTONII PERIDINIUM INCONSPICUUM PERIDINIUM PUSILLUM		o i i m i	111 00	inn fi	

TABLE C-2 (Continued)

CY AND PHYCE AE	-X: +	X-2	Y-1	Y-2	
CHROCCCCALES	• • •	• • •	• • •		
APHANDCAPSA DELICATISSIMA CHEGUCOCCOS EISPERSUS COELOSPERSUM KEUTZINGIANUM	111	N 1 1	111	111	
SIS	1 1 9 9	520	110	175	
MERCSMUPEDIA GLAUCA Merismopedia ténuissima Michocystis afrugindsa	••••	111	mmı	171	
CSCILLATORIALES	•••	••	•••		
ANABAENA SPIRCIDES OSCILLATORIA GEMINATA Spirulina laxissika	· · · · ·	1 4 W	m 20-	1 22 1	
TRICHODESMIUM SP	* * * 1	• • • 1	•	1	
BACILLARIOPHYTA (DIATOMS)	••••			• • • • •	
ACHNANTHES LANCECLATA N DUBIA ACHNANTHES LANCECLATA V DUBIA ANCHOEGNEIS SF	<b>○</b>   m	1 2 5	1 + 1	nın	
ASTEFICHELLA FERMESA ATHEYA ZACHARIAS Gecelneis Sp	4 1 M		115	IMO	
CYCLOTELLA STELLIGERA CYCLOTELLA SP CYMBELLA AFFINIS	™ n i	101	101	141	
CYPBELLA VENTRICCSA Cypbella Tumida Cypbella Sp		111	111	mli	
FRAGILAFIA CAFUCINA Fragilakia Crctcnensis Gyfcsigma SP	-01	115	40M	501	
MELOSIAA AWHICUA MELOSIAA DISTANS MELOSIPA GRANULATA M	16c + 287 + 190 +	137	1000	217	
MELOSINA GPANULATA V ANGUSTISŠIMA Melosira valians Navicula capitata	<u>.</u>	297	111	111	
MANICULA CRYPTCCEPHALA MANICULA RHYNCHCCEPPALA NAVICULA SP 1	• • • • • • • • • • • • • • • • • • •	• • • •	101	111	
NAVICULA SP 2 NAVICULA SP 3 NAVICULA SP 4		• • • •	NM I	111	
AVICULA SP S 112SCHIA ACI 112SCHIA PAL 172SCHIA SP	171 1	000 I	lon i	150	
MITZSCPIA SP 2 RHIZOSOL ENIA SP	• • •	•••	11	10	

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C-2

x-1 x-2 Y-1	111	22 + 30 + 5 - 5 - 5 - 5 - 5 - 5 - 5 - 5 - 5 - 5	52 + + + + + + + + + + + + + + + + + + +	+++	* • • • • • • • •	+ + + + + + + + + + + + + + + + + + +	### ### ###	* * * * * * * * * * * * * * * * * * *	****	· · · · · · · · · · · · · · · · · · ·	1647 + 1870 + 1606 +
	STEPHANDUISCUS ASTRAEA STEPHAN STEPHANCDISCUS ASTRAEA V. MINUTULA + SUFIRELLA SP	SYNEDRA ACTINASTROIDES SYNEUMA DELICATISSIMA SYNECHA HUMPENS	STNEDHA ULNA V DANICA + STNEDHA ULNA V DANICA + TABELLARIA FENSTRATA +	T GUADRISEPTATA +	EUGLENOPHYCEAE +	EUGLENA SP PHACUS SP 1 TRACHELOMONAS CREBEA	THACHELGHCNAS CBLCNGA TRACHELGHCNAS VCLVCCINA TRACHELGHONAS SP II	TRACHELOMONAS SP IV	UNIDENTIFIED TAXA ++++++++++++++++++++++++++++++++++	UNIDENTIFIED SPECIFS J + UNIDENTIFIED SPECIES M + + + + + + + + + + + + + + + + + +	TOTAL NUMBER OF ORGANISMS +

TABLE C-3

TAXONOMIC LIST OF VAAP PHYTOPLANKTON, AUGUST 11, 1975

CAVPTOPHYCEAE	۲-۲	N-1	S	<u>-</u>	E-1.	. [-3	B-1	4
CRYPTCMONAS SP RMCDOMONAS SP	ው <b>ዘ</b> ባ የት ነበ	175	79	31	E 2 E	239	159	159
• • • • • • • • • • • • • • • • • • • •	******	•	*******	‡ •		***	***	*
CHLCROPHYCEAE	**			•			•	
CHLCROCCCCALES					-			
ACANTLOSPHAERA ZACHAPIASI ACTINASTRUM MANTZSCHTI ANKISTHODESMUS FALCATUS	1 1 80	23	# 1 A	143	F15	M 1.7.4	47	2 2 1 8 9 9 9 9 9 9 9 9 9 9 9 9 9 9 9 9 9 9
CCELASTRUM CAMERICUM COLLASTHUM MICECECRUM COELASTRUM PROPOSCIDEUM	161	111	111	111	116	1 10 1	F = F	181
CRUCIGENIA APICULATA CRUCIGENIA TETRAFFETA DICTYOSPHAEPIUM PULCHELLUM	7 1 51	151	ត្	116	111	271	6 1 6 6	~1~
ECFINOSPHAFRELLA LIMMETICA FHANCEIA DROESCHERI GOLEMKINIA PAUCISPINA	111	161	F 1 8	116	166	F112	23 71	1 ~ 5
KIRCHNEWIELLA CCNTCRTA KIRCHNEWIELLA LUNARIS KIRCHNEWIELLA CPESA	1 1 85	111	1 80 EV	1 1 1	£ : 1	1 O 1	1 E	1 1 1
LACERHEIMIA SP Michactinium Pusillum Docystis Glueccystiformis	1 2 2 3 3 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5	1 2 1	31.	151	991	55 159	103	<b>~</b> 6 1
OCCYSTIS SOLITARIA PECIASTRUM CUPIFX PECIASTRUM SIMPLEX	1~1	, 111	1 1 1	511	<b>~ 1 1</b>	115	111	F 1 2
SCENEDESMUS ARUNDANS SCENEDESMUS ACUMINATUS SCENEDESMUS BIJUGA	23.7	1 1 10	1 1 1	<b>~</b> 1 1	211	31.	พ.⊶ พ.พ. เ	<b>₩</b>
SCENEDESMUS DENTICULATUS SCENEDESMUS GLADETCAUDA SCHMGERERTA SETTGERA	- œ I	1 (6) 1	155	161	161	22.1	103	186
TETHADESMUS WISCENSINENS! TETRAEDRUM LUNULA TETRAEGHUM MINIMUM	1 25 =	161	<b>~1 ~</b>	1831	116	23	***	116
TETRAEDHUN REGULAFE Tetraedron trigenum Treubayla setigenum	1 W 1	٠ <u>٠</u> ١	301	118	1 ~ 6 M	1 10 0	~ 6 M	135
TETRASPORALES				14				
ELAMATUTHRIX GELATINOSA SFHAERCCYSTIS SCHROERTERI	11	11	۱ ۸	<b>~</b> I	16	181	1 1	. 1

TABLE C-3 (Continued)

VOL VOC AL ES	Y-1 U-1	- -	S	F-	-3 ·	-5	8-1	Ø
LAMYDCMONAS SP NIUM FUPMOSUM HAERELLOPSIS SP	103	ţ11	295	103	916	271	N   -	327
DESMIDACEAE		•		•	• •	•		•
CLCSTERIUM SP CCSMARIEM PHASEOLUS EUASTRUM SP	167	116	<b>FIF</b>	111	111	1 ~ ~	166	161
STAURASTRUM LEPTECLADIUM ZYGMEMATELES	•	^	^	1	1	•		^
MDUGEOTIA SPP	1	•	•	ı	63	<b>*</b> :	1	•
AAN THOP HYCEAE	•	*			• •	<b>:</b>	•	•
OCY11 DOTET	1.1	1.1	• •	<b>M</b> I	1.1	<b>~</b> I	ı <del>.</del>	11
**************************************	•	• •	• • •	•	•	* •	• • •	<b>:</b>
CHRYSOCDCCUS PLSILLUM Dirllwych Ravaricum Synuka Sp	ŭ11	23.1	85 m e 85 m e	811	m H	<b>8</b>	nno 0 <b>4</b> h	2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2
DINEPHYCFAE	•				•	*	•	* *
CEFATIUM HIPUNDINELLA GLENDGINIUM ACICULIFERUM PEFIDINIUM CUNNINGTONII PEFIDINIUM INCONSPICUUM PEFIDINIUM PUSILLUM	161 M1	H11 PP	111 01	IIP MI	111 16	116 61	111 6m	111 106
************************	***	*****	•	•	•	•	•	:

