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Sustainable Carbon Dioxide Sequestration as Soil Carbon to Achieve Carbon Neutral Status for DoD Lands

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Abstract

Sequestration of atmospheric carbon dioxide in soils is a promising alternative for mitigation of atmospheric carbon dioxide (CO₂). The Department of Defense (DoD) owns significant land and water resources which can be managed to offset emissions. Accounting for this, sequestration could help DoD reach carbon neutrality. Many activities the DoD engages in for sustainable land management and training sustainment are conducive to soil carbon storage without even considering this as an important component; however, carbon storage could be greatly enhanced by increased understanding of optimal storage conditions and by making slight adjustments to existing practices. Land management techniques may require adjustments to maximize carbon storage while maintaining training and environmental quality. In order to achieve this, data gaps for estimating carbon fluxes need to be addressed so that accurate measurements can be taken. Unknown aspects of carbon storage as it relates to plant-soil-soil microbe interactions need to be investigated to maximize carbon storage while maintaining land use requirements. Geo-engineering concepts require further refinement to increase carbon storage in soils. These knowledge gaps are not insurmountable and could be addressed through focused research to maximize and accurately quantify carbon storage on DoD lands.

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Preface

The work reported herein was conducted at laboratories throughout the Engineer Research and Development Center (ERDC), including the Environmental Laboratory (EL), the Cold Regions Research and Engineering Laboratory (CRREL), and the Construction Engineering Research Laboratory (CERL) under Work Unit 33143. Significant contributions were also made by participants from Argonne National Laboratory (ANL) of the Department of Energy (DoE). This project was approved and funding was supplied by the Center for the Advancement of Sustainability Innovations (CASI).

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At the time of publication of this report Dr. Ilker Adiguzel was Director, CERL, Dr. Joseph L. Corriveau was Director, CRREL, and Dr. Beth Fleming was Director, EL.

COL Bryan S. Green was the Commander of ERDC and Dr. David W. Pittman was the Director of ERDC.

Unit Conversion Factors

Multiply	By	To Obtain
acres	4,046.873	square meters
cubic feet	0.02831685	cubic meters
cubic inches	1.6387064 E-05	cubic meters
cubic yards	0.7645549	cubic meters
degrees Fahrenheit	$(F-32)/1.8$	degrees Celsius
feet	0.3048	meters
gallons (U.S. liquid)	3.785412 E-03	cubic meters
hectares	1.0 E+04	square meters
inches	0.0254	meters
microns	1.0 E-06	meters
miles (U.S. statute)	1,609.347	meters
ounces (mass)	0.02834952	kilograms
ounces (U.S. fluid)	2.957353 E-05	cubic meters
pints (U.S. liquid)	4.73176 E-04	cubic meters
pints (U.S. liquid)	0.473176	liters
pounds (mass)	0.45359237	kilograms
pounds (mass) per cubic foot	16.01846	kilograms per cubic meter
pounds (mass) per cubic inch	2.757990 E+04	kilograms per cubic meter
pounds (mass) per square foot	4.882428	kilograms per square meter
pounds (mass) per square yard	0.542492	kilograms per square meter
quarts (U.S. liquid)	9.463529 E-04	cubic meters
square feet	0.09290304	square meters
square inches	6.4516 E-04	square meters
square miles	2.589998 E+06	square meters
square yards	0.8361274	square meters
tons (long) per cubic yard	1,328.939	kilograms per cubic meter
tons (2,000 pounds, mass)	907.1847	kilograms
tons (2,000 pounds, mass) per square foot	9,764.856	kilograms per square meter
yards	0.9144	meters

Abbreviations

AMF	Arbuscular mycorrhizal fungi
CASI	Center for the Advancement of Sustainable Innovation
CERL	Construction Engineering Research Laboratory
CO ₂	Carbon dioxide
CPR	Climate Preparedness and Resilience
CRREL	Cold Regions Research and Engineering Laboratory
DoA	Department of the Army
DoD	Department of Defense
DoE	Department of Energy
EL	Environmental Laboratory
EMF	Ectomycorrhizal fungi
EO	Executive Order
EPS	Exopolysaccharide
ERDC	Engineer Research and Development Center
ESM	Earth System Models
GAO	General Accounting Office
GHG	Greenhouse Gases
IPCC	Intergovernmental Panel on Climate Change
LULC	Land Use-Land Cover
MODIS	Moderate Resolution Imaging Spectroradiometer
NBCD	National Biomass and Carbon Dataset
NDVI	Normalized Difference Vegetation Index
NEON	National Ecological Observatory Network
NLCD	National Land Cover Database
NRCS	Natural Resources Conservation Service
NSF	National Science Foundation
ORC	Organizational Readiness to Change
ORIC	Organizational Readiness to Implement Change
PAR	Photosynthetically Active Radiation
RaCA	Rapid Carbon Assessment
SEIA	Solar Energy Industries Association

SOC	Soil Organic Carbon
SSURGO	Soil Survey Geographic Database
USACE	U.S. Army Corps of Engineers
USDA	United States Department of Agriculture

1 Overview of Carbon Sequestration

Federal agencies, including the Department of Defense (DoD), are accounting for greenhouse gas emissions. However, effects of landscape and land use management on soil emissions and soil sequestration (ecosystem services) are not yet considered completely in calculations towards meeting emissions reduction goals (Council on Environmental Quality, Federal Greenhouse Gas Accounting and Reporting Guidance 2016) and scientifically defensible estimates are lacking in many of these areas. The DoD is a major landowner and end-user of naturally vegetated lands with significant net carbon sequestration potential. Carbon offsets due to sequestration on these lands could be a major source of compliance for future requirements, if and when inclusion is allowed. Use of DoD land for sequestration could also contribute to the required goal of achieving “carbon neutral” status for DoD facilities and installations. While a significant body of literature and products (including predictive models) exist for carbon accounting, due to the highly variable and unique nature of DoD lands and land uses, the applicability of these models for DoD lands is not known. Also, land end-users often express reluctance to accept changes in soil and wetlands management strategies without supporting information due to potential large capital investments that may be required for the implementation of new land management techniques. Moreover, most accounting procedures only consider snapshot one-time estimates without estimating natural or disturbance modulated variability over time.

1.1 Carbon footprint of the DoD and impact on missions

The DoD is the largest energy using entity in the world (Solar Energy Industries Association (SEIA) 2013). Most of the DoD’s energy comes from fossil fuels. In 2011, it was estimated that the DoD consumed over 5 billion gallons of oil. Consequently, the DoD is also one of the most important sources of carbon emissions to the world’s atmosphere, with CO₂ emissions of 52.2 metric tons as estimated in 2010 (Walker 2011).

Carbon emissions affect DoD missions primarily because they relate to energy consumption and fossil fuel use. Lowering carbon emissions can be equated to lower fuel consumption, reducing costs, and making DoD entities less susceptible to supply issues associated with fossil fuels.

Carbon footprint can also affect DoD emissions by affecting global climate change. In many cases, climate change is causing, and is expected to further cause, increased frequency of drought and low precipitation. Water supply issues are contributing to worldwide conflict and may also affect water supply issues during military operations (CNA 2014). Climate change is also resulting in an increase of intensity of large storms. Climate change has already been linked to a 3 inch rise in sea level, which contributes to an increase in flood events worldwide (AECOM 2013), including recently in south Florida. In one case, 31 communities in Alaska are in danger of inundation and may have to be relocated (General Accounting Office (GAO) 2009). The costs of relocation or improvements are estimated by the U.S. Army Corps of Engineers (USACE) to range from \$150 to \$250 million per community (Herrman 2015; USACE 2006). Climate change is expected to affect wildlife habitat on Army installations, including those of threatened and endangered species (Lozar et al. 2011). It may also increase the frequency and intensity of Harmful Algal Blooms in USACE reservoirs (Medina et al. 2016) as well as potentially affecting potable water supplies for Army installations.

1.1.1 Executive Order 13693

In 2013, the Executive Office of the President of the United States released a Climate Action Plan, which identified climate change as a critical National issue and committed the United States to reducing its carbon footprint (Executive Office of the President 2013). On 19 March 2015, President Barack Obama issued Executive Order (EO) 13693, Planning for Federal Sustainability in the Next Decade (<https://www.whitehouse.gov/the-press-office/2015/03/19/executive-order-planning-federal-sustainability-next-decade>). This document focuses on energy sustainability and greenhouse gas emission reductions. The EO set a goal of 40% reduction in greenhouse gases by Federal entities within 10 years.

1.1.2 DoD approach to reducing carbon footprint

The DoD has already begun efforts to reduce its carbon footprint. Currently, these efforts fall under one of the following three options: reduction of fuel use due to changes in human behavior (see memorandum from Under Secretary of Defense urging Energy Awareness (<http://img.docstoccdn.com/thumb/orig/42642773.png>)), reduction of fuel use due to technological controls (for example, use of more efficient exterior lighting (Department of the Army (DoA) 2010), and reduction of fuel use due to

alternative energy, such as the use of solar energy (SEIA 2013). These methods have tremendous value and should definitely be encouraged and developed. However, the easily accomplished opportunities using these approaches has already been implemented in many cases. Further, these approaches cannot by themselves result in an installation that emits zero net carbon – a carbon neutral installation. This report proposes to explore carbon sequestration using the vast lands owned by the DoD. This approach could greatly decrease the carbon footprint for the DoD, and may legitimately open up opportunities for carbon neutral or even carbon negative installations.

