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| <b>6. AUTHOR(S)</b><br><br>Gary H. Dickinson  |                    |                                |                                   | <b>5d. PROJECT NUMBER</b>                                      |  |
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| <b>13. SUPPLEMENTARY NOTES</b>  |                    |                                |                                   |  |  |
| <b>14. ABSTRACT</b><br>The effect of environment on adhesion and biomineralization was assessed in the barnacle Amphibalanus amphitrite. Adhesive properties were robust to pH, but reduced pH resulted in an increase in shell growth and a decrease in shell toughness. Increased temperature moderated the effect of reduced pH. The response to salinity depended on larval-batch. The project sheds light on fundamental mechanisms of adhesion and biomineralization in barnacles and provides information that may be useful as the Navy assesses conditions that reduce the fouling load. |                    |                                |                                   |  |  |
| <b>15. SUBJECT TERMS</b><br><br>Baranacle, biofouling, adhesion, biomineralization, ocean acidification, shell mechanical properties  |                    |                                |                                   |  |  |
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| <b>a. REPORT</b>  | <b>b. ABSTRACT</b> | <b>c. THIS PAGE</b>            |                                   |  | <b>19b. TELEPHONE NUMBER (Include area code)</b> |

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**14. ABSTRACT.** A brief (approximately 200 words) factual summary of the most significant information.

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## Final Technical Report

**GRANT #:** N00014-14-1-0491

**PRINCIPAL INVESTIGATOR:** Dr. Gary H. Dickinson

**INSTITUTION:** The College of New Jersey, Department of Biology

**GRANT TITLE:** Assessing the effect of environment on barnacle settlement, adhesion and biomineralization

**AWARD PERIOD:** 15 April 2014 - 28 February 2018

**OBJECTIVES:** For barnacles grown under varying environmental conditions:

1. Assess adhesion strength and shell micromechanical properties
2. Characterize glue composition
3. Analyze the chemical composition and structure of shells
4. Quantify cyprid settlement intensity

**APPROACH:** We created conditions that emulated what the barnacle *Amphibalanus amphitrite* might experience in a typical estuary. Barnacle larvae (provided by Dr. Dan Rittschof at the Duke Marine Lab) were settled on T2 silicone coated panels either at Duke or in house at TCNJ. At TCNJ, barnacles were held in small aquaria with seawater equilibrated to the conditions of interest. Each batch of barnacles was exposed to experimental conditions for several months. Throughout experimental exposures, mortality and growth were monitored. At the conclusion of the exposure, adhesion strength of each barnacle was quantified using a digital force gauge, following ASTM standards. Base plate area, shell dimensions, shell mass and tissue mass were quantified. We then used a subset of animals to assess: 1) glue protein abundance and composition; 2) shell mechanical properties; and 3) shell structure and composition.

**ACCOMPLISHMENTS:** Over the award period, we conducted four long-term exposure assays assessing the effects of: 1) salinity alone; 2) pH alone; 3) the combination of pH and temperature; and 4) the combination of pH and salinity over two generations. Each exposure ran for several months and included hundreds of barnacles. We were also able to conduct a number of projects complementary to the award, which advanced our basic understanding of barnacle adhesion and mineralization.

**1) Salinity:** Juvenile barnacles (11 days post metamorphosis) were gradually acclimated to eight target levels of salinity, ranging from 10 – 45 in steps of 5, while keeping pH and temperature as consistent as possible. Barnacles were exposed for 16 weeks to these conditions. Over the course of the experiment, mortality was low and was not significantly affected by salinity, although tissue mass was reduced at salinities above and below 35 (the salinity at the site of breeding stock collection). Sexual maturity (determined by the presence of eggs) was not affected by salinity. At the conclusion of the exposure period, barnacle adhesion strength was assessed in shear following ASTM standards. The effect of salinity on adhesion strength was significant, but non-linear; adhesion strength was significantly lower in barnacles held at 35 & 45 as compared to those at 15. Trends in adhesion strength did not correlate with the presence of soft, thick (“gummy”) glue, and no effect of salinity was found on gummy glue production. SDS-PAGE analysis of glue proteins revealed a 130 kDa protein that was consistently more abundant at 35 ppt, as compared to lower salinity levels. In terms of biomimetalization, salinity did not exert a significant overall effect on shell mass, thickness, micromechanical properties, or base plate area.

**2) pH:** We assessed if a reduction in seawater pH, at levels predicted within the next 200 years would alter physiology, adhesion, and shell formation in *A. amphitrite*. Juvenile barnacles, settled on silicone substrates, were exposed to one of three levels of  $pH_T$ , 8.0, 7.8 or 7.5, for 13 weeks. We found that barnacles were robust to reduced pH, with no effect of pH on physiological metrics (mortality, tissue mass, and the presence of eggs). Likewise, adhesive properties (adhesion strength, morphology of the adhesive plaque, and the composition of glue proteins expressed) were not affected by reduced pH. Shell formation, however, was affected by seawater pH. Shell mass and base plate area were found to be higher in barnacles exposed to even moderately reduced pH; barnacles grown at  $pH_T$  8.0 exhibited approximately 30% lower shell mass and 20% smaller base plate area as compared to barnacles grown at  $pH_T$  7.5 or 7.8. Enhanced growth at reduced pH appears to be driven by the increased size of the calcite crystals that comprise the shell, with calcite crystals within the base plate approximately 95% smaller in barnacles grown at  $pH_T$  8.01 as compared to those grown at pH 7.5, and calcite crystals within the parietal plate approximately 35% smaller in barnacles grown at  $pH_T$  8.0 as compared to those grown at pH 7.5. Despite enhanced growth, mechanical properties of the base plate (but not the parietal plates) were compromised at the lowest pH level. Barnacle base plates at  $pH_T$  7.5 broke more easily and crack propagation, measured through microhardness testing, was significantly affected by seawater pH, suggesting that the toughness of base plates grown at the lowest pH level was reduced. Other shell metrics (plate thickness, relative crystallinity, and atomic disorder) were not affected by seawater pH, but magnesium content of the shells was reduced at pH 7.5. Hence, a reduction in pH resulted in larger barnacles but with base plates that cracked more readily.

**3) Combination of pH and temperature:** We conducted an 18-week multi-stressor experiment to assess the interaction of pH and temperature on barnacle physiology, shell formation and adhesion. Following conservative climate models, we exposed barnacles to ambient (26°) or elevated (30°) temperature at ambient (8.0) or reduced (7.8) pH, levels that are predicted for the year 2100. Mortality was not affected by pH or temperature, but tissue mass was reduced when both stressors (increased temperature and decreased pH) were combined. We found that neither pH nor temperature significantly affected barnacle adhesion strength or glue morphology, and the interaction of pH and temperature was not significant. Proteomic assessments of glue proteins from these barnacles are currently underway at NRL. Base plate area also was not significantly affected by pH or temperature independently, but in this case, the interaction of pH X temperature was significant. This suggests that the barnacle's response to pH in terms of shell growth depends on the exposure temperature. At ambient temperature (26°), base plate area was larger in barnacles grown at reduced pH, consistent with our previous observations. At elevated temperature (30°), however, this trend was absent with no difference in area between the two pH levels. Hence the response of barnacles to pH is temperature dependent. In terms of shell mechanical properties, barnacles were found to be robust to the moderate reduction in pH and increase in temperature assessed. Neither pH nor temperature exerted a significant impact on shell microhardness or crack propagation. Shells from all exposure conditions were structurally similar and in all cases shell plates were comprised of calcite crystallites. The shells were comprised of ~40% calcium by weight, and calcium content did not differ among treatments in either the base or parietal plates. Magnesium comprised ~0.4% of the shells by weight. There was a small, but significant, effect of temperature on both base and parietal plate magnesium content, with higher temperature resulting in slightly higher magnesium content. This minor

difference in magnesium content did not appear to manifest itself in altered mechanical properties.

**4) *Combination of pH and salinity:*** A multigenerational exposure was conducted to assess the interactive effects of pH and salinity over two successive generations. Larvae were collected from wild-type barnacles, juveniles were exposed to conditions of interest for 14 weeks, and then larvae were reared from these animals to produce a second generation. Two pH levels (8.0 and 7.5) and two salinities (35 and 15) were assessed. In the first generation, neither salinity nor pH significantly affected barnacle mortality. Tissue mass, however, was significantly lower in barnacles held at low salinity (at both pH levels) as compared to those at 35 ppt. Reduced tissue mass could indicate altered energy budgets due to salt and water balance. Base plate area was significantly affected by both salinity and pH. The base plate of barnacles at low salinity were significantly smaller, than those at full salinity; barnacles at low pH were significantly larger than those at ambient pH. Increased shell growth at low pH is consistent with previous exposures, suggesting that, at least for *A. amphitrite*, low pH (predicted in coming years) will result in larger barnacles. The strong effect of salinity observed, however, is somewhat surprising, given that in a previous exposure assessing varied salinity alone, salinity did not influence shell growth or properties. Unlike pH, the response to salinity appears to be strongly larval-batch dependent. For this exposure, neither salinity nor pH significantly affected barnacle adhesive properties. Both adhesion strength and the proportion of barnacles expressing atypical “gummy” adhesive varied among panels and was not driven by exposure conditions. The suite of proteins produced by these barnacles also did not differ among treatments.

We successfully obtained nauplii from all seawater conditions, and for all conditions at least some nauplii metamorphosed into cyprids. The viability of the cyprids, however, differed among seawater conditions. Few cyprids successfully settled for barnacles raised at low pH from either salinity and of the cyprids that settled, none of them survived the first week. Although pH did not affect mortality of juvenile barnacles in the first generation, it appears to have a strong impact on the viability of their offspring. Hence only the effect of salinity could be assessed in the second generation. Consistent with generation 1, reduced salinity did not affect adhesive strength or properties, but significantly reduced shell area and mass.

**5) *Other complementary work:***

We completed analyses of phosphoproteins in barnacle glue. Phosphoproteins were identified using phosphoprotein-specific gel staining, Western blotting, and immunohistochemistry. The presence and abundance of phosphoproteins were variable among barnacles, suggesting cyclical changes in the glue secretion overtime. Localization studies showed that phosphoproteins were associated with the capillary glue ducts. Mineralization assays on separated proteins showed that phosphoproteins can induce mineralization. Phosphoproteins form strong ionic bonds with mineral surfaces and could be involved in both the adhesion and biomineralization processes.

A new area of investigation was initiated to assess the effects of hydrophobic coatings on adult barnacle adhesion. Barnacles were reattached to hydrophilic glass, and to glass slides modified with a monolayer of PFOTS or MTS, making them hydrophobic. Surprisingly, barnacles adhered more strongly to the hydrophobic surfaces than to unmodified, hydrophilic glass slides. We attributed this trend to exclusion of water on hydrophobic coatings, which creates a protected environment for glue release thereby enhancing molecular interactions between the glue and the surface. For this project, we also developed a novel assay for assessing the amount of glue

remaining on a surface after barnacle removal. Residual glue after barnacle removal was stained with a general protein stain, the substrate was scanned with an optical scanner, and pixel density was quantified using ImageJ. We found a strong, positive correlation between staining density and adhesion strength. This assay could be useful for experiments where a force gauge might not be available, where the samples do not meet the ASTM standard for measurement of adhesion strength (e.g., if the barnacles are less than 5 mm basal diameter or if the base plate breaks upon removal) or when simply analyzing glue remnants in terms of adhesive characteristics.

Lastly, we leveraged techniques developed through the funded work to assess the effect of environment on shell formation in other crustaceans (specifically juvenile red and blue king crabs). Similar to what was found in *A. amphitrite*, shell mechanical properties were reduced with reduced pH in crabs. Importantly, we found that mechanical properties do not scale directly with mineral abundance, suggesting that the composition and organization of the organic matrix plays a strong role in determining shell mechanical responses.

**CONCLUSIONS:** Variations in environmental parameters (salinity, pH, and temperature) can significantly affect growth and development of the cosmopolitan barnacle *A. amphitrite*. The effect of salinity was strongly larval batch dependent and may vary based on genetics of the breeding population. Adhesive properties were consistently robust to pH, suggesting that the barnacle controls the pH of the adhesive interface, or that the suite of proteins present in the glue is insensitive to pH. Exposure to reduced pH, though, consistently resulted in an increase in shell growth. At least at the lowest pH assessed (7.5) this increase in shell growth was associated with a decrease in shell toughness. Increased temperature moderated the effect of reduced pH, suggesting an interactive effect of these two stressors. It is yet to be determined if such changes would alter the survival of *A. amphitrite* in the field, but changes in the abundance of this ecologically dominant species would undoubtedly affect the composition of biofouling communities.

**SIGNIFICANCE:** Overall, the project sheds light on fundamental mechanisms of adhesion and biomineralization in barnacles. It provides specific information how environmental factors affect barnacle growth and development, which may be useful as the Navy assesses conditions that may be favorable in terms of reducing the fouling load. Importantly, the project provided substantial training to students and young researchers (17 in total), prompting many of them to pursue STEM careers.

**PATENT INFORMATION:** No patents were filed for this project.

**AWARD INFORMATION:**

- Project PI, Gary H. Dickinson, tenured and promoted to Associate Professor.
- Project PI, Gary Dickinson, served as session co-chair, for a Biofouling Session at Aquaculture 2016 (Triennial Meeting of the National Shellfisheries Association), and currently serves as session chair, for the Macro and Integrative Biology Session at ICMCF 2018, June 24-29, 2018.
- Project PI, Gary H. Dickinson, awarded competitive teaching release time for research (“Support of Scholarly Activities” – SOSA) in 2015 and 2017, and competitive internal funding to support summer research students (TCNJ’s MUSE program) in 2017 & 2018 (total funding for these internal award ~\$41,000).

- Undergraduate researcher, Shrey Patel, awarded an NREIP internship to work with Chris Spillman's group at the Naval Research Laboratory; Undergraduate researchers Kyle Siegel and Ahmed Mahmoud accepted to summer research programs at Thomas Jefferson University and Stanford University.
- Undergraduate researcher, Kyle Siegel, accepted to a PhD Program at Northwestern University in Interdisciplinary Biological Sciences; Undergraduate researchers Julian Sison and Aparna Yarram accepted to Medical School.
- Undergraduate student Kyle Siegel received a Phi Kappa Phi Honors Society Faculty-Student Research Award, for his ONR related research (June 2016).
- Technician, Conall McNicholl, accepted to a PhD Program at Florida Atlantic University in Marine Science; Technician, Jessica Nardone, accepted to a Master's Program at Tufts University in Ecology, Evolution, and Behavior

## **PUBLICATIONS and ABSTRACTS** (for total period of grant):

### **A. Refereed Journal Articles:**

1. Nardone JA, Patel S, Siegel KR, Tedesco D, McNicholl CG, O'Malley J, Herrick J, Metzler RA, Orihuela B, Rittschof D, and Dickinson GH. Assessing the impacts of ocean acidification on adhesion and shell formation in the barnacle *Amphibalanus amphitrite*. *Frontiers in Marine Science*. *In review*.
2. Figueroa MA, Schablik JD, Mastroberte M, Singh LJ, and Dickinson GH. 2017. The effect of hydrophobic alkyl silane self-assembled monolayers on adult barnacle adhesion. *Marine Technology Society (MTS) Journal*. 51: 39-48.
3. Coffey WD, Nardone JA, Yarram A, Long WC, Swiney KM, Foy RJ, and Dickinson GH. 2017. Ocean acidification leads to altered micromechanical properties of the mineralized cuticle in juvenile red and blue king crabs. *J. Exper. Marine Biology and Ecology*. 495: 1-12.
4. Dickinson GH, Yang X, Wu F, Orihuela B, Rittschof D and Beniash E. 2016. Localization of phosphoproteins within the barnacle adhesive interface. *Biological Bulletin*. 230: 233-242.
5. Stafslie SJ, Sommer S, Webster DC, Bodkhe R, Pieper R, Daniels J, Wal LV, Callow MC, Callow JA, Ralston E, Swain G, Brewer L, Wendt D, Dickinson GH, Lim CS and Teo SLM. 2016. Comparison of laboratory and field testing performance evaluations of siloxane-polyurethane fouling-release marine coatings. *Biofouling*. 8: 949-968.
6. Lim CS, Dickinson GH, Sommer S, Teo SLM, Bodkhe R, Webster DC and Loo YY. 2015. A small scale waterjet test method for screening novel foul-release coatings *J. Coatings Technology & Research*. 12: 533-542.