TABLE C-3 (Continued)

CYANOPHYCE AE	۲-۲	Y-1 U-1 S F-1 (E-1 C-1 B-1	S	F-1	E-1	C-1	B-1	⋖
CMPDCDCCALES								
APPANDCAPSA DELICATISSIMA CHECOCCCUS DISPERSES CHECOCCCCUS MINUTES	922	23	111	6 M I	2 m 2 m	215	191	271 275 155
CUELOSPHAERIUM KEUTZINGIANUM DACTYLOCOCCOPSIS ACICULARIS DACTYLOCCCCOPSIS RAPHIOIOIDES	111	~ I &	116	116	111	811	0 N	15
GCMPHUSPMAERVA LACUSTRIS Marrichiella Sp Mercsmopedia glauca	687	159	181	<b>*</b> 1 1	24	0 0 0 1-	127	7 .
MEFISMOPEDIA TENUISSIMA	1.5	7	•	1	5	31	4.7	39
OSCILLATORIALES								
ANABAENA SP USCILLATORIA GEMINATA SPIRULINA LAXISSIMA	7 615 . 967	615 · 215 967 · 135	247	191.	191.		839, 1119, 1015 351 431 247	1015
TRICHODESMIUM SP	39	^	í	50	•	•	5	23
			****	****	*****	****	****	****

## BAC 11 LARIOPHYTA (DIATONS)

ACPNANTES LANCECLATA V DUBIA Anchofeneis SP Cocconeis SP	151	116	151	161	116	61 0 m	<b>~ 1 ~</b>	31
CYCLUTELLA SP FRAGILAMIA CAPUCINA FRAGILAMIA CRCTONENSIS	215	MK 1	160	5. 8.00 8.00	TH:	127 157	M → W NM →	16.
HELUSIFA AMEJGUA HELUSIFA DISTANS HELUSIFA GRANULATA N	សសា	90 H	191	101	70°1	41 188 188 188	1 25 3 4 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5	255 327 127
MELCSIRA GRANLLATA V ANGUSTISSIMA Navicula Phynchocephala Navicula SP 2		111	1 5 1		1 1 1		121	233
NAVICULA SP 3 NITESCHEA ACTCULARIS NITESCHIA PALFA	116	111	116	187	152	101	103	127
NITZSCHIA SINUATA V TABELLARÍA RHIZUSOLENIA SP STEPPANDDISCUS ASTRAEA V. MINUTULA	116	1 + 1	115	111	115	3.5	- 00 MMM	8 % W B
SYNEDA ACTINASTRCIDES Syneda delicatissima Synedwa kumpens		181	23 65	166	4-N F 10 m	111	116	0 - 0 0 0 0
STHEERA ULM V DANICA	31	~	15	15	•	4	23	23

TABLE C-3 (Continued)

EUGLFNOPMYCEAE	۲-۱	Y-1 U-1 S	S	F-1	F-1 E-1 C-1	C-3	B-1	A
EUGLENA SP Pracus SP I Pracus SP II	<u> </u>	NN1	<b>*</b> **1	*11	111	9 H	* 1 1	# 01
TRACHELCHINAS CREEFA THACHELUMONAS GIBBERUSA TRACHELCHCNAS CRECHGA	211	<b>~11</b>	111	23	<b>FII</b>	111	27.	1 1 1
THACHELOMCNAS VOLVOCINA IRACHELOMONAS SP III TRACHELOMONAS SP IV	- 191	~1~	110	~~!	' 'E	N 1 N	1 2 8	23.1
UNIDENTIFIED TAXA	*	•	*	•	•		•	
UNIDENTIFIED SPECIES M	1	ı	Ë	•	•	^	•	~
***************************************					•	•	•	
UNSPECIFIED TAXON	1	51		•	•	2	•	t
***************************************	•		•	•			•	•
TOTAL NUMBER OF DRGANISMS	3760	1614.	1614, 3076	1518	1518. 1339. 4138	4138	5150	5072
AUMBER OF TAKA	4	4.2	\$2	42	4.0	62	9	•
***************************************	•	•				•		•

U	<b>.</b>	- <del>0</del>			- !	<u>.</u>		
CRYPTCHONAS SP RHIDEMCNAS SP ***********************************	120	128.	128	336	288 · 640 ·	160	56 72	N 0 +
CHLORCPHYCEAE					÷			
CHLCPCCCCALES		•		:	•-			
ACANTEDSPHAERA ZACHABIASI Ankisteudeshus fractus Coclasifum Propuscideum	4 N O	40 B	7 7 9	1 4 1	24.	4 4 ©	841	24
COELASTHUM SPPAERICUM CHUCIGENIA APICULATA CRUCIGENIA GUADRATA	321	1 % 1	001	1 5 1	• • •	211	• • •	4 ® I
CRUCIGENIA TETRAPEDIA Dictrusphaerium Pulchellum Ecminosphaerella Limmetica		••1	4 ii 0 0 8	9 0 0	611	191	111	001
FRANCEIA DEDESCHERI Gelenkinia paucispina Kikchnemitla lunapis	* 1 1	32	481	0 0 0 0 0 0	• • •	101	1 2 2	1 1 10
KIFCHNEMITLA CPESA Lagemeimia suesalsa Lagemmeimia sp	919	32	• • •	101	• • •	111	• • •	
MICHACTINIUM PUSILLUM UUCYSTIS GLOFCCYSTIFORMIS UOCYSTIS SALITARIA	32	488	4 N Ø 4 Ø	100	9''	910	991	8 1 1
PECIASTRUM SIRADIATUM PECIASTRUM CUPLEX PEDIASTRUM SIMPLEX	1120	010	004	110	• • •	110	111	1 1 9
SCENEDE SMUS ABHINDANS SCENEDE SMUS ACLMINATUS SCENEDESMUS BERNARDII	N N	010	201	<b>0</b> 1 1	901	010	• • •	011
SCENEDESHUS ETJUGA SCENEDESHUS DENTICULATUS SCENEDESHUS GUADAICAUDA	<b>0.0</b> 0 <b>1.0</b> 0	010	9 1 9	202	2.0	32 .	& I &	40 1 40
SCENEDESMUS SFPRATUS SCHHUEGERIA SFTIGFPA TETRAEDMUN LUNULA			2 1 6	• • •	111	100	100	188
TETRAELMON MINIMUM TETHAEDMUN REGULAHE TETRAEDREN TRIGENUM	010	000	919	16 32	010	<b>60 1 1</b>	919	Ø 1 Ø
TREUBARIA SETICERUM	2E	32	•	24	1	1	32	9
TETRASFORALES Elakatothrix gelatingsa	•	0	•	•	•		•	91

TABLE C-4 (Continued)	¥	3-1	-	E-1	F-1	S	U-1	۲-۱
CHLAMYDURONAS SP GONIUM FORMOSUM PANDURINA CHARKOMIENS IS	400	336.	326	224	500	208 16	28.	. 1
SPWAERELLOPSIS SP	!	•	•	1	ŧ	ı	80	32
•••••••	•	•			•		•	:
DE SMI DACE AE								
CLOSTERIUM SP COSMARIEP PHASECLUS EUASTHUM SP	989	1 4 4	191	111	£) 40 t	. 6 6 I	1 10 1	141
STAURASTRUM LEPTCCLADIUM	•	•	•	ı	1	•	0	•
				•	i			
MOUGEOTIA SPP	•	•	32	1	72	ī	1	96
***************************	•	•			****	•	•	•
XANTHOPHYCEAE								
OPFICETION CAPITATUM PSEUDGIETRAEDRON NEGLECTUM	, 0 4	1 0	1 9	<b>; ©</b>	1 1	10	1 1	<b>4 1</b>
•••••••••••••••••••••••••••••	*	*				•	•	•
CHRYSOPHYCEAE							and the control of th	
CHROHULINA SP CHRYSCCOCCUS PUSILLUM DINCLHYEN BAVARICUM	328.	32.	100	32	900	900	101	011
DINCHRYON DIVERGENS Synura SP	10	19	1 %	1 1	• •		11	0 1
•		*				•	•	•
DINOPHYCEAE				-				
CERATIUM MIFUNDINFELA Glenguinium aciculiffrum Periuinium cunningtonii	1 1 5	201	110	112	, Ø 6	119	110	
PERIDINIUM INCENSFICUM PERIDINIUM PUSILLUM PERIDINIUM TATRICUM	401	\$*·	911	ØN I	• • •	<b>*</b> 11	×10	211
••••••••	•	*	•		*	*****		• • • • • • • • • • • • • • • • • • • •

TABLE C-4 (Continued)