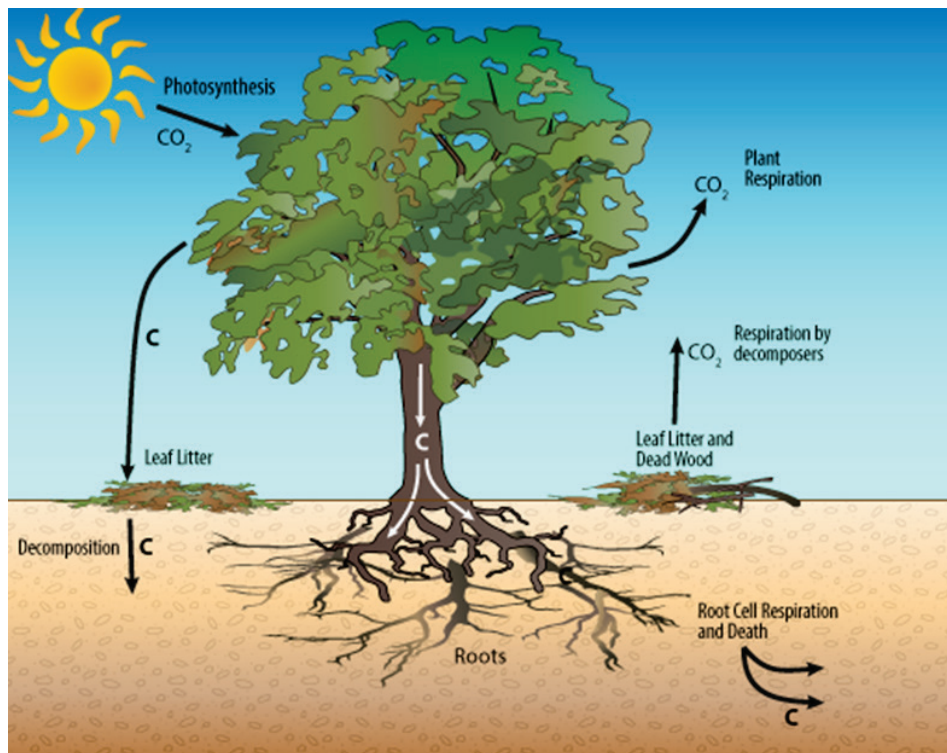
1.2 Soil and the carbon cycle

Carbon is a naturally-occurring element that is one of the key building blocks of all living things on Earth. Carbon is exchanged by photosynthesis, respiration, digestion, and decomposition. Carbon emissions refers to carbon that is released to the atmosphere by any process. It is the flux of carbon into and out of the environment that is the carbon cycle (Lal 2001).

Some of the carbon on earth is stored over long periods of time. The places where carbon is stored are called sinks, and include the deep ocean, rocks, fossil fuel deposits, soils, sediment at the bottom of lakes, reservoirs and wetlands; and woody vegetation (such as forested areas). Carbon storage is not permanent. For example, erosion can transport carbon from top soils by stream flow to reservoirs, where that carbon can become trapped in a new sink; the sediment at the bottom of the reservoir. A source is any living organism, or sink, that releases carbon to the active cycle; anything that accumulates and stores carbon can be a sink. Many critical components of the carbon cycle (such as forests) can serve as both sources (e.g., when burned) or sinks (when growing) for carbon, depending upon its state. Fossil fuels (coal, oil, and gas) are formed over long periods of time, and represent the slow, steady removal of a great deal of carbon from earth's atmosphere over millennia (sink). The burning of fossil fuels constitutes a very rapid release of this carbon over a handful of decades (source). At the same time, some important carbon sinks (such as tropical rainforests and coastal wetlands) have been degraded through development. Carbon is currently being added to the atmosphere faster than it can be absorbed by sinks.

The natural carbon cycle has remained relatively stable for thousands of years. Carbon enters the atmosphere as carbon dioxide (CO_2) from respiration and combustion processes. The atmospheric CO_2 is adsorbed by plants to make carbohydrates through photosynthesis. Plants grow through development of root systems and aboveground biomass. At the end of the growing season, these plant parts decay and are assimilated into the soil as organic matter (another form of carbon). The dead organic matter is food for microbes and fungi which respire some of the carbon as CO_2 back into the atmosphere. However, not all the carbon used by the microbes and fungi is respired; decomposition of organic matter can also result in the production of humic material. Humic material stores carbon in the soil on a geologic time scale. A summary of the carbon cycle is presented in Figure 1.

Figure 1. Simplified diagram of the carbon cycle through the atmosphere, soil and water (Credit: Valerie Martin, Technical Education Research Center (TERC)).



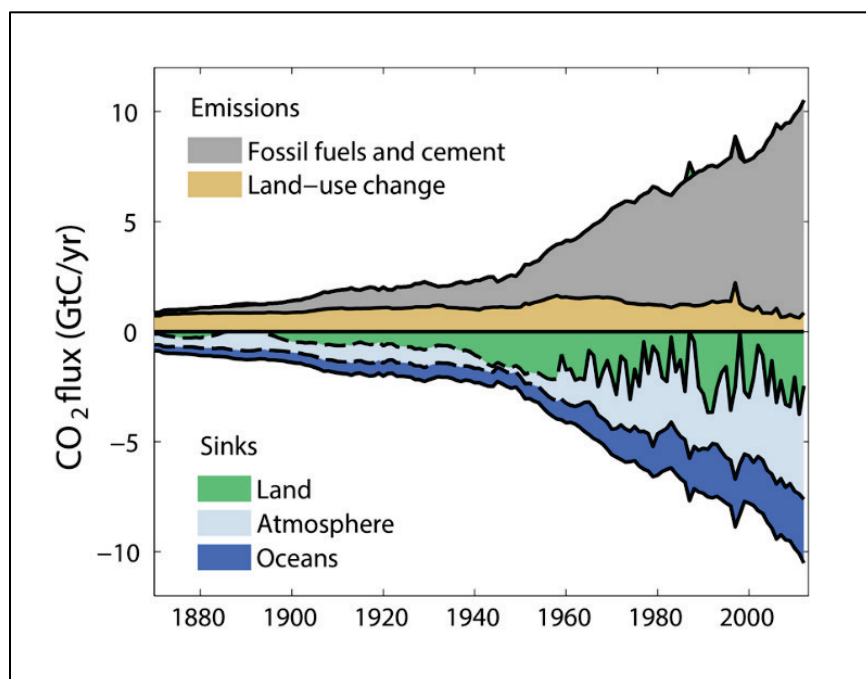
1.3 Soil sequestration of carbon

Carbon sequestration is the process of extracting carbon from the atmosphere and the active part of the carbon cycle and placing it in long term storage in natural or artificial carbon sinks. At the global scale, natural carbon sinks include the absorption of CO_2 by oceans and other

waterways through physiochemical and biological processes, photosynthesis by vegetation and sequestration of carbon in soil organic matter. This report addresses carbon sequestration on DoD lands and facilities within the continental U.S. It is therefore concerned with the terrestrial and aquatic portions of the carbon cycle. The major sinks of interest include soils, ecosystems, and reservoirs/lakes and wetlands (freshwater and marine), with an eye towards accounting for the amount of carbon that is currently sequestered in these sinks, and how much more could be sequestered in the future through changes in land management or facility use.

Land is a critical sink for carbon (Figure 2). The historical maximum concentration has been 2.8 trillion tons of organic carbon. The present concentration is 2.3 trillion tons. Carbon released from soil and not replaced is estimated to be between 200–500 billion tons. The hypothesis presented in this paper is that the soil carbon sink can be increased, and increased quickly. As the soil becomes a better carbon sink, the CO₂ levels in the atmosphere will decrease. Carbon in the ocean will also decrease as it equilibrates with the lower atmospheric CO₂.

Figure 2. Flux of carbon inputs and carbon sinks from 1880 to the present (Source: Canadell 2007, Global Carbon Project).



1.4 Study objectives

To better address EO 13693 directing all DoD installations to achieve carbon neutral status, and also meet the requirements of freshwater and coastal wetlands regulations, this Center for the Advancement of Sustainability Innovation (CASI) white paper will elucidate challenges and identify the data gaps to reducing atmospheric CO₂ levels through sequestration of carbon in natural soil sinks. This paper will explore the state of knowledge regarding carbon fluxes in the environment as they pertain to specific activities and requirements on DoD lands. Specific research areas will be identified to improve predictive models. Activities and requirements that influence carbon flux on DoD lands will be compared to existing general sequestration models to determine what differences exist and to what extent they differ. Also, investigative approaches to increasing carbon sequestration in soil that will engage land users, and clarify management and environmental drivers will be explored.

1.5 Research drivers

- EO 13693 (2015) directing reductions of carbon, with an overall goal of 40% reduction from 2008 levels by 2025.
- U.S. Army Net Zero Energy – Net Zero Energy is a goal for Army Installations to attempt to achieve a Net Zero energy status. This includes energy reductions, which impact carbon emissions. It can also involve the use of biomass to generate energy, which could be included in a sequestration program.

1.6 Stakeholders/Partners

It is important to identify the potential stakeholders within the DoD as the specific needs of each group will differ as they attempt to achieve carbon neutral status for the lands under their stewardship. That is, managers of active firing ranges will have different concerns from those of the DoD industrial base, or from those of primarily research facilities, or wetlands under USACE management. All DoD, generally, and USACE, particularly, are considered stakeholders in the development of techniques to enhance carbon sequestration in soil.

2 Measuring, Monitoring and Verifying Soil Carbon Change

Re-carbonization of the terrestrial biosphere by restoring carbon pools in Earth's soils, trees, and wetlands is an important step toward both mitigating and adapting to abrupt climate change. Carbon sequestration in terrestrial ecosystems has numerous co-benefits, such as the following:

- increasing net primary productivity
- advancing food security
- improving the quality and quantity of water resources
- enhancing biodiversity (Lal 2010).

Under the United Nations Framework Convention on Climate Change, land management activities that lead to change in carbon concentration of soils can be accounted as a carbon sink, or source, in generating and reporting greenhouse gas (GHG) inventories and using it for payment for ecosystem services through trading carbon credits (Intergovernmental Panel on Climate Change (IPCC) 2013). However, in order to have a successful carbon credit exchange and banking system, changes in carbon concentration of soils needs to be quantified and verifiable.