### **B. Books or Chapters**

1. Sokolova IM, Matoo OB, Dickinson GH, and Beniash E. Physiological effects of ocean acidification on animal calcifiers. 2016. In: Stressors in the marine environment: physiological and ecological responses and societal implications, Solan M and Whiteley N, Eds. Oxford University Press, Oxford, UK. Pp. 36-55.

### **C. Workshop and Conference Participation/Proceedings**

1. Dickinson GH, Nardone JA, Patel S, Siegel KR, Tedesco D, McNicholl CG, O'Malley J, Herrick J, Metzler RA, Orihuela B, Rittschof D. 2018. Assessing the impacts of ocean acidification on adhesion and shell formation in the barnacle *Amphibalanus amphitrite*. 47th Benthic Ecology Meeting, Corpus Christi, TX.

2. Patel S, Spillman CM, Wang C, Schultzhaus J, Dickinson GH, Orihuela B, Rittschof D. 2018. Building a barnacle cement protein profile using immunohistochemistry. 47th Benthic Ecology Meeting, Corpus Christi, TX.
3. Figueroa M and Dickinson GH. 2017. Characterization of adult barnacle adhesion upon reattachment to hydrophobic surfaces. American Vacuum Society 64th International Symposium, Tampa, FL.
4. Siegel KR, Nardone JA, Tedesco D, Patel S, Karra L, Orihuela B, Rittschof D and Dickinson GH. 2017. Examining the interactive effects of salinity and ocean acidification on the physiology of the barnacle *Amphibalanus amphitrite*. 46th Benthic Ecology Meeting, Myrtle Beach, SC.
5. Tedesco D, Nardone JA, Siegel KR, Patel S, Orihuela B, Rittschof D, Dickinson GH. 2017. Effects of ocean acidification and temperature on biomineralization and adhesion in the barnacle, *Amphibalanus amphitrite*. 46th Benthic Ecology Meeting, Myrtle Beach, SC.
6. Patel S, Nardone JA, Tedesco D, Siegel KR, Orihuela B, Rittschof D and Dickinson GH. 2017. Determining the impacts of ocean acidification on shell mechanics and structure in the barnacle *Amphibalanus amphitrite*. 46th Benthic Ecology Meeting, Myrtle Beach, SC.
7. Dickinson GH, Nardone JA, McNicholl CG, Siegel KR, Tedesco D, Patel S, Orihuela B, and Rittschof D. 2016. Barnacle adhesion and biomineralization in a changing ocean: assessing the effects of seawater salinity and pH. 18th International Congress on Marine Corrosion and Fouling, Toulon, France.
8. Nardone JA, McNicholl CG, Siegel KR, Tedesco D, Orihuela B, Rittschof D and Dickinson GH. 2016. Evaluating the effects of environment on barnacle adhesion and biomineralization. 45th Benthic Ecology Meeting, Portland, ME.
9. Dickinson GH, McNicholl CG, Nardone JA, Siegel KR, Tedesco D, Khan T, Orihuela B, and Rittschof D. 2016. Assessing the effects of salinity and acidification on barnacle adhesion, growth and biomineralization. Aquaculture 2016, Triennial Meeting of the National Shellfisheries Association, 108th Annual Meeting, Las Vegas, NV.
10. Figueroa MA, Shablik J, and Dickinson GH. 2016. Structural effects of hydrophobic coatings on barnacle adhesion. Aquaculture 2016, Triennial Meeting of the National Shellfisheries Association, 108th Annual Meeting, Las Vegas, NV.
11. Siegel KR, McNicholl CG, Tedesco D, Nardone JA, Orihuela B, Rittschof D and Dickinson GH. 2016. Effects of salinity on adhesive and shell properties in the barnacle *Amphibalanus amphitrite*. 45th Benthic Ecology Meeting, Portland, ME.
12. Tedesco D, Nardone JA, Siegel KR, McNicholl CG, Orihuela B, Rittschof D and Dickinson GH. 2016. Ocean acidification affects shell formation but not adhesion in the barnacle *Amphibalanus amphitrite*. 45th Benthic Ecology Meeting, Portland, ME.
13. Dickinson GH, McNicholl CG, Orihuela B, and Rittschof D. 2015. Assessing the effect of salinity on barnacle adhesion and biomineralization. 107th National Shellfisheries Association Annual Meeting, Monterey, CA.
14. McNicholl CG, Orihuela B, Rittschof D, and Dickinson GH. 2015. Assessing the effect of environment on barnacle adhesion and biomineralization. 3rd U.S. Ocean Acidification PI Meeting, Woods Hole, MA.
15. Sison J, Orihuela B, Rittschof D, and Dickinson GH. 2015. Characterizing Key Proteins that Contribute to Barnacle Adhesion. 107th National Shellfisheries Association Annual Meeting, Monterey, CA.



# Assessing the Impacts of Ocean Acidification on Adhesion and Shell Formation in the Barnacle *Amphibalanus amphitrite*

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### **Conflict of interest statement**

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest

### **Author contribution statement**

JN developed the seawater exposure system, collected and compiled data on seawater chemistry, physiology and adhesive metrics, and oversaw daily operations of the exposure. SP assessed shell thickness, mechanical properties, and crystallinity. KS and DT collected data on mortality, egg presence, adhesive properties, and shell mass and area. CM developed a prototype seawater exposure system and initiated a pilot study. JM, JH, and RM prepared shells and collected data on atomic disorder and RM analyzed atomic disorder data. BO and DR provided juvenile barnacles and algae and provided expertise on barnacle growth and assessment throughout the exposure. GD conceived of the experiments, analyzed final datasets, oversaw the experiment, and wrote the manuscript. All authors contributed to editing of the manuscript.

### **Keywords**

Barnacle, Adhesion, Biomineralization, ocean acidification, Shell formation, Mechanical Properties, Amphibalanus amphitrite, Balanus amphitrite

### **Abstract**

Word count: 350

Barnacles are dominant members of marine intertidal communities. Their success depends on firm attachment provided by their proteinaceous adhesive and protection imparted by their calcified shell plates. Little is known about how variations in the environment affect adhesion and shell formation processes in barnacles. Increased levels of atmospheric CO<sub>2</sub> have led to a reduction in the pH of ocean waters (i.e. ocean acidification), a trend that is expected to continue into the future. Here, we assessed if a reduction in seawater pH, at levels predicted within the next 200 years, would alter physiology, adhesion, and shell formation in the cosmopolitan barnacle *Amphibalanus amphitrite*. Juvenile barnacles, settled on silicone substrates, were exposed to one of three levels of pHT, 8.01, 7.78 or 7.50, for 13 weeks. We found that barnacles were robust to reduced pH, with no effect of pH on physiological metrics (mortality, tissue mass, and presence of eggs). Likewise, adhesive properties (adhesion strength and gross morphology of the adhesive plaque) were not affected by reduced pH. Shell formation, however, was affected by seawater pH. Shell mass and base plate area were higher in barnacles exposed to reduced pH; barnacles grown at pHT 8.01 exhibited approximately 30% lower shell mass and 20% smaller base plate area as compared to those at pHT 7.50 or 7.78. Enhanced growth at reduced pH appears to be driven by the increased size of the calcite crystals that comprise the shell. Despite enhanced growth, mechanical properties of the base plate (but not the parietal plates) were compromised at the lowest pH level. Barnacle base plates at pHT 7.5 broke more easily and crack propagation, measured through microhardness testing, was significantly affected by seawater pH. Other shell metrics (plate thickness, relative crystallinity, and atomic disorder) were not affected by seawater pH. Hence, a reduction in pH resulted in larger barnacles but with base plates that would crack more readily. It is yet to be determined if such changes would alter the survival of *A. amphitrite* in the field, but changes in the abundance of this ecologically dominant species would undoubtedly affect the composition of biofouling communities.

### **Ethics statements**

(Authors are required to state the ethical considerations of their study in the manuscript, including for cases where the study was exempt from ethical approval procedures)

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*This study was carried out in accordance with standard procedures for invertebrates.*

*In review*

1 **Assessing the Impacts of Ocean Acidification on Adhesion and Shell Formation in the**  
2 **Barnacle *Amphibalanus amphitrite***

3

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19 mechanical properties, *Amphibalanus amphitrite*, *Balanus amphitrite*

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23 **ABSTRACT**

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26 calcified shell plates. Little is known about how variations in the environment affect adhesion  
27 and shell formation processes in barnacles. Increased levels of atmospheric CO<sub>2</sub> have led to a  
28 reduction in the pH of ocean waters (i.e. ocean acidification), a trend that is expected to continue  
29 into the future. Here, we assessed if a reduction in seawater pH, at levels predicted within the  
30 next 200 years, would alter physiology, adhesion, and shell formation in the cosmopolitan  
31 barnacle *Amphibalanus amphitrite*. Juvenile barnacles, settled on silicone substrates, were  
32 exposed to one of three levels of pH<sub>T</sub>, 8.01, 7.78 or 7.50, for 13 weeks. We found that barnacles  
33 were robust to reduced pH, with no effect of pH on physiological metrics (mortality, tissue mass,  
34 and presence of eggs). Likewise, adhesive properties (adhesion strength and gross morphology of  
35 the adhesive plaque) were not affected by reduced pH. Shell formation, however, was affected  
36 by seawater pH. Shell mass and base plate area were higher in barnacles exposed to reduced pH;  
37 barnacles grown at pH<sub>T</sub> 8.01 exhibited approximately 30% lower shell mass and 20% smaller  
38 base plate area as compared to those at pH<sub>T</sub> 7.50 or 7.78. Enhanced growth at reduced pH  
39 appears to be driven by the increased size of the calcite crystals that comprise the shell. Despite  
40 enhanced growth, mechanical properties of the base plate (but not the parietal plates) were  
41 compromised at the lowest pH level. Barnacle base plates at pH<sub>T</sub> 7.5 broke more easily and crack  
42 propagation, measured through microhardness testing, was significantly affected by seawater pH.  
43 Other shell metrics (plate thickness, relative crystallinity, and atomic disorder) were not affected  
44 by seawater pH. Hence, a reduction in pH resulted in larger barnacles but with base plates that  
45 would crack more readily. It is yet to be determined if such changes would alter the survival of

46 *A. amphitrite* in the field, but changes in the abundance of this ecologically dominant species  
47 would undoubtedly affect the composition of biofouling communities.

48

49

50

In review

## 51 INTRODUCTION

52 Barnacles are dominant members of marine biofouling communities throughout much of the  
53 world's oceans. They settle and tenaciously adhere to nearly any inert surface in the  
54 marine environment, including ship hulls and maritime facilities, and once established can serve  
55 as a substrate for less tenacious species. This results in a tremendous cost burden for Naval and  
56 maritime industries in the form of coating application, cleaning and maintenance, as well as lost  
57 operational time (Callow and Callow, 2011; Schultz et al., 2011).

58  
59 The success of Balanomorpha barnacles depends on the firm attachment provided by their  
60 secreted adhesive and the protection imparted by their heavily calcified outer shell plates. Adult  
61 barnacles adhere using a largely proteinaceous glue which forms adhesive bonds with surfaces  
62 and cures (Walker, 1972; Naldrett, 1993; Kamino et al., 2000; Kamino, 2008). The glue is  
63 comprised of at least ten major proteins, which are thought to play differing but specific roles in  
64 the adhesion process, including displacement of water from the substratum, integrating the  
65 cement with the base plate, adsorption to the substratum, assembly, and curing (Kamino, 2016;  
66 So et al., 2016). Shell calcification in juvenile barnacles occurs soon after metamorphosis  
67 (LeFurgey et al., 1995; Metzler et al. *in prep*). In the genus *Amphibalanus*, barnacles are  
68 protected by six parietal (i.e. lateral) shell plates that sit atop a calcified basal plate (Pitombo,  
69 2004). Shell plates are composed of calcite, held within a matrix of chitin, acidic proteins, and  
70 sulfate-rich polymers (Fernandez et al., 2002; Rodriguez-Navarro et al., 2006; Khalifa et al.,  
71 2011). The base plate, and a narrow uncalcified growth region, are glued to the substrate. The  
72 calcified base plate is composed of layered structures, with grain size of calcite crystallites  
73 increasing with distance from the glue layer (Lewis et al., 2014; De Gregorio et al., 2015).

74  
75 Although we are beginning to understand the biochemical mechanisms involved in barnacle  
76 adhesion and shell formation, relatively little is known about how variations in the environment,  
77 for example in seawater pH, affect barnacle adhesion and shell formation. This topic is  
78 particularly relevant given current and predicted changes in the pH of ocean waters, a process  
79 known as ocean acidification (OA). Resulting from the absorption of CO<sub>2</sub> by the world's oceans,  
80 the pH of global surface waters has decreased by 0.1 pH units since the industrial revolution and  
81 is projected to drop a further 0.3 – 0.5 pH units by the year 2100 (Caldeira and Wickett, 2003;  
82 Doney et al., 2009). Such changes will be more extreme in coastal regions, due to a decreased  
83 buffering capacity of coastal waters and biological CO<sub>2</sub> production (Waldbusser et al., 2011;  
84 Baumann et al., 2015). Effects of reduced seawater pH or associated changes in carbonate  
85 chemistry (i.e. reduced calcium carbonate saturation states) have previously been found to affect  
86 adhesion and shell formation in other marine invertebrates. For example, in marine mussels  
87 reduced pH led to diminished attachment strength and changes in the expression of proteins that  
88 comprise the byssus and adhesive plaques (O'Donnell et al., 2013; Zhao et al., 2017). Alterations  
89 in calcification, growth, and shell properties resulting from decreased pH or calcium carbonate  
90 saturation states have been documented in a broad range of calcifying organisms (Kroeker et al.,  
91 2010; Kroeker et al., 2013).

92  
93 Given the general sensitivity of protein conformation to pH, and evidence of a reduction in  
94 marine mussel attachment with reduced pH, we hypothesized that the barnacle adhesive system  
95 is sensitive to seawater pH. Specifically, we predicted that barnacles grown under different levels  
96 of pH would vary in: 1) adhesive strength, and 2) gross morphology of the adhesive plaque (i.e.



97 whether the adhesive plaque was thin and transparent, or thick and opaque when grown on  
98 silicone: Berglin and Gatenholm, 2003; Wiegemann and Watermann, 2003; Holm et al., 2005).  
99 The effect of reduced pH (to  $\text{pH}_{\text{NBS}} 7.4$ ) on barnacle adhesion strength was tested previously in  
100 the barnacle *Amphibalanus amphitrite* by McDonald et al. (2009), but in this experiment,  
101 barnacles were grown on glass and all barnacles broke upon removal. Therefore, force  
102 recordings reflected mechanical properties of the shell rather than adhesive properties per se.  
103 Here, we assessed if pH affects barnacle adhesion strength when barnacles were grown on  
104 silicone coatings and measured following the ASTM “Standard test method for measurement of  
105 barnacle adhesion strength in shear” (ASTM, 2005). In this test, barnacles whose shells break  
106 during removal are excluded from analysis, and therefore force measurements are solely  
107 dependent on properties of the adhesive bond with the substrate and cohesive properties of the  
108 glue itself.

109

110 Interestingly, McDonald et al. (2009) observed an overall increase in exoskeleton calcification at  
111  $\text{pH}_{\text{NBS}} 7.4$  as compared to the ambient level of 8.2, despite weakened mechanical strength of the  
112 parietal plates in barnacles exposed to low pH. We predicted that alterations in the shell  
113 formation and maintenance processes in barnacles would occur even under more moderately  
114 reduced seawater pH. Specifically, we tested the shell size, mass, and plate thickness of  
115 barnacles exposed to seawater at moderately reduced  $\text{pH}_{\text{T}}$  levels of 7.78 and 7.50, values that are  
116 predicted global averages for oceanic surface waters in the years ~2100 and ~2200, respectively,  
117 and compared this to a current average oceanic level of  $\text{pH}_{\text{T}} 8.01$  (Caldeira and Wickett, 2003;  
118 Doney et al., 2009). Shell mechanical properties were assessed in both the base and parietal  
119 plates using microhardness testing. Further, using a combination of scanning electron

120 microscopy and FTIR spectroscopy, we tested if alterations in calcification or mechanical  
121 properties are driven by changes in ultrastructure, composition, crystallinity, or atomic disorder  
122 of the shell plates.