Y-1	2 1 E	<b>©</b>	240 320 320	•		186	. 102		321	91 0	119		111		. 9	<b>*</b> 01 <b>8</b>
U-1	ŭ ĥ	Ň	23			•	•	* *		6 2		•			N ==	
S	Ø Ø 1	<b>©</b> 11	112	•		110	. •	•	<b>0</b> 1 1	8 9 8 9	336	114	014	111	28 8 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2	9
F-1	W 40 60	<b>.</b> .	0.0	•		1 1 0	•	*	<b>60 ( )</b>	0 1 B	1 80 1	160	1918	110	32	•
E-1	889	<b>©</b>	160	•		118			119	112	240	111	• • •	111	100	91
L-3	N00	32.5	23.0	-,	-	014			3.0	. 56 72	400	218	32.	010	72	10
B-1	0 1 0 N 1 0	200	208. 32.	•		801	004		. 204 204	232	240	* 1 °	200	• • •	114	•
⋖	80 0 N	91	216	Ī		110	256	*	110	Ø I Ø	256 256 266	3 1 1	40	25 24 16	1 + 60	9
CYANDPHYCEAE	CHROCCCCALES APHANUCAPSA DFLICATISSIMA CHEGUCOCCUS DISPERSUS CPECUCCUCCUS MINUTUS	COFLOSPHAFHIUM KFUTZINGIAMUM DACTYLOCCCOPSIS ACTCULARIS DACTYLOCCCOPSIS RAPMIDIGIDES	GCPFMOSPMAERVA LACUSTRIS Mefcsmofedia (lauca Merismopedia tenuissima	MICAUCYSTIS SP	DSCILLATORIALES	ANABAENA SEIROLDES Anabaena Seiroldes Collingo a ceminata	PIRULINA LAXISSIMA RICHODESMIUM SP		ACPHANTPES LANCECLATA V OUBIA Anchulecheis Sp Cocconeis Sp	CYCLETELLA SP Cymbella affinis Fhagilaria cafucina	FRAGILAWIA CROTOWENSIS MELCSISA AMEICUA MELCSIRA DISTANS	MELCSIFA GRANULATA N MELCSIHA VARIANS NAVICULA RPYNCHOCEPHALA	NAVICULA SP 2 NITZSCHIA ACICULARIS NITZSCHIA PALEA	MITZSCHIA SINUATA V TABELLARIA RHIZOSCLENIA SF STEPMANGDISCUS ASTRAEA V. MINUTULA	SYNEDRA ACTINASTROIDES Synedra delicatissima Synedra Rumpfas	SYNEDRA ULNA V DANICA

TABLE C-4 (Continued)

EUGLENDPHYCEAE	ď	8-1	C-1	E-1	F-1	S	U-1	۲-۱
EUCLENA SP PHACUS SP I PHACUS SP I	311	210	<b>4</b> 1 1	011	\$11	861	811	111
PHACLS SF IV Thachelchoras creefa Thachelomchas Gibbfrosa	121	101	121	111	101	960	• • •	111
TRACHELDWORAS CRLOWGA THACHELDWORAS WOLVOCINA TRACHELOWONAS SP IV	1 48	118	112.	1 60	1 80 4	152,	1 : 2	101
TRACMELGHONAS SP VI	1	•	•	•	ı	•	•	•
·•••••••••••••••••••••••••••••••••••••			•		•	•	•	•
UNIDENTIFIED TAXA								
UNIDENTIFIED SPECIES M	•	0	2.	1	2.	8	9	•
······································	•		***	*****	:	•	•	•
UNSPECIFIED IN MASTER FILE								
UNSFECIFIED TAKEN UNSPECIFIED TAKEN	1 1	24	<b>t</b> 1	1 1	1 1	1 1	• •	10
·····			•	•				
TOTAL NUMBER OF ORGANISMS	3928	3928. 4200. 4344	4344	2536	2776	2776, 3088., 2272, 3352	2272	3352
NUMBER OF TAXA	•	9	67	*	•	5.	n •	*

TABLE C-5

VAAP PHYTOPLANKTON SHANNON-WEAVER SPECIES DIVERSITY INDICES, LAKE CHICKAMAUGA TENNESSEE, JUNE, 1975

Station	6/9	6/10	6/13	Mean
A	3.00	3.00	2.82	2.94
B-1	2.98	2.94	2.84	2.92
B-2	2.99	2.97	2.60	2.85
C-1	3.14	3.14	2.72	3.00
C-2	3.11	3.26	2.87	3.08
D-1	2.91	3.18	2.40	2.83
D-2	3.28	2.97	2.62	2.96
E-1	3.88	3.16	2.75	3.26
E-2	3.16	3.14	2.91	3.07
F-1	3.08	2.99	2.82	2.96
F-2	2.94	3.26	3.05	3.07
S	2.36	NM* 2.88 3.13 NM* 2.98 3.17 2.99 3.26 3.16	3.02	2.69
T-1	2.64		2.61	2.71
T-2	2.54		2.73	2.80
U-1	2.54		3.06	2.80
U-2	2.75		2.40	2.71
X-1	2.83		3.10	3.03
X-2	2.95		3.19	3.04
Y-1	2.75		3.25	3.09
Y-2	3.05		3.24	3.15

\*Not measured

TABLE C-6

VAAP PHYTOPLANKTON CELL DENSITIES (CELL/ML)
LAKE CHICKAMAUGA, TENNESSEE, 30NE, 1975

Station	6/9	6/10	6/13	Mean
А	1500	1811	1277	1529
B-1	2886	3342	1594	2607
B-2	3100	2260	2327	2562
C-1	2218	2981	1646	2282
C-2	2471	1545	1 383	1800
D-1	1520	1462	1 789	1590
D-2	1783	1661	1301	1582
E-1	1564	1810	1207	1527
E-2	1511	2177	1261	1650
F-1	1338	1870	849	1352
F-2	1816	1465	1071	1451
S	3251	N/A*	1258	2255
T-1	2175	1930	985	1697
T-2	2062	1428	872	1454
U-1	2565	NM*	1360	1962
U-2	1575	2386	1352	1771
x-1	1450	1757	1647	1618
X-2	1366	1456	1870	1564
Y-1	1306	1261	1666	1411
Y-2	1929	2133	1950	2004

<sup>\*</sup>Not measured

TABLE C-7

VAAP PHYTOPLANKTON, TOTAL NUMBERS OF SPECIES PER STATION, LAKE CHICKAMAUGA, TENNESSEE, JUNE, 1975

Station	6/9	6/10	6/13	Mean
Α	40	49	46	45
B-1	53	59	43	52
B-2	52	55	53 47	53 53
C-1	50	61	47	53
C-2	57	51	51	53 48
D-1	42	58	45 43	48
D-2	42	47	43	44
E-1	38	53	39 53	43
E-2	44	52	53	50
F-1	46	54	38	46
F-2	40	58	47	48
s	42	NM*	48	45
T-1	47	42	38	42
T-?	38	55	37	43
U-1	39	NM*	50	44
U-2	35	51	44	43
X-1	41	49	45	45
X-2	36	48	51	45
Y-1	38	55	53	49
Y-2	55	56	61	57

\*Not measured

TABLE C-8

VAAP PHYTOPLANKTON TOTAL NUMBERS OF SPECIES PER STATION, LAKE CHICKAMAUGA, TENNESSEE, AUGUST 1975

Station	8/11	8/12	8/13	8/14	8/15	Mean
A B1 C1 D2 E1 F1 S T2 U1 X1	64 66 62 65 45 42 52 52 42 76 48	76 77 57	NM* 71 63 73 48 53 52 51 57 71 43	76 61 52	64 68 67 52 47 49 57 66 43 57	64 68 64 68 47 48 54 61 47 61 46

<sup>\*</sup>Not measured

TABLE C-9

VAAP PHYTOPLANKTON CELL DENSITIES (CELLS/ML)
LAKE CHICKAMAUGA, TENNESSEE, AUGUST 1975

Station	8/11	8/12	8/13	8/14	8/15	Mean
A B1 C1 D2 E1 F1 S T2 U1 X1	5072 5150 4138 2186 1339 1518 3076 3226 1614 3524 3760	3239 2 <b>5</b> 76 4471	NM* 7064 5688 4943 2384 3520 3616 2821 2960 4378 3128	2949 4499 4177	3928 4200 4344 2396 2536 2776 3088 4038 2272 3181 3352	4500 5471 4723 3143 2086 2605 3260 3432 2282 3946 3413