2.1 Measuring soil carbon

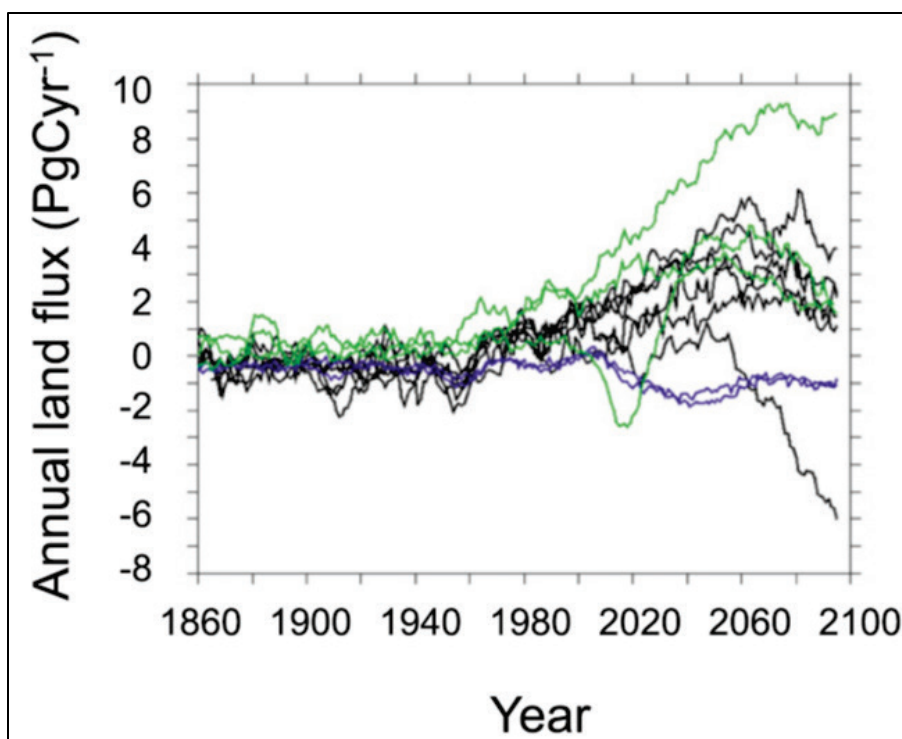
Measurements of soil carbon concentration in a laboratory have been conducted since the 1860s. The principal objective of such measurements has been the evaluation of soil fertility in the vegetation root zone as influenced by agronomic practices such as tillage, crop rotations, and fertilization. Both wet and dry combustion methods are used for these kinds of measurements. Conventional units of reporting soil organic carbon (SOC) and total carbon concentrations are percentages or gram-per-kilogram measurements on a dry-weight basis. Uncertainties in the data on measurements of carbon concentrations are caused by spatial variability in soil properties (e.g., soil bulk density).

In the context of evaluating soil carbon sink capacity of managed ecosystems, the objective is to quantify management-induced changes in the total SOC pool (Pg carbon per ha) and compute the rate of change with reference to baseline (Pg carbon per ha per year) over a landscape, farm, watershed, or regional scale (Lal 2010). Several broad field measurements

have been performed such as that reported for the entire lower United States by the Natural Resources Conservation Service, National Soil Survey Center (West et al. 2013). The Rapid Carbon Assessment (RaCA) project sampled for SOC and other soil variables, collecting 144,833 samples from the upper meter of 32,084 soil profiles at 6,017 randomly selected locations. Measurements were taken of both organic and inorganic carbon by visible and near infrared spectroscopy. Sites from the National Resources Inventory were used as the basis for random selection of sample sites stratified by soil group within RaCA region and by land use-land cover (LULC) classification within soil group. The RaCA regions were identical to the areas of responsibility of the Natural Resources Conservation Service (NRCS) Major Land Resource Area Regional Offices at the time of sampling. At each sample site, soil morphology and landscape characteristics were described and some limited information regarding vegetation and agricultural management was collected.

Land/soil carbon sensitivity to environmental change is a dominant source of uncertainty in predicting carbon-climate feedbacks (Burke et al. 2012; Friedlingstein et al. 2014). Figure 3 shows the increase in soil C flux detected after 1900 and the significant increases in the last half of the 20th century (Friedlingstein et al. 2014).

Figure 3. Soil organic carbon sensitivity to environmental change (Friedlingstein et al. 2014).



In Figure 3, the colored lines represent the summary of 11 specific climate models; blue/purple lines are for earth system models (ESMs) accounting for a terrestrial nitrogen cycle, green lines are for ESMs describing emissions of CO₂ due to land use change calculated as anthropogenic influence.

The carbon content of soils may increase or decrease with climate and land use change. The major issue is to assess whether these changes are detectable through soil monitoring efforts, taking into account the uncertainties caused by spatial heterogeneity, sampling methods, analytical errors and study scale (Lal 2010). The evaluation of the confidence with which changes in carbon content can be detected is important for implementing any soil or carbon monitoring program.

Soil properties, including temperature, moisture, texture, and SOC, vary vertically from surface to bedrock along the soil horizons. The SOC turnover rate decreases with depth; knowledge of the depth distribution of SOC, and its proper representation in models, is critical for predicting the response of subsoil carbon to environmental change and potential climate feedbacks (Trumbore 2009; Marin-Spiotta et al. 2011).

There are several mechanisms by which SOC can move through the soil profile, such as:

- Plant roots and root exudates (Jackson et al. 1996; Rasse et al. 2006). While root carbon flux is still poorly understood, what is known is that grasses have the shallowest root profiles, trees are intermediate and shrubs have the deepest root profiles. Also, the mean residence time of root-derived SOC is 2.4 times higher than that of shoot derived SOC (Rasse et al. 2005). The factors that affect root carbon flux remain to be elucidated.
- Bioturbation (Cernansky 2016; Kristensen et al. 2012). Earthworms, nematodes, termites and ants, for example, are responsible for incorporating large masses of surface organic matter, and nutrients, into the deeper layers of the soil. Soil fauna increase litter decomposition rates significantly in many regions of the country (Wall et al. 2008). Changes in climate have been reported to alter the nematode populations, which resulted in changes in litter decomposition rates. A follow-up study by Garcia-Palacios et al. (2013) and the Garcia-Palacios et al. corrigendum (2013), estimated that

- excluding soil fauna would reduce the decomposition rate by a global average of approximately 37%. Thus, a small change in atmospheric carbon could result in larger changes in soil carbon flux. The nuances in this feedback loop are one area that requires further research to fully understand its significance.
- Vertical transport of SOC through the following:
 - Cryoturbation, the mixing of different soil horizons due to freezing and thawing (Ping et al. 2015; Bockheim 2007)
 - Podzolization, the process by which the upper layers of soil become acidic through leaching of bases which are deposited in the lower soil horizons (White 2006; Lundstrom et al. 2000)
 - Preferential flowpaths such as tunnels left by dead roots and abandoned ant and termite colonies (Chabbi et al. 2009; Bundt et al. 2001).

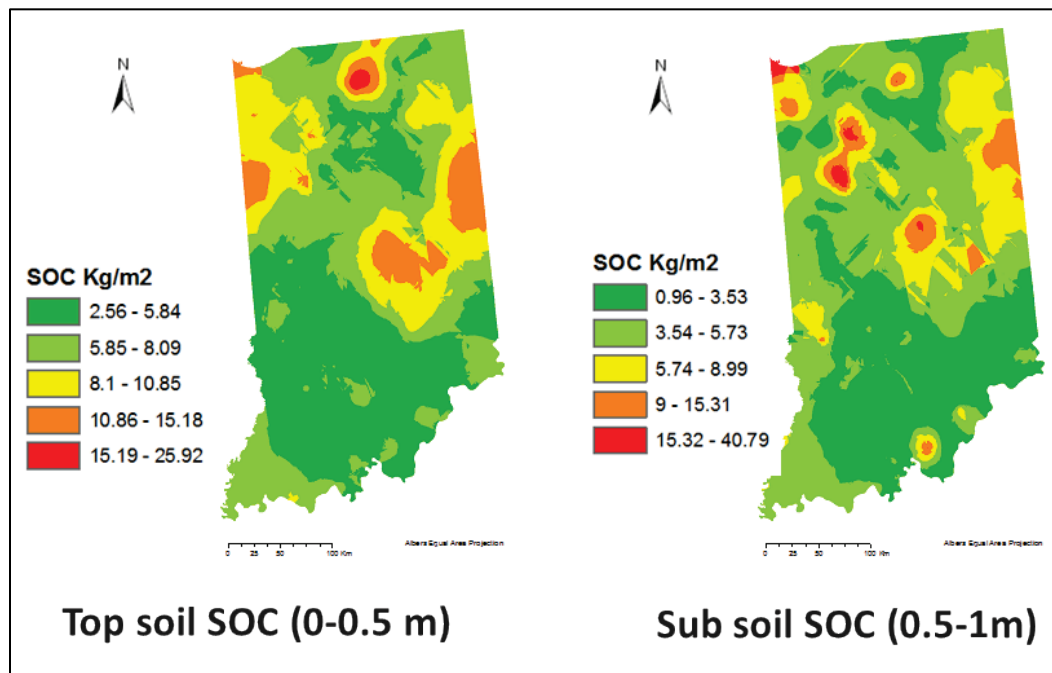
 - Human factors such as ploughing and ponding soils for cultivation.

This spatial variability of SOC content strongly influences the ability to detect changes, particularly at field-scale (Conant and Paustian 2002; Conant et al. 2003). Numerous field-scale studies have addressed this issue (for example, Conant and Paustian 2002; Garten and Wullscheleger 1999; Conant et al. 2003; Conen et al. 2003; Saby and Arrouays 2004; Smith 2004). Mishra et al. (2009) reported the depth distribution of SOC stocks in the state of Indiana including both agricultural and urban areas (Figure 4).

Estimates of soil carbon can also be made using vegetation classifications, vegetation biomass, vegetative growth, and soil carbon data. Vegetation can be classified using the National Land Cover Database (NLCD), a dataset from 2001 to 2011 at a 30 m resolution (Homer et al. 2015). The NLCD divides vegetation into very generalized categories: three forest types (deciduous, evergreen, mixed), shrubs/scrub, grasslands, woody wetlands, herbaceous wetlands, pasture, and cropland. Vegetation biomass is estimated using the National Biomass and Carbon Dataset (NBCD), that provides a 30 m resolution of estimated canopy height, standing biomass, and vegetation above-ground carbon stock estimates (Kellndorfer et al. 2012). Vegetative growth is estimated using Moderate Resolution Imaging Spectroradiometer (MODIS) and Normalized Difference Vegetation Index (NDVI) data. Both MODIS and NDVI

estimate vegetation based on absorbed photosynthetically active radiation (PAR). From this data, plant biomass and net primary productivity can be estimated to obtain relative growth rates and amounts of vegetation change over time. Both MODIS and NDVI provide monthly measurements; NDVI is at 1 km resolution and MODIS is at 500 m resolution. Soil carbon is estimated from the Soil Survey Geographic Database (SSURGO).

Figure 4. Estimating depth distribution of SOC stocks in Indiana when comparing top soil and sub soil concentrations.



2.2 Monitoring soil carbon

In this context, soil monitoring can be defined as taking repeated observations, measurement and evaluation of soils at a particular site along with related environmental or technical data, according to prearranged schedules in space and time, using standardized methods for data collection and analysis (such as provided by Donovan (2013)). Spatial distribution of collected soil samples should be representative of soil type, land use and climate conditions and the spatial arrangement of samples may be random or on a regular grid (Jones and Verheijen 2008).