123  
124 To address these predictions, we exposed juvenile barnacles, *Amphibalanus* (= *Balanus*)  
125 *amphitrite*, to one of three pH<sub>T</sub> levels (8.01, 7.78, or 7.5) for a total of 13 weeks. *A. amphitrite* is  
126 a cosmopolitan intertidal species, inhabiting tropical and semi-tropical waters, and both larvae  
127 and adults of this species have been found to tolerate reduced pH (McDonald et al., 2009).  
128 Therefore, in addition to the adhesion and shell formation metrics described, we also assessed if  
129 pH affects the general physiology of *A. amphitrite*. This was done by monitoring mortality  
130 throughout the exposure period and assessing tissue mass and the presence of eggs at the  
131 conclusion of the experiment. Overall, the goal of this work is to provide a comprehensive  
132 assessment of the effect of seawater pH on *A. amphitrite*; alterations in the presence or  
133 abundance of this common species could alter the composition of intertidal and biofouling  
134 communities.

135

## 136 **MATERIALS AND METHODS**

### 137 **Animal collection, larval culture, and experimental exposure**

138 Barnacle larvae were reared from field-collected adult barnacles following the methods of  
139 Rittschof et al. (1984). Adult *Amphibalanus amphitrite* were collected from dock pilings at the  
140 Duke University Marine Laboratory, Beaufort, NC, USA (34.7° N, 76.7° W). Larval culture was  
141 conducted using aged, natural seawater. Barnacle cyprid larvae were settled on T2 silicone-  
142 coated glass panels (15.2 x 7.6 cm) on July 22, 2015. Panels with settled barnacles were placed

143 individually in 1 L plastic bins (12.5 x 10.5 x 10.5 cm) filled with aged seawater. Barnacles were  
144 fed daily with a mix of *Skeletonema costatum* and *Dunaliella tertiolecta*. At 11 days post-  
145 settlement, barnacles were shipped overnight mail to The College of New Jersey (TCNJ; Ewing,  
146 NJ, USA). Immediately upon arrival, panels with juvenile barnacles were held individually in 1  
147 L plastic bins filled with Artificial Seawater (Instant Ocean, mixed to a salinity of 35). At TCNJ  
148 barnacles were fed daily with *Dunaliella tertiolecta*, and brine shrimp (*Artemia sp.*) were added  
149 to their diet at approximately 15 days post settlement. At approximately 30 days post-settlement,  
150 barnacles were large enough to feed exclusively on brine shrimp and addition of *Dunaliella*  
151 *tertiolecta* was discontinued. Throughout the experiment, a small number of barnacles were  
152 randomly culled to prevent overgrowth.

153  
154 Panels with juvenile barnacles were randomly assigned to one of three target pH<sub>T</sub> treatments,  
155 8.01, 7.78 or 7.50, with a total of 8 panels per pH treatment. Replicate panels were split evenly  
156 between two, replicate 5 gallon glass aquaria per pH treatment. Panels were placed into their  
157 assigned pH treatment on August 19, 2015 (26 days post-settlement). For each aquarium, a  
158 modified polypropylene test tube rack was used to hold the panels in place in an upright position.  
159 To prevent accumulation of organic matter in tanks and to enhance feeding, panels were  
160 removed from the aquaria once a day, 6 days a week, for feeding. Panels were placed  
161 individually in 1 L plastic bins (12.5 x 10.5 x 10.5 cm) that had been filled with seawater from  
162 the specific aquarium from which that panel had been taken. Barnacles were allowed to feed for  
163 approximately an hour per day and were always fed in excess. The presence of dead barnacles  
164 and those that had fallen off of panels was recorded during each feeding period and immediately  
165 removed from the panels. For the first week of the exposure, when barnacles were fed with both

166 algae and brine shrimp, water contained in feeding bins was discarded after feeding and this  
167 volume (~ 4 L) was replaced with freshly mixed seawater. Once barnacles were feeding  
168 exclusively on brine shrimp, seawater from two of the four bins per aquarium was run through a  
169 180 µm mesh filter to remove brine shrimp and particulate matter and added back to the  
170 respective tank. Seawater from the other two bins per aquarium was discarded and this volume  
171 (~ 2 L) was replaced with freshly mixed seawater. In all cases, new artificial seawater was added  
172 at the beginning of the feeding interval, to enable equilibration of seawater to the pH treatment.

173  
174 Throughout the experimental exposure aquarium water was tested weekly for ammonia, nitrite  
175 and nitrate using a saltwater aquarium test kit (API Saltwater Master Kit, Mars Fishcare North  
176 America). Levels of all nitrogen cycle components remained at negligible levels (detected at 0  
177 ppm) through the first 12 weeks of the experimental exposure. Within the last week of the  
178 experiment, nitrate levels rose slightly in most aquaria. At this point, half of the water from each  
179 aquarium was replaced with freshly mixed seawater, and after every remaining feeding, water  
180 from all four bins per aquaria was discarded and replaced with fresh seawater. The experimental  
181 exposure lasted for a total 13 weeks (91 days).

182

### 183 **Adjustment and monitoring of seawater conditions**

184 Experimental exposures at TCNJ were run in artificial seawater (Instant Ocean), mixed to a  
185 salinity of 35. Seawater was prepared in 32 gallon LLDPE (linear low density polyethylene)  
186 garbage cans, and continuous mixing was accomplished using a 1300 GPH aquarium pump  
187 (Marineland Maxi-Jet Pro water pump). Since Instant Ocean is formulated with total alkalinity  
188 (TA) above what is found in natural seawater, TA was reduced to ~2200 µmol kg<sup>-1</sup> SW by

189 addition of 12 M HCl (Lunden et al., 2014). The value of  $\sim 2200 \mu\text{mol kg}^{-1}$  SW was chosen to  
190 approximate typical TA values in Beaufort, NC where the barnacle broodstock was collected.  
191 The TA of artificial seawater was measured following SOP 3b (Dickson et al., 2007) on an  
192 automated titrator (Hanna Instruments, HI902) with 0.1 M HCl (Fluke #35335, certified  
193 volumetric) as a titrant, and values were checked against certified reference material from the  
194 Dickson Laboratory (Scripps Institution of Oceanography, La Jolla, CA). All TA samples were  
195 run at least in duplicate.

196  
197 Target pH and temperature levels were achieved using an automated aquarium control system  
198 (Apex AquaController, Neptune Systems), which functioned as both a pHstat and thermostat.  
199 Temperature was held at  $25^{\circ}\text{C}$  for all aquaria. Each aquarium was equipped with a temperature  
200 probe (Neptune Systems, Extended Life Temperature Probe) and a 50-watt submersible  
201 aquarium heater (Aqueon 06105 Pro). Heaters were activated through the aquarium control  
202 system when aquarium temperature dropped below the  $25^{\circ}\text{C}$  set point. Room temperature was  
203 always kept well below  $25^{\circ}\text{C}$ , preventing water temperature from rising substantially above this  
204 level. Each tank was also equipped with a pH probe (Neptune Systems, Lab Grade pH Probe).  
205 Seawater pH was brought to the set point for each aquarium (8.01, 7.78, or 7.50) by addition of  
206 pure  $\text{CO}_2$  gas (AirGas, food grade) or  $\text{CO}_2$ -free air. A  $\text{CO}_2$  regulator with an electronic solenoid  
207 (Azoo  $\text{CO}_2$  Pressure Regulator) enabled the aquarium control system to release  $\text{CO}_2$  when pH  
208 exceeded the set point.  $\text{CO}_2$ -free air was generated using an in-line soda lime column attached to  
209 a 1 L/min aquarium air pump (JW Pet Company, Fusion Air Pump 200) and was added to tanks  
210 when pH was below the set point. Both  $\text{CO}_2$  and  $\text{CO}_2$ -free air lines were split into two lines,  
211 which entered the water at opposite ends of the aquarium and each terminated with an airstone.

212 Water within each aquarium was continuously circulated using an 80 GPH submersible aquarium  
213 pump (Patuoxun 80 GPH Submersible Pump).

214  
215 Seawater conditions (salinity,  $\text{pH}_T$ , and temperature) were measured 6 days per week using a  
216 handheld multiparameter meter (YSI, Professional Plus). Measurements were taken each day  
217 immediately before removing barnacles for feeding. The multiparameter meter was calibrated  
218 monthly for salinity using conductivity standards (#3168 & 3169, YSI). pH was calibrated  
219 weekly against TRIS and AMP buffers, enabling determination of  $\text{pH}_T$  following SOP 6a  
220 (Dickson et al., 2007). A summary of seawater conditions is provided in Table 1 and reflects  
221 multiparameter meter readings. TA was measured weekly as described above. TA measurements  
222 were taken immediately after water collection from aquaria. Other components of the carbonate  
223 system were calculated using CO2Calc (Robbins et al., 2010). Calculations incorporated the  $\text{CO}_2$   
224 constant of Lueker et al. (2000),  $\text{KSO}_4^-$  constant of Dickson (1990), total boron of Lee et al.  
225 (2010), and air-sea flux of Wanninkhof (1992).

226  
227 To ensure consistency of temperature and pH monitoring among replicate aquaria, the aquarium  
228 control system was calibrated against the multiparameter meter readings. Temperature values  
229 were consistent to within  $0.05^\circ\text{C}$  among replicate tanks. If necessary, an offset was programmed  
230 for individual aquaria using the aquarium control system's software to ensure that control system  
231 readings matched that of the multiparameter meter. The aquarium control system would not  
232 allow calibration of pH probes with TRIS and AMP buffers. Instead, NBS buffers were used  
233 (#3821, 3822, & 3823, YSI) and an offset was programmed for each aquarium to ensure that  
234 aquarium control system readings matched  $\text{pH}_T$  readings of the multiparameter meter.

235

236 **Adhesion and growth assessments**

237 Adhesion strength in shear (critical shear stress) was measured in all barnacles after 13 weeks  
238 exposure. Adhesion testing followed ASTM D 5618-94 (ASTM, 2005). Barnacles were removed  
239 from panels using a hand-held digital force gauge (Shimpo, FGE-5X). Force values were  
240 discarded if damage to the barnacle shell or to the silicone panel occurred during removal. To  
241 enable determination of base plate area, base plate diameter was measured on each barnacle in  
242 two dimensions (along and perpendicular to the operculum) using a digital caliper. Removal  
243 force values were normalized to base plate area. The height of each barnacle was also measured  
244 with a digital caliper as the perpendicular distance from the bottom of the base plate to the top of  
245 the highest parietal plate. Lastly, the presence or absence of gummy, opaque glue (as described  
246 in Berglin and Gatenholm, 2003; Wiegemann and Watermann, 2003; Holm et al., 2005) was  
247 recorded for each barnacle.

248

249 Following assessments, barnacles from each panel were placed in shallow glass finger bowls  
250 with seawater taken from the aquarium in which that panel had been held. Within 72 hours of  
251 removal, barnacles were individually removed from water and dissected to remove the soft body  
252 from the shell plates. Opercular plates were removed from the soft tissue and were not included  
253 in subsequent analyses. Dissected barnacle bodies were placed individually on pre-weighed  
254 pieces of weigh paper, allowed to dry for 48 hours at room temperature and then dried in a  
255 vacuum oven at 45°C, 25 in. Hg. Dried barnacle bodies were weighed individually on an  
256 analytical balance with 0.02 mg precision (Metler-Toledo, XSE105DU) to determine tissue  
257 mass. During dissections, the presence or absence of eggs within the mantle cavity was recorded.

258 Eggs, if present, were removed from the mantle cavity but were not included in tissue mass  
259 measurements. Remaining barnacle shells (base and all parietal plates) were then cleaned  
260 thoroughly with water to ensure all tissue had been removed. Shells were dried overnight at room  
261 temperature, then dried in a vacuum oven at 45°C, 25 in Hg, for 24 hours, and then weighed  
262 individually on an analytical balance with 0.02 mg precision. Visibly damaged shells were not  
263 included in shell mass measurements.

264

### 265 **Structural and mechanical assessments**

266 All barnacle shells were inspected after shell mass measurements using a stereomicroscope  
267 (Leica, S8Apo) and any damage to the base plate that had not been observed by eye (e.g. micro-  
268 scale cracks or holes) was recorded. Two undamaged barnacle shells per panel (16 per pH  
269 treatment) were then randomly selected for structural and micromechanical assessments. These  
270 assessments required embedding and polishing of the shell as shown in Fig. 1. Individual shells  
271 were secured to the bottom of a 1.25 in. mounting cup (#198-10000, Allied High Tech Products)  
272 using superglue (Loctite, Control Gel). Shells were oriented with the opercular axis parallel to,  
273 and the base plate perpendicular to the bottom of the mounting cup. This orientation enabled a  
274 cross-section of both the base and parietal plates to be produced as successive layers of epoxy  
275 were removed during grinding and polishing. Mounting cups were filled with a two-part epoxy  
276 (#145-20000, Allied High Tech Products), which cured overnight. Embedded samples were  
277 ground to the mid-line of the barnacle on a manual grinding/polishing machine (Allied High  
278 Tech Products, M-Prep 5). Grinding and polishing followed the procedure described in Coffey et  
279 al. (Coffey et al., 2017).

280



281 Thickness of barnacle shell plates was measured on polished shell cross-sections using an  
282 upright reflected light microscope (Zeiss AxioScope.A1) equipped with a digital camera (Zeiss  
283 AxioCam 105 color). Images were taken of the entire shell cross-section at 2.5 X magnification.  
284 Typically, this required 2-4 images per sample, depending on the size of the barnacle. Using the  
285 camera's analysis software (Zeiss Zen 2), a 100 x 100  $\mu\text{m}$  grid was placed on each image.  
286 Thickness of the base or parietal plate was measured at each point the grid crossed the plate,  
287 resulting in about 40 measurements per shell plate. For the base plate, replicate measurements  
288 were averaged to determine the mean base plate thickness for each sample. At least two separate  
289 parietal plates were visible in each sample, and measurements were taken within all plates.  
290 Replicate measurements were combined among parietal plates, and values were averaged to  
291 determine the mean parietal plate thickness for each sample.

292  
293 Micromechanical properties were quantified using a microindentation hardness tester (Mitutoyo  
294 HM-200). Indents were made at 20 g load, 5 seconds dwell time. Testing was conducted in the  
295 base and one of the parietal plates, with 10 replicate indents made in each plate. For the base  
296 plate, 5 indents were made on each side of the plate, spaced about 200  $\mu\text{m}$  apart, starting at  
297 approximately 500  $\mu\text{m}$  from the distal edge of the plate on each of the two sides (Fig. 1A).  
298 Replicate indents were typically made throughout the length of a single parietal plate. Individual  
299 indents were measured directly on the hardness tester at 50 X magnification in two dimensions,  
300 and Vickers microhardness values were automatically calculated. An image of each indent was  
301 taken on the hardness tester using a digital microscope camera (Moticam 2.0MP), which enabled  
302 quantification of crack propagation (Fig. 1B). Crack length was determined as the radius of a  
303 circle emanating from the center of the indent and encompassing all visible cracks. Replicate

304 measurements within a shell plate were averaged to determine the mean microhardness and crack  
305 propagation for each shell plate.

306

### 307 **SEM imaging, elemental analysis, and calcite crystal area assessment**

308 After micromechanical testing, scanning electron microscopy (SEM) imaging was conducted on  
309 all polished shell cross-sections. Imaging was conducted on uncoated samples, at low vacuum  
310 (50 Pa), in back-scattered electron mode on a field emission SEM (Hitachi America, SU5000).

311 An accelerating voltage of 15 kV was used at a working distance of approximately 8  $\mu\text{m}$ . For  
312 each sample, images were taken within two separate regions of the base plate, and two regions of  
313 one of the parietal plates. In all cases, images were taken in close proximity (typically within 100  
314  $\mu\text{m}$ ) to the indents made during micromechanical testing. Images were taken at 1,000 and 5,000  
315 X magnification.