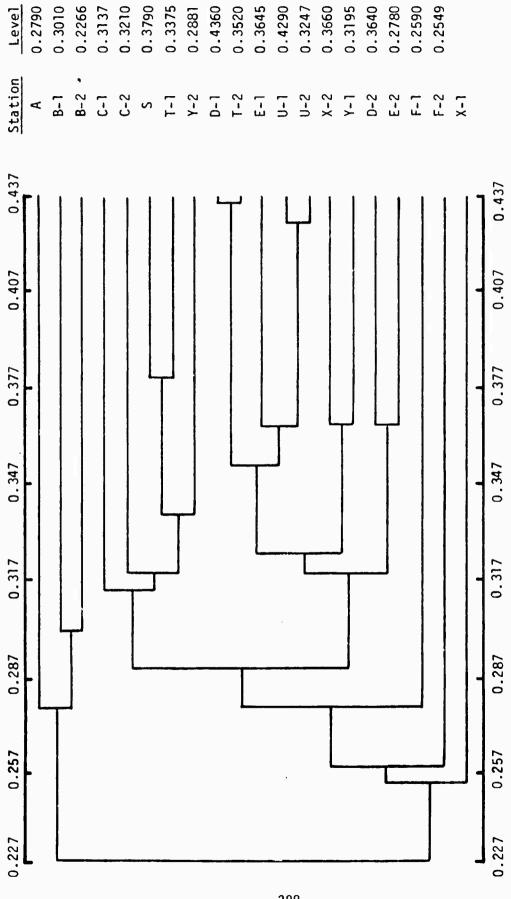
<sup>\*</sup>Not measured

TABLE C-10

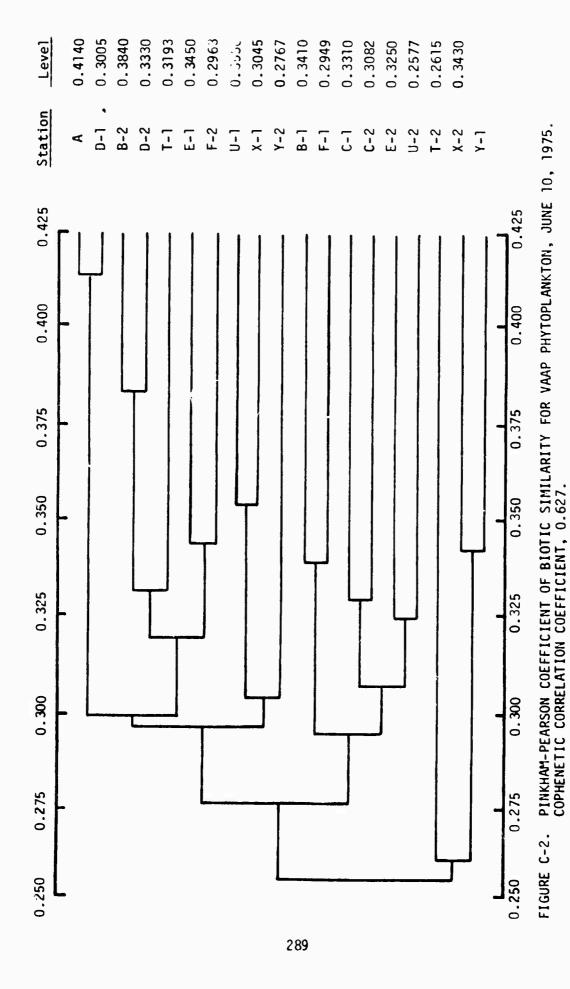
VAAP PHYTOPLANKTON SHANNON-WEAVER SPECIES
DIVERSITY INDICES, LAKE CHICKAMAUGA,
TENNESSEE, AUGUST 1975

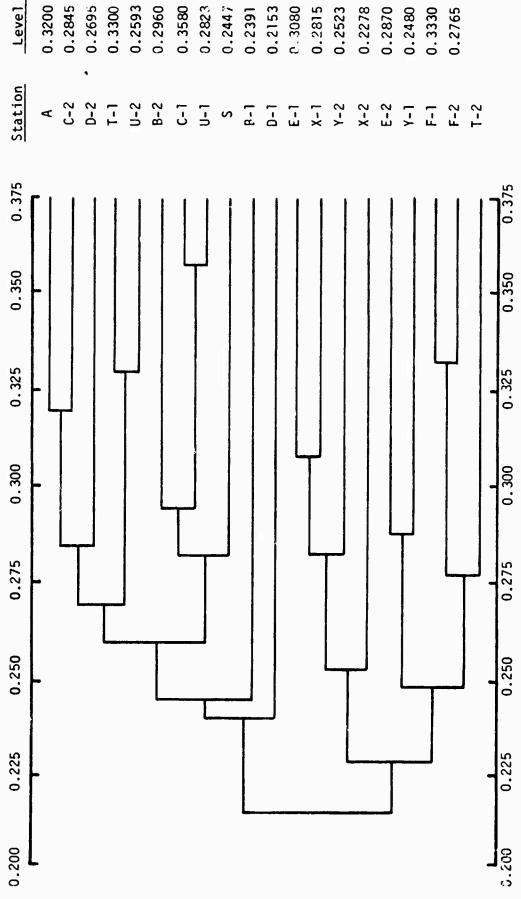
Station	8/11	8/12	8/13	8/14	8/15	Mean
A B1 C1 D2 E1 F1	3.38 3.39 3.19 3.33 3.30 3.15	3.53	NM* 3.46 3.38 3.33 3.00 3.14	3.34	3.37 3.42 3.47 3.20 3.10 2.95	3.38 3.42 3.35 3.35 3.13 3.08
S T2 U1	3.17 3.29 3.11	3.51	3.34 3.20 3.12	3.32	3.26 3.47 3.06	3.26 3.36 3.10
X1 Y1	2.91 2.67	2.92	2.77 2.82	2.08	2.58 2. <b>6</b> 9	2.65 2.73

<sup>\*</sup>Not measured



PINKHAM-PEARSON COEFFICIENT OF BIOTIC SIMILARITY FOR VAAP PHYTOPLANKTON, JUNE 9, 1975. COPHENETIC CORRELATION COEFFICIENT, 0.774. FIGURE C-1.

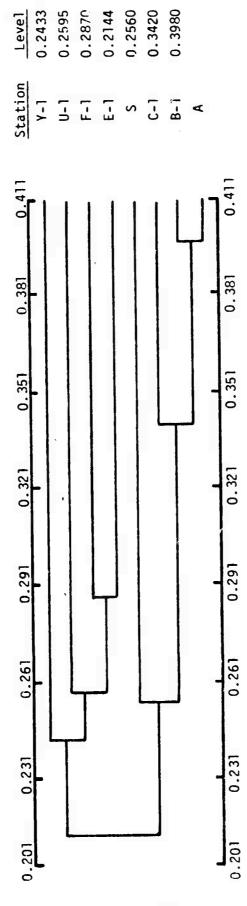




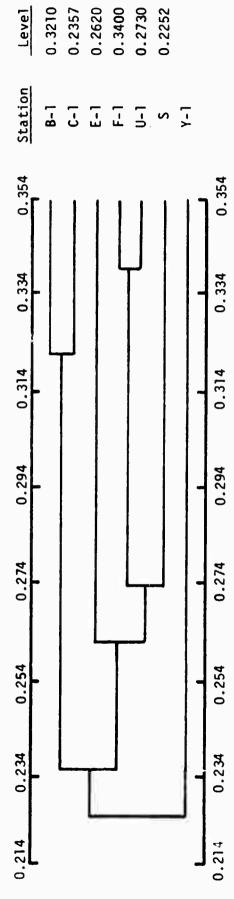
PINKHAM-PEARSON COEFFICIENT OF BIOTIC SIMILARITY FOR VAAP PHYTOPLANKTON, JUNE 13, 1975. COPHEMETIC CORRELATION COEFFICIENT, 0.561.

FIGURE C-3.

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PINKHAM-PEARSON COEFFICIENT OF BIOTIC SIMILARITY FOR VAAP PHYTOPLANKTON, AUGUST 11, 1975. COPHENETIC CORRELATION COEFFICIENT, 0.89. FIGURE C-4.



PINKHAM-PEARSON COEFFICIENT FOR BIOTIC SIMILARITY FOR VAAP PHYTOPLANKTON, AUGUST 13, 1975. COPHEMETIC CORRELATION COEFFICIENT, 0.746. FIGURE C-5.

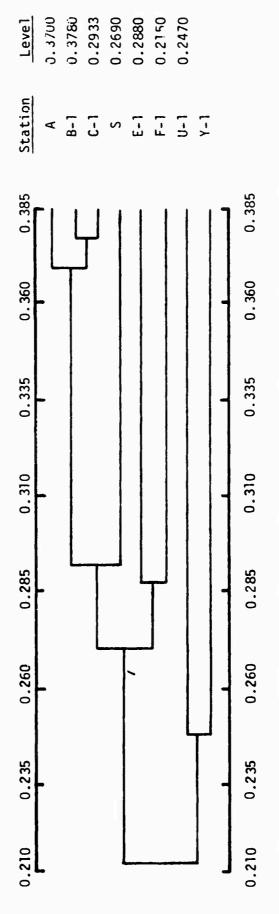


FIGURE C-6. PINKHAM-PEARSON COEFFICIENT OF BIOTIC SIMILARITY FOR VAAP PHYTOPLANKTON, AUGUST 15, 1975.