2.3 Verifying soil carbon models

Considering a single, hypothetical, study site, results from power analysis has shown that soils with low coefficients of variation (<25%) in SOC

require >100 samples to detect changes of 3% C from the background (Garten and Wullschleger 1999). For soils with a large coefficient of variation, prohibitively large numbers of samples are required (Conen et al. 2004). In this case, the study site could be revisited on a second and subsequent occasions, as this is the most efficient way of monitoring SOC change (Lark et al. 2006). For any particular monitoring study of a variable x , n sites are sampled at time t_0 and again at time t_1 . An estimate of the average change (\bar{A}) in x is calculated as;

$$\bar{A} = \sum_{i=1}^n (SOC_{i,t_0} - SOC_{i,t_1}) / n \quad (1)$$

Where, $SOC_{i,t}$ is the SOC measurement at site i at time t and n is the number of sites in the monitoring study. An estimate of the standard error of \bar{A} is $\sqrt{\frac{S_A^2}{n}}$ where S_A^2 is an estimate of the variance of the differences in SOC (de Gruijter et al. 2006).

Several insights can be reported after considering the results of several carbon mapping studies conducted from 2009 to 2014.

- A variety of mathematical functions (spline functions) are presently being used to describe the depth distribution of SOC.
- The most accurate results have been obtained when individual functions are fitted in each soil profile and used in a machine learning approach, as no single function can fit all profiles.
- These mapping studies have consistently found that prediction accuracy decreases with soil depth.
- It has been difficult to find appropriate environmental predictors of model (spline) coefficients for sub-soil carbon stocks.
- Validation R^2 values are substantially lower in comparison to calibration R^2 values, meaning carbon over-estimation by the current models.

There are data gaps in the land models that attempt to predict the depth distribution of SOC stocks. For example:

- Despite discretizing soils vertically for soil moisture and energy calculations, most coupled carbon-climate models typically represent SOC biogeochemistry with a single-layer (Todd-Brown et al. 2013; el Masri et al. 2015).

- Jenkinson and Coleman (2008) developed and added two parameters in RothC-26.3 to describe the subsoil biogeochemistry. One parameter moves organic carbon down the soil profile and a second parameter slows decomposition of organic matter at depth.
- Koven et al. (2013) implemented soil carbon and nitrogen vertical biogeochemistry in a community land model. Most models do not consider nitrogen along with the carbon stocks.
- In general, the land models (Century, RothC, TEM and CLM4.5) assume an exponential decrease of carbon with depth. This assumption has been proven wrong in several instances. Century and RothC represent SOC down to 1 m in depth. CLM4.5 represents SOC biogeochemistry up to 3.8 m in depth. However, tropical soils contain substantial SOC several meters deep.

U Mishra¹ is attempting to represent vertical biogeochemistry of SOC stocks in the Integrated Science Assessment model. The study explored a variety of methods to represent the vertical heterogeneity of SOC including soil type, land cover type, and ecoregion. Also, the study has reported that ecoregion-based stratification of the soil profile data provided the best fit to the observed data. The model now needs better representation of organic layer SOC and the impact of soil wetness.

2.4 Research needs

In summary, estimates of SOC stocks and their representation in models have started to include parameters for vertical heterogeneity.

Representing belowground biogeochemistry will make the most sense in terms of soil horizons (layers at depth). A knowledge gap remains concerning the carbon fluxes in different soil horizons in response to perturbations from cold, biological influence, and human intervention. A current share of uncertainty in model predictions due to soil carbon could be decreased by proper representation of both vertical and lateral heterogeneity. Accounting for both spatial and vertical heterogeneity of SOC is a prerequisite for proper estimates of SOC sequestration or release.

Also, classification of vegetation at only five levels of natural terrestrial vegetation in NLCD removes plant community level estimates, which have significant influences on soil carbon. Plant biomass is estimated from a single data point. Estimating plant activity using NDVI and MODIS is

¹ U. Mishra, Argonne National Laboratory, Lemont, IL pers. comm., July 2016.

troublesome, as algorithms to convert absorbance of PAR into plant growth and production have high uncertainties. Soil carbon data often comes from soil survey maps and soil sample analyses that were conducted for soil types nationwide. Survey maps have inconsistent boundary accuracy and chemical attributes, often have little coverage or replication, and often, only include soil surface horizons.

One option to improve carbon estimates is to directly measure them. Eddy covariance systems are currently in use across the United States to measure carbon flux. These systems are expensive to maintain and are primarily funded by the National Science Foundation (NSF) through the National Ecological Observatory Network (NEON). This network of ecological research stations collects significant environmental data. Partnering by DoD with this network could allow direct measurements of carbon flux on DoD lands, allowing a direct comparison between military training environments and ecosystem preserves for numerous environmental variables. Additionally, many NEON sites are located in close proximity to, and on similar ecosystems representing major DoD installations; this data would allow for direct comparisons (Figure 5). Many of these paired locations represent the major ecosystems of the United States (Table 1).

Figure 5. NEON sites and major DoD installations across the United States. Many DoD installations are in close proximity to NEON sites and occur on comparable ecosystems.

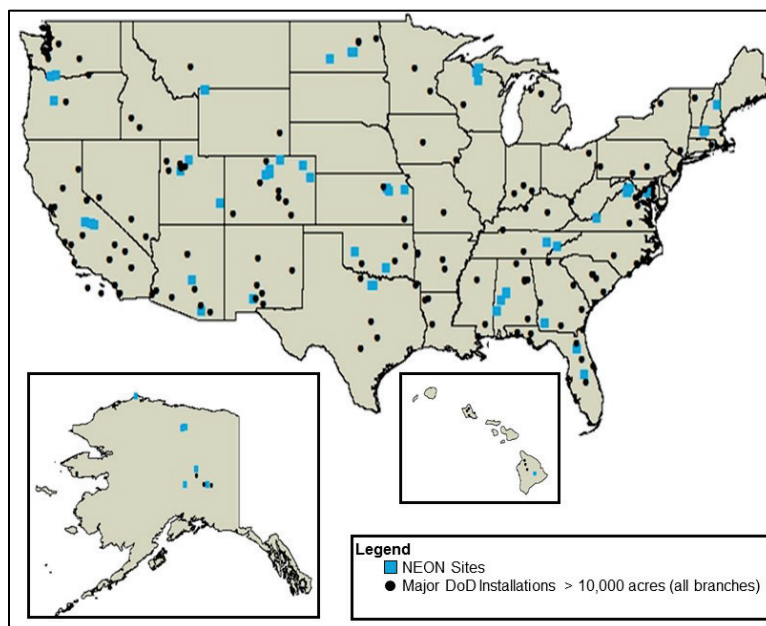


Table 1. Paired NEON sites and DoD installations in close proximity with comparable ecosystems.

NEON Site	DOD Installation	Distance (miles)	Ecosystem
Delta Junction	Fort Greely	0	Taiga
Konza Prairie	Fort Riley	22	Prairie
Santa Rita	Fort Huachuca	52	Desert
Jornada	Fort Bliss	52	Desert Grassland/ Shrubland
Unaqui-Ault	Dugway Proving Ground	62	Shrubland
LBJ National Grassland	Fort Sill	120	Grassland
Jones Ecological Research Center	Fort Benning	109	Longleaf Forest
Jones Ecological Research Center	Eglin Air force Base	162	Longleaf Forest
Oak Ridge National Lab	Fort Campbell	221	Oak-Hickory Forest
Oak Ridge National Lab	Fort Knox	227	Oak-Hickory Forest

3 Plant Processes Controlling Carbon Flux

3.1 Plant variation

Variation in vegetation occurring in ecosystems has a profound influence on soil carbon dynamics. This variation is a product of environment and interactions between plant species. While forests produce the greatest carbon stocks, open environments (such as, steppe, tundra, desert) produce the greatest *relative* amounts of soil carbon stocks (Table 2). This is due to a greater root:shoot ratio in these environments, where competition for soil resources is greater than for sunlight.

Table 2. Estimates of carbon stocks in ecosystems (Adams 2016).

Ecosystem type	Vegetation (kg C m ⁻²)	Soils (kg C m ⁻²)	Litter/Debris (kg C m ⁻²)	Total (kg C m ⁻²)
Giant Conifer Forest	35	26	12	73
Warm Temperate Forest	19	15	4	37
Cool Temperate Forest	16	14	3	33
Main Taiga	8	22	2	32
Southern Taiga	14	14	2	29
Tropical Montane Forest	13	13	2	28
Moist Steppe	1	25	0	26
Forest Steppe	1	22	1	24
Lowland Tundra	1	21	0	22
Temperate Woodland	10	10	2	21
Forest-Tundra	1	17	2	20
Open Boreal Woodland	5	13	2	19
Mediterranean Forest	10	8	1	19
Mediterranean Scrub	4	6	1	11
Temperate Scrub	5	5	1	10
Dry Steppe	1	8	0	8
Temperate Semi-Desert	.5	6	0	6
Steppe-tundra	.5	6	0	6
Montane/Dry Tundra	.5	5	0	6
Polar/Montane Desert	.1	0	0	.1
Temperate Desert	.1	0	0	.1

Root depth is important for soil carbon retention; deeper roots contribute greater proportions to long term soil carbon pools. Soils in the top 0.2 m contain 615 Gt of carbon with a mean residence time of a few hundred years, while soils in the top 3 m contain 2,344 Gt of carbon (more than all the carbon in the atmosphere) with a mean residence time of 2–10 millennia (Fontaine et al. 2007). This increase in residence time with soil depth is a result of slower decomposition due to fewer nutritional resources to support decomposition at increasing depth (Spohn et al. 2016). Rooting depth is primarily a function of plant type, but can also be influenced by climate and soil type, where total SOC increases with precipitation (surface soils) and clay content (deep soils) (Jobbagy and Jackson 2000), but appear to be more dependent on soil type than climate (Mathieu et al. 2015). Root:shoot ratios, rooting depth, root biomass, and root depth distribution all vary by ecosystem (Table 3). Generally, trees have the highest maximum rooting depths, followed by shrubs and finally herbaceous vegetation. Forbs are highly variable, but certain species can produce very deep roots. Also, perennials tend to have deeper roots than annual plants. The C₄ grasses (they produce a four-carbon sugar, oxaloacetate, during the Calvin cycle of photosynthesis) also have deeper roots than C₃ grasses (which produce two, three-carbon sugars during the Calvin cycle). The C₄ plants are more energy intensive and bring more CO₂ into the process. Examples of C₄ plants include corn or maize, sugarcane, sorghum and different millets, as well as switchgrass. C₃ plants include beans, rice, wheat, and woody trees.

Table 3. Root properties of ecosystems (Canadell et al. 1996; Jackson et al. 1996).