316

317 Elemental analysis was conducted on all samples at 1,000X magnification using an EDAX EDS  
318 detector (AMTEK Materials Analysis Division, Octane Plus). Imaging conditions resulted in a  
319 count rate of 5,000-10,000 counts per second. For each region, a total of five point spectra were  
320 taken across the region. Replicate spectra were averaged within and between the two regions per  
321 plate, to determine the mean elemental composition for each shell plate.

322

323 Individual calcite crystals were readily resolved at 5000 X magnification (Fig. 1C). The area of  
324 individual calcite crystals was determined on SEM images using image analysis software  
325 (ImageJ, Ver. 1.49). For each image, a total of 15 different calcite crystals were randomly  
326 selected for area determination. For selection, a 2  $\mu\text{m}^2$  grid was first placed on the image. Pairs

327 of random numbers were generated using a random number generator, and each pair was used as  
328 coordinates to identify a specific calcite crystal on the image. The perimeter of each identified  
329 crystal was traced by hand using the polygon tool in ImageJ and area within the traced region  
330 was automatically determined. One SEM image per shell plate was assessed, and replicate area  
331 measurements within the image were averaged to provide a mean crystal area for each shell  
332 plate.

333

### 334 **FTIR analysis**

335 FTIR spectroscopy was used to assess: 1) the polymorph of calcium carbonate present in  
336 barnacle shells; 2) relative crystallinity of shells and; 3) atomic disorder of shells. Spectra were  
337 collected for polymorph and relative crystallinity determination using a PerkinElmer Spectrum  
338 Two spectrometer and for atomic disorder using a PerkinElmer Spectrum 100 spectrometer. In  
339 all cases, individual barnacle shells that had been cleaned of soft tissue were powdered using a  
340 mortar and pestle. For these assessments, the base and all parietal plates were included for each  
341 sample, but the opercular plates had been removed. Powdered sample was placed directly on the  
342 instrument's ATR crystal and compressed to a uniform force with a built-in anvil. Spectra were  
343 taken at 4 wavenumbers resolution, with 32 scans per sample. Spectra were normalized and  
344 baseline corrected within the 600–2000  $\text{cm}^{-1}$  region. The ratio of  $\nu_2$  to  $\nu_4$  peak absorbance was  
345 used as a measure of crystallinity of the shells (Beniash et al., 1997). For atomic disorder  
346 assessments, each sample was ground into a coarse powder for the first spectrum. Following the  
347 first spectrum the sample was then ground with the mortar and pestle to make a slightly finer  
348 powder, and a second spectrum was acquired. This process was carried out 5-8 times, depending  
349 on how much the spectrum changed after each subsequent grind. Each spectrum was ATR

350 corrected and background removed before peak height measurements were acquired following  
351 Regev et al. (2010).

352

### 353 **Statistical analysis**

354 Statistical analyses were conducted using SPSS (V. 23, IBM Analytics). Categorical data  
355 (mortality, presence of eggs, expression of gummy glue) were assessed using chi-square tests.  
356 All other data were assessed using one-way ANOVA followed by Tukey HSD post-hoc testing.  
357 Prior to analyses, outliers were calculated for all metrics as values greater than three times the  
358 interquartile range above or below the third or first quartile, respectively, and were removed  
359 from the dataset. Following removal of outliers, assumptions of normality and equal variance  
360 were assessed using Shapiro-Wilk and Levene tests, respectively, and data were log transformed  
361 if necessary to meet these assumptions. If assumptions of normality or equal variance could not  
362 be met after log transforming data, a non-parametric Kruskal-Wallis test was used in place of the  
363 parametric ANOVA. In all cases, individual barnacles within a pH treatment were pooled among  
364 panels and tanks and treated as individual replicates. Testing for both panel and tank effects was  
365 conducted for metrics yielding a significant response to pH. No significant panel or tank effects  
366 were found.

367

## 368 **RESULTS**

### 369 **Seawater chemistry**

370 A summary of seawater chemistry over the 13 week exposure is provided in Table 1. pH targets  
371 were met in all treatments throughout the duration of the exposure. As expected, pCO<sub>2</sub> increased

372 with decreasing pH. Seawater remained supersaturated with respect to calcite for all treatments.  
373 Total alkalinity tended to be higher in pH 7.50 aquaria as compared to those at pH 7.78 and 8.01.

374

### 375 **Physiology**

376 Mortality of barnacles was low and variable throughout the experimental exposure and was not  
377 significantly influenced by pH treatment (chi-square:  $p > 0.05$ ). Cumulative mortality was 21.0,  
378 25.3 and 14.7% for the 7.50, 7.78, and 8.01 pH treatments, respectively. The value for the 8.01  
379 treatment excludes a single panel in which 85% of barnacles died during the third week of the  
380 exposure. The reason for this die-off is unknown, but given that this level of mortality was not  
381 observed in any of the other panels, it is unlikely that this was driven by pH treatment.

382

383 Barnacles were dissected at the conclusion of the experiment, enabling quantification of soft  
384 body tissue mass and identification of eggs within the mantle cavity. The effect of pH treatment  
385 on tissue dry mass was not significant (Table 2; Fig. 2). At the conclusion of the experiment,  
386 nearly all barnacles had eggs within the mantle cavity (96.4, 95.2 and 94.6% of barnacles in the  
387 7.50, 7.78, and 8.01 pH treatments, respectively); the proportion of ovigerous barnacles was not  
388 affected by pH treatment (chi-square:  $p > 0.05$ ).

389

### 390 **Adhesion**

391 Adhesion strength, measured as critical shear force, was assessed after the 13 week exposure.  
392 Adhesion strength was not affected by pH treatment (Table 2; Fig. 3). Likewise, the proportion  
393 of barnacles expressing opaque, gummy glue was not affected by pH treatment (chi-square:  $p >$

394 0.05; Fig. 3). Gummy glue was found on the base of most barnacles, with 72.9, 80.8 and 62.9%  
395 of barnacles exhibiting gummy glue in the 7.50, 7.78, and 8.01 pH treatments, respectively.

396

### 397 **Shell formation**

398 Assessments of shell growth and materials properties were conducted following adhesion assays.

399 Exposure pH was found to significantly affect barnacle shell mass and the area of the base plate

400 (Table 2; Figs. 2 & 4). Barnacles grown at pH 8.01 exhibited approximately 30% lower shell

401 mass and 20% smaller base plate area as compared to barnacles grown at pH 7.50 or 7.78. Height

402 of the barnacles, measured from the base plate to the highest parietal plate, was not affected by

403 exposure pH (Table 2; Figs. 4). Shell thickness was measured in both the base and parietal plates

404 on polished shell cross-sections (see Fig. 1C). Exposure pH did not significantly influence

405 thickness in either the base or parietal plates. SEM imaging of polished cross-sections revealed a

406 cobblestone like composite of calcite crystals (Fig. 1). Exposure pH was found to significantly

407 alter calcite crystal area in both the base and parietal plates (Table 2; Fig. 4). In the base plate,

408 calcite crystals were approximately 95% smaller in barnacles grown at pH 8.01 as compared to

409 barnacles grown at pH 7.50 or 7.78. The difference in calcite crystal area was less pronounced in

410 the parietal plates, but on average calcite crystals were 35 and 23% smaller in barnacles grown at

411 pH 8.01 as compared to barnacles grown at pH 7.50 or 7.78, respectively.

412

413 Qualitative assessments of barnacle base plates following adhesion testing and dissection

414 suggested that damage to the base plate (identified by the presence of cracks or holes) occurred

415 more often in barnacles exposed to pH 7.5 as compared to those exposed to pH 7.78 or 8.01.

416 Base plate damage was identified in 38.6% of barnacles grown at pH 7.50 as compared to 21.3%

417 and 18.8% at pH 7.78 and 8.01, respectively. Rigorous assessments of micromechanical  
418 properties were conducted on polished shell cross-sections. Mechanical testing revealed that  
419 microhardness was not affected by exposure pH in the base or parietal plates (Table 2; Fig. 5).  
420 Crack propagation resulting from microindentation, however, was significantly affected by pH  
421 when tested in the base plate (Table 2; Fig. 5). Within the base plate, cracks radiating from  
422 indents were, on average, 28% longer in barnacles exposed pH 7.50 as compared to those at pH  
423 8.01 (Fig. 5; Tukey HSD:  $p = 0.058$ ). Crack propagation was not affected by exposure pH when  
424 tested in the parietal plates.

425  
426 FTIR spectroscopy was used to assess the polymorph of calcium carbonate present in barnacle  
427 shells, as well as their relative crystallinity and atomic disorder. Spectroscopy was conducted on  
428 powdered shell samples and both base and all parietal plates were included in each sample. At all  
429 pH levels, identified peaks were characteristic of calcite with no other polymorphs of calcium  
430 carbonate present (Fig. 6). Relative crystallinity of the shells (assessed as the ratio of the  $\nu_2$  to  $\nu_4$   
431 absorbance) was not affected by pH (Table 2; Fig. 6). Atomic disorder, assessed through FTIR  
432 grinding curves, varied among samples and was not significantly affected by pH (Table 2; Fig.  
433 6).

434  
435 Elemental composition was assessed on polished shell cross-sections using EDS. Eight elements  
436 were identified in all samples: in order of abundance, Ca, O, C, S, Sr, Mg, Na, and Cl. The vast  
437 majority of the shell was comprised of Ca, O and C, with all other elements occurring at less than  
438 1 wt %. Exposure pH did not affect calcium content when assessed in the base or parietal plates  
439 (Table 2; Fig. 7). While no elements showed a significant effect of pH, magnesium content

440 tended to be lower (on average by 19%) in barnacles exposed pH 7.50 as compared to those at  
441 pH 8.01 (Fig. 7; Tukey HSD:  $p = 0.052$ ).

442

## 443 **DISCUSSION**

444 Barnacles are dominant members of marine intertidal communities and their success depends on  
445 both the firm attachment provided by their proteinaceous adhesive and the protection imparted  
446 by their calcified shell plates. Here we assessed if a reduction in seawater pH, at levels predicted  
447 within the next 200 years (i.e. to  $pH_T$  7.78 and 7.50, based on oceanic projections), would alter  
448 physiology, adhesion, and shell formation in the cosmopolitan barnacle *A. amphitrite*. Changes  
449 in the abundance of *A. amphitrite* could affect the composition of biofouling communities. We  
450 found that barnacles were generally robust to reduced pH, with no effect of pH on physiological  
451 metrics, and, contrary to our prediction, adhesive properties were not affected by reduced pH.  
452 Shell mass and base plate area were found to be higher in barnacles exposed to even moderately  
453 reduced pH, a trend that appears to be driven by increased size of the calcite crystals that  
454 comprise the shell. Although microhardness of the shell plates, a measure of resistance to  
455 permanent or plastic deformation, was not affected by pH, the length of cracks propagating from  
456 indents in the base plate was, suggesting that the toughness of base plates grown at the lowest pH  
457 level was reduced. Hence, a reduction in pH resulted in larger barnacles but with base plates that  
458 would crack more readily.

459

### 460 **Physiology**

461 Over the course of 13 weeks, we found no effect of seawater pH on cumulative mortality, tissue  
462 mass, or egg production in *A. amphitrite*. Assessments of ocean acidification responses in



463 juvenile or adult *A. amphitrite* are limited, but consistent with our observations, McDonald et al.  
464 (2009) found no effect of pH on *A. amphitrite* egg production after 11 weeks exposure to pH<sub>NBS</sub>  
465 7.4. Campanati et al. (2016) exposed *A. amphitrite* larvae to reduced pH (pH<sub>NBS</sub> 7.6) and tracked  
466 survival of juveniles settled from these larvae for 11 days. Mortality was actually reduced at low  
467 pH as compared to the pH<sub>NBS</sub> 8.2 control when assessed at 3 and 9 days post-settlement. At their  
468 final time point (11 days post settlement), there was no effect of seawater pH on juvenile  
469 survival.

470  
471 *Amphibalanus improvisus* (Pansch et al., 2014; Eriander et al., 2016) and *Elminius modestus*  
472 (Findlay et al., 2010a) have also been found to be robust to ocean acidification, as least in terms  
473 of survival and reproduction. Eriander et al. (2016) found no effect of reduced seawater pH (to  
474 pH<sub>NBS</sub> 7.7) on mortality in *A. improvisus* over 12 weeks, either when the pH reduction was kept  
475 at a stable level or when pH was allowed to fluctuate over the course of the day to mimic diurnal  
476 pH cycles. Pansch et al. (2014) similarly found no effect of reduced pH on *A. improvisus*  
477 mortality or reproduction (larval release rate) after 20 weeks in juveniles collected from a field  
478 site that shows natural variability in seawater pH (Kiel Fjord, Germany). In contrast though,  
479 juveniles collected from a field site with limited variability in seawater pH (Tjärnö Archipelago,  
480 Sweden) showed increased mortality at the lowest pH tested (pH<sub>NBS</sub> 7.2). Exposure of *Elminius*  
481 *modestus* to reduced pH (pH<sub>NBS</sub> 7.7) for 30 days did not affect mortality when tested at 14 or  
482 19°C (Findlay et al., 2010a).

483  
484 *Semibalanus balanoides*, a boreoarcitic species, appears considerably more sensitive to ocean  
485 acidification. Increased mortality has been shown in both short (30 days) and longer term (80 or

486 104 days) exposures to reduced pH (pH<sub>NBS</sub> 7.7-7.8)(Findlay et al., 2009; 2010a; Harvey and  
487 Moore, 2017). Barnacles in these studies were collected near the southern limit of the range for  
488 *S. balanoides*. Similar to what was shown for *A. improvisus*, environmental conditions of the  
489 collection site impact responses to ocean acidification (Findlay et al., 2010b). When Findlay et  
490 al. (2010) exposed *S. balanoides* collected near the northern limit of their range to reduced pH  
491 (to pH<sub>NBS</sub> 7.7 or 7.3) for 20 days, survival of juveniles did not differ from the pH<sub>NBS</sub> 8.1 control.  
492 Hence, barnacles show a mixed response to seawater pH. Individual responses vary by species as  
493 well as within a species based on local conditions of the breeding population.

494  
495 In general, crustaceans tend to be more tolerant of the effects of ocean acidification as compared  
496 to other taxa that build a mineralized shell (e.g. mollusks, corals: Kroeker et al., 2013) and such  
497 tolerance has been hypothesized to be due to their capacity for ion and acid-base regulation  
498 (Melzner et al., 2009; Whiteley, 2011). Populations that are routinely exposed to varying  
499 environmental conditions are likely to exhibit a considerable scope for physiological adjustment  
500 to a changing environment, making them well-adapted physiologically to changes in seawater  
501 pH (Wong et al., 2011; Pansch et al., 2012; Pansch et al., 2013). In *A. amphitrite* larvae for  
502 example, expression of energy metabolism related proteins was altered at reduced pH,  
503 illustrating the potential for proteomic plasticity in this robust species, and a possible mechanism  
504 for mediating the stress of ocean acidification (Wong et al., 2011). Variations among species and  
505 populations may stem from differences in ion and acid-base regulation ability, as well as  
506 differences in the ability to adjust energy metabolism (Whiteley, 2011).

507

508 **Adhesion**

509 We hypothesized that the barnacle adhesive system is sensitive to seawater pH, and based on this  
510 hypothesis, tested the prediction that barnacles grown under different levels of pH would vary in  
511 adhesive strength and gross morphology of the adhesive plaque. Previous work with marine  
512 mussels (*Mytilus trossulus*) found a reduction in both strength and extensibility of byssal threads  
513 when animals were exposed to reduced pH (tested at a range of pH<sub>T</sub> values from 8.1 to 7.5:  
514 O'Donnell et al., 2013). Such changes in mechanical properties of the byssal threads were  
515 calculated to lead to a 35-41% reduction in attachment tenacity, which could substantially affect  
516 a mussel's ability to anchor itself in high energy environments. O'Donnell et al. (2013) suggest  
517 that this response is due to sensitivity of DOPA to pH. DOPA is a major component of byssal  
518 threads that is involved in cross-linking and adhesion. A similar response was found in the  
519 mussel *Mytilus coruscus* (Zhao et al., 2017). Decreased pH (tested at a range of pH<sub>NBS</sub> values  
520 from 8.1 to 7.4) led to a decrease in byssal thread breaking force and toughness, as well as a  
521 reduction in the number of threads produced. Incorporating mechanical data with thread counts  
522 per mussel, Zhao et al. (2017) predicted a 60-65% reduction in attachment tenacity. Reduced  
523 seawater pH also led to significant alteration in expression of byssal thread related genes, thereby  
524 providing a mechanism that, in combination with direct effects of pH on DOPA chemistry, could  
525 explain the response of byssal threads to reduced pH.