APPENDIX D
Computational Methods

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#### COMPUTATIONAL METHODS

#### Community Analysis

#### Introduction

Biotic components of water quality are generally quantified by one-dimensional diversity indices when single samples or stations are examined, or two-dimensional coefficients of biotic similarity when sample/sample, station/station, or species/species comparisons are undertaken.

Diversity indices are mathematical expressions that describe the distribution of individuals within the community. There are a number of diversity expressions in use. In general, maximum diversity exists if each individual belongs to a different species. Minimum diversity exists if all individuals belong to the same species. An environmental parameter that influences community structure will also modify the diversity index. In cases where environmental stress may occur (such as competition among species, physiochemical limiting factors, or pollution), the community is reduced in the number of species present. Frequently, this reduction in the number of species is accompanied by an increase in the number of individuals of the remaining species, especially in the case of organic pollution. Environmental stress, therefore, tends to reduce the magnitude of diversity indices. One-dimensional diversity indices include the Shannon-Weaver Species Diversity, Evenness, and Simpson's Index of Dominance.

Coefficients of biotic similarity quantify the taxonomic overlap between two samples or stations. Most of these coefficients assume values between 0 and 1, where a value of 0 indicates no species overlap, and a value of 1 implies identical species composition. Morisita's Index of Faunal Affinity and the Pinkham-Pearson's Index of Biotic Similarity are measures of biotic similarity.

In this study data processing subsequent to manual taxonomic identification/confirmation was executed through the IBM 370/OS system at the Northeast Regional Data Center of the State University System of Florida (NERDC). Diversity indices and coefficients of similarity were calculated by proprietary FORTRAN IV routines. The phenograms were generated through application of the NT-SYS Numerical Taxonomy System developed by Rohlf, Kishpaugh and Kirk at Stony Brook (1974).

# Shannon-Weaver Species Diversity Index (H)

The Shannon Weaver Species Diversity Index,  $\overline{H}_e$  (Odum, 1971) is defined as:

$$H_{e} = \sum_{i=1}^{t} \frac{n_{i}}{N} \ln \frac{n_{i}}{N}$$

where  $n_i$  = total number of organisms persent as species  $\underline{i}$  t $N = \sum_{j=1}^{\infty} n_j$  = total number of organisms present in the sample

t = number of taxa present in the sample

He ranges from a minimum of 0.0, occurring when all organisms belong to the same taxon (no diversity), to a maximum of ln N, occurring where each organism present belongs to a unique taxon (maximum diversity).

The Shannon-Weaver Index is commonly expressed to other logarithmic bases, especially base 2 and base 10, and is easily converted by the following expression:

$$\overline{H}_{base_{\chi}} = \frac{\overline{H}_{e}}{\ln \chi}$$

#### Evenness (e)

If the organisms of a sample are uniformly distributed among the taxa present, the Shannon-Weaver Index assumes the value, In t, a condition of perfect evenness in the apportionment of individuals among species. The Index of Evenness, e (Odum, 1971), expresses the actual Shannon-Weaver Index as a fraction of this "ideal" value:

$$e = \frac{\overline{H}_e}{\ln t}$$
 (defined for t >1)

where  $H_e$  = actual Shannon-Weaver Species Diversity Index

t = number of taxa present in the sample

Evenness ranges from 0.0 (minimum evenness) to 1.0 (pc. fect evenness), and the calculated values are independent of the logarithmic base.

#### Simpson's Index of Dominance

The degree to which numerical dominance of a community is concentrated in one, several, or many species may be quantified by Simpson's Index, c (Odum, 1971):

$$c = \frac{t}{\sum_{i=1}^{n} \left(\frac{n_i}{N}\right)^2}$$

where  $n_i$  = number of individual organisms present as species  $\underline{i}$ 

 $N = \sum_{i=1}^{t} i^{i} = total no.$  of organisms present in the sample. t = number of taxa present in the sample

Simpson's Index ranges from 1/N, occurring when each organism represents a unique species (minimum dominance), to 1.0, occurring when all organisms represent the same single species (maximum dominance). In an evenly-dominated community, Simpson's Index assumes the value, 1/t, where t is the number of taxa observed in a sample --  $\overline{H}$  and e, for such a case, assume respective magnitudes of 1/t and 1/t. Simpson's Index is therefore inversely related to species diversity and evenness.

#### Pearson-Pinkham Index of Biotic Similarity (B)

Each of the previously discussed indices  $(\overline{H}, e, and c)$  quantify community structure with a sacrifice of taxonomic integrity important to paired comparisons between samples or stations. Such indices are incapable of distinguishing samples of similar gross community structure, but unlike taxonomic composition. That is, in computation, the i th species of one sample is not necessarily the same i th species of another sample.

This insensitivity to taxonomic overlap is surmounted by the Pearson-Pinkham Index of Biotic Similarity, B (Pearson and Pinkham, 1974) defined as:

$$B = \frac{1}{t} \sum_{i=1}^{t} \frac{Min (^{n}iA, ^{n}iB)}{Max (n_{iA}, ^{n}iB)}$$

where t = number of taxa considered

 $n_{iA}$  = number of organisms of species <u>i</u> present at Station A  $n_{iB}$  = number of organisms of species <u>i</u> present at Station B Min  $(n_{iA}, n_{iB})$  = the minimum value of the pair:  $n_{iA}, n_{iB}$  Max  $(N_{iA}, n_{iB})$  = the maximum value of the pair:  $n_{iA}, n_{iB}$ 

Biotic similarity is defined only for a paired comparison between two samples or stations. If two samples are characterized by identical taxonomic overlap (all species occur in identical abundance), the calculated index assumes a value of 1.0 (maximum similarity). Two samples possessing no species in common share an index of 0.0 (minimum or no similarity). The number of species considered, t, may include only those species observed in either or both of the two samples, or, if mutual absence is deemed important, may include species not necessarily present in either sample. If mutual absence is considered important, Min (0,0) - 1 and Max (0,0) - 1 in the computation of biotic similarity.

A biotic similarity index, B', between species may be defined on spatial and numerical occurrence by transposition of the axes in the preceding expression of station similarity:

$$B' = \frac{1}{k} \sum_{j=1}^{k} \frac{Min(^{n}j1, ^{n}j2)}{Max(^{n}j1, ^{n}j2)}$$

where: k = number of samples or stations considered

 $n_{j1}$  = number of organisms of species  $\underline{1}$  at Station  $\underline{j}$ 

 $n_{j2}$  = number of organisms of species  $\underline{2}$  at Station  $\underline{j}$ 

Min  $(n_{j1}, n_{j2})$  = the minimum value of the pair:  $n_{j1}, n_{j2}$ 

 $\text{Max} (n_{j1}, n_{j2}) = \text{the maximum value of the pair: } n_{j1}, n_{j2}$ 

This index likewise ranges from 0.0 (minimum similarity) to 1.0 (maximum similarity). B' may possess utility for grouping species according to environmental preference or pollution tolerance—that is, it may delineate "indicator organisms."

#### **Phenograms**

The quantification of similarity between paired stations, samples, or species by any of the previously-defined coefficients of similarity generates a diagonal matrix containing PC unique elements, where PC is calculated from the expression (Pearson and Pinkham, 1974):

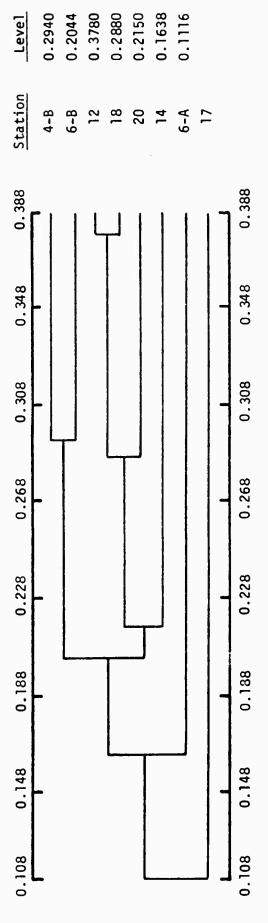
$$PC = \frac{S(S-1)}{2}$$

where: PC = number of unique paired comparisons

S = number of stations, samples, or species being compared

For a study comprising only 25 stations, a similarity matrix of 300 unique elements is produced. Evaluation and presentation of such a voluminous matrix is impractical without computer-aided analysis and graphic models.