Biome	Root: Shoot	Maximum Rooting Depth (m)	Root Biomass (kg m ⁻²)	% in Top 30 cm
Tropical Grassland/Savanna	0.7	15	1.4	57
Desert	4.5	9.5	1.2	53
Tropical Evergreen Forest	0.19	7.3	4.9	69
Sclerophyllous Shrubland	1.2	5.2	4.8	67
Temperate Coniferous Forest	0.18	3.9	4.4	52
Tropical Deciduous Forest	0.34	3.7	4.1	70
Temperate Deciduous Forest	0.23	2.9	4.2	65
Temperate Grassland	3.7	2.6	1.4	83
Cropland	0.1	2.1	0.15	70
Boreal Forest	0.32	2	2.9	83
Tundra	6.6	0.5	1.2	93

3.2 Rhizosphere processes

3.2.1 Plants

Up to 60% (higher for perennials than annuals) of the carbon fixed by plants through photosynthesis ends up in the soil, comprising 30–40% of the total SOC input while occupying only 2–3% of soil volume (Grayston et al. 1996; Jones et al. 2004). As plants grow, roots are constantly dying and being replaced. This root turnover is a primary source of soil organic matter, as the dead roots are decomposed by soil microbes. Root turnover is highly variable and dependent on a number of factors, including root size, environmental conditions, and plant type (Table 4).

Table 4. Estimates of root turnover by ecosystem (Lauenroth and Gill 2003).

Ecosystem	Root Turnover (Years)
Temperate Conifer Forest	0.08 - 2.57
Temperate Deciduous Forest	0.18 - 1.42
Boreal Forest	0.26 - 0.34
Temperate Grassland	0.47
Temperate Shrubland	0.44

As noted in Chapter 2 of this report, roots in the rhizosphere are also a significant source of bioturbation and carbon turnover. Tunnels left by dead roots contribute to higher moisture levels in the deeper soil horizons that are conducive to organic matter decomposition.

3.2.2 Bacteria

Rhizosphere microorganisms associated with roots and, roots in the boreal forest, specifically, have been shown in models to store up to 70% of the stored carbon in a chronosequence, a set of forested sites that share similar attributes but are of different ages (Clemmensen et al. 2013). Up to 10% of the total carbon fixed by plants through photosynthesis can be exuded by plant roots (Jones et al. 2004). These exudates are comprised of sugars and other polysaccharides; amino, fatty and organic acids; enzymes; sterols; vitamins; nucleotides; and numerous other secondary metabolites and miscellaneous molecules (Uren 2001). One of these exudates is a gel-like substance excreted into the soil surrounding the roots and known as an extracellular polysaccharides (EPS), or simply, a biopolymer. The natural functions of the EPS in the rhizosphere include surface adhesion between

soil particles, self-adhesion of cells into biofilms, formation of protective barriers, water retention around roots, and nutrient accumulation (Laspidou and Rittmann 2002), as well as pathogen resistance, stress response and protection, communication with soil microbes, and reduction of competing neighbors (Uren 2001). Addition of concentrated biopolymer to the soil results in increased seed germination success, increased above and below-ground biomass and reduced soil loss through erosion (Larson et al. 2016). These attributes tend to conserve soil carbon.

3.2.3 Fungi

Most plant species associate with beneficial root fungi, called mycorrhizal fungi. These fungi form secondary root structures that can access and chemically alter soil resources to increase nutrient availability to plants. Ectomycorrhizal fungi (EMF) associate primarily with conifers, although they form associations with many deciduous trees and other plants. Arbuscular mycorrhizal fungi (AMF) associate primarily with grasses and forbs, but also form associations with many trees and shrubs. EMF consist of thousands of fungal species that associate with a narrower range of hosts (around 2% of all plant species), while AMF consist of hundreds of fungal species that associate with a very wide range of hosts (around 80% of all plant families). EMF can function as free-living saprophytic fungi in the absence of plant hosts, while AMF cannot survive without a plant host association. EMF form sheaths around fine root tips and do not penetrate root cells, while AMF form vast hyphal networks within plant roots that penetrate root cells. Although there are immense differences between the two mycorrhizal types, they both exert substantial influences on soil carbon cycling.

AMF can produce over 100 m of hyphae per g of soil (Miller et al. 1995), can grow at a rate of more than a m per day per g soil, and have a lifespan of less than a week (Staddon et al. 2003). Further, these hyphae produce a substance, called glomalin, that binds soil particles together to increase soil aggregation, and alone can comprise more than a quarter of the SOC pool (Nichols 2003). Thus, AMF have a major role in SOC storage.

However, not all AMF are created equal. The order Glomerales represents the largest group of AMF, which generally provide the greatest benefits to their plant hosts but produces less soil hyphae and glomalin. The family *Gigasporaceae*, on the other hand, produces extensive soil hyphae that

produce greater amounts of glomalin and appear to be better adapted to soils low in nitrogen.

EMF hyphal production averages around 160 kg ha⁻¹ per growing season in the top 10 cm of soil across a range of forest types, with the highest production occurring in warm temperate forests such as exist in the southeastern United States (Ekblad et al. 2013). In a longleaf pine forest, EMF biomass was 300 kg ha⁻¹ with a production of 2800 kg ha⁻¹ and turnover around a month (Hendricks et al. 2016).

Across ecosystems in North America, forests are generally dominated by EMF-associated trees (Table 5) with understories dominated by AMF-associated plants. In grasslands and shrublands, AMF occur almost exclusively. Woodlands are co-dominated by EMF and AMF. EMF appear to be more effective at increasing carbon storage in soils through their ability to produce and release enzymes to degrade organic matter (Averill et al. 2014). This degradation increases nitrogen availability, which is then taken up by the EMF for plant utilization. Alternatively, AMF appear to acquire nitrogen through priming of soil bacteria, which increases mineralization of SOC (Brzostek et al. 2015; Paterson et al. 2016). However, the exact mechanisms and net impacts of these activities are not fully understood.

Table 5. Mycorrhizal status of common North American trees.

Genus	Common Name	Mycorrhizal Status
<i>Acer</i>	Maple	AMF
<i>Celtis</i>	Hackberry	AMF
<i>Fraxinus</i>	Ash	AMF
<i>Juglans</i>	Walnut	AMF
<i>Juniperus</i>	Juniper	AMF
<i>Liquidambar</i>	Sweetgum	AMF
<i>Liriodendron</i>	Tuliptree	AMF
<i>Platanus</i>	Sycamore	AMF
<i>Betula</i>	Birch	EMF
<i>Fagus</i>	Beech	EMF
<i>Carya</i>	Hickory	EMF
<i>Picea</i>	Spruce	EMF
<i>Pinus</i>	Pine	EMF
<i>Quercus</i>	Oak	EMF

Genus	Common Name	Mycorrhizal Status
<i>Tilia</i>	Basswood	EMF
<i>Alnus</i>	Alder	Both
<i>Crataegus</i>	Hawthorn	Both
<i>Populus</i>	Cottonwood/Poplar	Both
<i>Prunus</i>	Cherry	Both
<i>Pseudotsuga</i>	Fir	Both
<i>Salix</i>	Willow	Both
<i>Tsuga</i>	Hemlock	Both
<i>Ulmus</i>	Elm	Both

4 Land Management Effects on Soil Carbon Flux: Strategies to Maximize Carbon Sequestration in Soil

There are several major carbon soil sinks available on DoD installations and facilities, as shown in Figure 6. Many facets of land management efforts on military lands already promote soil carbon storage. Maintenance of desirable perennial native plant communities for soil stabilization, habitat quality and training realism tend to increase soil carbon. Management practices directed towards improving annual vegetation production, whether through forest growth or agricultural practices, also generally increase residue carbon input into the soil (Table 6). This results in greater soil aggregate stability and greater aggregate associated SOC levels. The soil has an increased capability of long-term carbon stabilization (Kong et al. 2005). However, more adjustments could be made to current practices that increase soil carbon storage without negatively impacting other requirements. These land management changes could significantly increase carbon soil sequestration.

Figure 6. Land sinks for atmospheric carbon available on many DoD installations on facilities.

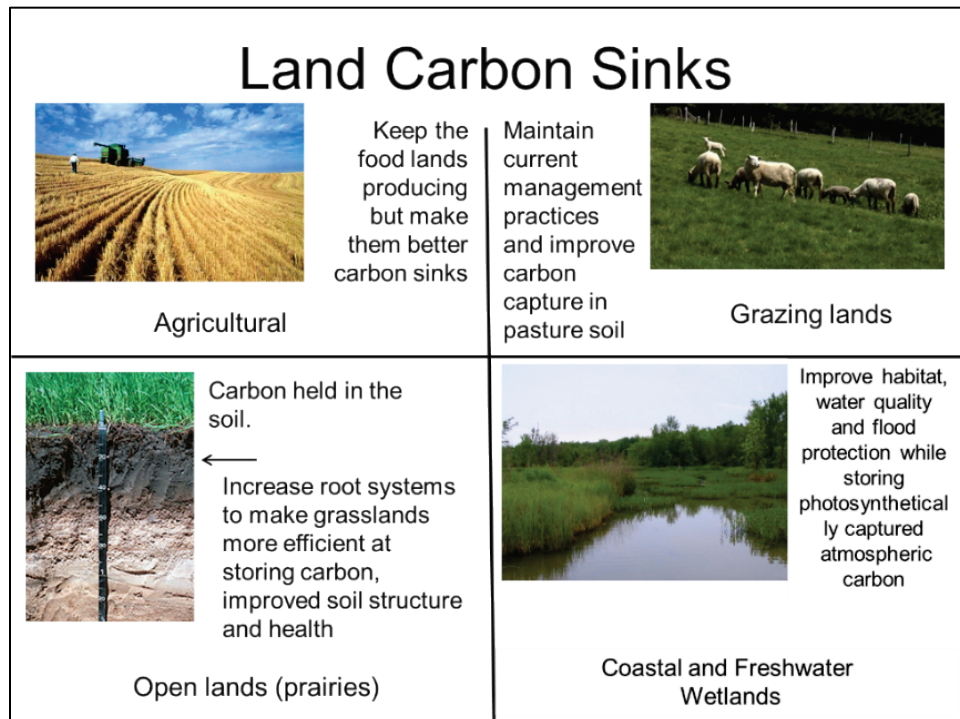


Table 6. Summary of inputs into soil that either increase/decrease soil carbon.