526

527 Data collected here did not support the hypothesis that the barnacle adhesive system is sensitive  
528 to seawater pH; neither adhesive strength (measured in shear following the ASTM standard for  
529 measurement of barnacle adhesive strength: ASTM, 2005) nor the gross morphology of the  
530 adhesive plaque (i.e. if the glue layer was thin and hard or thick and gummy) were affected by  
531 seawater pH. A number of factors may contribute to the observed difference in sensitivity to pH

532 between marine mussels and the barnacle (*A. amphitrite*) tested here. First, the Balanomorph  
533 barnacle adhesive system is fundamentally different from that of marine mussels in that the  
534 adhesive interface is relatively protected from the external environment. Glue is delivered  
535 directly to the substrate at the periphery of the base and parietal plates (Saroyan et al., 1970;  
536 Burden et al., 2014). Although the cured glue layer is partially hydrated (Barlow et al., 2009),  
537 there is no component of the adhesive system that is constantly exposed to seawater, as is the  
538 case for a mussel's byssal threads. Second, the chemistry of barnacle glue differs from that of  
539 marine mussels. Barnacle glue is composed of at least ten major proteins, which play differing  
540 but specific roles in the adhesion process (Kamino, 2016; So et al., 2016). DOPA, which is pH  
541 sensitive (O'Donnell et al., 2013), has not been identified in barnacle glue (Naldrett, 1993;  
542 Kamino et al., 1996; Naldrett and Kaplan, 1997), although evidence of oxidative activity and  
543 cross-linking has been found (Dickinson, 2008; Golden et al., 2016; Essock-Burns et al., 2017;  
544 So et al., 2017). The sensitivity of isolated barnacle glue proteins to altered pH has yet to be  
545 assessed. Last, reduced seawater pH may lead to changes in the suite of glue proteins expressed,  
546 either in terms of the specific proteins expressed or the relative abundance of these proteins.  
547 Such a mechanism has not been tested in barnacles, but hypothetically could compensate for  
548 altered structure and activity of individual glue proteins. Recent advances in sequencing of the  
549 barnacle glue proteome (Wang et al., 2015; So et al., 2016) will enable direct assessment of  
550 proteomic responses to varied seawater conditions.

551

552 In our study, barnacles were settled on silicone substrates, which enabled them to be removed  
553 intact. Individuals with broken shells were excluded from analysis, and therefore the response to  
554 pH described here reflects only adhesive and cohesive properties of the adhesive plaque. When

555 barnacles were settled on hydrophilic glass beakers, the force required to shear the barnacles  
556 from the glass was actually enhanced in barnacles exposed to reduced pH (McDonald et al.,  
557 2009). In this case, all barnacles broke upon removal. On hydrophilic substrates, mechanical  
558 properties of the shell plates are weaker than the adhesive bond between the glue layer and the  
559 substrate, and therefore shear removal measurements reflect integrity of the lower shell plates  
560 where the force is applied. McDonald et al. (2009) suggest that thickening of the growing edge  
561 of the barnacle was responsible for the observed difference in shear removal force. Clearly  
562 surface chemistry, and hence whether failure will occur within the adhesive layer or within the  
563 shell plates, will mediate the effect of seawater pH on attachment tenacity. Assessments of  
564 barnacle responses to varied pH when grown on substrates that naturally occur in the marine  
565 environment may shed light on if ocean acidification will affect attachment ability in the field.

566

### 567 **Shell formation**

568 Consistent with what was found previously for *A. amphitrite* (McDonald et al., 2009), reduced  
569 seawater pH resulted in elevated shell formation. We found that even a moderate reduction in  
570 seawater pH (to pH<sub>T</sub> 7.78) resulted in greater shell mass as compared to the pH<sub>T</sub> 8.01 control.  
571 After 13 weeks growth, barnacle base plates were larger at reduced pH. Thickness of shell plates,  
572 measured directly on cross-sectioned shells, was not affected by seawater pH, although thickness  
573 of the parietal plates did tend to increase with decreasing pH. Enhanced shell formation at  
574 reduced seawater pH has been documented in several other crustaceans (i.e. *Callinectes sapidus*,  
575 *Penaeus plebejus*, *Homarus americanus*: Ries et al., 2009). Given that bicarbonate ion  
576 concentration increases under ocean acidification (increased pCO<sub>2</sub>), enhanced growth in  
577 crustaceans may stem from their ability to utilize bicarbonate in the mineralization process

578 (Cameron and Wood, 1985; Whiteley, 2011; Roleda et al., 2012). Enhanced growth though, is  
579 not universal in crustaceans (e.g. mixed growth responses have been observed in *A. improvisus*  
580 and *S. balanoides*: Findlay et al., 2010b; Pansch et al., 2014) and likely depends on an  
581 individual's capacity to mitigate the stress of reduced pH (see physiology section above).

582

583 Scanning electron microscopy imaging of cross-sectioned barnacle shells revealed that the shell  
584 plates were composed of crystals, with dimensions on the order of 1  $\mu\text{m}$ . FTIR spectroscopy  
585 confirmed that these crystallites were composed of calcite with no other polymorphs of calcium  
586 carbonate present, and EDS spectroscopy identified both magnesium and strontium within the  
587 calcitic shell plates. Consistent with previous assessments of *A. amphitrite* shells, individual  
588 crystallites were irregular in shape, did not take on a uniform orientation, and the boundaries of  
589 larger crystals appear rough, suggestive of smaller crystallites on the surface of the larger  
590 crystals (Khalifa et al., 2011; Lewis et al., 2014). The organic matrix surrounding calcite crystals  
591 is composed of chitin, acidic proteins, and sulfate-rich polymers (Fernandez et al., 2002;  
592 Rodriguez-Navarro et al., 2006; Khalifa et al., 2011), and in *A. amphitrite* comprises  
593 approximately 2 wt% of the shell (Khalifa et al., 2011). We did not assess the organic matrix  
594 specifically in this study, but observations of a double peak at 1145  $\text{cm}^{-1}$  in the FTIR spectrum  
595 (taken on whole, crushed shells) are suggestive of sulfate-rich polymers within the organic  
596 matrix (Khalifa et al., 2011).

597

598 Although seawater pH did not alter the overall shape or orientation of calcite crystals, the size of  
599 individual crystals increased dramatically in barnacles grown at reduced seawater pH. Calcite  
600 crystals comprising the base plate were nearly twice as large in barnacles at reduced pH ( $\text{pH}_T$

601 7.50 or 7.78) as compared to the  $\text{pH}_T$  8.01 control. Seawater pH resulted in a graded response in  
602 parietal plates, with the largest calcite crystals at the lowest seawater pH, a trend that closely  
603 followed parietal plate thickness. Shell formation in barnacles is directed by cells of the mantle  
604 epithelium, which participate in deposition of organic matrix and calcium transport (Nousek,  
605 1984; Fernandez et al., 2002; Rodriguez-Navarro et al., 2006; Gohad et al., 2009). Crystal  
606 nucleation, structure, and orientation of calcite crystals is proposed to be controlled by the  
607 organic matrix (Fernandez et al., 2002; Khalifa et al., 2011).

608  
609 At present, the mechanisms driving such differences in calcite crystal size are unclear. Increased  
610 size of calcite crystals under reduced seawater pH may reflect differences in the process of  
611 organic matrix deposition by mantle cells or the rate by which matrix deposition occurs. Similar  
612 to what was observed here, an increase in the size of shell microstructures (folia) was observed  
613 in eastern oysters (*Crassostrea virginica*) when exposed to  $\text{pH}_{\text{NBS}}$  7.5 for 20 weeks (Beniash et  
614 al., 2010). In this case it was proposed that energy limitations could impede organic matrix  
615 deposition or cell division. Here though, we did not observe an effect of seawater pH on *A.*  
616 *amphitrite* soft tissue mass, and therefore it does not appear as though animals at reduced pH  
617 were functioning under an energy deficit. An alternative hypothesis is that reduced seawater pH  
618 led to differences in intracellular pH, which could alter the ability of cells to participate in the  
619 mineralization process. In another crustacean (the Tanner crab, *Chionoecetes bairdi*), a reduction  
620 in intracellular pH was observed in animals held at reduced pH, a response that was proposed to  
621 have implications on the shell formation process (Meseck et al., 2016). Direct assessments of  
622 intracellular pH in barnacles grown at reduced pH as well as further investigation into the role of

623 cells in the shell formation process in barnacles would be helpful in evaluating these  
624 mechanisms.

625

626 Changes in the shell formation process under reduced pH may have implications in terms of  
627 functionality of the shell. The base plates of barnacles grown at the lowest pH tested (pH<sub>T</sub> 7.50)  
628 tended to break more easily than those of barnacles grown at pH<sub>T</sub> 7.78 or 8.01. Rigorous  
629 mechanical testing supported this observation. When tested in the base plate, hardness, a  
630 material's ability to resist plastic or permanent deformation, was not affected by pH, but the  
631 cracks that propagated from mechanical tests were longer at low pH, indicating that reduced  
632 seawater pH led to a reduction in toughness. This response may partially be driven by the  
633 dramatic increase in calcite crystal size at low pH. Larger crystals would imply a lower ratio of  
634 organic to inorganic material within the shell. Organic matrix plays an important toughening role  
635 in biological materials, serving to trap and deflect cracks (Fratzl et al., 2007; Beniash et al.,  
636 2010; Meyers and Chen, 2014). Therefore, as the ratio of organic matrix to inorganic mineral  
637 decreases, crack deflection ability is also diminished. This mechanism, however, cannot fully  
638 explain the trends observed, given that animals at pH<sub>T</sub> 7.78 showed larger calcite crystal size but  
639 did not exhibit a reduction in base plate toughness.

640

641 Several additional factors could contribute to the observed reduction in base plate toughness.  
642 Assessment of the shell organic matrix was beyond the scope of this study, but given the ability  
643 of the organic matrix to trap and deflect cracks (Fratzl et al., 2007; Meyers and Chen, 2014)  
644 alterations in the composition or density of the organic matrix could also lead to changes in  
645 toughness. Considering the calcite crystals themselves, two factors could have influenced



646 mechanical properties. First, organic constituents (e.g. amino acids) can be occluded within  
647 biogenic calcite crystals and can dramatically impact mechanical properties via alterations in the  
648 crystal lattice (Cho et al., 2016; Kim et al., 2016). If the identity or quantity of these organic  
649 inclusions was altered at low pH, this could influence shell mechanical properties. Second,  
650 inclusion of magnesium into the calcite crystal lattice can have a major impact on mechanical  
651 properties of shells, with the addition of even small amounts of magnesium leading to enhanced  
652 mechanical properties (Kunitake et al., 2012; Long et al., 2014). Correspondingly, in the base  
653 plate of pH<sub>T</sub> 7.50 barnacles, we observed a trend ( $p = 0.056$ ) toward decreased magnesium  
654 content as compared to animals at pH<sub>T</sub> 7.78 or 8.01. Such a reduction in magnesium content  
655 could contribute to the observed reduction in toughness. At this point it is unclear if the reduced  
656 magnesium content at low pH is due to lower uptake and incorporation of magnesium, or  
657 increased dissolution of weakly bound magnesium (Findlay et al., 2009).

658  
659 A reduction in the force needed to break parietal plates was observed previously in *A. amphitrite*  
660 (McDonald et al., 2009) and *A. improvisus* (Pansch et al., 2013) after exposure to reduced pH.  
661 Force needed to break parietal plates was not assessed in this study, but at the micro-scale, we  
662 did not observe an effect of seawater pH on parietal plate hardness or crack propagation. Hence  
663 differences in force needed to break parietal plates may stem from structural changes in the  
664 plates (e.g. local dissolution of mineral: McDonald et al., 2009), rather than differences in their  
665 material properties. We did not observe damage or erosion of parietal plates for any of the  
666 treatments in this study, although alkalinity was consistently elevated at pH<sub>T</sub> 7.50, suggesting  
667 that some dissolution of shells at this pH may have occurred.  
668

669 Neither shell crystallinity nor atomic disorder, both measures of the level of structural order  
670 within a crystal at the atomic scale, were affected by seawater pH. Changes in formation  
671 conditions or the composition of molecules at the time of crystal formation could affect these  
672 metrics (Khalifa et al., 2011). Interestingly, although the mean values of these metrics were not  
673 significantly affected by seawater pH, variance around the mean was dramatically greater at  
674 reduced pH as compared to  $\text{pH}_T$  8.01 (e.g. for atomic disorder, standard error at reduced pH was  
675 4-5 times greater than at  $\text{pH}_T$  8.01). This suggests that reduced pH may increase variability  
676 among individuals in their ability to control the environment in which mineral forms. A similar  
677 response in terms of increased variability in crystal properties among individuals at decreased pH  
678 was observed previously in the marine mussel, *Mytilus californianus* (McCoy et al., 2018).

679

## 680 CONCLUSIONS

681 A reduction in seawater pH at levels predicted within the next 200 years (i.e. to  $\text{pH}_T$  7.78 and  
682 7.50, based on oceanic projections: Caldeira and Wickett, 2003; Doney et al., 2009) had little  
683 impact on physiological and adhesive metrics in the barnacle *A. amphitrite*. Shell growth,  
684 though, was significantly enhanced at reduced pH, while toughness of the base plate was  
685 diminished at  $\text{pH}_T$  7.50. If these changes impact the survival of *A. amphitrite* in the field is yet to  
686 be determined, but alterations in the abundance of this ecologically dominant species would  
687 undoubtedly affect the composition of biofouling communities. Given the economic impact of  
688 marine biofouling (Callow and Callow, 2011; Schultz et al., 2011), additional assessments of *A.*  
689 *amphitrite* under changing environmental conditions are warranted. Multi-stressor and  
690 transgenerational assessments, as well as experiments that test natural sources of mortality in

691 barnacles (e.g. predators or hydrodynamic stresses) would be especially helpful in predicting  
692 population level responses in *A. amphitrite*.

693

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698

#### 699 **AUTHOR CONTRIBUTIONS**

700 JN developed the seawater exposure system, collected and compiled data on seawater chemistry,  
701 physiology and adhesive metrics, and oversaw daily operations of the exposure. SP assessed  
702 shell thickness, mechanical properties, and crystallinity. KS and DT collected data on mortality,  
703 egg presence, adhesive properties, and shell mass and area. CM developed a prototype seawater  
704 exposure system and initiated a pilot study. JM, JH, and RM prepared shells and collected data  
705 on atomic disorder and RM analyzed atomic disorder data. BO and DR provided juvenile  
706 barnacles and algae and provided expertise on barnacle growth and assessment throughout the  
707 exposure. GD conceived of the experiments, analyzed final datasets, oversaw the experiment,  
708 and wrote the manuscript. All authors contributed to editing of the manuscript.

709

#### 710 **Conflicts of interest statement**

711 The authors declare that the research was conducted in the absence of any commercial or  
712 financial relationships that could be construed as a potential conflict of interest.

713

714 **Data availability statement**

715 The raw data supporting the conclusions of this manuscript will be made available by the  
716 authors, without undue reservation, to any qualified researcher.

717

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925

926 **TABLE AND FIGURE LEGENDS**

927

928 **Table 1.** Seawater chemistry parameters (means  $\pm$  SD). pH<sub>T</sub>, temperature, and salinity were  
929 measured six days per week (n = 76), total alkalinity was measured weekly (n = 12), all other  
930 parameters were calculated using CO2Calc.