Algorithms for clustering similarity matrices into two-dimensional, hierarchic relationships have been developed by numerical taxonomists (Sokal and Sneath, 1963). A technique frequently invoked by ecologists and generally regarded as introducing the least distortion into similarity relationships is the sequential, agglomerative, hierarchic, nonoverlapping clustering method (SAHN) using unweighted pair-groups with arithmetic averaging (UPGMA), described by Sokal and Sneath (1963). The product of this procedure is a branched diagram termed a phenogram (or dendrogram), illustrated below for a study of



PHENOGRAM OF PERIPHYTON ARTIFICIAL SUBSTRATE, JUNE 11-26, 1975. COPHENETIC CORRELATION COEFFICIENT, 0.927. FIGURE D-1.

diatom populations at six stations in the Holston River. This phenogram, like those contained in the current study, was generated directly by computer using the NT-SYS Numerical Taxonomy Package (Rohlf, Kishpaugh, and Kirk, 1974).

The horizontal scale or abscissa of the phenogram is graduated in the units of the similarity measure upon which the clustering was based -- in this case, the Pearson-Pinkham Biotic Similarity Index (mutual absence unimportant). Points of furcation (branching) between the horizontal stems, representing stations or groups of stations imply that the similarity between the two streams is at the coefficient value shown above the branch on the abscissa. The magnitude of similarity between stems is also shown to the right of the phenogram under the column heading, "Level;" these numbers give the exact similarity level at which each stem (station or group of stations) joins the stem below it. Stems are associated with their respective stations by labels to the right under the column heading "Station."

The magnitude of similarity between any two stations represented on the phenogram will, in general, differ from the corresponding magnitude given in the original similarity matrix. This arises as a consequence of the averaging necessary to recursively agglomerate the separate stations into a single, structured set containing all the stations. In the illustrative phenogram, the level of similarity between Stations 6B and 14 is shown to be 0.2044, whereas, in the original similarity matrix (not shown), the magnitude is given as 0.1720. The phenogram value is the arithmetic average of the original similarity indices of Stations 4B and 6B (the cluster containing Station 6B) respectively paired with Stations 12, 18, 20, and 14 (the cluster containing Station 14).

The degree of distortion resulting from the cluster analysis may be quantified by the cophenetic correlation coefficient,  $r_{coph}$ , defined as the product moment correlation coefficient computed between the elements of the original similarity matrix and the corresponding indices implied by the phenogram (Sokal and Sneath, 1963). High values of  $r_{coph}$  ( $r_{coph}$  >0.8 for fewer than 10 stations) indicate that the distortion introduced by the clustering procedure and depicted by the phenogram has not significantly masked the informational content of the original similarity matrix.

## Chemical Water Quality Analysis

## <u>Distance Coefficient</u>

The prior discussion has focused upon the numerical measures generally applied to biologic data. A somewhat more generalized approach to paired comparisons between stations, readily extended to the interpretation of chemical data, is the distance measure. The procedure treats stations as points in an n-dimensional hyperspace, where the n coordinates of a station are the values of the n chemical or chemical/biologic parameters considered. Analogous to the biotic similarity coefficients, a matrix of Euclidean distance coefficients is calculated from the expression (Sokal and Sneath, 1963):

 $\Delta_{AB} = \begin{bmatrix} n \\ \Sigma \\ i=1 \end{bmatrix} (X_{iA} - X_{iB})^2$ 

where  $\Delta_{\mbox{\scriptsize AB}}$  = the Euclidean distance between Stations A and B

n = the number of chemical or chemical/biologic parameters
 considered

 $X_{iA}$  = the magnitude of the <u>i</u> th parameter at Station A

 $X_{iB}$  = the magnitude of the  $\underline{i}$  th parameter at Station B

Clustering is then executed by grouping together station pairs possessing low distance coefficients, that is, stations close to one another in Euclidean hyperspace.

Difficulty in considering parameters of widely different magnitudes and ranges is overcome by normalization of all parameters to standard variables,  $Z_{i,\lambda}$ , with zero mean and unit variance.

$$Z_{iA} = \frac{X_{iA} - \overline{X}_{i}}{S_{i}}$$

where  $Z_{iA}$  = the standardized magnitude of parameter  $\underline{i}$  at Station A

 $X_{iA}$  = the measured magnitude of parameter  $\underline{i}$  at Station A

 $\overline{X}_{i}$  = the mean measured magnitude of parameter  $\underline{i}$  (all stations considered)

 $S_i$  = the standard deviation of parameter i (all stations considered)

In computation,  $\rm Z_{iA}$  and  $\rm Z_{iB}$  respectively replace  $\rm X_{iA}$  and  $\rm X_{iB}$  in the expression for  $\rm \Delta_{AB}$  .

The magnitude of the Euclidean distance,  $\Delta_{AB}$ , increases for any pair of stations as the number of parameters considered is increased. To eliminate this dependence, an average distance,  $\underline{d}_{AB}$ , may be calculated (Sokal and Sneath, 1963):

$$d_{AB} = \sqrt{\frac{2}{\Lambda_{AB}/n}}$$

where  $d_{\mbox{\scriptsize AB}}$  = average distance between Stations A and B

 $\Delta_{AB}$  = Euclidean distance between Stations A and B

n = number of chemical or chemical/biclogic parameters considered. For standardized, independent, normally-distributed parameters, the expected value of  $\underline{d}_{AB}$  converges to  $\sqrt{2}$  as  $\underline{n}$  approaches infinity, (Sokol and Sneath, 1968).

## Biologic Sampling Requirements

Estimation of biologic community structure in natural substrates is confounded by the oft-noted heterogeneity or spatial patchiness of organisms. Sampling of such populations should be conducted so as to provide both an indication of the degree of heterogeneity and some (albeit hypothetical) mean measure of standing crop and structure to allow quantitative comparison of sampling zones or stations.

A biologic community may be considered to possess base population characteristics (density, constituency) governed by gross controlling macrophenonema (i.e. munitions wastes) to which are superposed population variations of lesser magnitude (the apparent random error). The sampling objective is realized when a minimum area or volume is collected such that the error caused by random variations is acceptably small.

In practice, the minimum sampling requirements are generally unknown at the time of collection, unless the investigator has had the benefit of prior studies or preliminary field surveys. If prior information is unavailable, the investigator may choose to bracket the likely requirements and rely upon subsequent detailed laboratory analyses at representative stations to provide that information -- the costs of additional sample collection is usually insignificant relative to the basic expense of a site visit.

One approach to the laboratory determination of minimum sampling requirements is to collect and analyze replicate sets of samples at select stations, each set constituting a unique sampling area or volume. A mean population parameter (diversity, standing crop) may then be plotted against sample area or volume analyzed, bracketed by the calculated standard deviations or confidence limits. Sample size is determined by locating that minimum area or volume where the slope of the plotted data approximates zero and is bracketed by acceptable error limits.

A disadvantage of this procedure is the requirement for collecting and identifying independent replicates of each sample size considered. For an illustrative case, the investigator might collect triplicate sample sets comprised of 1.2, 1.8, 2.3, and 2.9 ft<sup>2</sup> of streambed material if he were studying macrobenthic sampling requirements. These represent a total of 24.6 ft<sup>2</sup> of bottom sediment area and 43 grabs of a 9" x 9½ (60 lb) Ponar dredge or 98 grabs of a 6" x 6" Ekman dredge, both standard benthic samples. Aside from being physically abusive and expending substantial amounts of costly taxonomic identification time, such a program might require disruption of more substrate area than exists in a particular sampling zone.

A modified procedure, applied to this study, utilized the recombination of subsets of the same sample set to estimate mean Shannon-Weaver diversity for any particular sample size. This allowed the determination

of minimum sampling requirements with much greater economy of collection and identification at a sacrifice, however, of precise error limits. For the illustrative case of the prior paragraph, one set of samples totaling 2.9 ft<sup>2</sup> -- 5 hauls of the Ponar or 12 hauls of the petit Ekman dredge could be collected. A plot of mean diversity versus number of dredge hauls (corresponding to varying substrate areas) would be prepared. Mean diversity,  $H_{\nu}$  for  $\chi$  dredge hauls would be calculated as:

$$\overline{H}_{X} = \Sigma H_{X}^{j}$$

where  $k = c(m,x) = \frac{m!}{x!(m-x)!}$  = the number of combinations of m dredge hauls taken x at a time

m = the total number of dredge hauls collected at a sampling site

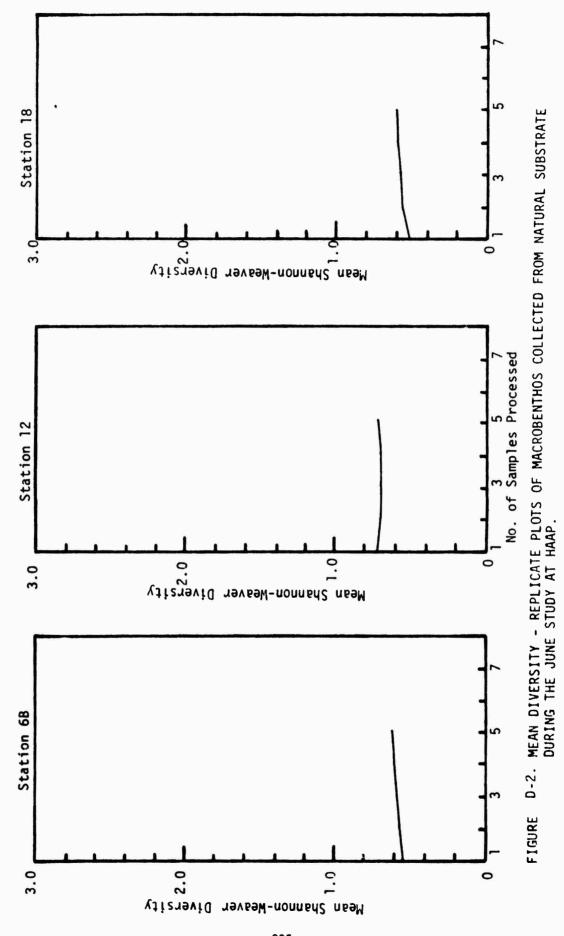
H<sup>j</sup> = the Shannon-Weaver diversity based upon the cumulative taxonomic data of a particular combination, j of x dredge hauls.