Carbon Inputs	Effect on Soil Carbon (Gain/Loss)
Increase in root mass	Decreased soil erosion (carbon gain) Increase rhizosphere activity (carbon gain/ carbon loss)
Decrease in root mass	Increased soil erosion (carbon loss) or re-deposition of C at depth (carbon gain)
Increase in soil microbial communities	Increase soil respiration (carbon loss)
Increase in microbial exudates	Increase in soluble soil carbon leaching to groundwater (carbon loss, but not to the atmosphere)
Increase in ectomycorrhizal fungi (EMF)	Increased generation of soil humic material (carbon gain)
Increase in bioturbation and soil carbon spatial distribution by soil fauna (nematodes, ants, earthworms, etc)	Increase in decomposition of soil surface litter (carbon gain)
	Carbon from decomposition respired by bacteria (carbon loss) or converted to soil humics (carbon gain)
Use of soil amendments (biochars, biopolymers)	Increase carbon input into soil (carbon gain), decrease erosion (carbon gain)

However, some carbon inputs could result in either a gain or a loss of soil carbon. The factors that determine gain/loss are not always well understood and provide much uncertainty in modeling efforts.

4.1 Forestry

The RaCA by the NRCS found that forests are the LULC class showing the highest mean SOC stocks (West et al. 2013). Old growth forests continue adding carbon to the soil even at old ages, averaging 0.13 kg C per m² soil per year (Luysaert et al. 2008). Further, shoot carbon in forests is estimated to increase 1–2% over the previous year growth (Alexandrov 2007). Timber harvesting can cause soils to become net sources of carbon for decades. While forest management is often economically driven, adjusting harvesting schedules to maximize carbon production and minimize losses could dramatically reduce soil carbon mineralization and increase sequestration of atmospheric carbon. Managing forests to maximize ecosystem services, such as carbon storage, could be achieved just by comparing common practices to determine their impacts on important, yet overlooked, ecosystem characteristics.

Ongoing research on military lands is comparing forestry management methods to identify their impacts on mycorrhizal fungi that play a vital

role in soil carbon storage¹. Preliminary data (Anna 2009)(R. Busby¹ pers. comm., July 2016) suggests that fires have a much greater impact on AMF than EMF, and herbicide utilization to control unwanted woody vegetation facilitates increases in AMF density. Removal of understory vegetation has shown mixed effects on mycorrhizal fungi. Some forests show no effect of removal on SOC content, while others show significant losses¹. Additional research is needed to understand how specific management activities affect SOC accumulation and other ecosystem services.

4.2 Grazing/Agriculture

Perennial plants produce greater root biomass and increase SOC compared to annual plants. Also, perennial crop plants negate the need for annual soil tillage that releases tremendous volumes of soil carbon. Soil carbon storage on agricultural outlease lands could be increased by:

- Converting these lands to perennial cropping systems,
- Shifting to biomass fuels production to reduce GHG emissions from energy production, and by
- Adding additional carbon to soil through waste paper mulch or biopolymer soil amendment once every few years (frequency not yet established).

Grazing also impacts SOC storage through grazing intensity, rotation timelines, and type of vegetation. However, since grazing lands are already in a perennial cropping system, there are fewer management options for change and the potential to increase SOC storage is lower.

4.3 Prescribed Burning and Wildfire

Fires have variable impacts on soil carbon. Some fires increase soil carbon while some decrease it. This appears to be site specific and dependent on a number of factors such as timing, intensity and frequency (Hurteau and Brooks 2011). Fires in native North American grasslands increase primary production and maintain the dominance of desirable perennial vegetation, leading to overall increases in soil carbon stocks. Forest wildfires however, often leave soil bare and vulnerable to erosion and the concomitant carbon loss (Delwiche 2009). Erosion control soil amendments along with desired perennial grass seeds may be a method of reducing soil loss through

¹ R. Busby, USACE ERDC CERL, Champaign, IL pers. comm., July 2016.

erosion after wildfires (Larson et al. 2016). Because fires are common and sometimes necessary on military lands (either controlled or accidental), additional data regarding how to incorporate soil carbon management into fire management plans could improve carbon sequestration while maintaining the desired environmental properties provided by fires.

4.4 Invasive plant species

Invasion of non-native plant species can alter SOC in multiple ways. First, many are annuals or short-lived perennials that displace longer lived native perennial plants that devote more carbon to root production. Second, plant invasions reduce biodiversity, which decreases SOC. Finally, non-native plant species generally don't associate as heavily with mycorrhizal fungi (Busby et al. 2011). This disassociation, known as invasional meltdown, degrades mycorrhizal fungi in soils that are important for soil carbon storage, healthy soils, and native plant communities. In areas where invasive annual plants replace native perennial plants, such as the cheatgrass invasion in the Intermountain West, soil carbon losses are considerable (Koteen et al. 2011).

4.5 Biodiversity

Maintenance of diverse plant communities on military lands is promoted for a number of reasons, including habitat, resilience, and training realism. An additional benefit is increased SOC (Fornara and Tilman 2008). However, not all plant functional groups are associated with high SOC. Replacement of C4 grasses in prairies with C3 grasses reduces overall SOC (Fornara and Tilman 2012). Soils under evergreen trees have higher SOC than deciduous trees (Setälä et al. 2016). Species within functional groups also vary in their ability to accumulate and store SOC (Ahmed et al. 2016). Thus, an increased understanding of which plant species and groups store more SOC in different environments could be used in land management to increase soil carbon storage.

4.6 Soil amendments

4.6.1 Biochar

Biochar is the byproduct of gasification and pyrolysis systems. This byproduct can be useful as both a soil amendment to improve soil health and for increasing SOC. Biochar has the potential to increase soil carbon through the following:

- Direct addition of carbon to the soil
- Increasing the density and diversity of rhizosphere bacteria and fungi
- Decreasing in soil carbon loss through erosion.

However, different sources of material and different production processes affect the biochar properties and how they interact with different soil types (Ahmedna et al. 2000; Fang et al. 2014; Sigua et al. 2014). Additional research is necessary to determine optimal process parameters for what type of biochar is desirable, where, when, and how much to apply to soils, including logistics for production, and application at an operational scale.

4.6.2 Paper waste

Paper waste at DoD installations is frequently pulverized to a particle size that renders it unfit for recycling. The military produces large amounts of this waste material, which is often incinerated or landfilled. Current research is evaluating this waste product as a soil amendment to increase soil carbon in training lands. Adding waste materials such as paper, with a high carbon:nitrogen ratio, ties up soil nitrogen and promotes establishment of perennial vegetation that increase soil carbon through increased root growth.

Paper waste has the potential to increase soil carbon through the following:

- Decrease of soil carbon loss through erosion
- Increasing the biomass available for decomposition by soil fauna and microbes
- Increasing below ground biomass.

4.6.3 Fertilizer

Adding fertilizer to training areas occurs routinely when rehabilitation with seeding occurs. Fertilizing healthy stands of vegetation has the capacity to increase soil carbon, but can also lead to degraded ecosystems due to reductions in plant biodiversity and promotion of undesirable vegetation. Thus, understanding where this amendment can yield positive effects and where it is unwarranted is needed for effective utilization.

Proper fertilization increases soil carbon through the following:

- Increase in above ground biomass
- Increase in below-ground biomass.

4.6.4 Biopolymer

Some beneficial soil microbes produce and excrete exopolysaccharides (EPS) into the soil around roots. The EPS alter the soil to produce an environment that favors both plant and microbial growth (Bais et al. 2006). By performing the EPS production industrially (Newman et al. 2010, Patent No. 7,824.569) and applying the concentrated EPS to soils, the soil modification that would naturally occur over hundreds of years of leguminous plant/microbe EPS production can be performed in hours.

The EPS of *Rhizobium tropici* has been field-tested in upland situations and has been shown to decrease erosion on slopes, such as berms and levees (Larson et al. 2012, 2016). In addition, soil amendment with the biopolymer encourages rapid seed germination and root development of desirable grasses over that of invasive species (Larson et al. 2012). It also aids in drought resistance of the plants which means lower maintenance costs for the installation (Muller 2015). Agriculturally, the biopolymer amendment is associated with increased crop yields¹. The effect of a single application performed during normal soil preparation/seeding practices has been observed to last for several years².

Biopolymer amendment increases soil carbon through the following:

- Reduced soil carbon loss through erosion
- Direct addition of biopolymer carbon to the soil
- Increase in above-ground biomass
- Increase in below-ground biomass.

4.7 Nitrogen deposition

Nitrogen deposition from vehicle emissions, air and water pollution is a widespread and growing concern that is not easily managed. However, the impacts of this addition to soils must be understood and managed to

¹ S. Sevinc, unpublished communication.

² S. Larson, USACE ERDC EL, Vicksburg, MS, pers. comm., July 2016. Field study research was conducted in Edwards, MS.

effectively maximize soil carbon storage. Increased nitrogen favors competitive, less desirable, annual plants and weeds, alters plant litter quality, and accelerates nutrient cycling in soils. Ectomycorrhizal diversity in soils has also been shown to drop in areas where nitrogen is deposited due to plants' ability to acquire nitrogen without the cost to associate with their symbiont. In certain instances, this loss of diversity has had detrimental impacts on plant communities, as subsequent droughts have resulted in high plant mortality and community shifts because the symbionts are not present to improve water acquisition, another beneficial function of these fungi (Allen et al. 2010). The reduced vegetative mass could allow an increase in soil carbon loss through erosion. Better ecosystem model development incorporating nitrogen inputs (e.g., fertilizer, deposition), atmospheric carbon levels, and land management will provide more realistic predictive model outputs (Clemmensen et al. 2013).

5 Wetland Management Effects on Soil Carbon Flux

The carbon sequestered in freshwater or coastal wetlands sediment is known colloquially as *blue carbon*. Coastal vegetation includes both marshes, seagrass and mangrove stands (Figure 7). Wetlands are a dynamic system of carbon sequestration and emission (i.e., carbon flux). Issues that affect sequestration of carbon in sediments include the identification of factors which might limit carbon emission from these sediments as well as those factors which might increase carbon sequestration (Downing et al. 2008).

Figure 7. Examples of coastal vegetation: marshes (top left), mangrove stand (bottom left) and beach grass (right).



Table 7 compares the sequestration potential of coastal vegetation to that of terrestrial forests. In this summary, sequestration is defined as the burial and storage of carbon in the soil/sediment. Although the carbon saturation potential of wetlands is considered low compared to forests, the sequestration rate and carbon permanence are very high (FAO 2005; MacDicken 2015). A later study also stated that, among LULC classes, wetlands have the highest mean SOC stocks while rangelands have the lowest (West et al. 2013). However, the total area of wetlands is much lower than that of forests and loss of wetlands continues to increase. The high potential for self-expansion implies that if this loss trend is reversed, the plant growth would continue with minimal assistance, and carbon

sequestration would also increase. The total wetland acreage managed by the USACE, combined with DoD-managed freshwater and, particularly, coastal wetlands, has a very high potential for long-term sequestration of atmospheric carbon.