931

932 **Table 2.** Assessment of the effects of seawater pH on physiology, adhesion, and shell formation  
933 in the barnacle *Amphibalanus amphitrite*. ANOVA results are shown for all metrics except shell  
934 dry mass, crystallinity, parietal plate thickness, and parietal plate crack propagation, where  
935 Kruskal-Wallis tests were applied. Significant p-values are shown in bold.

936

937 **Fig. 1. A.** Assessments of barnacle shell properties were conducted on polished cross-sections of  
938 individual barnacle shells, which exposed both the base and parietal plates. **B.** Mechanical testing  
939 was conducted using microindentation, which results in a diamond shaped indent and induces  
940 crack formation (arrows). Image was taken under polarized light, revealing organization of

941 calcite crystals. C. Shell ultrastructure was assessed using SEM. Shells are comprised of a  
942 cobblestone-like network of calcite crystallites, the area of which can be measured using image  
943 analysis software.

944

945 **Fig. 2.** Tissue and shell mass of barnacles, *A. amphitrite*, exposed to one of three levels of pH<sub>T</sub>  
946 for 13 weeks (mean ± s.e.m.). Groups marked with different letters are significantly different as  
947 shown by Tukey HSD post-hoc analysis. N is indicated on each bar and represents individual  
948 barnacles.

949

950 **Fig. 3.** Adhesion strength (critical shear stress) and gummy glue expression of barnacles, *A.*  
951 *amphitrite*, exposed to one of three levels of pH<sub>T</sub> for 13 weeks (mean ± s.e.m.). N is indicated on  
952 each bar. For adhesion strength, this represents individual barnacles. Gummy glue expression  
953 was calculated as the percent of barnacles per panel with any visible gummy glue.

954

955 **Fig. 4.** Shell structural assessments of barnacles, *A. amphitrite*, exposed to one of three levels of  
956 pH<sub>T</sub> for 13 weeks (mean ± s.e.m.). Groups marked with different letters are significantly  
957 different as shown by Tukey HSD post-hoc analysis. N is indicated on each bar and represents  
958 individual barnacles.

959

960 **Fig. 5.** Micromechanical assessments of barnacles, *A. amphitrite*, exposed to one of three levels  
961 of pH<sub>T</sub> for 13 weeks. Mean ± s.e.m. is shown for microhardness and crack propagation. N is  
962 indicated on each bar and represents individual barnacles. pH significantly affected crack

963 propagation in the base plate (ANOVA:  $p < 0.05$ ). Bottom images show indents and cracks  
964 (arrows) for the base plate of a pH 8.01 barnacle (left) and 7.50 barnacle (right).

965

966 **Fig. 6.** Representative FTIR spectra of a powdered barnacle shell with peaks characteristic of  
967 calcite (left). The ratio of  $\nu_2$  to  $\nu_4$  peak absorbance was used as a proxy of crystallinity (center),  
968 and FTIR grinding curves were used to assess atomic disorder (right). Mean  $\pm$  s.e.m. is shown  
969 for crystallinity and atomic disorder, and pH did not significantly affect either metric (ANOVA:  
970  $p < 0.05$ ). N is indicated on each bar and represents individual barnacles.

971

972 **Fig. 7.** Calcium and magnesium content of the shells of barnacles, *A. amphitrite*, that were  
973 exposed to one of three levels of  $\text{pH}_T$  for 13 weeks (mean  $\pm$  s.e.m.). N is indicated on each bar  
974 and represents individual barnacles.

975 **TABLES**

976 **Table 1.** Seawater chemistry parameters (means  $\pm$  SD).  $\text{pH}_T$ , temperature, and salinity were  
 977 measured six days per week ( $n = 76$ ), total alkalinity was measured weekly ( $n = 12$ ), all other  
 978 parameters were calculated using CO2Calc.

|  | <b>Treatment</b>  |                   |                    |
|--|-------------------|-------------------|--------------------|
|  | <b>8.01</b>       | <b>7.78</b>       | <b>7.50</b>        |
| <b>pH<sub>T</sub></b>  | 8.01 $\pm$ 0.03   | 7.78 $\pm$ 0.04   | 7.50 $\pm$ 0.04    |
| <b>Temperature (°C)</b>  | 25.0 $\pm$ 0.1    | 25.0 $\pm$ 0.1    | 25.0 $\pm$ 0.1     |
| <b>Salinity</b>  | 35.8 $\pm$ 0.5    | 35.7 $\pm$ 0.5    | 35.8 $\pm$ 0.5     |
| <b>pCO<sub>2</sub> (<math>\mu\text{atm}</math>)</b>                        | 412.8 $\pm$ 35.5  | 810.4 $\pm$ 119.5 | 1747.7 $\pm$ 263.4 |
| <b>DIC (<math>\mu\text{mol kg}^{-1}</math> SW)</b>                         | 1889.6 $\pm$ 41.3 | 2015.5 $\pm$ 49.7 | 2330.6 $\pm$ 92.5  |
| <b>HCO<sub>3</sub><sup>-</sup> (<math>\mu\text{mol kg}^{-1}</math> SW)</b> | 1687.5 $\pm$ 40.5 | 1871.7 $\pm$ 53.5 | 2203.9 $\pm$ 89.8  |
| <b>CO<sub>3</sub><sup>2-</sup> (<math>\mu\text{mol kg}^{-1}</math> SW)</b> | 190.5 $\pm$ 14.3  | 120.9 $\pm$ 17.1  | 77.4 $\pm$ 10.4    |
| <b>Total alkalinity (<math>\mu\text{mol kg}^{-1}</math> SW)</b>            | 2166.6 $\pm$ 47.7 | 2174.7 $\pm$ 49.6 | 2394.1 $\pm$ 86.3  |
| <b><math>\Omega_{\text{Calcite}}</math></b>                                | 4.55 $\pm$ 0.33   | 2.89 $\pm$ 0.41   | 1.85 $\pm$ 0.25    |

979

980

981

982

983 **Table 2.** Assessment of the effects of seawater pH on physiology, adhesion, and shell formation  
 984 in the barnacle *Amphibalanus amphitrite*. ANOVA results are shown for all metrics except shell  
 985 dry mass, crystallinity, parietal plate thickness, and parietal plate crack propagation, where  
 986 Kruskal-Wallis tests were applied. Significant p-values are shown in bold.

987

| Parameter                   | df         | Test Statistic | p            |
|-----------------------------|------------|----------------|--------------|
| <i>Physiology</i>           |            |                |              |
| Tissue dry mass             | 134        | 1.860          | 0.160        |
| <i>Adhesion</i>             |            |                |              |
| Adhesive strength           | 168        | 1.363          | 0.259        |
| <i>Shell formation</i>      |            |                |              |
| <i>Whole shell</i>          |            |                |              |
| <b>Dry mass</b>             | <b>2</b>   | <b>8.366</b>   | <b>0.015</b> |
| Height                      | 172        | 0.938          | 0.393        |
| Crystallinity               | 2          | 1.088          | 0.581        |
| Atomic disorder             | 9          | 1.021          | 0.408        |
| <i>Base plate</i>           |            |                |              |
| <b>Area</b>                 | <b>171</b> | <b>6.016</b>   | <b>0.003</b> |
| Thickness                   | 46         | 2.019          | 0.145        |
| Microhardness               | 47         | 0.367          | 0.695        |
| <b>Crack propagation</b>    | <b>46</b>  | <b>3.244</b>   | <b>0.049</b> |
| <b>Calcite crystal area</b> | <b>47</b>  | <b>16.736</b>  | <b>0.000</b> |
| Ca content                  | 47         | 0.094          | 0.911        |
| Mg content                  | 47         | 3.083          | 0.056        |
| <i>Parietal plate</i>       |            |                |              |
| Thickness                   | 2          | 2.642          | 0.267        |
| Microhardness               | 47         | 0.354          | 0.704        |
| Crack propagation           | 2          | 1.108          | 0.575        |
| <b>Calcite crystal area</b> | <b>47</b>  | <b>7.080</b>   | <b>0.002</b> |
| Ca content                  | 47         | 0.076          | 0.927        |
| Mg content                  | 47         | 0.620          | 0.542        |