Error limits estimated for  $\overline{H}x$  are based upon k-l degrees of freedom. Since k-l approaches zero as x approaches m, the total number of dredge hauls collected should be somewhat greater than the expected minimum number of dredge hauls required to obtain a reasonably constant estimate of the population diversity.

For macrobenthos both artificial and natural substrate samples were taken. The artificial substrate samplers (Hester-Dendy samplers) were disassembled in the field and preserved on a plate by plate basis. Hence, a replicate consisted of a single plate. For three different sampling sites 15 or 16 plates were counted and tabulated. Utilizing a computer to minimize data processing time, combinations of replicates were pooled utilizing 1, 2, 3, etc. total replicates. Mean pooled Shannon-Weaver values are shown in Figure D-2. In all cases it can be seen that Shannon-Weaver values increased as sample size (total number of replicates pooled) increased up to about seven samples. Addition of more samples to the pool beyond that point had little or no effect on the mean Shannon-Weaver value. Based on these results, seven plates (pooled) were considered to be sufficient to obtain a reasonable estimate of the Shannon-Weaver diversity for the remaining sampling sites.

For macrobenthos in natural substrates the identical procedure was utilized to show that five dredge samples would be sufficient (see Figure D-3).

For diatom populations on artificial substrates (glass slides) this procedure showed five slides to be sufficient (see Figure D. 4.)



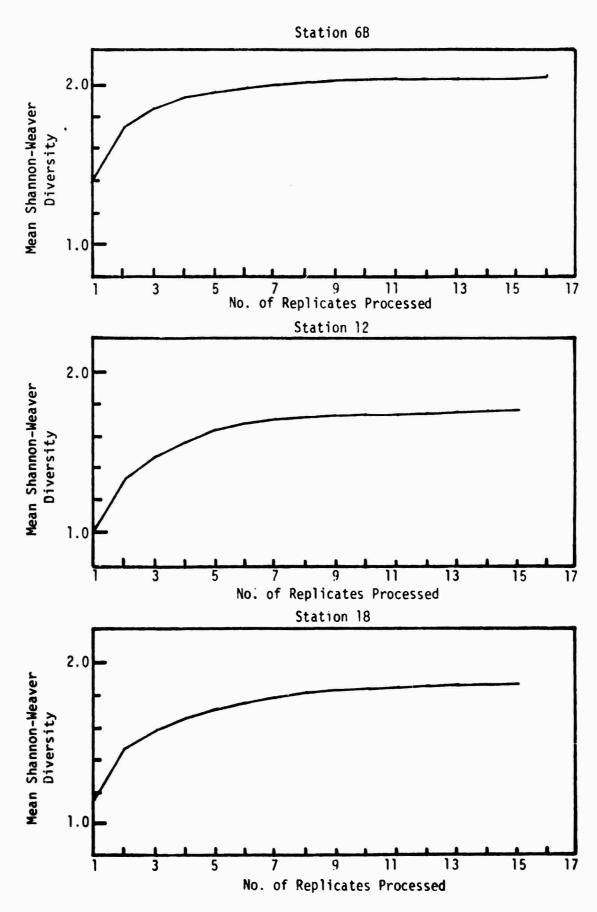


FIGURE D-3. MEAN DIVERSITY - REPLICATE PLOTS OF MACROBENTHOS COLLECTED FROM ARTIFICIAL SUBSTRATE DURING THE JUNE CTUDY AT HAAP.

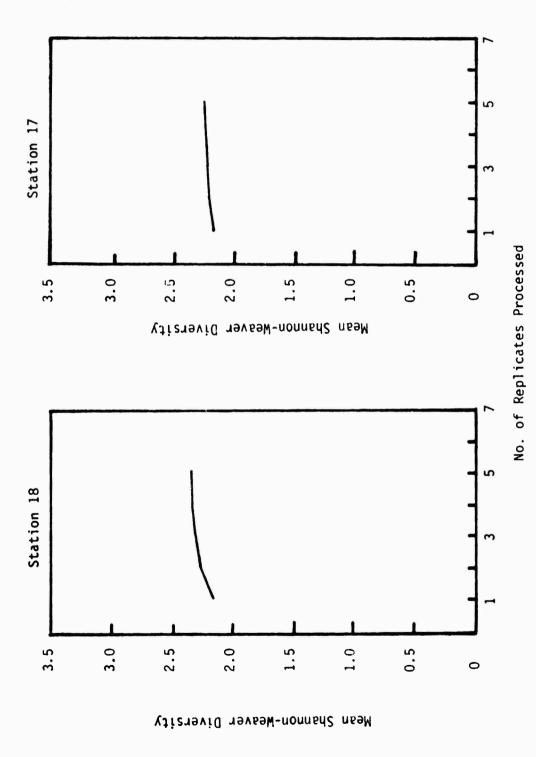


FIGURE D-4. MEAN DIVERSITY - REPLICATE PLOTS OF DIATOMS COLLECTED FROM ARTIFICIAL SUBSTRATE AT THE HAAP SITE AFTER 4-WEEKS INCUBATION - JUNE - JULY 1975

APPENDIX E
SAMPLING MATRICES

# LIST OF TABLES

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E-1	SAMPLING PARAMETERS FOR JUNE SURVEY, VOLUNTEER ARMY AMMUNITION PLANT	310
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E-4	TIME OF DAY SAMPLED - AUGUST TRIP	315

TABLE E-1 SAMPLING PARAMETERS FOR JUNE SURVEY, VOLUNTEER ARMY AMMUNITION PLANT

No Wake	Yes	0 % -	Yes	00-	5 5	co	00	0 0	000	00
Y-2	·	040	Yes	0 m m	Yes	<u></u> 80	v 0	20	₩ W W	00
<u>/-1</u>	Yes	S 22	9 <u></u>	000	Yes	<b>ω</b> ω	വവ	ო	582	00
x-2	Yes	ოსი	Yes	0 m 4	Yes Yes	18	သည	3.5	532	00
x-1	Yes	0.00	8	000	Yes	18 16	5	3.5	535	
n-2	Yes	000	No No	000	Yes	00	വവ	3.5	23.5	00
3	Yes	w <b>4</b> 0	2	000	Yes	12	5.5	5.2	582	00
<b>T-2</b>	Yes	000	2 €	000	Yes	12	9.0	3 2	536	00
1-1	Yes		Yes	285	Yes No	13	22	3.5	582	00
S	Yes	- 22	Yes	388	Yes	18	2.2		592	
F-2	Yes	O 22 M		282	Yes		 	5.2	533	00
II.	Yes	mm0 	2	000	Yes			3.02	23*	
E-2	Yes	00 m			Yes				N 90	
<u>E-1</u>	N O	•••	2		Yes	18	22	ω ω	533	
1D-2	Yes	اد ما ما اد ما ما		000	Yes	18				
0-1	Yes	~~~~ ~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~		288	Yes	18	~~~~~		988	00
2-2	Yes	o	<b>=</b>	000	Yes		<u></u>		577	0.0
2	Yes	 	Yes Yes	264	Yes		<b>₽</b>	 	10 m m	- 2
B-2 C-1	Yes			200	Yes Yes Yes Yes No No			ω m	2 m Q	00
13-1	Yes	~ w ~	e	200		18			900	00
V	Yes	2.50	Yes	ww4	Yes Yes	18	വവ	ທີນ	588	
	PERIPHYTON Artificial Substrate 2-Week Incubation Collected	Analyzed Organisms Biomass Chlorophyll	4-Week Incubation Collected	Analyzed Organisms Biomass Chlorophyll	Natural Substrate Collected Analyzed	MACROINVERTEBRAT_5 Artificial Substrate Collected Analyzed	Matural Substrate Collected Analyzed	PLANKTON Collected Analyzed	CHEMISTRY - WATER Group A Solids Sulfates	Metals A B

\*One munitions sample broken in shipment.