Table 7. Comparison of carbon sequestration characteristics of coastal vegetation and terrestrial forests.

Characteristic	Coastal vegetation	Terrestrial forest
Sequestration rate, avg (gC m ⁻² yr ⁻¹)	High ^{1,2} Marsh 210 ³ Mangrove 139 Seagrass 83 ⁴	Lower ^{1,2} Tropical 2 Temperate 1–12 Boreal 1–2
Sequestration permanence	High ¹⁻⁴	Low ^{1: 2}
Fire risk	None to low ²	High ²
Carbon saturation potential	Low ¹⁻⁴	High ^{1: 2}
Area	Low ^{1: 2}	High ⁵
Recent loss rate and trend	1–5% yr ⁻¹ , increasing ^{1: 2: 7}	0.8% yr ⁻¹ , stable or decreasing ⁸
Self-expansion potential (via unassisted clonal expansion)	High/rapid ^{4: 9}	Low

¹ Laffoley and Grimsditch (2009)

² Nellemann et al. (2009)

³ Chmura et al. (2003)

⁴ Duarte et al. (2005, 2010)

⁵ IPCC (2013)

⁷ Polidoro et al. (2010)

⁸ FAO Global Forest Resources Assessment (2005)

⁹ Liu et al. (2007)

As an example, the USACE, and other public and private organizations, are currently spending large amounts of money over several years to restore Gulf of Mexico barrier islands, coastal harbors, vegetated storm barrier areas and brackish and freshwater wetlands (United States Department of Agriculture (USDA), Natural Resources Conservation Service 2003, 2011, Alabama Dune Restoration Project Phase 1 2012). Re-vegetation of these saline and brackish areas is a challenge. However, according to Restore the Earth's proprietary EcoMetrics model, \$1 of investment in wetlands restoration produces \$9 of value (<http://restoretheearth.org>). That value comes in the form of everything from tangibles like water and carbon credits, sustainable timber harvests, and hunting licenses to more difficult-to-value benefits like strengthened social bonds from fishing trips (Boynton 2016). Considering this return-on-investment, it becomes economically advantageous to consider wetlands restoration (Enwright et al. 2016).

6 Instigating and Maintaining Changes in Land Management that Promote Carbon Sequestration

Change is a necessary part of life, but change is difficult. Up to 70% of all change initiatives fail (Ashkenas 2013). Increasing the mass of atmospheric carbon that is sequestered in soil will require changes in facility land management protocols at both the personal and the organizational level. Attempts to implement new programs, practices, or policies in organizations often fail because leaders do not establish sufficient organizational readiness for change (Kotter 1996). While change management experts have suggested some strategies for creating readiness for change within an organization, there has been limited research in this area.

Disruptive, transformative technologies are being introduced at an accelerating pace that increases individual fear and resistance to change (Juma 2016). Transparency, inclusiveness, and caution in the handling of scientific uncertainty are critical elements of public trust (Juma 2016). How they are managed can increase or decrease the public resistance to technological innovation (change). Objections to change can fall into one, or more, of the following categories:

- Intuitive responses – reflects patterns of behavior that rely on deep evolutionary roots of fears and phobias, such as fears of new foods.
- Vested interests – a clash of competing economic world views and moral values.
- Intellectual challenges –includes philosophical objections to new technologies such as the manipulation of nature.
- Business models – changes can face strong opposition and include some vested interests.

Therefore, successful change requires inclusive innovation for acceptance of new controversial technologies. Ultimately, the goal is to manage public risk perception and foster trust between the public and companies/institutions that introduce innovative technologies. Inclusive innovation means:

- More timely scientific assessments of the benefits and risks of new technologies,

- Faster adjustment of social institutions to keep pace with the new technologies,
- Greater public awareness and citizen engagement,
- Inclusive economic changes including training in the new technologies by public institutions, equitable management of intellectual property rights, and creation and support of long-term technology partnerships, and
- Building local capabilities in technology adoption and development.

6.1 Organizational readiness for change

Figure 8 summarizes Organizational Readiness to Change (ORC), as adapted from Weiner (2009). The ORC is a multi-level, multi-faceted, shared psychological state in which organization members feel committed to implementing an organizational change and confident in their collective abilities to do so (Weiner 2009). Change commitment reflects the shared resolve to implement change. Change efficacy refers to this shared belief in the collective capability to organize and execute the courses of action involved in change implementation. Organizational readiness varies as a function of how much individual members value the change and how favorably they view the task demands, resource availability and situational factors involved in the change. Kotter (1996) suggests that failure to establish sufficient readiness within the organization accounts for 50% of all unsuccessful large-scale organizational change efforts.

6.1.1 Change commitment

According to research of Weiner (2009), members of the organization commit to implementation of change for the following reasons:

- They “want to” – They value the change on a personal level.
- They “have to” – They have little choice about the change.
- They “ought to” – They feel obligated to change.

Members whose commitment to change is based on “want to” rather than “have to” or “ought to” motives exhibit both more cooperative behaviors, such as volunteering, and also championing behavior, such as promoting the value of the change to others (Weiner 2009).

Figure 8. Factors involved in determining Organizational Readiness for Change (ORC).



Strategies to enhance change commitment include the following:

- Highlighting the discrepancies between current and desired performance levels
- Encouraging dissatisfaction with the status quo
- Creating an appealing vision of the future state of the organization.

When an organizations members are more likely to initiate change, they exert more effort and persistence in improving the situational factors (such as information availability, task demands, resources) that oppose that change. These members show increased cooperative behaviors in this regard. Those members who see the change as undesirable often avoid the steps required to implement the change, resist change efforts and may even sabotage change efforts (Kotter and Schlesinger 2008).

6.1.2 Change efficacy

Weiner (2009) discussed organizational change efficacy as a shared belief by members of the organization in their collective capability to implement change. It is a cognitive appraisal of the individuals in a group where members assess, acquire, assimilate, and integrate information that addresses these basic questions:

- Task demands:
 - What are the mechanical steps to success?
 - Do we know what it will take to implement this change effectively?
- Resource availability (human, financial, material, and information):
 - Do we have the resources we need to implement this change effectively?
 - Situational factors (sufficient time, internal political environment)?
 - Can we implement this effectively given the current situation we face?

Strategies to enhance change efficacy include promoting positive assessments of task demands, available resources and other situational factors.

6.1.3 Change valence

Shea et al. (2014) defined change valence as a hypothesized determinant of change commitment. That is, do the members of the organization value the change? Do they believe the change is needed, important, beneficial, or worthwhile? Is the change a part of the core value system of the organization?

How much the change is valued is more important than why it is valued. An unknown in change valence is whether the reasons for supporting a change must match across individual organizational members. Positive change valence resulting from individuals' disparate reasons might be just as potent a determinant of change commitment as change valence resulting from commonly shared reasons.

The true measure of change valence, then, is the question (Shea et al. 2014)

“Regardless of their individual reasons, do organizational members *collectively* value the change enough to commit to its implementation?”

Increasing the change efficacy of an organization increases the change valence and, thus, increases the organizational readiness to *implement* change (Shea et al. 2014).

6.1.4 Change factors

In addition to the factors mentioned previously that affect change efficacy, factors concerning the organization itself may affect an individual's commitment to change (Shea et al. 2014). These include the organization's current culture, policies and procedures, past experience with implementing change, its resources that it can allocate to implementing a change, and the structure of the organization. These directly impact training and information sharing, task demands, and resource availability, all considered critical factors of successful change.

6.2 Organizational Readiness to Implement Change (ORIC)

Organizational Readiness to Implement Change (ORIC) is a psychometric assessment tool developed by Shea et al. (2014). The assessment helps organizations understand how ready they are to implement change. This directly affects how likely a major change might be accomplished in a complex, multi-level organization. The test is a measure of the change valence and change efficacy of an organization. Its use highlights areas where the organization can improve its performance if change is to be implemented successfully. For example, are a number of responses indicating overwhelming task demands on limited resources or information? This will discourage change commitment by even the positive members and decrease change valence. Increasing information and resources dedicated to the change and shifting personnel responsibilities to equitably share task demands will decrease the negative perceptions, increase the change valence and increase the chance of successful change implementation. The ORIC test (adapted from Shea et al. 2014) is presented as Appendix A.

6.3 Change for soil carbon sequestration and the DoD

Understanding DoD readiness for meaningful and significant carbon footprint change from previously accepted protocols and mindsets is an essential element of achieving a statistically significant, sustainable, carbon neutral stance. The organizations' commitment to that mandate must be assessed and re-assessed as carbon neutral-promoting activities

and technologies are advanced. The ORIC evaluation test (Shea et al. 2014) in Appendix A could be used over time to measure the DoD's continuing commitment to carbon neutrality.

The ORIC test would be ideal for these assessments because it is:

- Brief
- Change independent
- Group referenced
- Amenable to multiple respondents from the entire organization
- Inter-rated agreement
- Valid and reliable.

The ORIC evaluation would assess the change valence (commitment) of both specific DoD facilities, and across the DoD generally, to the idea of carbon neutrality. The evaluation itself might change a number of “have to” and “ought to” individuals to “want to” persons. The ORIC evaluation would also point out weak areas supporting change efficacy. This evaluation might change the DoD's assessment of its task demands, resources and other situational factors. Changes in these areas might contribute to increased chances for successful change implementation. The ORIC data may be used to alter presentations and information sharing and rollout initiatives to maximize the likelihood of successful change.

7 Research Areas to Maximize Carbon Sequestration on DoD Lands

Stockmann et al. (2013) attempted to summarize the “state-of-the-science” concerning sequestration of SOC. Stockmann, and other researchers, agree that it is possible to increase the concentration of carbon stored long-term in soil/sediment as humic material. Potential measures to increase soil carbon include the following:

- Increase the above-ground biomass to include those species that most effectively promote soil carbon sequestration
 - Increase forest lands
 - Increase perennial plants
- Increase the soil root mass
- Increase the concentration of EPS in the soil,
- Increase the ratio of EMF to AMF in the soil
- Increase total area of wetlands (freshwater, brackish and saline)
- Decrease loss of soil carbon through erosion.

Some specific examples are discussed in the following sections.