988

Figure 1.JPEG

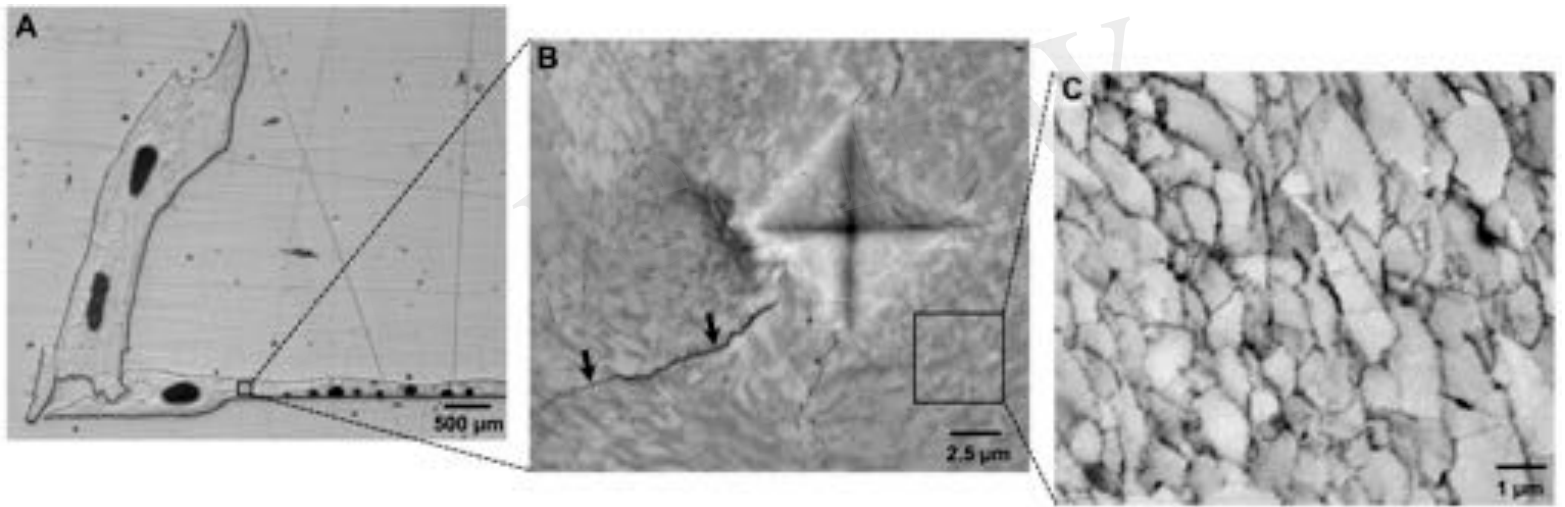


Figure 2.JPEG

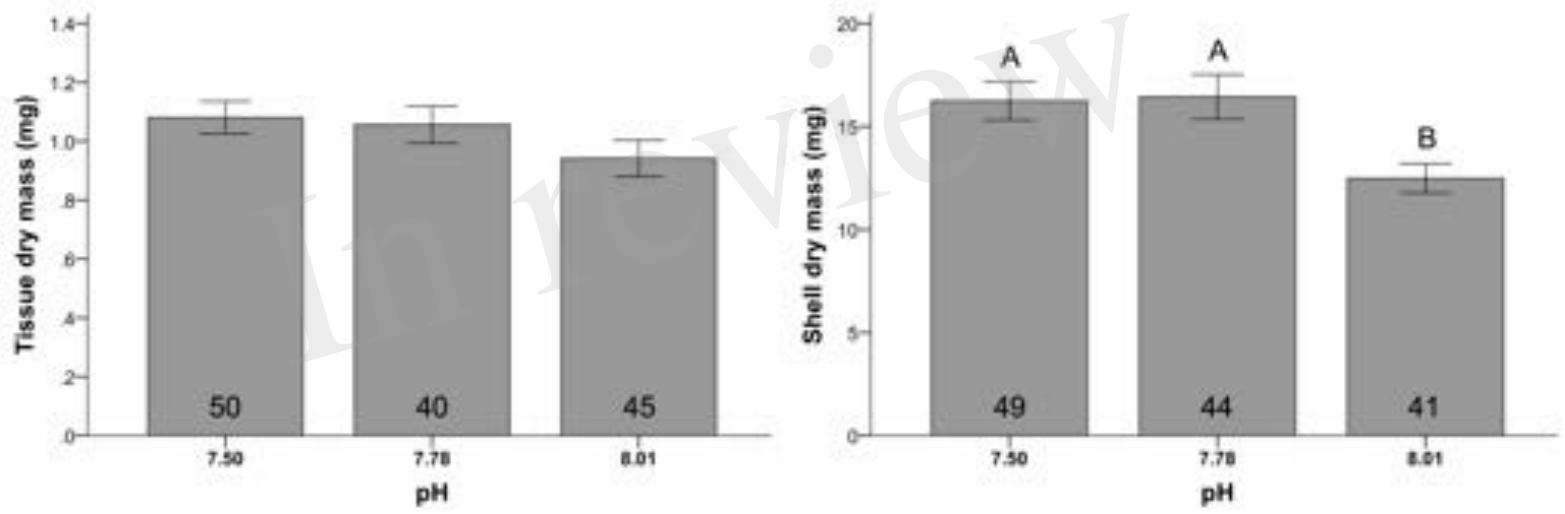




Figure 3.JPEG

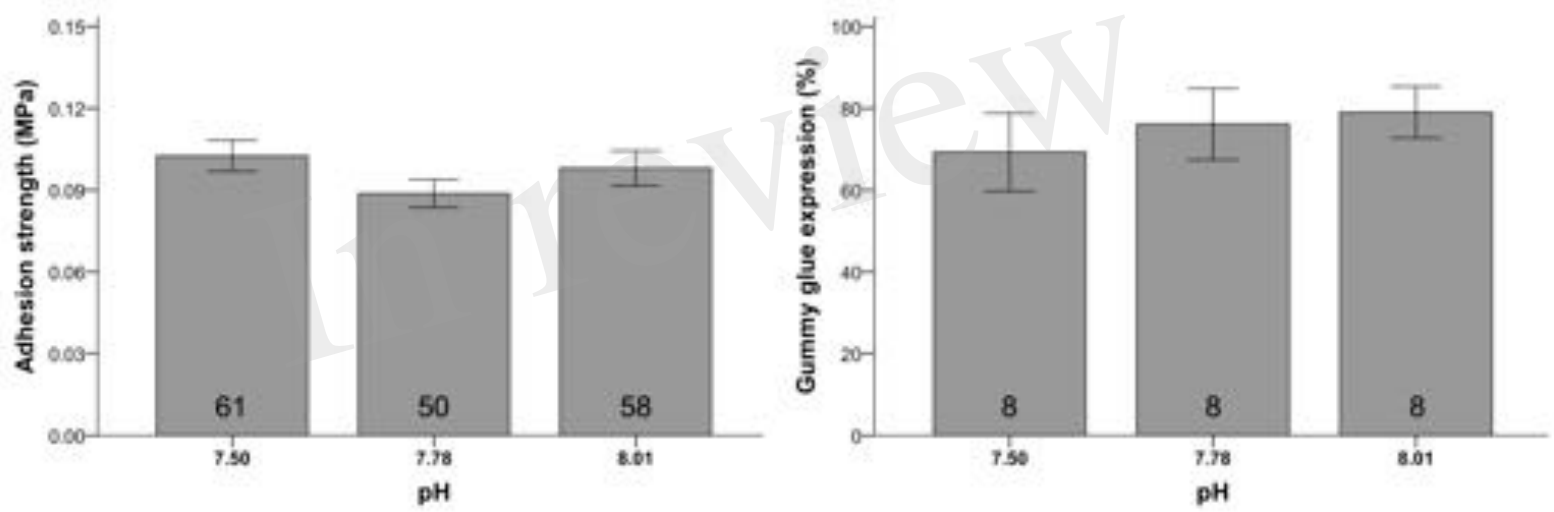


Figure 4.JPEG

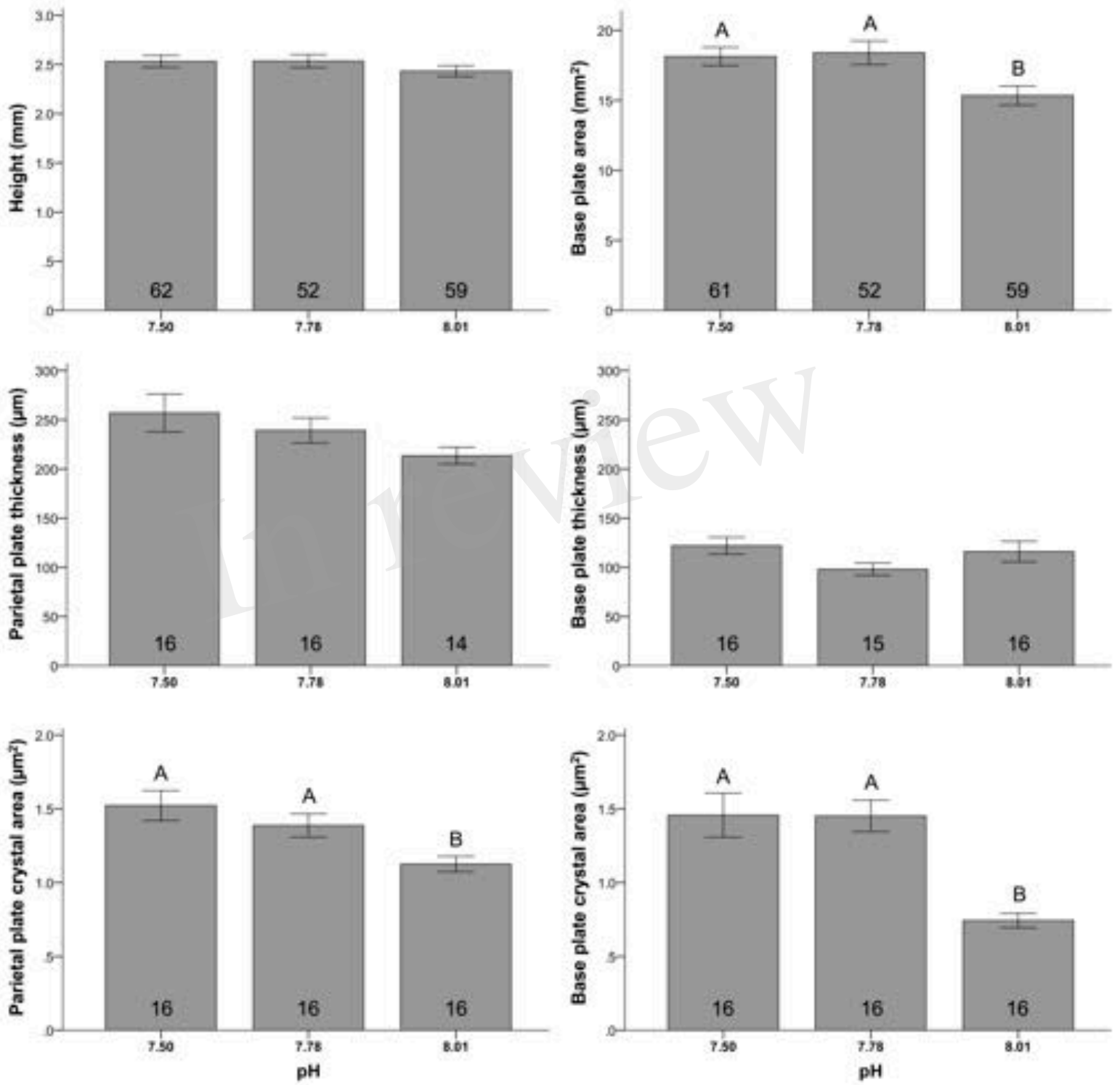


Figure 5.JPEG

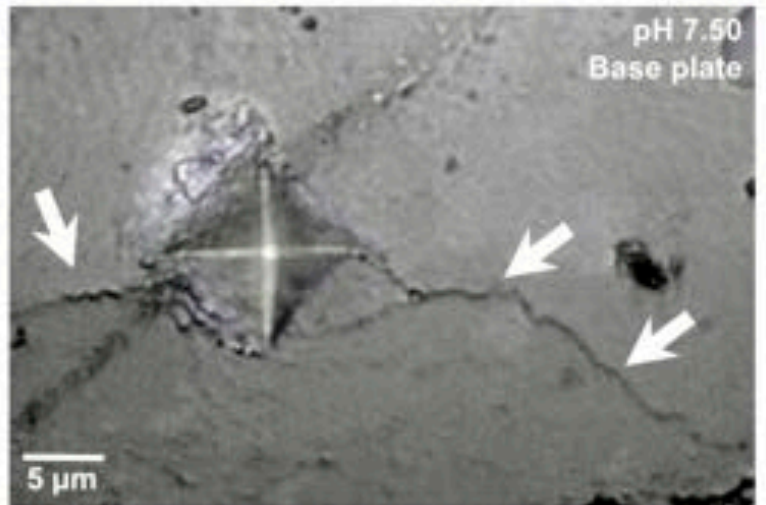
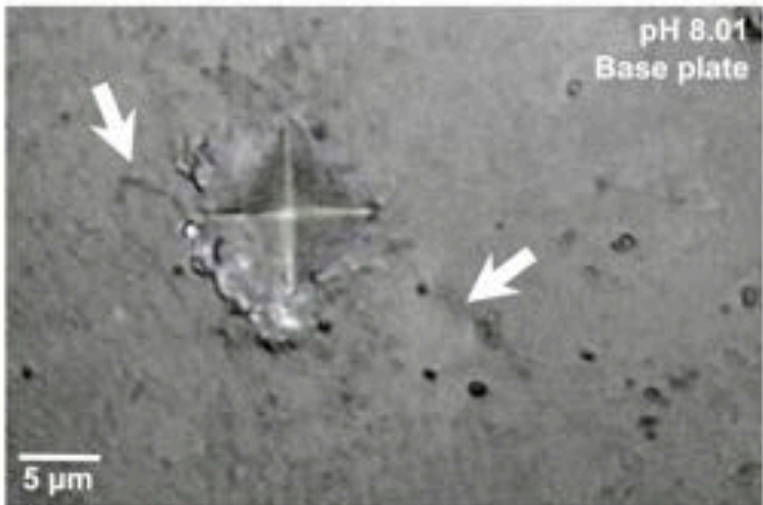
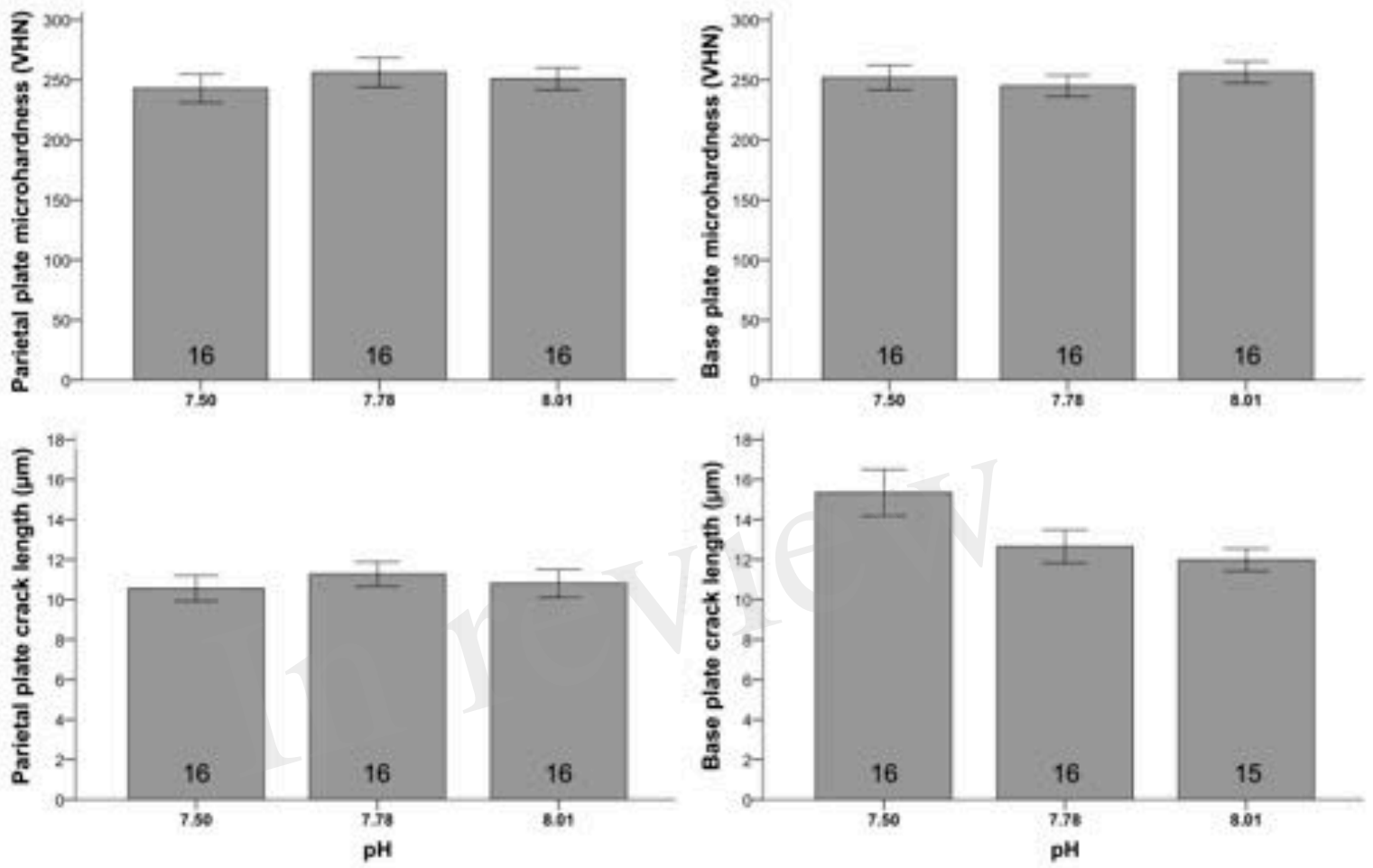


Figure 6.JPEG

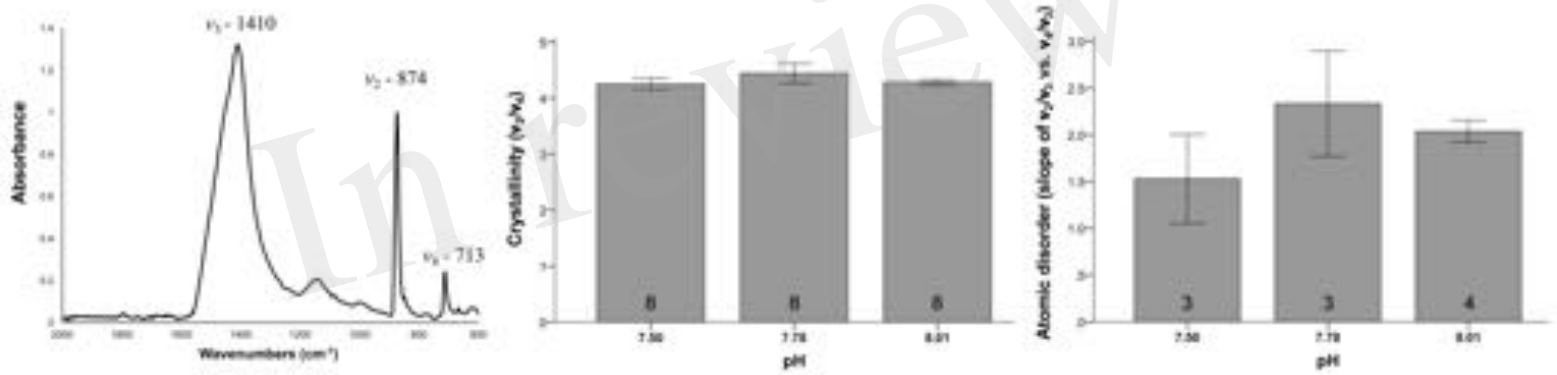
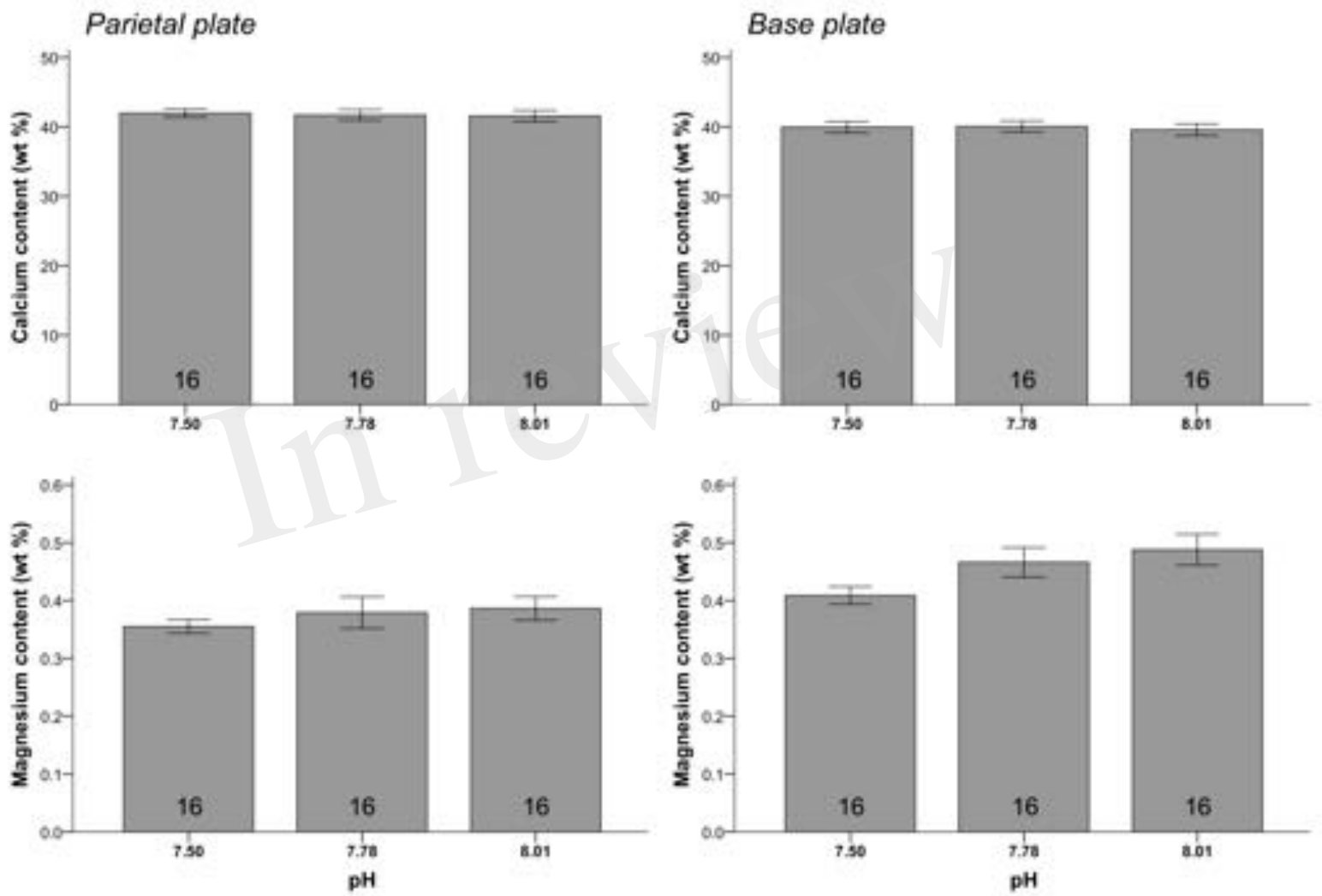


Figure 7.JPEG



# Monday Morning, October 30, 2017

source by Oregon Physics that replaced, for the first time, Duoplasmatron on ADEPT1010 Dynamic SIMS System by Physical Electronics.

The FIB sample preparation procedure is discussed in detail and the first back side SIMS results compared to the front side depth profiles.