Ţ

TABLE E-1 (CONTINUED)

	A	B-1	B-1 B-2 C-1		2-5	<u></u>	2-0		E-2	1	F-2	S	1-1	T-2 U	ח-ח	U-2 X		X-2 Y-	<del>`</del>	-5	No Wake
CHEMISTRY - SEDIMENTS Nitrite Nitrogen Total P COD		_		-				<del></del>	_		<b>—</b>		_					_			0
Volatile Solids Jotal Solids	2	2	2	2	<del></del>		2		2	2	2	2			_	7			· _		0
Zn, Pb, Cu	2	2	2	2	_	_	2	_	2	2	2	2	_	_	_	2	_	7	_	_	0
	2	2	2	_	_	_	2	_	_	_	_	_	_	_	0	_	_	_	_		0
Nitrate Nitrogen	2	2	2	0	-	_	2	_	_	pr	_	0	_	_	_	_			_		0
Munitions		_		_	_	0	_	0	_	0	_	_	_	0	C	_	_	0	0	0	0
	2	_		0	_	0	0	0	0	0	_	0	0	0	0	0	0	0	_	0	0
Mn, Cd, Fe	_	_	_	_	_	_	Ė	_	_	_	2	_	_	_	_	_	_	2	_	_	0
	_	_		_	_	0	_	0	_	_	_	_	C	0	0	_	0	_	0	0	0
	2	_	_	2	_	_	_	_	_	2	_	_		_	_	_	2	~	_		0

METALS B	ਜੂ <b>ਰ</b>
METALS A	Cd. Cu Ni. Zn Cr
GROUP A	Hitrogen Forms Total Phosphorus COD TOC Munitions Alkalinity Hardness Chloride

SAMPLING PARAMETERS FOR AUGUST SURVEY, VOLUNTEER ARMY AMMUNITION PLANT

	A	B-1	<b>B-</b> 2	0-1	C-2		D-2 E	<u> </u>	-2 F		-2 S		-	-2 -2	<del> </del>	-x -	1 x-2	<u>-</u>	۲-2	No Wa	ške
PERIPHYTON Artificial Substrate									·		<del></del> , -,	······································		······································		· · ·					
2-Week Incubation Collected	Yes	Yes	Yes	Yes	ves.	Yes	Yes	Yes	(es	No	es*	es	es 🔻	es Ye	es	Ye	SYes	s Yes	Yes	Yes	40
Analyzed Organisms Biomass	0 0	~ m	0	0	0	N 60	0 0	0 0	0	00	0 0	7	2 %	28	3.8	3.2	0.5	3 0	3.5	0 0	~ ~
Chlorophyll	0	_	0	_	_	m	0	0	0	0	0	n	0	_	2	0				_	
4-Week Incubation Collected	Yes	Yes	No Yes		Yes	ON ON	Yes	No	es	N <sub>O</sub>	No Y	es Y	es Ye	<u>&gt;</u>	es Ye	Ye	s Yes	 8	Yes	Yes	40
Organisms	2 6	7	00	2	2 5	00	7 0	00	~ <	00	00	2 4	2 0	0,	~ ~	20	-25	-0		· · ·	0:
Chlorophyll	· –	- ;	00	- m	0 <b>4</b>	0	00	00	00	00	00	0.4	2 2		2 0	00			7 m		٠.
Natural Substrate Collected Analyzed	Yes	Yes	Yes	Yes	Yes N	Yes	Yes	Yes Y No	es ≺ No	es ≺ No	es ≺ No	es Y	es Yo	es Ye	es Ye	ss Yes	s Yes o No	Yes No	Yes	0 0 Z Z	0.0
MACROINVERTEBRATES Artificial Substrate Collected Analyzed	18	18	18	18	180	00	18	18	00	18	180	99	18	<u>8</u> 0	18 7		7	8 18	<u> </u>	O O	0.0
Matural Substrate Collected Analyzed	ນນ	ည	0	വവ	0.02	വ	വവ	ഉ	0 0	വവ	0.02	ည	വവ	0.02	0 2	4 25	22		സസ		0.0
PLANKTON Collected Analyzed	2	ကက	5	3.5	0	0	n n	3.5	ω <b>0</b>	ന വ	0	9 8	2 C	2 2	വ വ	0 22	2,7	32	0.0	00	0.0
CHEMISTRY - WATER Group A TOC	5+	დ 4	ზ 4	rv 4	<del>ر</del> 4	+ 4	2 4	ი 4	<b>υ 4</b>	<del>ر</del> 4	+5	r. 4	დ 4	<del>د</del> ک	rv 4	ro <b>4</b>				00	<b>~</b> ~
up B fates Cu,	. m 2 0	753	-123	000	0 0 3	. m 0/-	. 600	000	. m 2 0	. m 0 0	. 620	. 620	0 5 3	. m 2 0	m 2 0	. 620	0 5 3	22 3	000		
Pb, Ni, Zn	2	ຕ້	**0	2	2++	**0	0	0	0	2	0	2	0	0	0	0					_

\*Alternate sampling site "F-Buoy" was utilized for this parameter. \*\*One Hi sample. +One munitions sample broken in shipment. ++No Zn samples.

312

TABLE E-2 (CONTINUED)

	4	8-1	B-1   B-2   C-1		2-5	D-1	D-2 E	E-1	E-2	F-1 F	F-2	S	T-1	T-2 L	1-1	( z-n	x-1	x-2 1	۲۲	Y-2	No Wake
CHEMISTRY - SEDIMENTS Nitrite-Nitrogen Total P																	0			_	c
COD Volatile Solids Total Solids	2	<del></del>	2		- 7			para.									_	<del></del>	• –		0
TKN Nitrate-Nitrogen	7	_	<del></del>		2		2		_			_				_			_		0
flunitions	_	_	_	0	_	0	_	0	0	0		_	0	0	0	0	0	0	0	0	0
Zn, Pb, Cu Mn, Cd, Fe, Cr <sup>+</sup> 6, Ni			0		0		0	_	0		0			0		0	_	0		c.	c
Нд		0	_	0	-0		0	0	0	0	0		_	0		0		0	0	0	0

Mitrogen	Solids
Phosphorus	Chloride
cou Munitions Alkalinity	Hardness

TABLE E-3
TIME OF DAY SAMPLED
JUNE TRIP

Chatian	6 (0 ) 75	Da		6/12/75	1 6/12/75
Station	6/9/75	6/10/75	6/11/75	6/12/75	6/13/75
А	1255	1127	1125	0958	0956
B-1	1315	1142	1144	1019	0944
B-2	1307	1134	1139	1009	0950
C-1	1328	1421	1155	1115	0933
C-2	1337	1433	1115	1124	0927
D-1	1725	1452	1053	1035	0914
D-2	1715	1446	1105	1028	0921
E-1	1 735	1504	1044	1058	0907
E-2	1 745	1511	1037	1049	0900
F-1	1902	1533	1016	1152	0843
F-2	1852	1523	1027	1140	0851
S	1835	0928	1205	0858	1036
T-1	182 <b>4</b>	0953	1219	0914	1023
T-2	1815	0942	1215	0923	1030
U-1	1805	1002	1233	0932	1016
U-2	1755	1100	1227	0943	1011
X-1	1110	1602	0924	1232	0812
X-2	1040	155 <b>3</b>	0914	1220	0800
Y-1	1122	1610	0934	1241	0819
Y-2	1132	1619	0944	1253	0825

TABLE E-4
TIME OF DAY SAMPLED
AUGUST TRIP

Station	8/11/75	8/12/75	Date 8/13/75	8/14/75	8/15/75
А	1005	1413	0845	1050	0920
B-1	1020	1425	0903	1032	0938
B-2	1030	1438	0913	1040	0929
C-1	1055	1458	0935	1015	0955
C-2	1105	1445	0925	1000	0945
D-1	1 335	1108	1155	0940	1013
D-2	1 348	1115	1145	0950	1004
E-1	1405	1055	1130	0928	1032
E-2	1418	1040	1120	0915	1024
F-1	1448	1015	1050	0850	1140
F-2	1433	1028	1105	0905	1130
S	0840	1 300	1035	1110	1118
T-1	0900	1315	1015	1122	1108
T-2	0910	1325	1025	1130	1102
U-1	0927	1335	1000	1150	1052
U-2	0940	1348	0950	1140	1042
X-1	1615	0918	1342	1402	0810
X-2	1555	0905	1330	1350	0825
Y-1	1628	<b>0930</b>	1352	1414	0840
Y-2	1540	0850	1315	1424	0852

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