7.1 Stimulation of sequestration using amendments.

Common commercially available products (i.e., fertilizers and water holding additives like perlite) have shown promise in increasing storage of soil carbon. New products can be developed for this result (such as new biopolymers to improve re-vegetation of salt tolerant species and industrialization of glomalin as a soil amendment). Further, military activities generate unique waste materials (such as biochars or pulverized paper waste) that can be utilized to increase soil carbon in an environmentally friendly manner. Waste management can generate useful soil amendments such as compost or sludge digestates. However, knowledge of where, when, and how much to apply of these amendments is largely lacking. Focused research in this area would result in significant advances in net zero waste and energy objectives while contributing to soil carbon storage.

Through geoengineering and an understanding of the natural symbiotic relationship between plants and soil microbial communities in the rhizosphere, soils can be rapidly improved with regards to root structure, soil microbial activity, and root carbon exudation. This would lead to increased soil carbon storage and ultimately a healthier soil and soil structure. Costs associated with the production and use of soil amendments could potentially be offset by the following:

- Increased crop yields and reduced crop loss when used in agricultural soils
- Decreased soil loss through erosion when used for berm and levee construction and maintenance
- Increased grazing efficiencies when used with livestock industries
- Increased habitat, flood protection, and water quality when used in wetlands.

7.2 Stimulation of plant growth in marginal soils

Brackish wetlands can be very productive, but increasing plant growth in these areas is not well studied. Several EPS producing bacteria are being investigated for use in these marginal, nutrient-poor, and saline environments. There are plant species tolerant of salt that grow as well in saline soils as they do under more normal soil conditions. These plants also have associated rhizosphere bacteria that produce biopolymers and biofilms that appear to protect the plant from salt toxicity (Feagin et al. 2009; Qurashi and Sabri 2011). These include *Rhizobium leguminosarum* associated with Strawberry clover, several *Bradyrhizobium* sp. associated with Birdsfoot trefoil, and *Mesorhizobium loti*. These bacteria produce exopolymers that can be grown in bioreactors. These biopolymers should be evaluated for seed germination and development in marginal soils.

7.3 Eco-Engineering

Military lands are already managed to balance training and testing needs with environmental stewardship. However, little consideration is given to maximizing soil carbon content. Fortunately, many of these perspectives are mutually inclusive. Training doctrine and land management require healthy plant communities. Additional knowledge on what traits of healthy plant communities promote the greatest carbon storage and where this knowledge can be applied will have little effect on current training

requirements while having a potentially large impact on soil carbon storage. Examples include the following:

- Identifying areas where agricultural outleasing can be altered to increase soil carbon storage
- Prioritizing invasive species management efforts towards those that have the most negative impact on soil carbon
- Substituting species in seed mixtures to include greater and deeper root growth
- Identifying soils and other prime areas where carbon storage can be significantly increased
- Optimizing controlled burns to favor soil biological processes that promote carbon storage
- Inclusion of soil ecosystem services in timber management plans.

One of the most overlooked aspects of ecosystems is the biology of soil, including the management of mycorrhizal fungi that are vital in multiple ecosystem processes. In grasslands and shrublands, the AMF dominate (Sabais et al. 2012). Identifying mycorrhizal associations that increase glomalin and extra-radicle hyphae production would increase carbon storage. In forests, where EMF dominate (Pena et al. 2010), identifying the ideal ratio of ectomycorrhizal to arbuscular mycorrhizal hosts would increase soil carbon storage. An improved understanding of which associations between fungi and plants concomitantly promote a desirable training environment and maximize soil carbon storage would yield simple strategies with large payoffs.

7.4 The use of biomass for energy production

It may be possible to couple soil carbon sequestration with alternative energy production to increase the carbon benefit to DoD installations. Biomass at installations may be able to be collected to be used to produce energy. This includes methods like direct incineration, gasification, pyrolysis or anaerobic digestion to produce burnable gases or fuels, as well as extraction of oils and fats for biofuel production (Medina et al. 2003; Medina et al. 2015 is a discussion of approaches for wastewater residuals, but can also be applied for other biomass forms).

7.5 ERDC resources and funding vehicles

There are several avenues for collaboration and funding research into closing these data gaps, including the following:

- The DoD agencies of SERDP and ESTCP
- The USACE Climate Preparedness and Resilience Program
- The NSF NEON program.

8 Conclusions

The USACE has a long history of developing solutions for the nations and the world's most challenging engineering issues. From critical advances in military engineering to protection from floods and promoting navigation, solving the large and difficult challenges through the application of innovative engineering solutions on a large scale has often fallen on the Corps of Engineers. There is a belief that one of the greatest engineering challenges of the future will be the reduction of atmospheric CO₂ levels.

In 2014, in conjunction with the Institute for Water Resources, the USACE Climate Preparedness and Resiliency Community of Practice (CPR CoP) was created within the Engineering and Construction CoP. The purpose of the new CPR CoP was

“to increase the individual and organizational abilities and technical competence within USACE to plan, design, engineer and execute work that enhances the climate preparedness and resilience of its partners and stakeholders.”

UASCE climate change preparedness and resilience information (2014) can be accessed at <http://www.corpsclimate.us>.

Compared to engineering projects such as solar mirrors and pumping atmospheric carbon deep into the earth, soil carbon sequestration through changes in land management strategies is one of the few atmospheric carbon reduction efforts that could be implemented relatively quickly, over a large scale, and potentially at low cost. This CASI white paper has attempted to summarize the major impacts of DoD land use management methods on carbon flux and the human element of strategy implementation. This information can be used to better guide planning and accounting of environmental compliance activities. Additionally, while existing products and models are incapable of accurately estimating carbon fluxes on military lands, the unique nature of these activities should warrant focused research to develop accurate sequestration models. This proactive approach will help determine what future follow-on efforts might be necessary to fully understand carbon fluxes on military lands, including potential planning and policy guidance and research and

development efforts. Recommendations put forward by this white paper identify specific research needs in these areas.

8.1 Is it possible to achieve carbon neutral status on DoD managed lands and facilities?

Yes, with qualifications. While small changes in the SOC stock could result in significant impacts on the atmospheric carbon concentration (Stockmann et al. 2013), land management change may not happen quickly. Considering just a single land management change, such as the use of biopolymers as a soil management tool, how effective could this approach be for reducing atmospheric carbon? The University of Manchester carbon sequestration calculator (<http://dbkgroup.org/carbonsequestration/rootsystem.html>) was used to consider this question on a global scale:

- The amount of carbon as CO₂ currently in the atmosphere is approximately 750 Pg, and that in the soil is approximately 1,500 Pg.
- If the global land area for crops is 2,300 Mha and that for grassland is 2,300 Mha, and an extra 1.0 m depth of roots are grown, and they take up 1.0 % (by mass at an equivalent carbon density of 1,000 kg m⁻³) of the relevant soil volume, then the extra amount of carbon that could be sequestered by the above land areas is 10 kg.m⁻² = 100 t.ha⁻¹.
- This equates to 230 Pg (230,000 M tons), if the carbon in the roots and other sequestered carbon are re-respired over a period of two years.
- As 1 Pg is equivalent to 0.51 ppmv of atmospheric CO₂, this would decrease the CO₂ in the atmosphere by 118 ppmv; the current value of 385 ppmv would decrease to 267 ppmv. This represents a 31% reduction in atmospheric CO₂ after two years.

Obviously there are numerous assumptions leading to the calculated results, and significant effort would be required to produce a more meaningful prediction of the reduction of atmospheric CO₂. Beyond that, the inclusion of reductions of atmospheric CO₂ levels would need to be incorporated into advanced climate models in order to evaluate the technology's actual effectiveness. However, the possibility of such a reduction should merit the further evaluation of changes in land management strategies to increase soil carbon for climate protection.

8.2 Could DoD facilities be useful for carbon banking?

Yes, but with larger qualifications. The size of these areas could yield substantial carbon offsets. However, a carbon banking system has not yet been set up in the United States. Further, it would be necessary to explore if there are any regulatory restrictions to DoD facilities participating in carbon trading programs.

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Appendix A: ORIC Evaluation

ORIC Change Valence

Complete the sentence “People who work here.....” by circling the most appropriate response for your organization.

	1 Disagree	2 Somewhat disagree	3 Neither agree or disagree	4 Somewhat agree	5 Agree
1	...are committed to implementing this change.				
	1	2	3	4	5
2	...are determined to implement this change.				
	1	2	3	4	5
3	...are motivated to implement this change.				
	1	2	3	4	5
4	...will do whatever it takes to implement this change.				
	1	2	3	4	5
5	...want to implement this change				
	1	2	3	4	5

ORIC Change Efficacy

Complete the sentence “People who work here ...” by circling the most appropriate response for your organization.

	1 Disagree	2 Somewhat disagree	3 Neither agree or disagree	4 Somewhat agree	5 Agree
1	...feel confident that they can keep the momentum going in implementing this change.				
	1	2	3	4	5
2	...feel confident that they can manage the politics of implementing this change.				
	1	2	3	4	5
3	...feel confident that the organization can support people as they adjust to this change.				
	1	2	3	4	5
4	...feel confident that the organization can get people invested in implementing this change.				
	1	2	3	4	5
5	...feel confident that they can coordinate tasks so that implementation goes smoothly.				
	1	2	3	4	5
6	...feel confident that they can keep track of progress in implementing this change.				
	1	2	3	4	5
7	...feel confident that they can handle the challenges that might arise in implementing this change.				
	1	2	3	4	5

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14. ABSTRACT Sequestration of atmospheric carbon dioxide in soils is a promising alternative for mitigation of atmospheric carbon dioxide (CO ₂). The Department of Defense (DoD) owns significant land and water resources which can be managed to offset emissions. Accounting for this, sequestration could help DoD reach carbon neutrality. Many activities the DoD engages in for sustainable land management and training sustainment are conducive to soil carbon storage without even considering this as an important component; however, carbon storage could be greatly enhanced by increased understanding of optimal storage conditions and by making slight adjustments to existing practices. Land management techniques may require adjustments to maximize carbon storage while maintaining training and environmental quality. In order to achieve this, data gaps for estimating carbon fluxes need to be addressed so that accurate measurements can be taken. Unknown aspects of carbon storage as it relates to plant-soil-soil microbe interactions need to be investigated to maximize carbon storage while maintaining land use requirements. Geo-engineering concepts require further refinement to increase carbon storage in soils. These knowledge gaps are not insurmountable and could be addressed through focused research to maximize and accurately quantify carbon storage on DoD lands.					
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