**11:40am AS+BI+MI-MoM11 Phase Quantification of Mixed TiO<sub>2</sub> Powders by X-ray Photoemission Valence Band Analysis and Raman Spectroscopy, Paul Mack, T.S. Nunnay, Thermo Fisher Scientific, UK; R.G. Palgrave, University College London, United Kingdom of Great Britain and Northern Ireland**

Titanium dioxide is one of the most studied materials in surface science. It has applications in heterogeneous catalysis, dye-sensitized solar cells, bone implants and self-cleaning windows. Many polymorphs of TiO<sub>2</sub> are known to exist but only two occur naturally in abundance: rutile and anatase. Rutile is the more thermodynamically stable form but anatase is more energetically favourable when forming nanoparticles at atmospheric temperature and pressure. The anatase polymorph has been recognised as more photoactive than rutile, although recent research indicates that the greatest photovoltaic efficiencies are achieved in devices that contain a mixture of anatase and rutile. The degree of mixing between two polymorphs influences other material properties, such as catalytic activity. This raises the question: how can one determine the polymorph ratio in a sample that contains a mixture of anatase and rutile?

Quantitative phase analysis of anatase-rutile mixtures by two experimental methods is presented in this work. Spectra of pure reference anatase and rutile were acquired X-ray Photoelectron Spectroscopy (XPS) and Raman spectroscopy. These spectral shapes were then used to fit similar data from mixed phase samples. XPS and Raman spectroscopy give information from different depth regions in a sample. The surface sensitive character of XPS yields a surface phase fraction of anatase and rutile. Mixed phase samples were prepared from high and low surface area anatase and rutile powders. In this work, the surface phase fraction of anatase was found to be linearly correlated with photocatalytic activity of the mixed phase samples, even for samples with very different anatase and rutile surface areas.

## Biomaterial Interfaces Division

### Room 12 - Session BI-MoM

#### Engineering a Paradigm Shift in Control of Microbes and Fouling

**Moderators:** Joe Baio, Oregon State University, Daniel Barlow, US Naval Research Laboratory

**8:20am BI-MoM1 Characterization of Adult Barnacle Adhesion Upon Reattachment to Hydrophobic Surfaces, Manuel Figueroa, G. Dickinson, The College of New Jersey**

Although a wide range of environmentally friendly surface coatings can reduce biofouling on marine structures, there is still not a fundamental understanding of barnacle adhesion upon reattachment. This study assessed the effect of hydrophobicity on adhesion in the barnacle *Amphibalanus amphitrite*, an abundant and widespread biofouler. Self-assembled monolayers were made on glass slides from alkyl silanes with methylated and fluorinated terminal groups to produce hydrophobic surfaces. Coated and uncoated glass slides underwent a 2-week barnacle reattachment assay. Barnacles were removed using a force gauge and critical shear stress was calculated for each substrate. Following reattachment assays, a Coomassie Blue G250 protein stain was used to quantify the amount of glue remaining on substrates by measuring pixel density with ImageJ software on glue scans.

Critical shear stress was found to be significantly higher for both hydrophobic surfaces as compared to the hydrophilic uncoated glass, and correspondingly the density of residual glue was higher on hydrophobic surfaces. Given that hydrophobic substrates can exclude water from the surface, they may provide a protected environment for glue release that is favorable for adhesive bond formation with the substrate as well as inter and intramolecular bonding within the glue layer. Critical shear stress showed a strong positive correlation with residual glue density, suggesting barnacle release occurs primarily via cohesive failure. Scanning electron microscope micrographs depicted a diverse mixture of features in the glue remnants depending on the coating and its location under the base plate. These features, which included a sponge-like matrix, globular structures, viscous fingering and nanoscale fibers contribute to adhesion strength. The design of marine coatings must continue to consider the nanoscale

topography as an essential attribute to reducing biofouling as well as the ability of a coating to exclude water from the surface.

**8:40am BI-MoM2 Constructing and Deconstructing the Barnacle Adhesive Interface, C.R. So, K.P. Fears, US Naval Research Laboratory; H. Ryou, ASEE Research Fellow at US Naval Research Laboratory; D.E. Barlow, D.H. Leary, J.A. Wollmershauser, C.M. Spillmann, Kathryn Wahl, US Naval Research Laboratory**

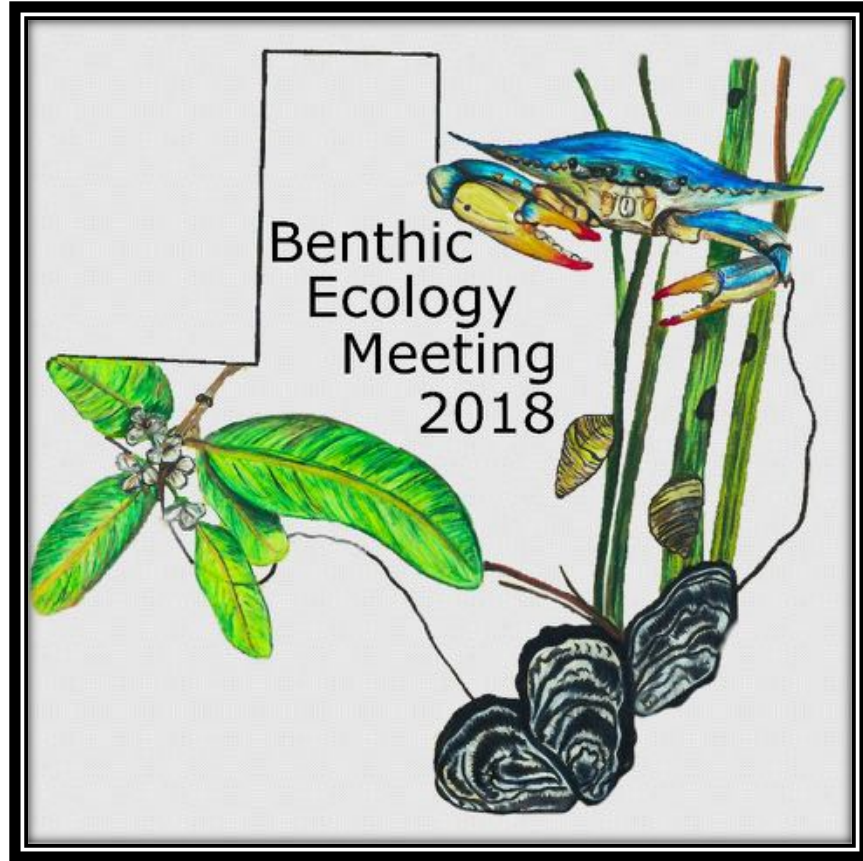
Barnacles are sessile marine arthropods that live and reproduce on nearly any available surface in the ocean. They adhere via a thin adhesive layer developed through a multistep secretory process synchronized with growth via molting. Unlike other arthropods, the combination of expansion, molting and protein secretion within the narrow adhesion interface leads to a nanofibrillar protein layer manipulated by shear stresses, protected by calcite, and containing a cocktail of chemically active molecules and proteins. Here we use *in vivo* imaging, mechanics, and spectroscopy of barnacle growth and development, coupled with mass spectrometry and proteomics to reveal much about the biophysics and biochemistry of barnacle adhesion. We will discuss the role of interfacial processes, self-assembly, amino acid composition, and chemical manipulation in the construction and function of the adhesive.

**9:00am BI-MoM3 Live Confocal Microscopy of *Balanus Amphitrite* Reveals Anti-Fouling Strategy of a Marine Fouler, Kenan Fears, US Naval Research Laboratory; B. Orihuela, D. Rittschof, Duke University Marine Laboratory; K.J. Wahl, US Naval Research Laboratory**

The adhesion of hard foulers (e.g., barnacles and tubeworms) has plagued the maritime community for as long as mankind has been setting sail. Since the biological processes responsible for adhesion occur at buried interfaces, elucidating the mechanisms by which foulers adhere is challenging. Through the use of multiple fluorescent probes, peptides, and antibodies, we have been able to discern an unprecedented level of detail about biological processes that occur at the interface between acorn barnacles (*Balanus Amphitrite*) and the underlying substratum during the barnacle growth cycle. Barnacles secrete a lipidaceous substance around the outside of their shell, prior to expansion that dislodges microorganisms and biofilms to present a cleaned surface. During molting, epithelia cells build a new interfacial cuticular layer, which becomes autofluorescent as it is sclerotized, above the existing cuticle whose degradation coincides with the exuviation of the main body's cuticle. Rather than being directly secreted onto the substrate, nanostructured barnacle cement accumulates in the space in between the new and old cuticle. As the barnacle expands, the cuticular layers are stretched and pulled around the outside of the side plate. The strain causes the old cuticle to randomly tear, allowing the new cuticle to deposit cement into the interface as it is dragged across the substrate. Furthermore, antibody staining allowed us to spatially and temporally identify where different cement proteins are present. These results illustrate that the methodologies we have developed to break down and analyze barnacle cement collection are yielding a more accurate representation of the proteins at the buried interface, and providing insight on their roles which will lead to improved strategies to both combat and mimic barnacle adhesives.

**9:20am BI-MoM4 Considering the Consequences of a Paradigm Shift in Biofouling Management, Daniel Rittschof, B. Orihuela, Duke University; K. Efimenko, J. Genzer, NC State University**

Present Fouling Management Strategies that use long-lived, broad-spectrum biocides are not sustainable because they alter ecosystem services and threaten food security. As globalization continues, human populations increase and wild fisheries collapse there will be increasing pressure and genuine need for less environmentally damaging approaches. A question that should be asked up front for any new fouling management technology is what are the environmental, food security and human health consequences if a technology gains market share. Information on impacts of industrial grade components, acute and chronic toxicity, breakdown and non-toxic biological effects such as teratogenicity, carcinogenicity and environmental steroid activity should be evaluated. This presentation looks at a few of the details of basic silicone coatings which have had their components purified and then tested for acute toxicity, impacts on a hydrolytic enzyme and teratogenicity. Some components like catalysts and small cyclics are extremely toxic. Other components impact enzyme activity, some inhibit activity other potentiate activity. Teratogenicity assays are so sensitive that even effects of medical grade silicones can be demonstrated. This information needs to be taken as preliminary factual information that can be used by engineers in developing risk benefit analysis.



47<sup>TH</sup> Annual Benthic Ecology Meeting

# ABSTRACT BOOK

March 27-30, 2018

Corpus Christi, TX

Hosted by:

Harte Research Institute for Gulf of Mexico Studies,  
Texas A&M University – Corpus Christi



### **Assessing the impacts of ocean acidification on adhesion and shell formation in the barnacle *Amphibalanus amphitrite***

Dickinson, G. H.\*<sup>1</sup>; Nardone, J. A.<sup>1</sup>; Patel, S.<sup>1</sup>; Siegel, K. R.<sup>1</sup>; Tedesco, D.<sup>1</sup>; McNicholl, C. G.<sup>1</sup>; O'Malley, J.<sup>2</sup>; Herrick, J.<sup>2</sup>; Metzler, R. A.<sup>2</sup>; Orihuela, B.<sup>3</sup>; Rittschof, D.<sup>3</sup>

<sup>1</sup>Department of Biology, The College of New Jersey

<sup>2</sup>Department of Physics and Astronomy Colgate University

<sup>3</sup>Duke University Marine Laboratory, Marine Science and Conservation

Barnacles are dominant members of intertidal communities. Their success depends on firm attachment provided by their adhesive and protection imparted by calcified shell plates. We assessed if reduction in seawater pH would alter physiology, adhesion, and shell formation in the barnacle *Amphibalanus amphitrite*. Juvenile barnacles were exposed to one of three levels of pH<sub>T</sub>, 8.01, 7.78 or 7.50, for 13 weeks. We found that barnacles were robust to reduced pH, with no effect of pH on mortality, tissue mass, or egg presence. Likewise, adhesive strength and morphology were not affected by reduced pH. Shell formation, however, was affected by seawater pH. Shell mass and base plate area were higher in barnacles exposed to reduced pH. Enhanced growth at reduced pH was driven by increased size of calcite crystals that comprise the shell. Despite enhanced growth, barnacles grown at pH<sub>T</sub> 7.5 had base plates that broke more easily, suggesting reduced shell toughness. Shell thickness, crystallinity, and atomic disorder were not affected by pH. It is yet to be determined if such changes would alter survival of *A. amphitrite* in the field, but changes in abundance of this ecologically dominant species would undoubtedly affect composition of biofouling communities. Authors acknowledge ONR funding.

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### **Direct and indirect effects of disturbance and eutrophication on SW Florida seagrasses**

Douglass, J.\*; Sang, A.

Florida Gulf Coast University

Shallow estuaries and lagoons along the SW coast of Florida have historically supported extensive seagrass beds, but anthropogenic pressures increasingly threaten these vulnerable habitats. This presentation reviews the status of SW Florida seagrass ecosystems, highlighting alarming declines in the Caloosahatchee Estuary and Estero Bay. The causes of these declines are discussed and categorized as: 1) disturbances stemming from the interaction of extreme weather events and altered watershed hydrology, 2) chronic stresses from nutrient loading and water quality degradation, and 3) indirect effects of these stressors mediated by benthic algae, epifauna, infauna, macrograzers, and microbes. We identify gaps and uncertainties in our current understanding of the mechanisms of seagrass decline in the region, and we describe ongoing and anticipated studies investigating these mechanisms.

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### **Building a barnacle cement protein profile using immunohistochemistry**

Patel, S. <sup>\*,1,2</sup>; Spillmann, C. M. <sup>1</sup>; Wang, C. <sup>1</sup>; Schultzhaus, J. <sup>1</sup>; Dickinson, G. H. <sup>2</sup>; Orihuela, B. <sup>3</sup>; Rittschof, D. <sup>3</sup>

<sup>1</sup> Naval Research Laboratory

<sup>2</sup> Department of Biology, The College of New Jersey

<sup>3</sup> Duke University Marine Laboratory

Acorn barnacles are able to adhere to a range of surfaces using secreted cement. In the barnacle *Amphibalanus amphitrite*, secretion of proteinaceous cement occurs at the periphery of the basal plate through a complex process where biomineralization and cuticle formation occur. Previously, cured cement was partially solubilized and analyzed where 19, 43 and 114 kDa cement proteins were identified and polyclonal antibodies prepared against them. The purpose of this experiment was to determine where these cement proteins are localized within the barnacle. Transverse sections of barnacle tissue were prepared, beginning just above the cement-substrate interface and continuing to the operculum and exposed to antibodies. No cement proteins were found in exterior muscle bundles or penis. Positive staining against Aacp43 was found within testes and in longitudinal canals, which run through parietal plates of the shell. Staining against Aacp19 was found lining the interior side of mantle tissue, and in granular-like structures in longitudinal canals. Antibody for Aacp114 tended to aggregate, so the staining profile was inconclusive. Overall, we demonstrate positive staining against cement proteins in barnacle sections, indicating their presence in areas other than the region in contact with the underlying interface. Authors acknowledge support from NRL. S.P. acknowledges NREIP.

*Presenting author contact info: patels111@tcnj.edu*

### **Nutrient loads explains spatial checker boarding of oysters and seagrass along the US Gulf of Mexico coastline**

Patrick, C. J. <sup>\*,1</sup>; Lister, J. <sup>1</sup>

<sup>1</sup>Department of Life Sciences, Texas A&M University Corpus Christi

While areas of the eastern United States have experienced large declines in biogenic foundation habitat over the last century, the estuaries of the Gulf Coast have some of the best remaining oyster reef habitat in North America and extensive seagrass beds. Using gulf coast estuaries, federal databases and geospatial information about foundation habitats, infauna, estuary characteristics, climate, nutrient loads, and upstream watershed land use were summarized to build a structural equation model of factors driving biogenic habitat distributions ( $P = 0.069$ , Chi-Square = 30.666,  $df = 38$ ). Oysters and seagrass covaried among estuaries due to differential responses to nitrogen loads and air temperature. Effects also cascaded to the density and richness of the infaunal assemblages. Nutrient loads predicted water column chl *a*. Oyster coverage was positively related with chl *a*, possibly because of its food value. In contrast, chl *a* reduced water clarity and was negatively related with seagrass coverage. Temperature effects may be a proxy for hyposaline conditions during summer. Oysters were lower in high salinity, while seagrass were not due to successional turnover from *Thalassia testudinum* to *Halodule wrightii* and *Syringodium filiforme*. These patterns demonstrate a macroscale interaction between nutrient loads, climate, and living resources.

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